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Article

The Variation of Bioproperties of Mineral Waters used in Balneotherapy after Contamination with Coliform Bacteria

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Abstract: The presence of fecal indicator bacteria in therapeutic mineral waters (TMWs) represents an important factor that can alter their chemical properties and therapeutic effect, correlated with an increased health risk for patients. The purpose of study was to evaluate the interrelationship between the level of contamination with coliform bacteria/*Escherichia coli* determined by the culture-dependent method and the concentrations of major ions, trace elements and gases in mineral waters that are used for curative or prophylactic purposes. Coliform bacteria showed a degree of survival modulated by the physical-chemical profile of TMWs, their viable count being statistically correlated with T°C, pH, TDS, EC and major ions (Mg²⁺, Ca²⁺, Fe²⁺, Cl⁻, HCO₃⁻, SO₄²⁻). The prevalence of *E.coli* bacteria was not sensitive to the oscillation of the values of T°C; pH; TDS; EC of TMWs, but on the other hand, for the concentration variations of Cl⁻, HCO₃⁻, total mineralization from the composition of lakes/bathing pools, a positive correlation was identified with its numerical density. For sulfurous TMWs, the results showed a negative correlation between H₂S concentration and the level of CB or *E.coli*, and in carbonated TMWs, CO₂/CB values varied directly proportionally, while *E.coli* evolved numerically independently of CO₂ concentration fluctuations, as expected, being aerobe/facultative anaerobic bacteria.

Keywords: therapeutic mineral waters; contamination; coliform bacteria; chemical composition of mineral waters; bioproperties of therapeutic waters

1. Introduction

Balneotherapy is an important component of Physical and Rehabilitation Medicine, which uses natural factors in spa resorts for therapeutic purposes, to modulate the symptoms of numerous diseases, and also representing a non-pharmacologic alternative, easily accepted by patients, used for both curative and preventive purposes [1]. The use of medicinal natural resources combined with the healing properties of the climate contributes not only to the reduction of treatment time for many diseases but also helps improve the therapeutical results [2]. Spa therapy has a long tradition in Europe, being recognized as a complementary therapy in treatment of rheumatological, dermatological or gynecological pathological, and also having an impact in metabolic and

orthopaedic pathologies.[3–5]. However, at the same time, hydrotherapy carried out in spas with microbiologically unsafe water can represent a health risk, especially for the elderly and immunocompromised patients, which represent the target group of population to whom balneotherapy is addressed [6]. Depending on the soil from which they origin, the mineral waters have different therapeutic indications, their biological properties also coming from their own microorganisms, very specific to the mineralization of these waters [7]. The results of numerous previous studies have proven the ability of mineral waters to positively influence the functionality of the digestive system, to eliminate various metabolic disorders and to regulate hormonal and immunological processes [8]. Studies have also shown the beneficial effect of balneotherapy on the quality of life of people with rheumatoid arthritis, an effect induced by both mineral water bathing procedures and peloidotherapy [9]. Coliform bacteria belong to the Enterobacteriaceae family, including species of several genera: *Citrobacter*, *Enterobacter*, *Echerichia*, *Hafnia*, *Klebsiella*, *Serratia* and *Yersinia*. The presence of coliform microorganisms in water attests to the contamination of faecal origin, but also indicates the possibility of contamination with other pathogenic bacterial species such as *Shigella* spp., *Salmonella* spp, *Yersinia* spp. or *Vibrio cholerae* [10], as well as problems that may intervene in the transport system of water [11]. The numerical density of coliform bacteria in water is determined by the complex interactions between biological, physical and chemical parameters, these interactions being controlled by factors that differ from one aquatic environment to another [12]. In recent years, an increase in the number of coliform bacteria in surface waters (lakes or water reservoirs) has been observed, but these high concentrations were rather correlated with an increase in temperature during the warm season, than with an induced faecal contamination with coliforms and intestinal enterococci [13]. *E. coli* is a human and animal commensal that reaches the water almost exclusively through human and animal fecal waste eliminated in the environment. These bacterial cells can also accumulate in the sediment, a niche which favors the horizontal gene transfer, including resistance genes to different antibiotics and virulence genes, which explains the existence of *E.coli* pathotypes [14,15]. Coliforms are Gram negative and non-sporogenes species as all family, with fermentative properties, producing acids and gas in special culture media, after a 48h of incubation at optimal temperature of 35-37°C. *E.coli* was first recognized as a human pathogen in 1982 and was divided into two main categories: commensal an pathogenic, the last one producing intestinal/diarrheal and extraintestinal infections; it can survive in water, depending on environmental conditions, which is why it is considered to be a bacterial indicator of a low sanitary quality of food and water, indicating their contamination with human and/or animal waste. Contamination can significantly change the chemical properties of groundwater, disrupting the general balance of ecosystems and causing economic losses because the respective waters become unfit for human consumption [16]. Environmental factors such as the texture of sand and gravel in the aquifer or an underground environment with accumulations of precipitation, have been correlated with a low or medium level of faecal coliform pollution, while a high level of faecal pollution was frequently identified in tap water from humanized areas, where there are also large populations of animals and birds [17]. Usually present in the lower intestinal tract of warm-blooded animals, *E. coli* can be released outside the host, thus ending up in waste water, drinking water or recreational water. Although for a long time *E.coli* was considered an indicator of recent faecal contamination, its ability to adapt to different natural habitats can create confusion in its use as a faecal indicator [18]. The abundance of *E.coli* can also vary depending on the dynamics of the human population and the degree of urbanization, socio-economic circumstances and extreme weather conditions (floods or drought) [19].

2. Materials and Methods

The study, carried out in the period 2019-2021, aimed to evaluate the contamination with coliform bacteria (Enterobacteriaceae that express β -D-galactosidase enzyme) and *E.coli* (with simultaneous β -D-galactosidase and β -glucuronidase activity) of the mineral waters used in curative or prophylactic purpose, in spa resorts, and the impact of this microbiological pollution on their physical and chemical properties. The analyzed samples were taken from hydro-mineral sources of

underground origin (springs, boreholes, wells) or above ground (natural lakes, swimming pools arranged for bathing in spas) that were in operation at the time of sampling. The physical parameters that condition a normal biological activity of microorganisms were monitored: temperature, pH, amount of salts (TDS) and, implicitly, electrical conductivity. To determine the pH and electrical conductivity, the electrometric method was used according to the method standards in force [20,21]. To define the chemical characterization of therapeutic mineral waters contaminated with coliform bacteria/*E.coli*, 14 chemical parameters were quantified (5 cations, 6 anions, 2 gases (CO_2 , H_2S), total mineralization) considered to have a primary role in defining their chemical profile that is responsible for the therapeutic effect. The quantitative determinations were made according to the method standards in force for natural mineral waters and the obtained results were expressed in $\text{mg}\cdot\text{l}^{-1}$. Since the level of nitrate and nitrite concentrations can affect human health but also aquatic organisms [22], nitrogen compounds (NH_4^+ , NO_3^- , NO_2^-) were also monitored in parallel with the evolution of the numerical densities of coliform bacteria and *E.coli* from TMWs. The samples were taken from natural lakes and bathing ponds from depths of 0.5 m; from springs, boreholes or wells, the sampling procedure was carried according to SR EN ISO 19458:2007 [23], with measures taken in order to avoid external contamination. The abundance of coliform bacteria and *E.coli* was quantified by culture-based method using the membrane filtration technique (cellulose ester membrane, 0,45 μm pore size) and the culture medium used was Microinstant Chromogenic Coliform Agar which was obtained by Scharlab S.L. The β -galactosidase positive colonies were counted as coliform bacteria-not *E.coli* and β -galactosidase positive/ β -glucuronidase positive colonies were counted as *E.coli*. To confirm the identification of *E.coli* colonies, indole production was checked by adding the Kovacs reagent. The number of *E.coli* and coliform bacteria was obtained by counting the specific colonies developed on culture media according to SR EN ISO 9308-1:2015/A1:2017 [24]. The results of microbiological analysis were expressed in colony-forming units (CFU) per unit volume. The obtained results were processed in GraphPad Prism 9 version 9.5.1 and statistically analyzed by simple linear regression, multiple linear regression, 2 way ANOVA and Pearson's correlation.

3. Results

Microbiological contamination with coliform bacteria β -galactosidase+ (CB β -galactosidase+) and *E.coli* was monitored in 52 samples of TMWs, 44.23% being underground therapeutic waters (spring-type natural sources (15) or artificially drilled (13 water drills, 1 well) and 55.77% above-ground therapeutic bathing waters (natural therapeutic lakes (11 samples) and bathing pools from spa resorts (12 samples)). CB β -galactosidase+ contaminated the TMWs regardless of the type of source, while *E.coli* bacteria was mainly identified in natural lakes or bathing pools introduced in the spa circuit (86.96%), always in association with the pollution indicator - total coliforms (Figure 1).

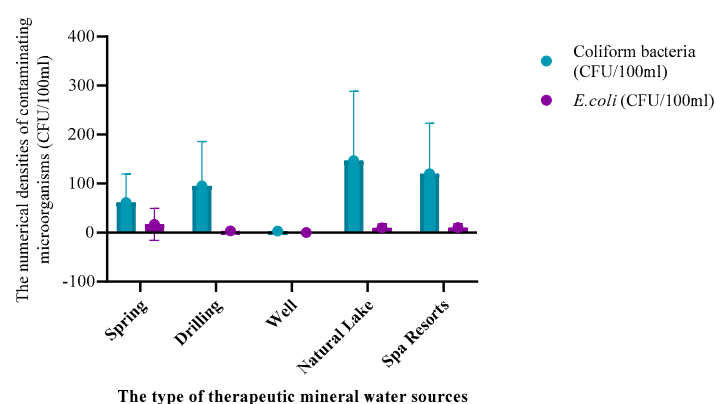


Figure 1. The numerical variation of coliform bacteria/*E.coli* in different types of contaminated TMWs sources.

The physical and chemical characteristics of the TMWs included in the study are presented in Table 1. They were specific to each analyzed sample, while their total degree of mineralization (total dissolved solids) recorded values that were included in the range of (0.25-165.8) g·dm⁻³.

Table 1. Statistical data of underground and surface therapeutic mineral water (TMWs) samples.

Parameters	Underground TMWs (spring, well, drilling) N=29	Surface TMWs (natural lake, bathing basin) N=23
Physical	Min/Max/SD	Min/Max/SD
Temperature (°C)	4.5/47.6/11.34	9.7/30.7/5.53
pH	6.04/8.2/0.58	6.08/8.94/0.95
Total dissolve solids (g/l)	0.24/14.94/4.19	9.15/149.61/41.01
Electrical Conductivity (mS/cm)	0.19/16.76/2.6	4.2/194.3/57.87
Chemical		
Total Mineralization (g/l)	0.25/27.44/6.13	8.11/165.80/45.91
Cations		
Na ⁺ (g/l)	0.012.7/1.03/2.29	1.95/59.32/17.09
Ca ²⁺ (mg/l)	16.1/234.1/53.32	32.1/3935.7/770.22
Mg ²⁺ (mg/l)	0.5/79.7/33.38	25.3/7898.1/2166.16
Fe ⁺² (mg/l)	0.05/6.1/1.33	0.05/2.1/0.49
NH ₄ ⁺ (mg/l)	0.06/12.2/2.58	Abs*
Anions		
Cl(g/l)	0.018/15.61/3.27	0.027/92.18/25.98
I(mg/l)	0.1/6.6/0.46	0.1/1.2/0.22
NO ₂ -(mg/l)	0.06/0.1/0.06	Abs
NO ₃ -(mg/l)	0.06/12.5/3.41	2.5/20.1/3.94
HCO ₃ -(mg/l)	48.8/4148.1/991.97	146.4/1189.1/268.2
SO ₄ ²⁻ (mg/l)	2.4/652.3/167.3	28.4/59800.1/11979.2
Gas		
H ₂ S(mg/l)	Abs/122.1/28.26	Abs
CO ₂ (mg/l)	Abs/1364.1/452.05	Abs
Microbiological		
Coliform bacteria (β-galactosidase+) (CFU/100ml)	1/267/76.73	6/414/116.97
<i>Escherichia coli</i> (β-galactosidase+, indol+) (CFU/100ml)	Abs/96/17.56	Abs/28/8.31

*abs=absent.

CB β-galactosidase+ and *E.coli* were identified in 17 samples of oligomineral TMWs (with <1g·dm⁻³ mineral substances), in 11 samples of TMWs with medium mineralization (1-15 g·dm⁻³ mineral substances) and in 24 concentrated or very concentrated TMWs (with 15-150 g·dm⁻³ mineral substances) (Figure 2a,b).

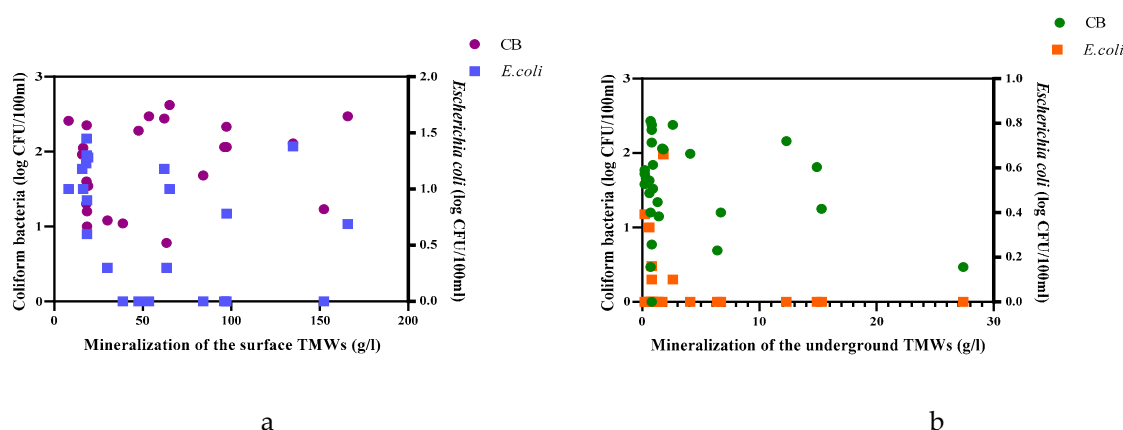


Figure 2. The distribution of coliform bacteria/*E.coli* (CFU/100ml) in underground/surface TMWs with different degrees of mineralization: (a) surface TMWs; (b) underground TMWs.

The maximum density ($4.14 \cdot 10^2$ CFU $\cdot 100\text{ml}^{-1}$) of CB β -galactosidase+ populations was recorded in Lake Ursu-Sovata (exploitation point), where the water mineralization value was $65210.8 \text{ mg}\cdot\text{l}^{-1}$ and the lowest, $0.1 \cdot 10$ CFU $\cdot 100\text{ml}^{-1}$ in the sample taken from F4011, Băile Felix (Bihor), with a mineralization of $814.8 \text{ mg}\cdot\text{l}^{-1}$.

The maximum level of *E.coli* was identified in Izvorul 2 Moinești (Bacău), with a mineralization of $1420.4 \text{ mg}\cdot\text{l}^{-1}$, while the minimum numerical density of $0.1 \cdot 10$ CFU $\cdot 100\text{ml}^{-1}$ was recorded in the samples from Izvor 3 Moinești (Bacău), Izvor 3 Slănic Moldova (Bacău), Izvor 11 Olănești (Vâlcea), Izvorul de Vest Vatra-Dornei (Suceava), F1-4714 Carei (Satu Mare), whose degree of mineralizations was measured and included in the interval (254.9 - 14941.1) $\text{mg}\cdot\text{l}^{-1}$.

The physical parameters in the contaminated underground TMWs evolved differently compared to the numerical levels of the target polluting microorganisms; thus, according to the results of the statistical analyses, either there was an interdependence of the CB β -galactosidase+ levels, the $T^\circ\text{C}$ values ($F=19.11$; $p=0.0002$) and the pH ($F=26.14$; $p<0.0001$) (Figure 3) or these micropollutants had an independent evolution in this type of ecosystems, regardless of the variation of TDS ($F=3.075$; $p=0.0905$) and EC ($F=2.533$; $p=0.1227$) values.

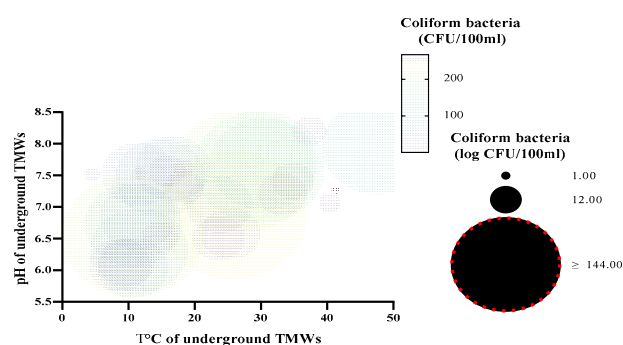


Figure 3. The behavior of coliform bacteria (variation in numerical densities) according to pH and $T^\circ\text{C}$ value of underground TMWs.

In the same type of samples (underground TMWs), the numerical densities of the *E.coli* populations evolved independently of the variation in the values of all the monitored physical parameters: $T^\circ\text{C}$ ($F=0.7135$; $p=0.4054$), pH ($F=1.775$; $p=0.1935$) (Figure 3b), TDS ($F=0.1987$; $p=0.6592$) and EC ($F=0.1419$; $p=0.7093$).

For surface TMWs, the statistical results reflected a behavior of CB β -galactosidase+ populations depending on the variation of their physical parameters: $T^\circ\text{C}$ ($F=29.85$; $p<0.0001$), pH ($F=25.37$; $p<0.0001$), TDS ($F=10.46$; $p=0.0038$) si EC ($F=15.70$; $p=0.0007$) (Figure 4a). *E.coli* was present in

numerical densities that varied proportionally only with the values recorded for T°C (F=18.09; p=0.0003), pH (F=19.24; p=0.0002) and TDS (F= 4.774; p=0.0398) (Figure 4b).

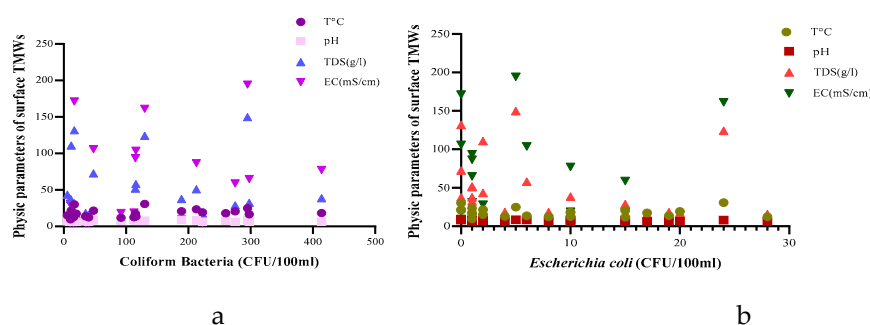


Figure 4. – The relationship between coliform bacteria(a)/*E.coli*(b) and physical parameters of the surface TMWs.

In sulfurous underground TMWs (with a concentration ≥ 1 mg/l of titratable sulfur in the form of H_2S , HS^- , S^{2-}), waters whose bioproperties have not yet been fully understood exactly, in the context of their healing capacity [25], the variation of H_2S concentrations was not associated with the population oscillations of CB β -galactosidase+ (p=0.1886) nor with those of *E.coli* bacteria (p=0.5258) (Figure 5a).

For carbonated TMWs (with a concentration ≥ 1000 mg/l of CO_2), the level of CO_2 concentration was correlated with the numerical densities of CB β -galactosidase+ (p=0.0287), while the populations variations of *E.coli* were not influenced by the CO_2 concentration fluctuations in the underground TMW (p=0.9999)(Figure 5b).

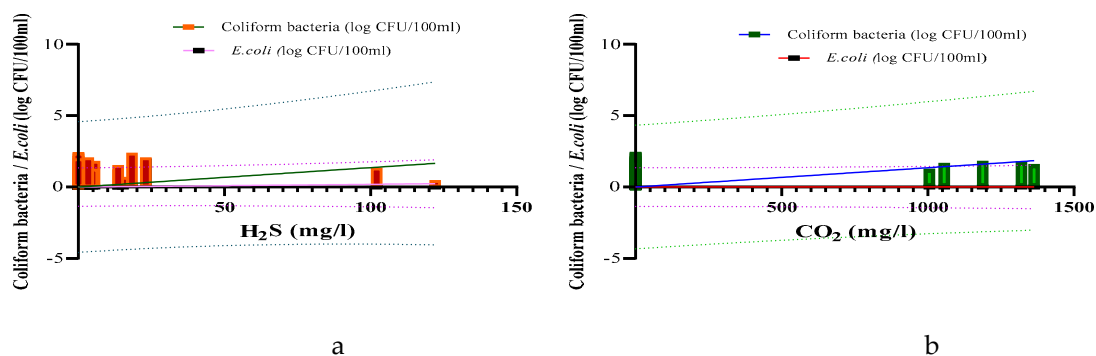


Figure 5. – The relationship between the H_2S (a)/ CO_2 (b) content of underground TMWs and the level of coliform bacteria/*E.coli*.

Since in carbonated TMWs, the iron ion is frequently found at a high concentration level, the combined effect of the two chemical constituents on the incidence of *E.coli* and coliform bacteria was studied comparatively, observing an independent evolution of these microorganisms (p=0.9680; p=0.9806) compared to the variation of the two chemical indicators in the composition of the analyzed waters (Figure 6).

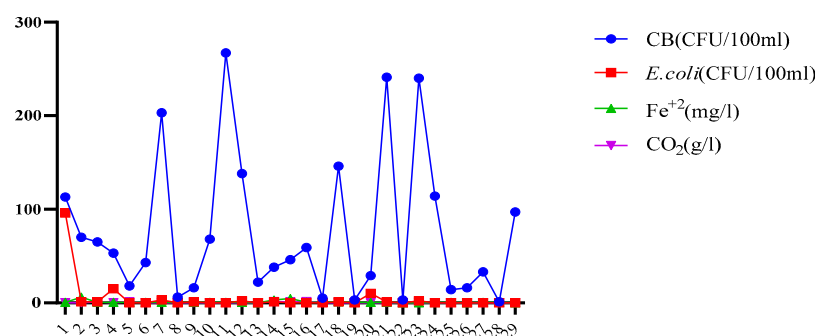


Figure 6. – The relationship between the content of $\text{CO}_2/\text{Fe}^{2+}$ and the level of coliform bacteria/*E.coli*.

Regarding the intervention of nitrogen compounds in modulating the presence and survival of CB β -galactosidase+ and *E.coli*, the statistical results indicated a correlation of the degree of pollution induced by total coliforms with the level of concentrations of NO_3^- ($p=0.0030$), NO_2^- ($p<0.0001$) or with the pH value ($p<0.0001$) from contaminated underground TMW (Figure 7).

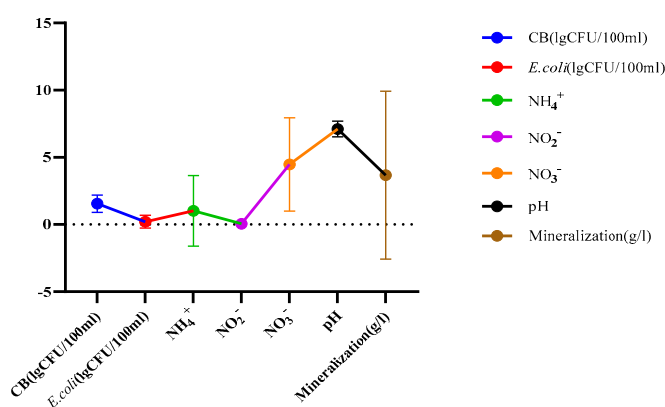


Figure 7. – The relationship between pH, mineralization and the content of nitrogen compounds in underground TMWs and the level of coliform bacteria/*E.coli*.

For underground TMWs, the major ions whose concentration values were correlated with the level of CB β -galactosidase+ were Mg^{2+} ($p=0.0010$), Ca^{2+} ($p=0.0002$), HCO_3^- ($p=0.0273$) (Figure 8a), while the numerical variations of *E.coli* evolved independently of the quantitative variations of Na^+ ($p=0.8521$), Mg^{2+} ($p=0.4812$), Ca^{2+} ($p=0.083$), Fe^{2+} ($p=0.6383$) and NO_3^- anions ($p=0.1341$), Cl^- ($p=0.8685$), I^- ($p=0.88180$), HCO_3^- ($p=0.494$) (Figure 9a).

In surface TMWs contaminated with CB β -galactosidase+, the ions that registered concentration oscillations were Mg^{2+} ($p=0.0169$), Fe^{2+} ($p=0.0061$), Cl^- ($p=0.0015$), HCO_3^- ($p=0.0002$), SO_4^{2-} ($p=0.0227$) (Figure 8b), while the numerical oscillations of *E.coli* were dependent on the levels of Cl^- ($p=0.0302$), HCO_3^- ($p=0.0054$) anions, as well as the degree of mineralization of this type of water ($p=0.0325$) (Figures 9b).

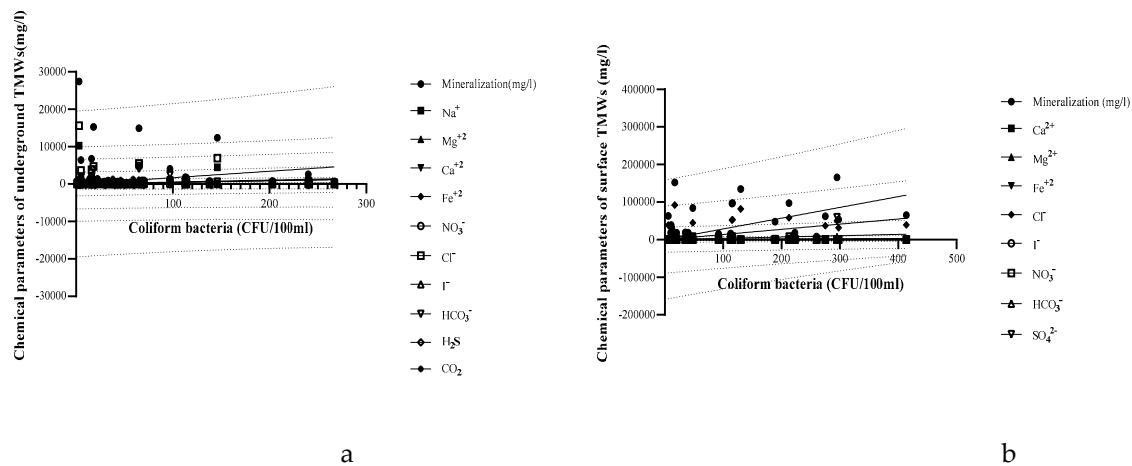


Figure 8. The effect of the chemical composition of the underground (a)/surface TMWs (b) on the incidence of coliform bacteria (CFU/100ml).

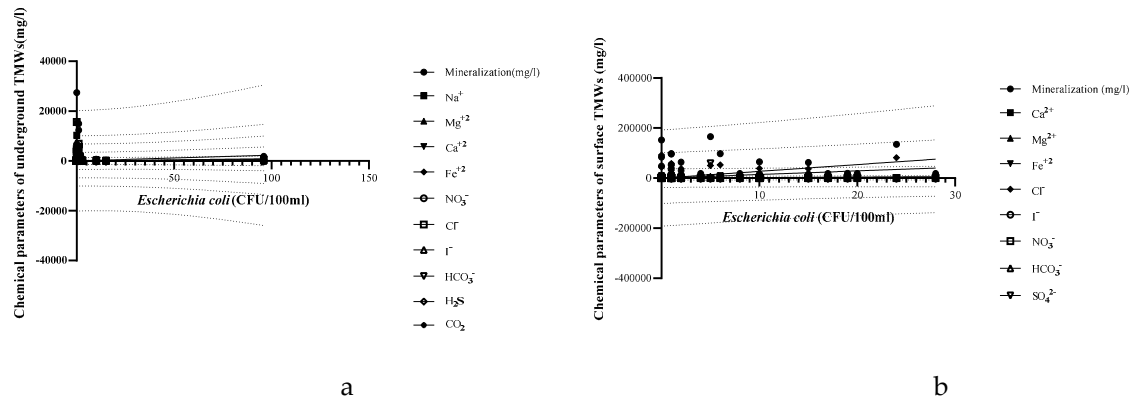


Figure 9. The effect of the chemical composition of the underground (a)/surface TMWs (b) on the incidence of *Escherichia coli* (CFU/100ml).

Figures 10 and 11 express the correlation matrix between CB β -galactosidase+/*E.coli* and the physicochemical parameters of underground and surface TMWs; significant correlations were identified between different parameters for a value of $\alpha=0.05$.

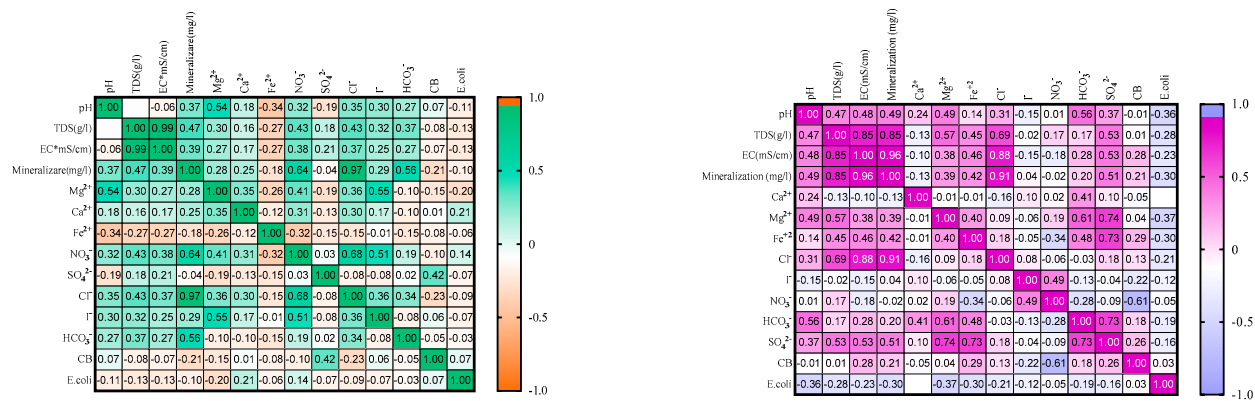


Figure 10. Pearson's correlation matrix for physicochemical parameters and coliform bacteria/*E.coli* in underground/surface therapeutic water.

4. Discussion

The behavior of CB β -galactosidase+ in surface TMWs that we observed corresponds to the already known trends according to which, in surface waters, the presence and population densities of coliform bacteria are more likely to be affected by physical and climatic factors (water temperature, precipitation, leaks, solar radiation, dissolved nutrients, competition with other bacteria or other physico-chemical conditions) and in groundwater, where geochemical processes (dissolution, hydrolysis, precipitation, adsorption and ion exchange, redox) control the water chemistry, and the presence of coliforms is controlled by these processes [26].

For *E.coli*, the results obtained outlined a similar evolution to the data of previous studies that identified the stressful effect of temperature, pH, UV light and nutrient availability on the viability of this bacterium from faecal-polluted waters [27] or the significant predictive value of precipitation on high concentration levels of *E.coli* [28].

Identifying the independence of the numerical variations of coliform bacteria and *E. coli* compared to the H₂S concentrations in underground or surface TMWs represents an important element in the management of the therapeutic use of these waters, knowing that this gas is considered to be an important biological mediator involved in the processes of vasodilatation, neuromodulation or anti-inflammatory processes [29], with beneficial effects on diabetes, glucose metabolism [30] but also with a significant antiviral activity on a large range of RNA enveloped viruses [31,32] and even with a significant decreasing effect on the expression of the transmembrane protein serine 2 (TMPRSS2), one of the two main host proteins involved in the SARS-Cov2 infection process [33].

For carbonated TMWs, the results obtained confirm a behavior already known from previous studies that presented a tendency to reduce the abundance of faecal coliforms and that of faecal streptococci at high CO₂ concentrations in spring and borehole groundwater [34], a rapid numerical decrease of *E.coli* was also observed in the bottled carbonated mineral waters, due to the decrease in the pH value, induced by the formation of carbonic acid and the cellular effects caused by CO₂ [35].

The interdependent evolution of the abundance of coliform bacteria and that of the concentrations of nitrogen compounds in underground TMWs reconfirmed the effectiveness of using these indicators (fecal bacteria/nitrates) in the screening of micropollutants from affected water wells or boreholes of poor sanitation conditions [36] but also the need to regulate this ion and its metabolites (nitrite, nitric oxide and N-nitroso compounds) which have a potentially negative effect on health [37].

The identification of the numerical variations of coliform bacteria and *E.coli* depending on the degree of mineralization/salinity of the surface TMWs (natural lakes or bathing pools introduced into the spa circuit), reflects the dependence of the two variables, otherwise there are previous studies that show a negative correlation between the density of total and faecal coliforms with the salinity of recreational waters [38].

On the other hand, it is already known that there is a significant correlation between the rhythm of anthropogenic activities combined with precipitation and the level of total and faecal coliform groups [39]. The results obtained, which reflect the higher numerical densities of *E.coli* and coliform bacteria in the allochthonous microbiota of surface TMWs (therapeutic lakes and pools used for hydrotherapy), could therefore represent the effect induced by the combined intervention of several pollution factors (leakage of seepage water, spills of agricultural fertilizers, household waste or excrement of permanent or seasonal avifauna) that can affect this type of ecosystems at a given time.

The European directives in force regarding the quality of bathing water express the enumeration of bacteria indicator for faecal pollution, but not for the identification of sources of contamination, nor for the variation in the inactivation rates of potentially pathogenic microorganisms, in the ecological contexts of origin, which are dependent on factors such as temperature, solar radiation, salinity [40]. This aspect could however prevent a better management of the use of natural therapeutic factors, and implicitly, an effective protection of patients.

5. Conclusions

The intervention of faecal pollution indicator bacteria in the microbiota of mineral, underground or surface waters, which are used in the treatment bases of spa resorts, for curative or prophylactic purposes, was reflected differently by their physico-chemical characteristics.

Coliform bacteria (β -galactosidase+), showed a degree of survival modulated by the structural profile of TMWs, their viable count being statistically correlated with both physical parameters ($T^{\circ}C/p=0.0002/<0.0001$, $pH/p<0.0001/<0.0001$; $TDS/p=0.0038$; $EC/p=0.0007$) as well as with major chemical indicators (Mg^{2+} ($p=0.0010/0.0169$), Ca^{2+} ($p=0.0002$), Fe^{2+} ($p=0.0061$), Cl^{-} ($p=0.0015$), HCO_3^{-} ($p=0.0273/0.00012$), SO_4^{2-} ($p=0.0227$)).

The prevalence of *E.coli* bacteria was not sensitive to the oscillation of the values of the physical parameters ($T^{\circ}C$ $p=0.4054$; pH $p=0.1935$; TDS $p=0.6592$; EC $p=0.7093$) regardless of the origin of TMWs, but only to the variation of certain chemical indicators in the composition of lakes/bathing pools (Cl^{-} ($p=0.0302$), HCO_3^{-} ($p=0.0054$), and total mineralization ($p=0.00325$)), for which we obtained a positive correlation with its numerical density.

For sulfurous TMWs (with >1 mg/l titratable Sulfur), the results showed a negative correlation between H_2S concentration and the level of CB ($p=0.1886$) or *E.coli* ($p=0.5258$), and in carbonated TMWs (with ≥ 1000 mg/l CO_2), CO_2/CB values varied directly proportionally ($p=0.0287$), while *E.coli* evolved numerically independently of CO_2 concentration fluctuations ($p=0.9999$).

The degree to which microbiological contamination affects the bioproperties of TMWs, represents a problem that requires a permanent assessment throughout their route of adduction, transport and use in spa resorts. The moment when microbiological pollution occurs, its intensity, the type of polluting microorganism, the quantitative variations of the chemical elements involved in bacterial metabolism and/or those with therapeutic action, are factors that must be monitored and identified in order to preserve the chemical identity of TMWs and, implicitly their expected beneficial effect, but also minimize the human health risk of infections.

Author Contributions: Conceptualization, G.G. and L.G.; methodology, L.G. and B.M.; software, G.G. ; validation, G.G. L.G. and E.H.; formal analysis, G.G. and E.H.; investigation, L.G.; resources, L.G.; data curation, L.G.; writing—original draft preparation, G.G. and L.G.; writing—review and editing, V.L. and E.H.; visualization, G.G. and B.M.; supervision, L.G.; project administration L.G. G.G. and V.L. All authors have read and agreed to the published version of the manuscript.

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Data Availability Statement: The data presented in this study are available upon request from the corresponding author.

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