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Article

Photodegradation of a Broad-Spectrum Antibiotic Azithromycin Using H2O2 under UV Irradiation

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Abstract: The photodegradation of azithromycin present was carried out in water using H2O2 under UV irradiation. The reaction variables considered in this study were the amount of H2O2 solution and the initial concentration of azithromycin to evaluate the performance of the photodegradation process. The azithromycin degradation was not observed in the dark during stirring for 20 minutes. The study showed efficient photodegradation of azithromycin using H₂O₂ as an oxidant in the presence of UV irradiation. The azithromycin degradation was altered significantly by the pH of the irradiated solution. The degradation was found to be low at an acidic pH and showed an increasing trend as the pH changed to basic. The azithromycin degradation was increased with a higher amount (higher concentration) of H2O2. The degradation of azithromycin decreased with a higher concentration of azithromycin in the reacting solution. The highest degradation of AZT was achieved in 1 hour using a 1.0 ppm AZT solution containing 3 ml of H₂O₂. The experimental data obtained was well-fitted to zero-order reaction kinetics. The results of this study were found superior when compared with those reported in the literature, with photocatalysis using nanomaterials and photolysis using irradiation and H2O2. The UV/H2O2 system was found to be quite efficient for the photodegradation of azithromycin, and this system can be applied to degrade other organic pollutants present in industrial wastewater. To this point, such a study has not been reported specifically for AZT photodegradation.

Keywords: azithromycin; UV irradiation; photodegradation; HPLC determination; H2O2; zero-order reaction kinetics

1. Introduction

Researchers all over the world are trying to eliminate organic pollutants from wastewater using several techniques such as adsorption, photolysis, photocatalysis, electrochemical conversion, etc. These organic pollutants consist of dyes, pharmaceuticals, surfactants, etc. Organic dyes have been degraded in water by photocatalysis using different types of nanocomposites. For example, photocatalysts have been used to eliminate methylene [1,2], methyl orange and Eriochrome black T [3], reactive red 250 dye [4], methylene blue and methyl orange [5], rhodamine B and crystal violet [6], methyl orange and alizarin red [7], and congo red and methylene blue [8]. Similarly, pharmaceuticals present in wastewater have been eliminated by photocatalysis. Hussain et al. [9] used a photo-Fenton process using MnO2 and moving bed biofilm reactor treatment followed by an ozonation process and reported more than 90% removal rates for Ibuprofen and Ofloxacin. Likewise, photocatalysis has been used for the degradation and elimination of azithromycin (AZT) antibiotic that has been used successfully and widely in the treatment of COVID-19, as well as against several other infections. AZT belongs to a class of medications called macrolide antibiotics. Because of its wide use in the treatment of coronavirus conditions as well as in other infectious diseases, its concentration in wastewater has been quite significantly increased in recent years and thus poses health hazards for humans and other living organisms. One of the effective ways to remove AZT from the environment is photodegradation either through photolysis or photocatalytic degradation. Several investigators have used a variety of semiconductor nanomaterials to eliminate AZT and other related drugs. Over the years, several types of nanomaterial-based catalysts have been synthesized

and used under variable conditions of pH, and AZT concentration with or without oxidants. These nanomaterials contain titanium, iron, zinc, molybdenum, cobalt, and other metals mostly in oxide form alone or with other metals. Other research studies have been conducted using UV or visible irradiation only (photolysis) and have employed H₂O₂ as an oxidant. Some of these studies are reported here. Sharma et al [10] used Cu₂O-TiO₂ nanotubes for the photodegradation of AZT and achieved complete degradation within 90 min in a solution having 100 µg/mL AZT concentration and 1.5 g/L of nanotubes at pH 7 under visible light irradiation. Ling et al [11] synthesized g-C₃N₄/Fe-MCM-48 nanosize composite with a 3-D structure and used it for photocatalytic ozonation of AZT under simulated solar light. The degradation achieved was almost 100% within 11 min. Shajahan and AbuHaija [12] used CuFe₂O₄, binary CdS/CuFe₂O₄, and ternary g-C₃N₄/CdS/CuFe₂O₄ nanocomposite, for photodegradation of AZT using visible light source. The ternary nanocomposite exhibits the highest degradation efficiency of 85% within 90 min. Mehrdoost et al [13] utilized PAC/Fe/Ag/Zn nanocomposites for photodegradation by UV irradiation and achieved 99.5% degradation of AZT (10 mg/L) at pH 9 within 120 min irradiation time using 0.04 g/l of PAC/Fe/Ag/Zn. Mohammed et al [14] used ZnO/Co₃O₄ nanocomposites and reported 71.7% degradation of AZT within 80 minutes under visible light irradiation. Shukla et al [15] reported a 40% degradation of AZT (10 mg/L) using ZnO nanoparticles under UV irradiation. Kumar et al [16] reported 98.4% AZT photodegradation in 90 min under visible light using an Ag@Bi4OsI2/SPION@calcium alginate catalyst. Ospino-Atehortúa et al [17] used persulfate and simulated sunshine to photodegrade AZT. After 120 min of irradiation, the photodegradation of AZT was about 70%. Furthermore, it was revealed that AZT removal was enhanced at a higher pH (basic). Tenzin et al [18] utilized SrTiO₃ for the photodegradation of AZT under UV irradiation. The photodegradation at pH 2 was 55% within 150 min of irradiation, while at pH 12, it was 98% within 90 min. Li et al [19] used MoS₂-GO composites for photocatalytic degradation of AZT (100 mg/L) under visible light and reported 87% degradation within 3 hours. Rueda-Márquez et al [20] used natural solar irradiation with immobilized TiO2 and achieved degradation of AZT of more than 80% in urban wastewater effluents. Sayadi et al. [21] used 1 g/L of GO@Fe₃O₄/ZnO/SnO₂ nanocomposites and achieved 90% of AZT (30 mg/L) within 120 min of UV-C irradiation at pH 3. Cizmic et al [22] investigated the use of nanostructured TiO₂ film (comprised of pure anatase) and achieved degradation using UV-C irradiation at pH 10. Naraginti et al. [23] used nanorod ternary nanocomposite ZrO₂/Ag@TiO₂ for effective AZT degradation and achieved 90% of AZT (20 mg/L) within 8h exposure to visible light radiation. Biancullo et al [24] investigated the efficiency of TiO2-based photocatalytic treatment using UVA-LEDs for wastewater samples spiked with AZT and reported complete degradation of AZT within 1 hour using 1 g/L of TiO₂. Luo et al [25] used La-TiO₂/AC nanocomposites for photocatalytic degradation of AZT under UV irradiation in wastewater and reported that at pH 4 and AZT concentration of 10 mg/L, irradiation for 90 min degraded 95.6% of AZT.

Some investigators have reported the use of photolysis using UV or visible irradiation with or without oxidants. Chen [26] performed photolysis of AZT under visible and UV light irradiation and concluded that UV irradiation provided higher degradation (82.6%) than visible light (53.1%) using an AZT concentration of 10 mg/L at pH 4. Cano et al [27] used simulated sunlight radiation and H₂O₂ to photodegrade AZT. The AZT degradation was almost complete after 120 min of photolysis. Voigt et al [28] achieved photoinduced degradation of AZT using UVC/VUV-irradiation. Tong et al [29] reported photolysis of AZT under simulated solar radiation.

The literature cited above has shown a variety of nanomaterials used for AZT elimination from water and the effect of H₂O₂ on the photodegradation of AZT. We have utilized in this study UV irradiation with H₂O₂ to perform AZT degradation without using any solid expensive catalytic nanomaterials. When H₂O₂ is exposed to light irradiation, photolysis of hydrogen peroxide occurs and produces hydroxyl radicals (•OH), which are highly reactive. These highly reactive hydroxyl radicals created during this process, oxidize and break down a variety of organic contaminants. The idea of using H2O2 as an oxidant and not using any expensive catalysts was also to economize the research study and to achieve the research goals with minimum chemicals, materials, effort, and time.

2

These features of this study are very important during the scaling up of this photodegradation process.

The parameters studied were the pH of the irradiated solution and variable concentrations of H₂O₂ and AZT in the solution. First, the pH of the solution was varied to determine the effect of pH on the degradation efficiency of AZT. Then keeping the optimized pH constant (at which maximum AZT degradation was achieved), the concentration of H₂O₂ was varied followed by varying the concentration of AZT. **Figure 1** shows the photos of AZT in solid powder form and its molecular structure.



Figure 1. The AZT in solid powder form (white color) along with its molecular structure.

2. Materials and Methods

2.1. Materials

The chemicals used in this study were 98.0% purity azithromycin dehydrate (molecular formula C₃₈H₇₂N₂O_{12•}2H₂O and molecular weight 785.0) and 30% hydrogen peroxide (H₂O₂) solution, purchased from Sigma Aldrich. NaOH and HCl were obtained from Alfa-Aesar and were used to maintain the pH of the irradiated solution. All chemicals used were of analytical grade. All the AZT-containing solutions were prepared in drinking water.

2.2. Methods

The AZT solutions of 1.0, 3.0, and 5.0 ppm concentration were prepared in drinking water. Fifty ml of AZT solution from each concentration was taken in separate glass beakers and a known amount of 30.0% H₂O₂ solution was added. The mixture was agitated for 20 minutes in the dark to check for AZT degradation with H₂O₂ in the dark. Then, the mixture was irradiated with UV light under continuous stirring. Figure 2 shows an aqueous solution containing AZT and H₂O₂ on a magnetic stirrer cum hot plate, which was irradiated with UV light having 500 W/m² intensity, 230 V input voltage, and 50/60 HZ. The external temperature of the sample was maintained at 40 °C during irradiation. After every 20 minutes, 1.0 ml of the solution was sampled from the beaker and analyzed for AZT concentration by the HPLC system. There was no need to separate the solid catalyst from the sampled solution using a centrifuge since no solid catalyst was used in the degradation process. The pH of the solution was adjusted using dilute solutions of NaOH and HCl. The separation of the components was accomplished on a reverse phase C₁₈ column (5.0 µm, 250.0 mm × 4.6 mm) installed in an HPLC analytical system (Waters Corporation) coupled with a triple quadrupole mass spectrometer. The mobile phase was optimized to achieve fast separation of the components of the solution at a flow rate of 1.5 mL/min. An isocratic methanol/buffer mobile phase at the ratio of 90:10 v/v provided the separation of the components [30]. The components of the solution were eluted from the column and converted into mass spectra, which were matched with the MS library for identification. Using this method, all irradiated solutions were HPLC chromatographed and the amounts of unconverted AZT present in the solutions were determined. The concentration of AZT present in the irradiated solutions was the unconverted AZT amount that provided the percent degradation of AZT. The parameters studied were, the pH of the solution (3, 6, and 9), the concentration of H₂O₂(1, 2, and 3 ml), and the concentration of AZT (1, 3, and 5 ppm). For determining

3

4

the effect of pH, the AZT concentration was 1 ppm, 60 min irradiation time, UV irradiation intensity 500 W/m^2 , and the amounts of H_2O_2 added were 1, 2, and 3 ml. Once the pH of the solution was optimized, it was kept and the concentration of H2O2 was varied followed by varying the concentration of AZT.



Figure 2. An aqueous solution containing AZT and H₂O₂ on a magnetic stirrer cum hot plate, irradiated with UV light having 500 W/m² intensity, input voltage 230 V, and 50/60 HZ.

3. Results and Discussion

3.1. Effect of Solution pH on the AZT Degradation Efficiency

Figure 3 shows the result of varying the pH of the irradiated solution on the percent photodegradation of AZT. The degradation was low at a pH of 3 (acidic). The degradation was increased upon increasing the pH to 6 and it was further increased at pH of 9 (basic). The complete degradation of AZT was achieved at pH 9 using a 50 ml solution having an AZT concentration of 1.0 ppm, added with 3 ml of 30% H₂O₂ solution, which was irradiated with UV light for 60 minutes. At higher pH in the basic region, the degradation of AZT increased due to the higher concentration of OH radicals produced which attacked the AZT molecules more frequently and converted them into degraded products and finally into CO₂ and H₂O [27].

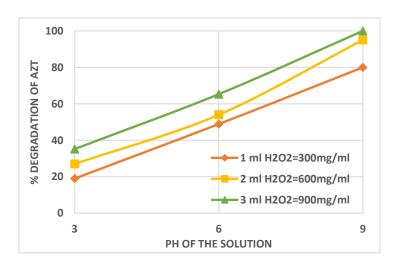


Figure 3. Percent degradation of AZT as a function of pH of the solution at different H₂O₂ concentrations. Experimental conditions: 50 ml solution of AZT 1.0 ppm concentration, 60 min irradiation time, UV irradiation intensity 500 W/m², 30%H₂O₂ solution added, 1, 2, and 3 ml.

3.1. Effect of Amount/Concentration of H₂O₂ in the Reacting Solution

The effect of the amount of 30%H₂O₂ (1, 2, and 3 ml) in the photocatalytic degradation of AZT was recorded. It was witnessed that the AZT degradation was increased with increasing amounts of H₂O₂ solution and with increasing UV irradiation time. **Figure 4** shows the percent degradation of AZT for variable amounts of H₂O₂ using three different initial AZT concentrations (1, 3, and 5 ppm). The reason was that a higher amount of H₂O₂ provided a higher amount of OH radicals generated due to the interaction with UV light irradiation. As a result, more and more AZT molecules were attacked by highly reactive OH radicals and experienced breakdown and thus the percent AZT photodegradation was higher.

3.2. Effect of Initial AZT Concentration on the AZT Degradation

The initial AZT concentration was varied from 1 to 3 and 5 ppm to observe the effect on the percent degradation of AZT. As shown in **Figure 4**, at a low AZT concentration of 1ppm (**Figure 4A**), the highest and complete degradation of AZT was observed. **Figure 4A** showed that the 3 ml H₂O₂ solution added provided the highest and complete (100%) degradation of AZT in a solution having 1ppm AZT, in 60 min of UV irradiation followed by 2 ml H₂O₂ solution, which provided 95% AZT degradation. The 1 ml H₂O₂ yielded 80% AZT degradation. The AZT percent degradation was low at 3 ppm initial AZT concentration which was further reduced in the case of 5 ppm AZT concentration (**Figure 4B and 4C**). The full breakdown of AZT was promoted greatly with a greater concentration of 30%H₂O₂ solution. Reaction solutions containing 1 and 3 ppm of AZT showed a greater rate of AZT degradation than those containing 5 ppm of AZT. Using 3 mL of 30%H₂O₂ solution, the maximum and highest deterioration trend was noted for the 1 ppm AZT solution.



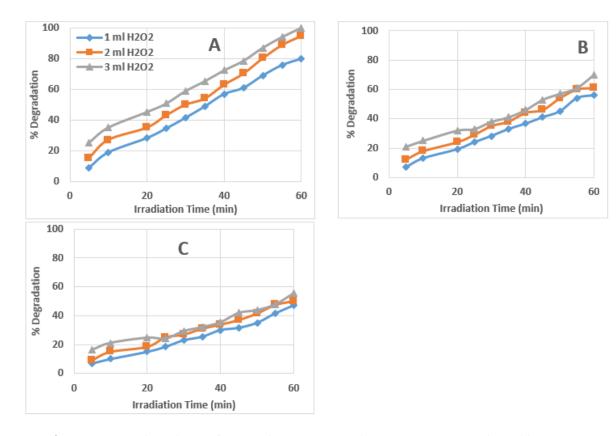


Figure 4. Percent degradation of AZT with respect to initial AZT concentration and variable amount of H₂O₂ in 50 ml of irradiated solution, **A**=1.0 ppm AZT solution, **B**=3.0 ppm AZT solution, **C**=5.0 ppm AZT solution. (Irradiated AZT solution volume 50 ml, pH 9, and UV irradiation intensity 500 W/m²).

3.3. Kinetics of AZT Photodegradation

To decide on the kinetics of the AZT photodegradation reaction, three plots were developed which are as follows: (1). C_t/C_o versus t (If the plot is linear, then it is a zero-order kinetic reaction), rate = k; (2). Ln C_t/C_o versus t (If linear, then it is a first-order kinetic reaction), rate = $k[C_t]$; (3). $1/C_t/C_o$ versus t (If linear, then it is a second-order kinetic reaction), rate = $k[C_t]^2$.

Among the three plots (**Figures 5, 6, and 7**), the plot of C_t/C_0 vs. t (**Figure 5**) yielded the best straight-line equation in which the slope of the equation represented the rate constant (k). This endorsed that the kinetics of the AZT photodegradation reaction using the UV/ H_2O_2 reaction system is zero order.



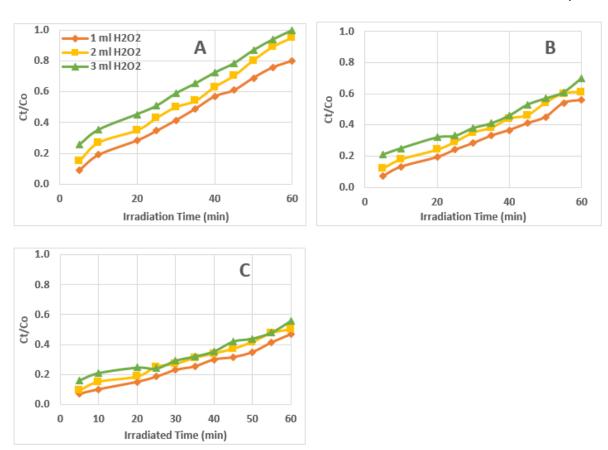


Figure 5. Plots of Ct/Co vs irradiation time (t) for the zero-order reaction kinetics. **A=**1.0 ppm AZT solution, **B=**3.0 ppm AZT solution, C=5.0 ppm AZT solution. (Irradiated solution volume 50 ml, pH 9, and UV irradiation intensity 500 W/m²).



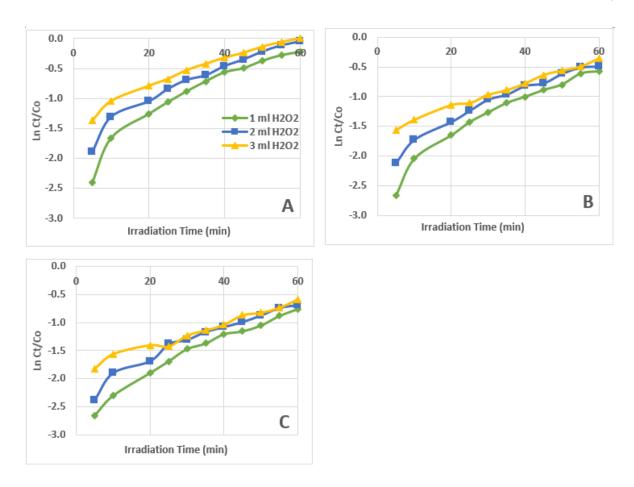


Figure 6. Plots of Ln (Ct/Co) vs irradiation time (t) for the first-order reaction kinetics. **A**=1.0 ppm AZT solution, **B**=3.0 ppm AZT solution, **C**=5.0 ppm AZT solution. (Irradiated solution volume 50 ml, pH 9, and UV irradiation intensity 500 W/m²).

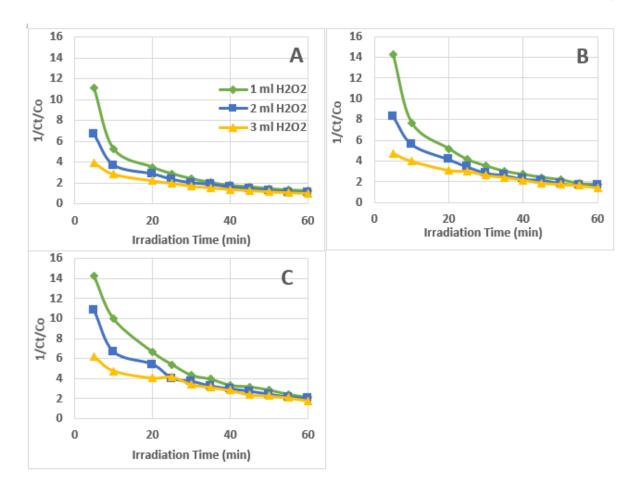


Figure 7. Plots of 1/(Ct/Co) vs irradiation time (t) for the second-order reaction kinetics. **A**=1.0 ppm AZT solution, **B**=3.0 ppm AZT solution, C=5.0 ppm AZT solution. (Irradiated AZT solution volume 50 ml, pH 9, and UV irradiation intensity 500 W/m²).

Table 1 displays the values of the correlation coefficient (R²) and the rate constant (k) obtained for the degradation of AZT under UV irradiation for the zero, first, and second orders of reaction. The correlation coefficient values indicate that the data fit well with the zero-order kinetics. However, the correlation coefficient values obtained for first-order kinetics are also close to zero-order ones. Here a point can be made that the reaction kinetics may be between the zero and first order.

Table 1. Correlation coefficient (R²) and rate constant (k) values for the degradation of AZT under UV irradiation for the zero, first, and second orders of reaction.

	Zero Order Kinetics (C_t/C_o) vs t		First Order Kinetics Ln (C₁/C₀) vs t		Second Order Kinetics $1/(C_t/C_o)$ vs t	
1 ppm AZT solution (50ml)	R ²	k, min ⁻¹	R ²	k, min ⁻¹	R ²	k, min ⁻¹
1 ml 30% H ₂ O ₂	0.9968	0.0130	0.9050	0.0352	0.6569	0.1304
2 ml 30% H ₂ O ₂	0.9912	0.0142	0.9418	0.0301	0.7418	0.0789
3 ml 30% H ₂ O ₂	0.9974	0.0134	0.9671	0.0235	0.8589	0.0460
3 ppm AZT solution (50ml)	R ²	k, min-1	R ²	k, min-1	R ²	k, min-1
1 ml 30% H ₂ O ₂	0.9925	0.0089	0.9308	0.0341	0.7107	0.1730
2 ml 30% H ₂ O ₂	0.9923	0.0092	0.9559	0.0282	0.8164	0.1028
3 ml 30% H ₂ O ₂	0.9760	0.0085	0.9936	0.0210	0.9491	0.0563
5 ppm AZT solution (50ml)	R ²	k, min ⁻¹	R ²	k, min ⁻¹	R ²	k, min ⁻¹
1 ml 30% H ₂ O ₂	0.9848	0.0070	0.9647	0.0325	0.8150	0.1878
2 ml 30% H ₂ O ₂	0.9908	0.0073	0.9462	0.0281	0.7872	0.1295
3 ml 30% H ₂ O ₂	0.9848	0.0070	0.9819	0.0210	0.9405	0.0716

R^2 = Correlation Coefficient, k = Rate Constant

3.4. Photodegradation Mechanism

The photodegradation of AZT with H₂O₂ under UV irradiation was supposed to proceed by the following reactions.

First, the H₂O₂ under UV irradiation produced OH radicals (OH•) which are quite reactive species and attacked the AZT molecules. This resulted in the fragmentation of AZT molecules into different species that are further oxidized and converted into intermediates and then finally into CO₂, H₂O, and NO_x. This reaction scheme advocates that the critical part is the generation of hydroxyl radicals from the photolysis of H₂O₂ that reacts with the AZT molecules. In addition, the O₂• generated also attacked the AZT molecules and degraded them into intermediates and finally into CO₂, H₂O, and NO_x. The scheme of AZT photodegradation using H₂O₂ under UV irradiation is illustrated in **Figure 8**.

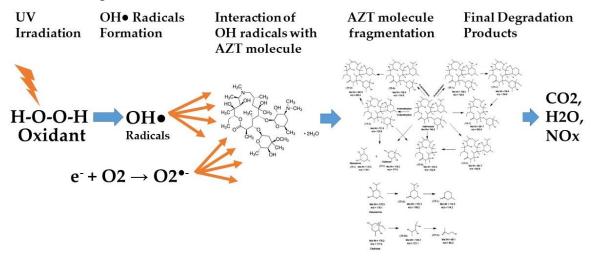


Figure 8. Scheme of AZT photodegradation using H₂O₂ under UV irradiation.

3.5. Comparison of this Work with the Literature

Table 2 presents a comparison of the photodegradation efficiency of this research work with the literature-reported results. The comparison is made with those results produced with and without catalysts. The results of this study were found quite excellent and showed 100% degradation in 1 hour when compared with those reported in the literature, both photocatalysis using nanomaterials as well as photolysis using light irradiation and/or H₂O₂.

Table 2. The photodegradation efficiency of this study compared with the literature-reported results.

No.	Catalyst	Process conditions	Degradation Efficiency	Reference
1.	10%Cu ₂ O/TiO ₂ nanotubes, 1.5 g/L	Visible light irradiation, AZT 100 μg/mL, pH 7	100% in 1.5h	[11]
2.	PAC/Fe/Ag/Zn, 0.04 g/L	UV irradiation, pH 9, AZT 10 mg/L, pseudo-first-order kinetic	99% in 2h	[13]
3.	Ag@Bi ₄ O ₅ I ₂ /SPION@calcium alginate	visible light (300W Xe lamp), AZT 10 mg/L, 0.3 mg/mL	98% in 1.5h	[16]
4.	K ₂ S ₂ O ₈ 5.0 to 80.0 mg/L	Simulated solar irradiation 30 min (1.5 KW xenon lamp, 290-800nm), 50.0 mL solution, AZT 1.0 mg/L, pH 5	70% in 2h	[17]
5.	SrTiO₃, 30 mg	UV irradiation, AZT 20 mg/L, pH 12, pseudo-first-order reaction kinetics	99% in 4h	[18]
6.	(a). MoS ₂ (b). MoS ₂ -GO	Visible light irradiation, AZT 100 mg/L	(a). 75% in 3h (b). 87% in 3h	[19]
7.	TiO ₂ P25, 227 ng/L (UV), 250 ng/L (solar)	UV irradiation 55 Wm ⁻² , Solar 33W m ⁻² , domestic wastewater	52% (UV), 87% (solar)	[20]
8.	GO@Fe ₃ O ₄ /ZnO/SnO ₂ , 1g/L,	UV-C Irradiation, AZT 30 mg/L, pH 3	90% in 2h	[21]
9.	ZrO ₂ /Ag@TiO ₂	Visible light irradiation, AZT 20 mg/L (50 ml solution)	90% in 8h	[23]
10.	La-TiO ₂ /active carbon	UV irradiation, pH 4, AZT 10 ppm	96% in 1.5h	[25
11.	Fe (III)-oxalate	UV irradiation, AZT 10 mg/L, pH 4,	83% in 2h	[26]
12.	35% H ₂ O ₂ , 482.0 ppm as oxidant	Simulated sunlight irradiation 500 W/m², AZT 1 ppm (50 ml solution), pH 9	100% in 2h	[27]
13.	No catalyst, only UV irradiation,	Xenon arc lamp (power 500 watts, UV power: 765 W/m², Wavelength 300–800 nm), AZT 1.0 μ g/L, pH 7, Temp. 40 °C	90% in 7days	[31]
14.	7.5 mg/L FeSO ₄ + 27.5 mg/L H ₂ O ₂	Simulated solar irradiation, UV power 50mW/cm², Wavelength 290–800 nm, AZT 3 mg/L, pH 3	92% in 0.5h	[32]
15.	Gd3+ doped BiVO ₄ , catalyst 2g/L	UV-LED irradiation, Wavelength 370 nm, power 4.65 mW/cm², domestic wastewater	63% in 3h	[33]
16.	No catalyst, only UV irradiation,	UV irradiation power 163 mW/cm², AZT 110 mg/L, pH 7. Temp. 65°C	73% in 0.8h	[34]
17.	Solar Photo-Fenton catalyst, Fe ²⁺ , 20 mg/L, H ₂ O ₂ 50 mg/L	Solar UV power 2.65 ± 0.68 mW/cm ² , Reactor volume $2x15$ L, AZT 25 ng/L, pH 7	24% in 3h	[35]
18.	3 ml of 30% H ₂ O ₂ solution	UV irradiation intensity 500 W/m², AZT 1.0 ppm (50 ml solution), pH 9.0, zero-order reaction kinetics	100% in 1h	This work

Conclusion

This study shows efficient photodegradation of AZT using H₂O₂ as an oxidant in the presence of UV irradiation. The effects of variable concentrations of AZT and H₂O₂ present in the irradiated solution have been demonstrated successfully. The degradation of the AZT was not observed in the dark in the presence of H₂O₂. The AZT degradation was affected drastically by the pH of the irradiated solution. The AZT degradation was found low at acidic pH and showed an increasing trend as the pH changed to basic. The highest AZT degradation was observed at pH 9. The percent degradation of the AZT was decreased with a higher concentration of AZT in the reacting solution. The degradation of the AZT was found to increase with a higher amount (higher concentration) of H₂O₂. The experimental data obtained was well fitted to the plot of C₁/C₀ vs t and thus verified that the AZT photodegradation with H₂O₂ under UV irradiation followed the zero-order reaction kinetics. The correlation coefficients obtained for the pseudo-first-order kinetics are also quite matching with the zero-order kinetics. It may be argued that the kinetics of the reaction could be somewhere in

12

between the zero-order and pseudo-first order. It is predicted that the OH radicals (OH•) and oxygen radicals (O2•) generated during the breakdown of H₂O₂ attack the MB molecules and produce MB radicals (MB•). After going through an oxidation process, these MB radicals are converted to intermediate fractions, and then finally produce CO2 and H2O. The results of this study were found quite better when compared with those reported in the literature, both using photocatalysis with nanomaterials as well as using photolysis using irradiation and H₂O₂. The UV/H₂O₂ system was found quite efficient for the photodegradation of AZT without the use of any solid catalyst and this system can be applied to degrade other pollutants and drugs present in the industrial wastewater. Such a study has not been reported so far for AZT photodegradation. Future studies are recommended to be conducted using continuously stirred reactors for AZT elimination from real wastewater containing other pollutants as well as using UV irradiation with different wavelengths and intensities. Other oxidants and activators can also be tested to see their effect on the degradation of AZT and other pollutants present in wastewater.

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