The Kerch Peninsula in Transition: A Comprehensive Analysis of Land Use and Land Cover Changes over Thirty Years

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Abstract: This study presents an in-depth analysis of land use and land cover change on the Kerch Peninsula over a period spanning three decades. Utilizing convolutional neural networks alongside satellite imagery analysis, we have mapped and quantified the changes in land use and cover, revealing significant trends and transformations within the peninsula's landscape. The research aims to elucidate the interplay between anthropogenic activities, climatic variations, and policy interventions in shaping land cover dynamics, thereby providing insights into the environmental and socio-economic impacts of these changes. Our findings indicate a marked increase in urban expansion at the expense of natural ecosystems, including forests and wetlands, underscoring the urgent need for sustainable land management strategies. The study highlights the role of agricultural intensification and expansion in altering the peninsula's ecological balance and emphasizes the critical importance of integrating change analysis and prediction. By leveraging advanced remote sensing and GIS technologies, our analysis not only enhances the understanding of the complex dynamics driving land use and land cover change but also showcases the potential of predictive modeling in forecasting future land cover scenarios. The implications of this study extend beyond the Kerch Peninsula, offering valuable lessons for managing land use changes and conserving natural landscapes in similar regions globally.

Keywords: land use and land cover change; Kerch Peninsula; convolutional neural networks; satellite imagery; urban expansion; sustainable land management

Introduction

Understanding the dynamics of Land use and land cover change (LULCC) is crucial for grasping the intricate relationship between human activities and environmental changes. This understanding not only illuminates the patterns and drivers behind changes in our landscapes and ecosystems but also underpins the development of sustainable land management, conservation strategies, and policy formulations necessary to navigate towards sustainable development [1]. The importance of LULCC lies in its capability to inform decision-making processes, offering insights into how to mitigate climate change impacts and promote environmental stewardship.

The study of LULCC models reveals their powerful role in explaining and predicting the interactions between human activities and environmental changes. These models, by delineating the causes and consequences of land cover dynamics, become indispensable tools for supporting better land management decisions [2]. Furthermore, in the context of urban development, LULCC modeling serves as a cornerstone in urban planning. It aids in integrating policy frameworks into the planning process, thereby facilitating sustainable urban expansion and management [3].

The advancements in remote sensing and GIS have significantly bolstered the assessment techniques for LULCC. These technologies offer enhanced accuracy and efficiency in monitoring land cover changes, thereby contributing to more effective environmental management practices [4]. Despite these advancements, implementing LULCC in models that contribute to climate change assessments presents challenges. These include the over-simplification of anthropogenic land-use activities and the pressing need for integrated modeling frameworks. Such frameworks are essential

for a nuanced understanding of land-use climate interactions, which in turn, could enhance the efficacy of climate change mitigation and adaptation strategies [5].

In the comprehensive analysis presented in our previous work [6], we delved into the intricacies of LULCC on the Kerch Peninsula over the last three decades. Utilizing advanced satellite imagery and spatial modeling techniques, including Landsat 5, 7, and 8 data, we employed convolutional neural networks (CNN) to achieve enhanced classification accuracy of various LULCC classes. This methodological innovation allowed for a detailed examination of the region's dynamic landscape transformations, revealing significant trends such as the notable increase in urbanized areas and the concurrent decline in natural ecosystems like forests and wetlands. These findings underscore the profound impact of sustained urban growth and agricultural expansion on the peninsula's ecological fabric, necessitating informed land management and conservation strategies to mitigate these effects.

Expanding upon this foundational work, our objective now evolves to include not only a nuanced analysis of LULCC over different periods but also to incorporate change analysis and prediction. This progression aims to dissect the underlying environmental and socioeconomic drivers catalyzing these land cover changes. By integrating additional layers of analysis, we endeavor to construct a more holistic understanding of the forces shaping the Kerch Peninsula's landscape.

The motivation behind broadening our research scope stems from the need to address the intricate web of factors influencing LULCC. It's increasingly clear that changes in land use and cover are not merely the result of natural ecological processes but are deeply intertwined with human activities and climatic shifts. For instance, our analysis indicated that the area of croplands has fluctuated significantly over the years, influenced by political dynamics, climate change, and urban development. Similarly, grasslands and barren lands have shown variable trends, reflecting the impact of human activity and climatic factors on these ecosystems.

Incorporating change analysis and prediction into our research will allow us to identify potential future trends in LULCC on the Kerch Peninsula, informed by both past patterns and current drivers. This approach is essential for developing effective land management policies and conservation strategies that can adapt to changing environmental conditions and socioeconomic needs. By understanding the likely directions of future land cover changes, stakeholders can better prepare for and mitigate the adverse impacts of these transformations on the region's ecological systems and human communities.

Our expanded research will employ a combination of quantitative methods, including advanced spatial modeling and predictive analytics, to analyze the interactions between environmental and socioeconomic drivers of LULCC. Through this comprehensive analysis, we aim to provide valuable insights into the sustainable planning and management of land resources on the Kerch Peninsula, contributing to the broader field of environmental science and policymaking. This endeavor not only builds upon our previous findings but also opens new avenues for understanding and addressing the complex challenges of land use and cover changes in a changing world.

The advancements in remote sensing and GIS have been pivotal in enhancing the study of LULCC, providing an array of tools and technologies to improve the accuracy, efficiency, and depth of analysis. These advancements are illustrated through a variety of examples and case studies that demonstrate their significant contributions to LULCC analysis.

Remote sensing technology has revolutionized the field of LULCC studies by providing valuable insights into environmental monitoring. The use of Landsat imagery has become increasingly popular due to its cost-effectiveness and availability. Piekkoontod [7] demonstrated the effectiveness of combining Landsat imagery with high-resolution imagery from unmanned aerial vehicles (UAVs) to classify LULCC classes with high accuracy. The integration of UAV data with Landsat imagery not only reduces costs but also provides immediate access to high-resolution data, making it a valuable tool for real-world applications. Li [8] proposed a hybrid approach that integrates Google Earth imagery with Landsat data to improve the accuracy of land cover mapping. By combining deep convolutional neural network-based classifiers with traditional machine learning classifiers, the study achieved a classification accuracy of 84.40%, surpassing previous state-of-the-art accuracy. This approach not only enhances the discrimination of land cover types but also reduces misclassifications

between certain vegetation types and improves the identification of impervious areas. Moon [9] evaluated the quality of land surface phenology (LSP) information derived from PlanetScope imagery by comparing it with Harmonized Landsat 8 and Sentinel-2 imagery. The study found high temporal correlation and agreement between LSP metrics derived from different data sources, highlighting the potential of integrating features of datasets at different resolutions to produce more reliable land cover maps. This comparison provides valuable insights into the quality and character of LSP metrics derived from different sources. Overall, the integration of Landsat imagery with other high-resolution datasets and advanced algorithms has significantly enhanced the accuracy and reliability of LULCC studies, providing valuable insights into environmental monitoring and land cover mapping.

The use of tools such as the MOLUSCE plugin in QGIS has allowed researchers to analyze spatiotemporal changes and predict future land use patterns [10]. Previous studies have shown that modeling LULCC can provide valuable insights into the rate of change and essential components of land use and land cover [11]. Additionally, the integration of tools like MOLUSCE with other techniques such as SVM and Sentinel-2 imagery has enabled researchers to analyze the effects of LULCC on various phenomena like flood simulation [12]. The MOLUSCE plugin has been widely used in research to analyze and predict changes in land use and land cover. For example, a study on the Pakhal Lake area in Telangana utilized the MOLUSCE plugin to detect LULCC and predict future changes in the landscape [13]. Similarly, cellular automata-based artificial neural network model incorporated the MOLUSCE plugin to assess changes in various LULC classes over time [14]. Furthermore, the MOLUSCE plugin has been instrumental in monitoring and simulating landscape changes over long periods. Studies have shown that by using MOLUSCE in QGIS, researchers can model decadal LULCC patterns and understand the expansion and contraction of different land cover classes [15]. This tool has also been used for predicting LULCC along the Boteti-Thamalakane river system, highlighting the importance of understanding land use changes in specific regions [16].

Machine learning techniques, including artificial neural networks and deep learning, have shown potential in accelerating and advancing LULCC studies [17]. The use of deep learning in LULCC studies has been particularly highlighted in a study conducted in Adama District, Ethiopia, where deep learning was employed for the first time to analyze LULCC patterns [1]. Furthermore, machine learning algorithms such as random forests, support vector machines, and deep neural networks have been applied in numerous studies for LULC/LULCC classification, showcasing the effectiveness of these methods in analyzing land cover changes [18]. Additionally, the integration of machine learning models with remote sensing data, such as Landsat satellite images, has enabled the mapping and analysis of LULCC patterns in various regions [19]. Overall, the combination of machine learning techniques and remote sensing data has provided new evaluation tools for studying LULCC, allowing for the efficient analysis of large datasets and the extraction of valuable information on land cover changes.

The impact of different driver variables LULCC has been widely explored. These variables are integral to understanding the dynamics of LULCC as they provide insights into the biophysical and anthropogenic factors driving these changes. The rationale for selecting these variables in LULCC studies is grounded in their ability to indicate changes in vegetation cover, urbanization levels, moisture content, and climatic conditions, which are crucial for assessing environmental impacts and guiding sustainable land management practices. NDVI is widely used to assess vegetation health and changes. Studies have shown a negative correlation between NDVI and land surface temperature (LST), indicating that areas with dense vegetation tend to have lower temperatures, highlighting the importance of vegetation in regulating urban climates and mitigating the urban heat island effect [20]. SMI and NDWI are crucial for monitoring moisture availability and water bodies, respectively. These indices help in understanding the hydrological changes resulting from LULCC, such as the reduction of water bodies due to urban expansion, which can significantly affect local climates and biodiversity [21]. NDBI is a key indicator of urbanization and built-up areas. It has been positively correlated with increases in LST, demonstrating the role of urbanization in contributing to higher temperatures and the urban heat island phenomenon [22]. Changes in species density directly reflect alterations in

habitat quality and biodiversity due to LULCC. Studies leveraging remote sensing data have linked habitat fragmentation and loss with LULCC, emphasizing the need to incorporate biodiversity considerations in land use planning [23]. Population data are critical in assessing anthropogenic pressures on land resources. Rapid population growth often leads to increased demand for housing and agricultural land, driving LULCC and affecting environmental sustainability [24]. Climate variables such as air temperature and precipitation are both drivers and consequences of LULCC. Changes in land cover, such as deforestation and urbanization, can alter local climate patterns, leading to variations in temperature and precipitation, which in turn influence land cover states [25].

Integrating diverse datasets and predictive modeling in LULCC analysis is expected to make significant contributions to our understanding and management of land cover changes, especially in regions like the Kerch Peninsula. This integration allows for a comprehensive analysis that combines satellite imagery, climate data, population statistics, and ecological information to uncover the multifaceted drivers of land cover change. It moves beyond simply observing changes to understanding why these changes occur, considering economic, policy, demographic, and climatic influences.

The use of predictive modeling, informed by this rich dataset, enables researchers and planners to forecast future land cover scenarios. This forward-looking capability is crucial for planning sustainable development, implementing effective conservation strategies, and mitigating adverse impacts on the environment and local communities. By anticipating how different factors might shape the landscape, stakeholders can make informed decisions that balance development needs with environmental preservation.

Furthermore, this approach can reveal novel insights into the specific conditions and processes driving land cover changes in the Kerch Peninsula. It supports evidence-based planning by providing a detailed, data-driven foundation for decision-making. This not only aids in addressing current environmental and socio-economic challenges but also in preparing for future changes. Ultimately, the integration of diverse datasets and predictive modeling stands to significantly enhance our ability to manage land resources sustainably, protect biodiversity, and support the well-being of local communities in the Kerch Peninsula and beyond.

These driver variables are selected in LULCC studies due to their comprehensive representation of the ecological, hydrological, and anthropogenic factors influencing land cover dynamics. By integrating these variables, researchers can better understand the complex interactions between human activities, vegetation cover, water resources, and climate, facilitating the development of strategies for sustainable land management and conservation.

2. Materials and Methods

2.1. Research Area

The Kerch Peninsula, located at the eastern edge of the Crimean Peninsula, presents a complex and dynamic landscape characterized by its unique geological, biological, and climatic attributes (Figure 1). This area has been shaped by a myriad of natural processes and, in recent times, has been significantly influenced by human activities. The peninsula serves as a pivotal region for both industrial and recreational development, offering diverse scenarios for future growth. Understanding the intricate balance between the peninsula's natural systems and its economic development is crucial for crafting sustainable management and conservation strategies [26].
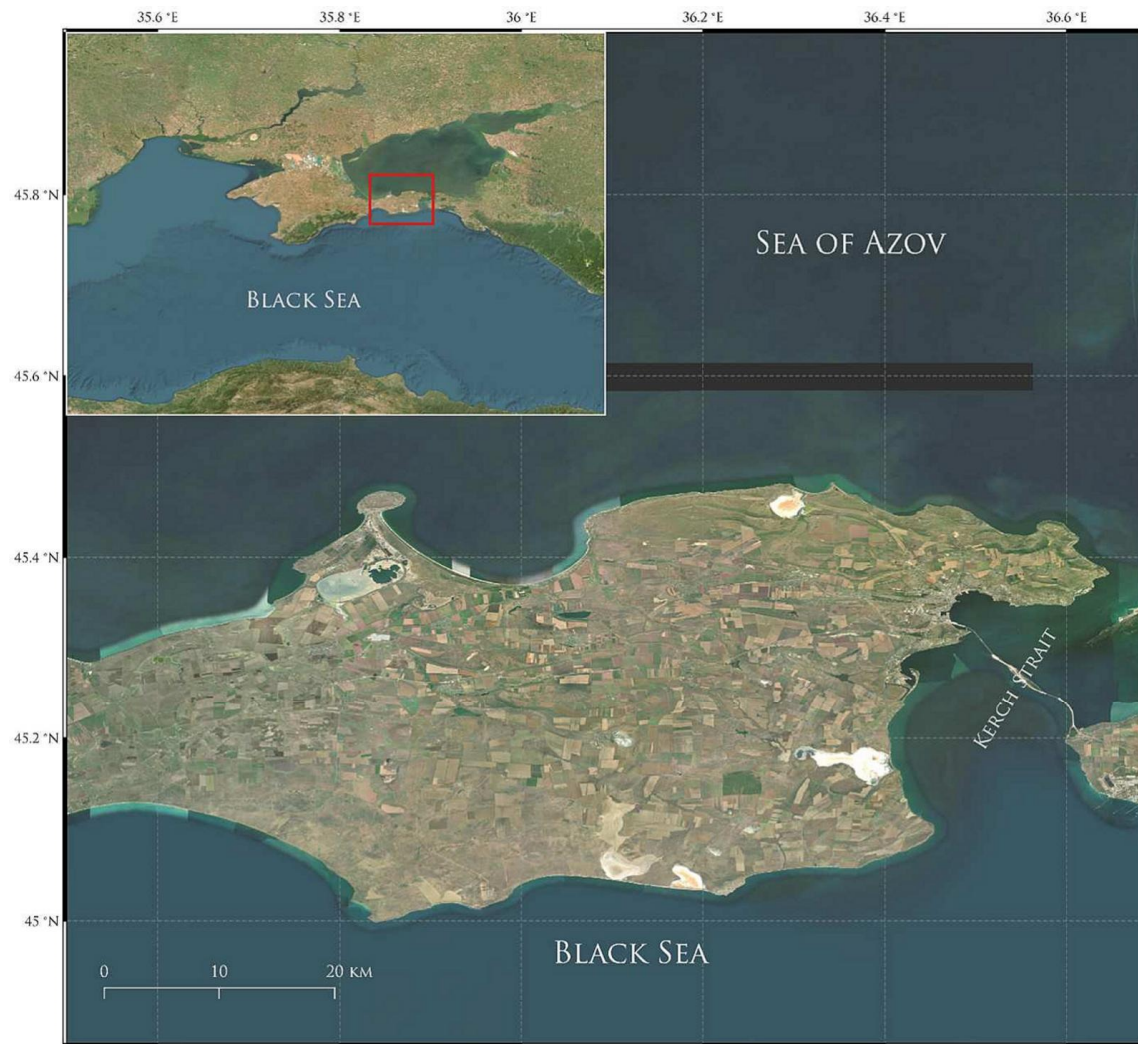
Figure 1. Schematic map of the research area.

Geologically, the Kerch Peninsula is distinguished by its stratified rock formations, which are the result of various geological epochs. These formations not only dictate the relief and soil types found across the region but also influence its mineral wealth, making it a focal point for mining activities, particularly for iron ore, limestone, clay, sand, and table salt [27]. The peninsula's relief is generally characterized by steppe hills, with a notable division into gentler southern and western parts and hillier northern and eastern sections. This varied topography plays a significant role in the distribution and diversity of local flora and fauna, contributing to the peninsula's ecological richness.

Climatically, the Kerch Peninsula experiences a dry, moderately hot continental climate, which, coupled with its strategic location between the Black Sea and the Sea of Azov, makes it an attractive destination for tourism and recreation [28]. The average annual air temperature and precipitation levels influence not only the natural vegetation and wildlife but also the human activities that can be sustainably supported in the region.

The region's water resources are under considerable strain due to both natural factors and human consumption. Water supply challenges, exacerbated by aging and often inadequate infrastructure, pose significant hurdles for both residential use and industrial activities [29]. Additionally, the peninsula's energy supply and waste management systems are areas of concern that require immediate attention to prevent adverse environmental impacts and ensure sustainable development [30].

Modern landscapes of the Kerch Peninsula are a testament to the interplay between its natural endowments and anthropogenic influences. The ongoing economic development, particularly with

the construction of the Kerch bridge, is poised to alter the region's logistical and economic landscape, further intensifying the anthropogenic pressures on the natural ecosystems. This evolving dynamic underscores the need for integrated landscape management approaches that consider both the natural and economic subsystems of the peninsula. Such strategies should aim to harmonize economic growth with the preservation of the peninsula's natural heritage, ensuring that development is both equitable and ecologically sustainable [31].

2.2. LULCC Analysis and Prediction Using Molusce

MOLUSCE operates on the premise that land use changes are driven by a combination of biophysical, environmental, and socioeconomic factors. The plugin employs various algorithms, including cellular automata (CA), artificial neural networks (ANN), and logistic regression, to simulate changes based on these drivers [11].

The general equation representing the change prediction model in MOLUSCE can be expressed as:

$$P_{i,j}(t + 1) = f\left(X_{i,j}, C_{i,j}(t)\right)$$

where $P_{i,j}(t + 1)$ is the probability of change in cell (i, j) at time $(t + 1)$, $X_{i,j}$ represents the set of driver variables for cell (i, j), $C_{i,j}(t)$ is the current land cover state of the cell at time (t). The function f embodies the predictive model derived from historical data analysis.

The first step in LULCC modeling with MOLUSCE is the preparation and importation of spatial datasets. For the Kerch Peninsula, this involves historical and current LULC maps and driver variables - NDVI, NDWI, NDBI, SMI, population density, species distribution, air temperature, and precipitation data formatted as raster layers.

Model calibration is crucial for ensuring accurate predictions. MOLUSCE allows users to perform this by analyzing the relationship between land cover changes and driver variables [32]. This involves statistical analysis to identify significant predictors and understand their influence on LULCC. The calibration process might employ logistic regression or ANN, depending on the model chosen.

With calibrated models, MOLUSCE can simulate future land cover changes. The plugin uses the CA algorithm for spatial simulations, where the state of each cell is determined by its current state, the states of its neighbors, and the influence of driver variables. The CA model is defined as:

$$L_{i,j}(t + 1) = g\left(L_{i,j}(t), N_{i,j}(t), D_{i,j}\right)$$

where $L_{i,j}(t + 1)$ is the land cover state of cell (i, j) at time $(t + 1)$, $L_{i,j}(t)$ is the current land cover state, $N_{i,j}(t)$ represents the states of neighboring cells, $D_{i,j}$ are the driver variables affecting the cell.

2.3. LULC Data

In this study, the collection of LULC data was fundamental for understanding the historical and current landscape configurations and for training the predictive models. The LULC data was obtained from Landsat 5, 7 and 8 satellite imagery, utilizing the thematic mapper sensor [6,33]. This sensor captures earth observation data in seven spectral bands, which include both visible and infrared spectra, crucial for distinguishing different land cover types.

The satellite images were sourced from the United States Geological Survey (USGS) archives, providing a temporal snapshot of the study area from 1990 to 2019. The Landsat program, known for its long-term global observations, was chosen due to its extensive historical archive and the ability to provide consistent, repeatable measurements of the Earth's surface, facilitating the analysis of LULC changes over time.

To process and analyze these images, the geemap Python library [34] was employed, which interfaces with the Google Earth Engine platform, enabling robust and scalable geospatial analysis. This approach allowed for the extraction of median composite images over the specified time frame, reducing cloud cover and atmospheric disturbances to produce clear and reliable LULC data.

Given the lack of comprehensive field study data for the period under investigation, ancillary data sources were utilized to supplement satellite imagery. These included archival cadastral data, topographic maps from the "Genshtab" for the years 1990-1991, and scholarly articles. This historical data provided essential context for the classification process, offering insights into land use patterns, and aiding in the interpretation of satellite-derived observations.

To classify the territory of the Kerch Peninsula, a sample of 12580 points was manually annotated based on the interpretation of high-resolution satellite data available on Google Earth Pro's Historical Imagery feature. These points were then categorized into distinct LULC classes, which included water bodies, urban areas, agricultural lands, and various vegetation covers, among others. The sample points were subsequently divided into a training set and a testing set, maintaining a 70:30 ratio, to ensure a rigorous validation of the machine learning models applied for LULC classification.

The preparation of this LULC dataset formed the foundational base for the change detection analysis and future LULC predictions, which are central to the understanding of landscape dynamics and the potential impacts of environmental and anthropogenic influences within the Kerch Peninsula.

2.4. Driver Variables

NDVI is a remote sensing index calculated from the visible and near-infrared light reflected by vegetation [35]. It's a measure of the difference between the near-infrared (which vegetation strongly reflects) and red light (which vegetation absorbs), scaled between -1 and 1. High NDVI values indicate healthy, dense vegetation, while low values signify sparse or non-vegetated areas. NDVI data are derived from the Landsat series of satellites, specifically Landsat 5, 7, and 8. These satellites have been capturing Earth's images for decades, providing multispectral images that are used to calculate NDVI. The methodology involves capturing reflected light in both the red and near-infrared spectrums, processing these data to compute NDVI for each pixel, thereby enabling detailed analysis of vegetation patterns over time. NDVI is pivotal in assessing vegetation health and cover, essential for understanding changes in land use, especially in agricultural expansion, deforestation, and urbanization [36]. It helps in monitoring changes over time, indicating shifts in vegetation density that might be due to various land management practices or natural phenomena [37].

NDWI uses green and near-infrared light to maximize the presence of water bodies while minimizing the presence of vegetation and built-up areas [38]. It's particularly effective in highlighting water features in the landscape, helping to differentiate them from other land cover types. Like NDVI, NDWI is calculated using the multispectral imagery provided by the Landsat satellites. The specific bands used for NDWI are designed to highlight the reflectance differences between water and other surfaces, enabling the identification and analysis of water body dynamics over time. NDWI is essential for identifying and monitoring water bodies within land cover types, which is crucial for managing water resources, studying wetland changes, and assessing the impacts of urbanization on hydrological systems [39,40].

NDBI leverages the reflective properties of built-up materials in the short-wave infrared spectrum to highlight urban areas [41]. It contrasts these areas with the surrounding land cover, making it a valuable tool for detecting and monitoring urban expansion. Using Landsat satellite imagery, NDBI is calculated by exploiting the spectral reflectance characteristics of built-up materials, which differ significantly from natural vegetation and water bodies. This index uses specific spectral bands that are sensitive to the physical properties of urban materials, facilitating the analysis of urban land cover changes [42]. NDBI is crucial for understanding urbanization patterns and the conversion of natural landscapes into built-up areas. It aids in tracking the growth of urban areas and infrastructure development, which are key aspects of land cover change [42].

SMI estimates the moisture content in the soil using remote sensing techniques. It's calculated based on the reflectance properties of the soil as captured by satellite imagery, indicating the level of soil moisture available for vegetation and crops [43]. SMI data is derived from the analysis of Landsat imagery, using algorithms that consider the reflective characteristics of wet and dry soils. This involves the comparison of different spectral bands to estimate the moisture content, providing a spatially explicit measure of soil moisture across landscapes. SMI provides insights into soil moisture

levels, essential for agricultural planning, drought assessment, and understanding the land-atmosphere interactions. It helps in identifying areas at risk of desertification or those experiencing changes in hydrological conditions due to land cover changes [44].

Global Human Settlement Layer (GHSL) provides detailed information on human settlements globally, derived from satellite imagery and census data [45]. It offers insights into the spatial distribution of population, urbanization patterns, and the growth of settlements over time. GHSL utilizes high-resolution satellite imagery combined with demographic and census data to map and quantify human settlements. This integrated approach allows for the creation of detailed, georeferenced population density maps, facilitating analysis of human impacts on land cover changes [46]. Population density and distribution data from GHSL are used to correlate human activity with changes in land use and cover. Understanding where people live in relation to land cover types is vital for assessing the impact of human settlements on land degradation, urbanization, and conservation areas [47].

Data on species distribution is crucial for assessing the impacts of land cover change on biodiversity, including habitat loss, fragmentation, and shifts in species distributions. This information helps in conservation planning and understanding the ecological consequences of land use changes [48]. GBIF aggregates data on the occurrence of species across the globe, including information on plants, animals, fungi, and microorganisms. It provides a comprehensive view of biodiversity by compiling records from observations, collections, and research findings. GBIF collects data from a wide range of sources, including museums, research institutions, and individual observations. This data is then standardized and made available through an online platform, offering access to species occurrence records and distribution maps [49].

Climate variables such as air temperature and precipitation are integral to understanding the climatic drivers of land cover change. They influence vegetation patterns, water availability, and soil conditions, thereby affecting land use practices and ecosystem health. ERA5 provides global atmospheric reanalyses of the past, offering hourly data on a wide range of climate variables, including air temperature and precipitation [50]. It represents state-of-the-art climate data, providing detailed information on weather and climate conditions. ERA5 data are generated by combining vast amounts of historical observations with modern forecast models, using a process known as data assimilation. This results in a globally complete and consistent dataset that captures the dynamics of the Earth's atmosphere, land, and ocean systems over time.

2.5. General Scheme of LULCC Analysis

The study commenced with the collection of baselines LULC data ($LULC_{t_0}$) and follow-up LULC data ($LULC_{t_1}$), which were obtained from previous studies [6,33]. These data sets were crucial for identifying the initial land cover state and subsequent changes (Figure 2).
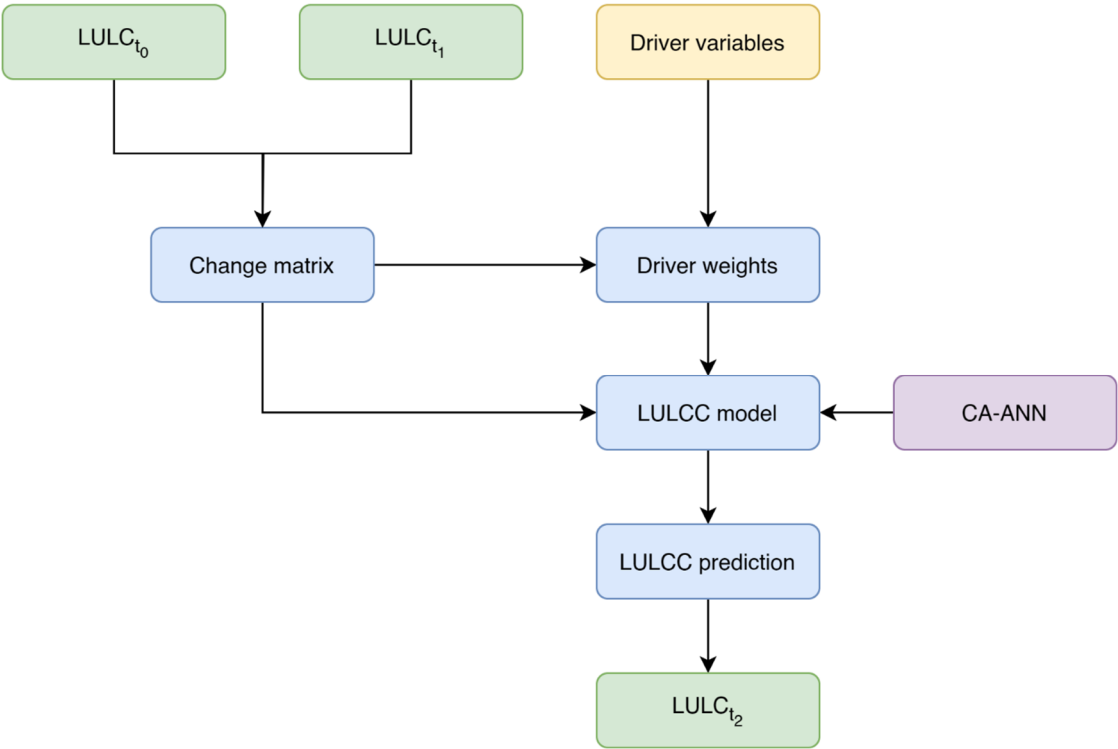
Figure 2. General scheme of LULCC analysis on Kerch peninsula.

Upon establishing the LULC states, we constructed a change matrix, which served as the core analytical framework for deciphering the transitions between different LULC types from time t_0 to t_1. The transitions within this matrix were instrumental in understanding the historical dynamics of land cover and setting the stage for predictive modeling.

Concurrently, we identified and integrated a set of driver variables believed to influence LULC dynamics. These included various environmental and socioeconomic factors such as vegetation indices, climate data, and demographic information. Each driver variable was assigned a weight based on its relative influence on LULC changes, determined through regression analysis and expert consultation.

With the change matrix and weighted driver variables at hand, we fed this information into a LULCC model, which synthesized the data to elucidate the complex interactions and quantify the impact of each driver on land cover changes. This model was designed to capture the nuances of the regional LULCC process and was essential for establishing a baseline understanding of the driving forces behind the observed changes.

To advance from analysis to prediction, we employed a CA-ANN model that integrates the robustness of cellular automata for spatial modeling with the predictive power of artificial neural networks. This hybrid model leveraged spatially explicit data from the change matrix and the quantitative force of the driver weights to forecast future LULCC scenarios.

The culmination of this pipeline was the generation of a predicted LULC state ($LULC_{t_2}$), representing the future configuration of the landscape under various potential scenarios informed by the modeled drivers. This predictive insight is pivotal for planning and conservation efforts, offering a foresight into the potential trajectories of the Kerch Peninsula's landscape evolution. Through rigorous validation against additional ground-truthed data and iterative model refinement, the LULCC prediction aimed to achieve a high degree of accuracy and reliability, thus serving as a valuable tool for decision-makers and stakeholders involved in land management and environmental protection.

At the outset of an investigation into the predictive modeling of LULCC using the CA-ANN approach, it is paramount to establish the foundational assumptions upon which such models are

constructed. The predictive capacity of these models hinges on a wide array of driver variables, ranging from demographic shifts and economic trends to environmental factors and policy frameworks. These variables are woven into the model to project the trajectory of land cover changes, providing a scenario-based forecast grounded in the current understanding of these driving forces. The fidelity of these models to actual outcomes is contingent upon the consistent influence of the included drivers. When predictions align closely with observed LULC changes, it validates the model's assumptions and the continued influence of the drivers as they were understood at the time of the model's creation. This congruence reinforces the utility of such models in planning and decision-making processes. However, when there is a divergence between the predicted and the actual LULC maps, it points to the intervention of factors outside the model's scope or changes in the behavior of the variables involved. Such deviations necessitate a thorough analysis to unearth the underlying causes, which may include unforeseen policy decisions, shifts in economic conditions, unexpected human activities, or novel technological advancements that alter land use patterns. This discrepancy between expectation and reality is not merely a reflection of the model's limitations but is also indicative of the dynamic and often unpredictable nature of human-environment interactions.

3. Results

The latter part of the 20th century, particularly the decade from 1990 to 1999, saw the Kerch Peninsula undergo significant shifts in LULC, which can be attributed to a complex interplay of socio-economic transformations and climatic variability. The disintegration of the Soviet Union ushered in a period of economic transition, marked by industrial decline and the contraction of arable land, which led to profound alterations in the landscape. From a socio-economic perspective, the 1990s were characterized by a de-intensification of agriculture due to economic downturns. The arable land in Crimea by the mid-1990s had diminished to 1154 thousand hectares, which reflects the broader trend of agricultural land abandonment witnessed across many post-Soviet landscapes. Such transitions in land use have often led to land cover changes, including the succession of previously cultivated areas into early-seral stages or grasslands, which in turn may evolve into shrublands or forests over time if left uncultivated. Concomitantly, climatic shifts became particularly influential. A warming trend evident since the late 1980s has led to increased winter and spring temperatures, significantly affecting the LULC. These warmer temperatures contributed to an increased incidence of late frosts, such as the catastrophic frosts of 1999, which obliterated burgeoning agricultural yields. These late frosts, paired with the previous warm spells, suggest an increased phenological vulnerability of the peninsula's vegetation, potentially leading to a more delicate and fluctuating agricultural landscape.

The transition matrix (Table 1) for the 1990-1999 period presents a quantitative synthesis of these dynamics, indicating the probabilities of land cover changes from one class to another. Notably, the matrix reveals substantial inertia within the croplands class, but also significant transitions to grasslands. This suggests that while certain agricultural areas remained stable, a notable portion reverted to grassland states, potentially due to abandonment or following practices. Urban expansion was another discernible trend, with transitions from high vegetation and barren lands to urban areas suggesting increasing urban sprawl and industrial land development.

Table 1. Transition matrix for the 1990-1999 period.

From/To	Croplands	Grass Lands	High Vegetation	Barren Lands	Shrublands	Bare Soils	Urban Areas	Water Bodies
Croplands	0.353	0.307	0.003	0.019	0.141	0.158	0.016	0.002
Grass Lands	0.213	0.324	0.001	0.005	0.290	0.160	0.006	<0.001
High Vegetation	0.339	0.244	0.105	0.012	0.103	0.193	0.004	0.001

Barren Lands	0.014	0.003	<0.001	0.436	0.004	0.008	0.057	0.478
Shrublands	0.259	0.300	0.029	0.003	0.247	0.159	0.002	<0.001
Bare Soils	0.301	0.317	0.001	0.008	0.225	0.145	0.003	0.001
Urban Areas	0.017	0.015	0.016	0.523	0.140	0.038	0.184	0.067
Water Bodies	0.003	0.001	<0.001	0.026	0.002	0.006	0.004	0.959

Class dynamics presents changes in land cover from 1990 to 1999 in the study area.

The data of class dynamics (Table 2) underscores a substantial increase in croplands of 8.98% over the five-year period, suggesting intensified agricultural activity. Conversely, there was a sharp decline in high vegetation of 2.26%, which could be due to deforestation or land development pressures. Shrublands saw a reduction by 5.14%, indicating potential land degradation or conversion to other uses. Urban areas expanded by 0.36%, which is consistent with global trends of urbanization. Water bodies also increased slightly, perhaps due to changes in hydrological management or natural fluctuations.

The LULC maps for the 1990-1994 and 1995-1999 (Figure 3) periods reveal a spatial distribution dominated by expansive cropland areas, especially in the central regions of the peninsula, indicative of an agriculturally driven landscape. The interface between croplands and grasslands likely delineates areas in transitional states, either set aside from cultivation or evolving into higher-successional stages due to reduced agricultural activity.

Table 2. LULC dynamics on the Kerch peninsula in 1990-1999.

Class Name	1990-1994 (km²)	1995-1999 (km²)	Δ (km²)	1990-1994 (%)	1995-1999 (%)	Δ (%)
Croplands	493.39	753.76	+260.37	17.02	26.00	+8.98
Grass Lands	866.85	857.47	-9.38	29.91	29.58	-0.32
High Vegetation	102.29	36.79	-65.50	3.52	1.26	-2.26
Barren Lands	45.97	48.60	+2.63	1.58	1.67	+0.09
Shrublands	793.13	643.90	-149.23	27.36	22.21	-5.14
Bare Soils	508.28	437.26	-71.03	17.53	15.08	-2.45
Urban Areas	11.30	21.83	+10.52	0.39	0.75	+0.36
Water Bodies	76.99	98.59	+21.60	2.66	3.40	+0.74
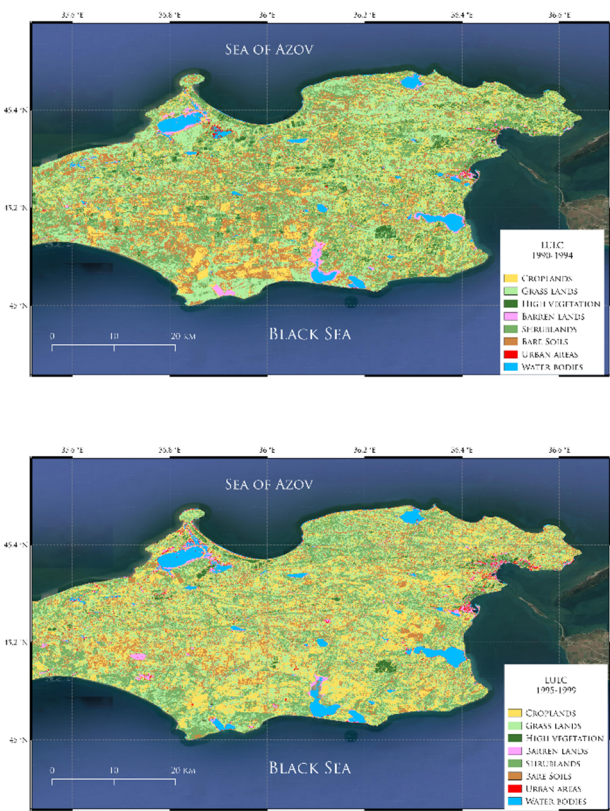
Figure 3. LULC maps of the Kerch peninsula for 1990-1994 and 1995-1999 periods.

The LULC changes observed during this period are not merely the outcome of natural successional processes but are also significantly shaped by human activities, including the strategic socio-economic decisions that influenced land management practices. The interrelation between these anthropogenic factors and the natural climatic variability produced a dynamic and evolving LULC pattern on the Kerch Peninsula. The region's LULC transformations, especially the shrinkage in high vegetation and the concomitant growth in urban areas, resonate with global trends but are uniquely modulated by the local socio-economic and environmental contexts. The shift in LULC on the Kerch Peninsula thus exemplifies the broader transitions occurring across the post-Soviet space, where ecological, economic, and climatic forces converge to reshape the landscape, reflecting the peninsula's adaptation to both its internal stressors and the broader post-Cold War restructuring.

During the period from 1995 to 2004, the Kerch Peninsula witnessed a complex mosaic of LULC transformations (Table 3). The substantial decline in croplands, as seen in the reduction of agricultural areas by 438.51 km², can be directly associated with the socio-economic turbulence of the post-Soviet transition.

Table 3. LULC dynamics on the Kerch peninsula in 1995-2004.

Class Name	1995-1999 (km²)	2000-2004 (km²)	Δ (km²)	1995-1999, %	2000-2004, %	Δ, %
Croplands	753.76	315.25	-438.51	26.00	10.87	-15.13
Grass Lands	857.47	1160.78	+303.31	29.58	40.04	+10.46
High Vegetation	36.79	153.66	+116.87	1.26	5.30	+4.04
Barren Lands	48.60	39.65	-8.95	1.67	1.36	-0.31

Shrublands	643.90	640.06	-3.84	22.21	22.08	-0.13
Bare Soils	437.26	488.43	+51.17	15.08	16.85	+1.77
Urban Areas	21.83	20.81	-1.02	0.75	0.72	-0.03
Water Bodies	98.59	80.08	-18.51	3.40	2.76	-0.64

Economic instability and shifts in agricultural policy led to land abandonment or a reorientation of land use, potentially coupled with a growing recognition of the importance of sustainable land management practices. In this context, the increase in grasslands by 303.31 km² may be interpreted as either a sign of agricultural land being set aside and gradually reverting to a more natural state or a deliberate policy shift towards extensive farming practices. The remarkable tripling of high vegetation areas by 116.87 km² points to a concerted effort towards afforestation or the natural succession of abandoned lands towards forested areas, a response perhaps to both economic incentives and environmental concerns. The slight increase in bare soils, despite being modest, raises questions about soil conservation and the long-term sustainability of land use on the peninsula. At the same time, the modest decrease in urban areas and water bodies could reflect a stagnation in urban development or a re-evaluation of hydrological infrastructure, possibly influenced by the changing economic priorities of the era.

These shifts are visually corroborated by the LULC maps of the respective periods (Figure 4), which exhibit a contraction of the green spaces associated with croplands and an expansion of grasslands and high vegetation zones. These spatial transformations highlight a period of transition for the Kerch Peninsula, with the reconfiguration of the agricultural landscape being a standout feature.
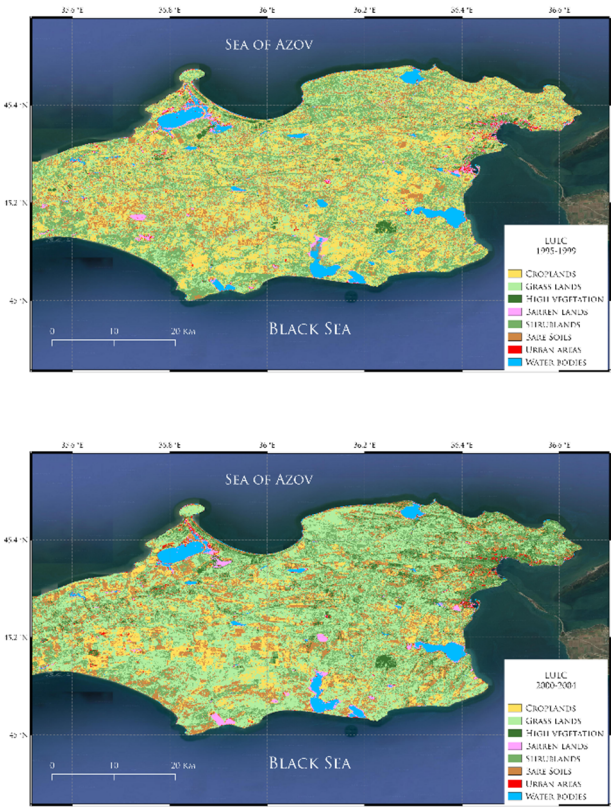
Figure 4. LULC maps of the Kerch peninsula for 1995-1999 and 2000-2004 periods.

The transition matrix for 1995-2004 (Table 4) shows that the transition from croplands to grasslands and bare soils underscores the dynamic changes within the agricultural sector. Concurrently, the robustness of grasslands, as evidenced by their high probability of remaining

unchanged, could indicate a landscape in equilibrium or a stabilization of land use practices that favor grassland conservation or extensification.

Table 4. Transition matrix for the 1995-2004 period.

From/To	Croplands	Grass Lands	High Vegetation	Barren Lands	Shrublands	Bare Soils	Urban Areas	Water Bodies
Croplands	0.180	0.365	0.050	0.001	0.163	0.239	0.002	<0.001
Grass Lands	0.102	0.456	0.041	<0.001	0.235	0.164	<0.001	<0.001
High Vegetation	0.003	0.037	0.582	<0.001	0.368	0.011	<0.001	0.000
Barren Lands	0.101	0.016	0.139	0.319	0.019	0.118	0.211	0.077
Shrublands	0.047	0.490	0.026	0.001	0.327	0.109	<0.001	<0.001
Bare Soils	0.121	0.403	0.073	0.001	0.208	0.191	0.003	<0.001
Urban Areas	0.154	0.026	0.149	0.056	0.011	0.313	0.288	0.003
Water Bodies	0.002	0.001	0.004	0.211	0.001	0.007	0.005	0.771

High vegetation's stability suggests a landscape recovering from previous uses, possibly benefiting from less human intervention and more protective environmental policies. The transition of barren lands into urban areas and water bodies could be attributed to both the expansion of urban frontiers into less fertile lands and efforts to manage water resources more effectively, considering the ecological and economic value of these areas.

The LULC dynamics on the Kerch Peninsula from 2000 to 2009 (Table 5) marked by significant ecological shifts and human activities, saw notable alterations in land use patterns, underscored by quantitative changes (Table).

Table 5. LULC dynamics on the Kerch peninsula in 2000-2009.

Class	2000-2004 (km²)	2005-2009 (km²)	Δ (km²)	2000-2004 (%)	2005-2009 (%)	Δ (%)
Croplands	315.25	289.35	-25.90	10.87%	9.97%	-0.89%
Grass Lands	1160.78	1416.13	+255.35	40.01%	48.81%	+8.80%
High Vegetation	153.66	69.55	-84.12	5.29%	2.40%	-2.89%
Barren Lands	40.03	70.99	+30.97	1.38%	2.44%	+1.06%
Shrublands	640.07	552.86	-87.20	22.07%	19.09%	-3.00%
Bare Soils	488.43	388.52	-99.92	16.84%	13.39%	-3.44%
Urban Areas	20.82	21.08	+0.26	0.72%	0.73%	+0.01%
Water Bodies	81.70	92.25	+10.56	2.82%	3.18%	+0.36%

Croplands showed a considerable reduction, decreasing from 315.25 km² in 2000-2004 to 289.35 km² by 2005-2009, marking a decline of 25.90 km². This 0.89% decrease in the cropland area could be indicative of a shift away from intensive agriculture, possibly due to changing economic conditions that made certain agricultural practices less viable or due to policy-driven efforts towards more sustainable land use practices. The reduction in croplands also reflects the broader socio-economic transformations during this period, including the impact of climatic changes on agricultural productivity. Grass lands, conversely, saw an increase of 255.35 km², expanding from 1160.78 km² to 1416.13 km². This 8.80% increase highlights a significant transition, likely driven by the abandonment of agricultural lands, which subsequently reverted to natural vegetation, or by deliberate changes in land management practices favoring the conservation of grassland ecosystems. The growth in grasslands suggests a landscape adapting to both the withdrawal of human agricultural activity and possibly benefiting from conservation policies aimed at preserving these ecosystems. High vegetation areas witnessed a sharp decrease of 84.12 km², plummeting from 153.66 km² to just 69.55 km². This decline points to deforestation or the conversion of forested areas for other uses, such as agriculture or urban development. The loss of high vegetation areas underscores the pressure on forest ecosystems during this decade, possibly exacerbated by the demand for timber, expansion of agricultural lands, and urban sprawl. Barren lands saw a marginal increase, rising by 30.97 km² from 40.03 km² to 70.99 km². This suggests a slight intensification of land degradation processes or the conversion of other land types into barren lands due to activities such as extraction or construction, highlighting concerns over soil health and land stability. Shrublands and bare soils also experienced changes, with shrublands decreasing by 87.20 km² and bare soils by 99.92 km². These shifts could reflect natural succession processes where shrublands evolve into other vegetation types or the impact of land use changes that either encourage or discourage the presence of these land covers. Urban areas remained relatively stable, with a slight increase of 0.26 km², reflecting gradual urban expansion in line with global urbanization trends, albeit at a slower rate. This stability suggests a balance between development pressures and efforts to conserve natural landscapes. Water bodies increased marginally by 10.56 km², possibly reflecting changes in hydrological management, climate-induced variations in water availability and efforts to enhance water conservation.

The LULC changes from 2000 to 2009, particularly the transition from croplands to grasslands and the decrease in high vegetation areas, can be attributed to a combination of factors, including climatic changes, economic shifts, and policy decisions affecting land use. The late frosts of 1999 and significant agricultural damage from 2000 to 2004 highlight the vulnerability of the region's agriculture to climatic extremes. Additionally, the increase in high temperatures and the delay in May frosts by 10-15 days due to warmer winter and spring months underscore the climatic challenges facing the region.

In the transition from 2000-2004 to 2005-2009 (Table 6), there was an increase in Grass Lands of about 8.80%, suggesting a continuing trend of cropland conversion to natural vegetation or a change in land management practices. This might be due to agricultural land abandonment or the transition towards more extensive forms of farming. High Vegetation areas saw a substantial decrease of almost 2.90%, which could indicate deforestation or conversion to other land use types, possibly for agricultural expansion or urban development.

Table 6. Transition matrix for the 2000-2009 period.

From/To	Croplands	Grass Lands	High Vegetation	Barren Lands	Shrublands	Bare Soils	Urban Areas	Water Bodies
Croplands	0.216	0.405	<0.001	0.050	0.123	0.205	<0.001	<0.001
Grass Lands	0.081	0.601	0.001	0.007	0.194	0.116	<0.001	<0.001
High Vegetation	0.037	0.328	0.184	0.009	0.291	0.104	0.047	0.001

Barren Lands	0.019	0.003	0.008	0.483	0.012	0.111	0.011	0.353
Shrublands	0.069	0.488	0.062	0.001	0.250	0.110	0.019	<0.001
Bare Soils	0.150	0.464	0.001	0.021	0.171	0.191	0.001	<0.001
Urban Areas	0.155	0.032	0.001	0.547	0.017	0.238	0.007	0.003
Water Bodies	<0.001	0.000	0.000	0.042	<0.001	0.003	0.001	0.953

Barren Lands increased slightly, possibly due to natural soil degradation processes or the conversion of other land types into barren land due to extraction activities or construction. Conversely, Bare Soils saw a decrease, indicating possible soil conservation efforts or land cover changes due to revegetation or development projects. Urban Areas remained relatively stable with a slight increase, reflecting a gradual urban expansion that is in line with the global trend of increasing urbanization, although at a lower rate. Water Bodies saw a small increase, possibly due to climatic variations affecting water levels or changes in water management practices.

Spatial analysis of the LULC maps for 2000-2004 and 2005-2009 (Figure 5) shows these changes. From 2000 to 2005, we observe a noticeable contraction of cropland areas. This reduction can be interpreted as either a response to lessened agricultural activity or a shift towards different land uses such as fallow land, which could later evolve into grass Lands.
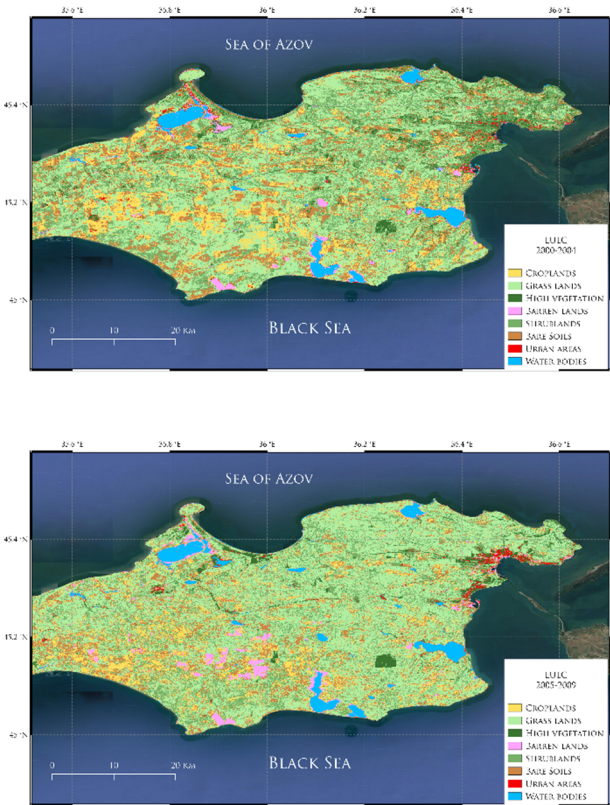
Figure 5. LULC maps of the Kerch peninsula for 2000-2004 and 2005-2009 periods.

The substantial increase in grass lands corroborates this theory, as it may indicate the natural progression of fields left uncultivated or the conscious conversion of land for conservation purposes. It's also possible that these areas represent a transition towards more extensive, low-intensity agricultural practices, reflecting a regional shift in farming strategies possibly driven by economic or

policy changes. High vegetation areas have markedly decreased, which could be attributed to both anthropogenic and natural factors. Deforestation for timber, clearing for agriculture, urban expansion, or natural disturbances such as fires could contribute to this decline. Meanwhile, an increase in grass lands and a decrease in bare soils could indicate a maturing of ecological succession processes, where grasses and shrubs begin to colonize previously bare or disturbed lands. The slight uptick in barren lands is intriguing. This could suggest new developments in areas that were previously unused or the degradation of existing land covers due to factors like construction activities, mining, or soil erosion. The stability of shrublands and the moderate increase in urban areas reflect a slow yet steady urbanization process, which may be linked to population growth and economic development on the peninsula. However, this urbanization does not appear to be at the expense of green spaces, as often seen in other regions undergoing rapid urban expansion. Water Bodies exhibit a small increase, which may reflect changes in water management, such as the creation of reservoirs, or natural variations in water levels. This stability is crucial for maintaining regional biodiversity and water resources for agricultural and human consumption.

In examining the LULC transitions, we delve into a period of continued environmental change and development on the Kerch Peninsula. The LULC class statistics (Table 7) and spatial analysis (Figure 6) of the 2005-2014 period offer a comprehensive overview of these changes, highlighted by a decrease in grass lands by 4.82%, signaling potential agricultural encroachment and a reevaluation of land management practices favoring the use of areas previously designated as grasslands for agricultural purposes. This period saw croplands expand by approximately 191 km², suggesting an intensification or expansion of agricultural activities, possibly driven by broader economic trends or policy incentives aimed at enhancing agricultural productivity.

Table 7. LULC dynamics on the Kerch peninsula in 2005-2014.

Class	2005-2009 (km²)	2010-2014 (km²)	Δ (km²)	2005-2009 (%)	2010-2014 (%)	Δ (%)
Croplands	289.52	480.63	+191.10	9.97%	16.57%	+6.59%
Grass Lands	1416.27	1276.56	-139.71	48.82%	44.00%	-4.82%
High Vegetation	69.55	83.54	+13.99	2.39%	2.88%	+0.48%
Barren Lands	71.08	47.29	-23.78	2.44%	1.63%	-0.82%
Shrublands	552.96	415.80	-137.16	19.05%	14.33%	-4.77%
Bare Soils	388.54	501.75	+113.21	13.39%	17.29%	+3.90%
Urban Areas	21.08	18.72	-2.36	0.73%	0.65%	-0.08%
Water Bodies	92.25	76.96	-15.29	3.18%	2.65%	-0.53%

Conversely, the observed reduction in grass lands from 1416.27 km² to 1276.56 km² might reflect a reallocation of land resources, where grasslands were converted to support agricultural expansion or subjected to urban development. The slight increase in high vegetation of 0.48%, although modest, signals a gradual but positive shift towards the recovery or planting of forested areas, possibly indicative of environmental conservation efforts or natural forest regeneration initiatives. This change, from 69.55 km² to 83.54 km², although not vast, is significant in the context of ecological restoration and land management practices on the peninsula.

The decline in barren lands by 0.82% and shrublands by 4.77% could be attributed to a variety of factors, including soil conservation initiatives that transform barren landscapes into more productive land covers or the natural succession processes where shrubs give way to other vegetation forms. This suggests an improvement in landscape management and a possible increase in land cover diversity, enhancing the ecological stability of the region.
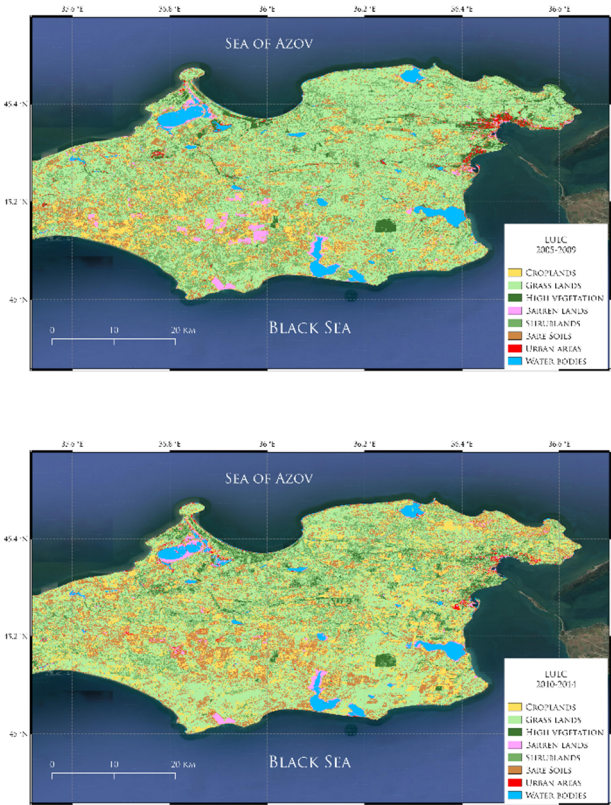
Figure 7. LULC maps of the Kerch peninsula for 2005-2009 and 2010-2014 periods.

Urban areas showed a slight decrease, which, within the global context of increasing urbanization, suggests a localized trend of urban consolidation or potential decay. This reduction, albeit slight from 21.08 km² to 18.72 km², could indicate a deliberate policy shift towards limiting urban sprawl, promoting urban greening initiatives, or perhaps demographic changes affecting urban growth patterns. Water bodies also experienced a decrease, which might be attributed to natural variations in precipitation and evaporation rates, changes in water management practices, or land reclamation efforts for agricultural or urban development. This slight reduction, from 92.25 km² to 76.96 km², underscores the challenges in maintaining water resources amidst changing land use patterns and climatic conditions.

The transition matrix (Table 8) further solidifies these observations. Croplands, once a dominant class, ebbed significantly, suggesting either a decline in agricultural exploitation or a transformative rest period leading to the resurgence of natural grasslands. This is indicated by the marked flow from croplands to grass lands, which could point to a fallow strategy in farming or perhaps a strategic shift towards extensive grazing practices or conservation initiatives that allow for ecological recovery. Grass lands, displaying a robust self-renewal, still surrendered a significant fraction to shrublands, underscoring a potential narrative of land abandonment or natural succession. This transition could also signify a landscape recovering from agricultural use, moving towards a more complex ecological state with increased biodiversity. Conversely, high vegetation, representing the dense arboreal tapestry, saw a notable fraction succumb to the advance of Shrublands, reflecting possible deforestation or the conversion of forested areas for other uses. Nonetheless, a significant part of this class showed resilience, persisting through the time span, potentially due to conservation efforts or less aggressive land-use policies. The slight rise in barren lands, areas often denuded of significant vegetation, indicated either a loss of productive land to degradation or an intentional setting aside of land for future use. Meanwhile, the gradual transformation of barren lands into bare soils might suggest soil erosion processes at play, or the reversion of previously used lands to a more primal state. Urban areas, which usually burgeon with the advance of time, portrayed a narrative twist as

they integrated back into natural land classes such as grass lands and high vegetation. This suggests a rare phenomenon of urban contraction or the greening of urban spaces, reflecting a community and policy-driven preference for urban sustainability and natural aesthetics. The water bodies, typically a static feature in land cover analyses, held their ground with little variation. However, a slight encroachment by barren lands hints at the impacts of intense agricultural practices, infrastructural developments, or the shifting patterns of water management.

Table 8. Transition matrix for the 2004-2014 period.

From/To (Class)	Croplands	Grass Lands	High Vegetation	Barren Lands	Shrublands	Bare Soils	Urban Areas	Water Bodies
Croplands	0.232	0.400	0.007	0.008	0.073	0.272	0.006	0.001
Grass Lands	0.165	0.500	0.009	0.001	0.162	0.161	<0.001	<0.001
High Vegetation	0.016	0.072	0.455	0.002	0.417	0.031	0.004	0.002
Barren Lands	0.189	0.068	0.001	0.311	0.005	0.302	0.105	0.019
Shrublands	0.142	0.490	0.046	0.002	0.175	0.143	0.001	<0.001
Bare Soils	0.221	0.413	0.008	0.009	0.098	0.232	0.017	0.001
Urban Areas	0.009	0.482	0.354	0.005	0.073	0.052	0.023	0.002
Water Bodies	<0.001	0.006	0.001	0.172	<0.001	0.002	0.007	0.812

From 2010 to 2014, and then to 2015-2019, the Kerch Peninsula's landscape was subject to significant changes coinciding with Crimea's political transition and the subsequent shutdown of the North Crimean Canal. This period, while marked by challenges, also presented an opportunity to reflect on positive adaptations and resilience within the region's land use and management.

Despite the constriction in water resources post-2014, which significantly impacted agricultural practices, croplands experienced only a slight decrease. This suggests an adaptation towards more water-efficient agricultural methods or a shift to crops less dependent on large water inputs. Meanwhile, the area under grass lands remained remarkably stable, indicating the robustness of these ecosystems or a potential increase in rain-fed pasturelands compensating for the reduction in irrigated fields (Table 9).

Table 9. LULC dynamics on the Kerch peninsula in 2010-2019.

Class	2010-2014 (km²)	2015-2019 (km²)	Δ (km²)	2010-2014 (%)	2015-2019 (%)	Δ (%)
Croplands	480.59	262.46	-218.13	16.57%	9.05%	-7.52%
Grass Lands	1276.55	1270.07	-6.48	44.01%	43.79%	-0.22%
High Vegetation	83.54	23.38	-60.16	2.88%	0.81%	-2.07%
Barren Lands	45.35	71.39	+26.04	1.56%	2.46%	+0.89%
Shrublands	415.80	486.42	+70.62	14.33%	16.77%	+2.44%

Bare Soils	501.71	702.91	+201.20	17.30%	24.24%	+6.94%
Urban Areas	18.40	50.63	+32.24	0.63%	1.75%	+1.11%
Water Bodies	78.04	32.70	-45.34	2.69%	1.13%	-1.56%

The high vegetation class saw a notable reduction, possibly due to the re-purposing of land for agriculture to compensate for the loss of water-intensive crops or changes in forestry management. Yet, it is the resilience and slight expansion of barren lands that demonstrate a noteworthy adaptation to the new hydrological realities, possibly through the introduction of land management practices aimed at mitigating soil degradation and promoting conservation. Shrublands and bare soils show a dynamic interplay, with shrublands increasing, possibly due to natural succession processes that fill the gaps left by reduced agricultural use. Bare soils also increased, suggesting a response to the loss of irrigated lands, with these areas possibly being set aside for future development or reclamation projects. Urban areas expanded modestly, reflecting continued growth and development in the region, despite the broader geopolitical changes. The ability of urban areas to continue to develop in this period highlights the adaptive capacity of local governance and urban planning in the face of significant challenges. Water bodies experienced a decrease, likely a direct result of the canal shutdown. However, the region's response, potentially through improved water conservation and management strategies, underscores a commitment to overcoming the constraints posed by the new water resource dynamics.

The LULC maps for 2010-2014 and 2015-2019 (Figure 8) visually capture these changes. The 2010-2014 map shows a diversified landscape, with a rich tapestry of croplands and natural vegetation. In contrast, the 2015-2019 map reveals adjustments in land use, with increased barren and shrubland areas and a contraction in high vegetation zones, reflecting both the challenges and the adaptive strategies implemented in the wake of the canal closure.
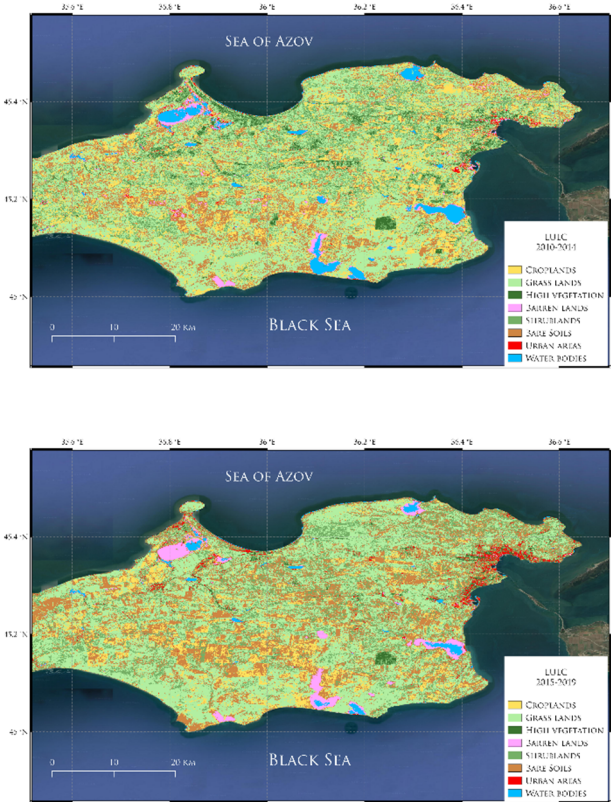
Figure 8. LULC maps of the Kerch peninsula for 2010-2014 and 2015-2019 periods.

The transition matrix for this period (Table 9) reveals the probabilities of each land cover type evolving into another, shedding light on the adaptive strategies in place. The matrix showed that, while some croplands converted to grass lands and bare soils, there was also a significant retention, reflecting the continued importance and resilience of agriculture in the region. Furthermore, the transition from grass lands to shrublands and from high vegetation to both shrublands and bare soils indicates a landscape in ecological flux, potentially moving towards a new equilibrium. Urban areas saw a positive increase, which might be linked to the strategic development of infrastructure and housing to support the region's population and economy.

Table 9. Transition matrix for the 2010-2019 period.

From/To (Class)	Croplands	Grass Lands	High Vegetation	Barren Lands	Shrublands	Bare Soils	Urban Areas	Water Bodies
Croplands	0.120	0.434	0.001	<0.001	0.168	0.273	0.004	0.000
Grass Lands	0.069	0.522	0.001	<0.001	0.167	0.227	0.014	<0.001
High Vegetation	0.029	0.200	0.145	0.001	0.333	0.158	0.133	<0.001
Barren Lands	0.079	0.022	0.001	0.567	0.029	0.239	0.046	0.016
Shrublands	0.033	0.493	0.023	<0.001	0.236	0.205	0.010	<0.001
Bare Soils	0.186	0.342	0.001	0.001	0.125	0.333	0.012	<0.001
Urban Areas	0.164	0.037	0.005	0.065	0.138	0.199	0.391	0.001
Water Bodies	0.004	0.002	<0.001	0.558	0.003	0.018	0.008	0.407

3.1. Landscape Metrics Analysis

Shannon diversity index. The Shannon diversity index for the Kerch Peninsula, as depicted in the graph (Figure 9), reveals a narrative of landscape transformation over a span of two decades, influenced by socio-political dynamics and environmental policies.

Figure 9. Dynamics of the Shannon diversity index for the Kerch peninsula in 1990-2019.

Starting in the mid-1990s, there's a discernible downtrend in the index, indicating a gradual homogenization of the landscape. During this time, agricultural intensification or the expansion of urban areas may have led to a simplification of the landscape, reducing the variety of land cover types. This decline reflects a broader trend where economic development often comes at the cost of ecological diversity. The most striking feature of the graph is the precipitous fall in the Shannon Index between 2005 and 2010. This period likely corresponds to substantial land-use changes such as the conversion of varied natural habitats to agriculture or urban sprawl. The change suggests a concentration of land use activities, possibly driven by policy shifts or economic imperatives that favor certain types of land cover over others. Post-2010, there is a recovery in the diversity index, peaking around 2015, which could signify efforts to diversify land use possibly through environmental conservation programs or a deliberate policy-driven approach to balance agricultural, urban, and natural spaces. This increase could be indicative of positive environmental stewardship where the region saw a resurgence of various natural habitats or the implementation of sustainable land management practices. However, the index takes a downturn once more in the latter half of the decade, pointing to another phase of landscape consolidation. This could be tied to the geopolitical changes in the region, such as the reunification with the Russian Federation, and the subsequent alterations in land management, including the impact of the North Crimean Canal shutdown. The resulting water scarcity would have undoubtedly affected land use, potentially leading to the abandonment of certain agricultural practices and a consequent reduction in landscape diversity.

Throughout these years, the fluctuating Shannon index underscores the Kerch Peninsula's sensitivity to both ecological shifts and human impacts. It suggests a landscape caught between the push of human economic activity and the pull of environmental conservation, with each force leaving its imprint on the diversity of the region's mosaic of land covers.

Contagion index. Observing the graph (Figure 10) representing the contagion index for the Kerch Peninsula, we can discern significant variations over the years, reflective of underlying land-use dynamics.
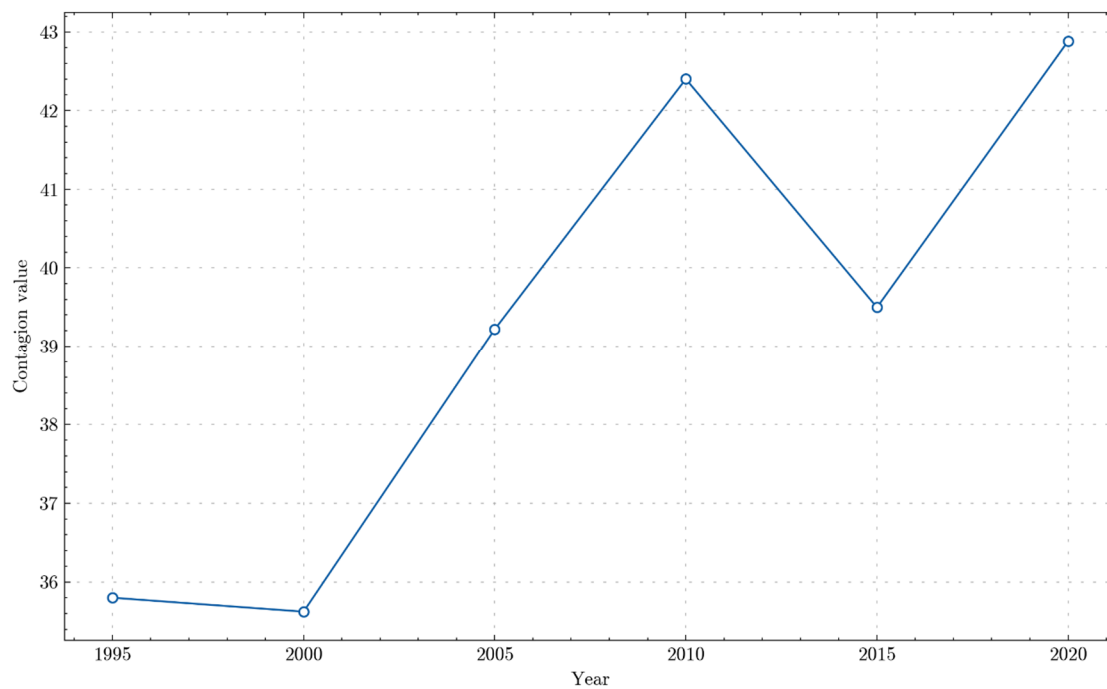
Figure 10. Dynamics of the contagion index for the Kerch peninsula in 1990-2019.

In the early phase, from 1995 to 2005, there's a steady increase in the contagion index. This progression suggests a trend toward larger, more contiguous patches of the same land cover type, likely indicative of agricultural expansion or the consolidation of natural areas. Such a trend could be associated with policy initiatives aimed at boosting agricultural efficiency or possibly the result of large-scale land management practices promoting homogeneous habitats, like forest plantations or nature reserves. Between 2005 and 2010, the contagion index peaked, which could signal the apex of land cover consolidation. At this point, the landscape could be characterized by extensive, uninterrupted expanses of certain land cover types, such as vast fields of a single crop type or large areas of natural vegetation, potentially reducing edge effects and habitat fragmentation. From 2010 to 2015, there is a dramatic dip in the index. This sharp decline might be attributed to a significant restructuring of the landscape, possibly due to the fragmentation of previous land cover types into smaller, more disparate patches. This could result from various factors such as land redistribution, changes in agricultural practices due to economic shifts, or the introduction of more diverse land use regulations that promote mixed-use landscapes. Interestingly, after 2015, the contagion index once again rises sharply, suggesting a reconsolidation of the land cover types. In the context of the Kerch Peninsula, this could be interpreted as a positive response to the challenges brought on by geopolitical changes and water shortages. The landscape might have witnessed a reorganization where agricultural lands were redefined, potentially leading to the emergence of larger, more water-efficient agricultural patches or the regrouping of natural habitats as a part of ecological restoration efforts.

The fluctuations in the contagion index across these three decades underscores the dynamism of the Kerch Peninsula's landscape. The shifts in the index capture the region's adaptability and resilience in the face of significant environmental and political transformations, reflecting an ongoing negotiation between human activity and the need for sustainable land management and environmental conservation.

Entropy. The entropy metric in landscape ecology is like the Shannon diversity index in that it measures the diversity of LULC types and their distribution across the landscape. However, entropy specifically gauges the randomness in the distribution of these patches. Higher entropy values indicate a more complex and diverse landscape configuration, where land cover types are more equally distributed and there's a higher level of disorder or randomness. Conversely, lower values

suggest a more ordered landscape with less complexity, possibly dominated by fewer land cover types.

Analyzing the entropy graph for the Kerch Peninsula from 1995 to around 2020 (Figure 11), we observe an initial gradual decline. This suggests the landscape was becoming less complex and more orderly, potentially due to the consolidation of land cover types. This might be the result of the expansion of monoculture in agriculture or the growth of uniform urban areas that reduce landscape complexity. The significant drop in entropy around 2005 indicates a rapid simplification of the landscape, which could correspond with intense land development or agricultural expansion that homogenized the land cover types. The uniformity implied by this drop could also be related to land policies or economic activities that promoted extensive areas of similar land uses, potentially reducing habitat heterogeneity, and increasing vulnerability to environmental changes. After 2010, there's a noticeable increase in entropy, reaching a peak around 2015. This suggests a reversal of the previous trend, with the landscape becoming more diverse and complex. The recovery in entropy could be related to changes in land management that encouraged a variety of land uses, such as the introduction of diverse crop types, the abandonment of less profitable agricultural land which then returned to natural states, or policies aimed at enhancing environmental conservation and land use planning. The subsequent decline post-2015 could reflect another period of simplification, potentially exacerbated by the geopolitical changes in the region, including the rejoin of Crimea to Russian Federation and the subsequent loss of water resources. The entropy decrease in this phase might indicate a transition towards less diverse land uses because of the new constraints, leading to more uniformity in the landscape.

Figure 11. Dynamics of the entropy index for the Kerch peninsula in 1990-2019.

Overall, the entropy trend highlights the adaptive response of the Kerch Peninsula's landscape to both internal management decisions and external geopolitical forces. The complexity of the landscape, as measured by entropy, offers insights into the resilience of the region's ecological and socio-economic systems in facing change and managing diversity.

Proportion of the landscape. The proportion of landscape (PLAND) metric quantifies the percentage of the landscape occupied by each LULC class (Table 10).

Table 10. Dynamics of the PLAND on the Kerch peninsula in 1990-2019 in %.

Class name	1990-1994	1995-1999	2000-2004	2005-2009	2010-2014	2015-2019
Croplands	17.00	26.00	10.86	9.98	16.56	9.05
Grass Lands	29.87	29.58	40.00	48.82	43.98	43.80
High Vegetation	3.53	1.27	5.29	2.40	2.88	0.81
Barren Lands	1.68	1.71	1.39	2.45	1.64	2.46
Shrublands	27.33	22.21	22.05	19.06	14.32	16.77
Bare Soils	17.52	15.08	16.83	13.39	17.29	24.24
Urban Areas	0.39	0.75	0.72	0.73	0.65	1.75
Water Bodies	2.67	3.41	2.86	3.18	2.70	1.13

In 1990-1994, croplands and grass lands were already prominent features of the landscape, with substantial coverage of shrublands as well. Notably, high vegetation, which represents forested areas, and bare Soils, also comprised a significant portion of the landscape. By the period 1995-1999, there was a noticeable increase in the proportion of croplands, which could be due to intensified agricultural practices or the conversion of other land cover types to agriculture. Conversely, high vegetation saw a decrease, possibly because of deforestation or land clearing for agriculture. In 2000-2004, a marked increase in the proportion of grass lands was observed, possibly reflecting a shift from intensive agricultural practices to more extensive, possibly grazing-oriented land use. This is consistent with the decreased proportion of croplands, suggesting some agricultural land might have been abandoned or converted back to natural states, increasing the heterogeneity of the landscape. By 2005-2009, grass lands dominated the landscape, perhaps a consequence of continued land-use changes or natural succession on abandoned agricultural lands. Meanwhile, croplands saw a further reduction, and urban areas remained a small but consistent feature, reflecting controlled urban growth. In 2010-2014, croplands increased again, reflecting a possible resurgence of agricultural activity, perhaps due to new agricultural policies or practices following the reunification. However, by 2015-2019, there is another reduction in croplands, alongside a significant increase in bare soils. This dramatic increase could be a direct consequence of the water shortages following the North Crimean Canal shutdown, leading to the abandonment of water-intensive crops and an increase in non-vegetated, potentially degraded lands. Throughout these years, the proportions of grass lands remain relatively stable, suggesting that grassland ecosystems have been less affected by changes in land use policies or environmental conditions. The continual presence of urban areas, although small, indicates ongoing but measured development within the region.

The fluctuating proportions of each LULC class across the 30-year span reflect the Kerch Peninsula's dynamic response to a combination of environmental challenges and regional policies. The landscape's adaptability is evident in the shifting PLAND values, showcasing how land use has evolved to meet changing socio-economic needs while also responding to environmental constraints.

Number of patches. The number of patches (NP) metric in landscape ecology provides an understanding of fragmentation or subdivision of each LULC class across a landscape. It indicates how many discrete spatial units or 'patches' exist for each class. Higher NP values suggest a more fragmented landscape with smaller, isolated patches of a given class, while lower values suggest larger, more contiguous patches (Table 11).

Table 11. Dynamics of the number of patches on the Kerch peninsula in 1990-2019.

Class name	1990-1994	1995-1999	2000-2004	2005-2009	2010-2014	2015-2019
Croplands	47831	47037	22842	33214	44249	19944
Grass Lands	44573	53443	26078	19135	26656	24728
High Vegetation	15038	2792	18034	3727	7333	2179

Barren Lands	2412	3774	2013	3918	3314	1498
Shrublands	34804	41173	41501	52831	30776	43533
Bare Soils	39807	42192	38257	47167	40987	45437
Urban Areas	2584	5227	4297	3392	4371	7750
Water Bodies	639	690	603	502	625	509

In 1990-1994, the Kerch Peninsula exhibited high NP values for croplands, grass lands, and bare soils, indicating a fragmented agricultural landscape and a patchy distribution of non-vegetated areas. Shrublands and high vegetation also showed significant fragmentation. From 1995-1999, there was an increase in the number of patches for most classes, especially notable for grass lands and bare soils. This could reflect changes in land management practices, leading to a more heterogeneously managed landscape or the subdivision of agricultural land. Notably, high vegetation patches decreased dramatically, which could be indicative of consolidation efforts or deforestation leading to fewer but larger forest patches. In 2000-2004, croplands patches decreased substantially, perhaps due to consolidation of farmland or a shift towards larger agricultural holdings. Meanwhile, high vegetation patches increased significantly, possibly due to reforestation initiatives or natural forest regeneration leading to more but smaller patches of forested areas. By 2005-2009, the number of patches increased for shrublands, bare soils, and particularly for croplands, indicating a reversal from previous consolidation trends, possibly due to land abandonment or the diversification of agricultural practices. In 2010-2014, the NP for croplands soared, reflecting a further fragmentation of agricultural lands. This could be due to the subdivision of land for different uses or the impact of policy changes. Urban areas also saw an increase in the NP, suggesting ongoing urban expansion and development into smaller discrete units. By 2015-2019, the number of patches for croplands decreased once again, indicating a possible reconsolidation of agricultural lands, perhaps as an adaptive response to changing environmental conditions such as water scarcity. Conversely, shrublands and bare soils saw their highest number of patches in the 30-year span, which may reflect land cover changes due to natural succession on abandoned lands or the spread of non-vegetated areas due to the water crisis.

Patch density (PD). In 1990-1994, the Kerch Peninsula had relatively high PD values for croplands and grass lands, and even higher for bare soils, indicating a highly fragmented landscape with numerous small patches. This could be a result of varied agricultural practices or the existence of many small natural areas within a matrix of other land uses (Table 12).

Table 12. Dynamics of the number of patches on the Kerch peninsula in 1990-2019.

Class name	1990-1994	1995-1999	2000-2004	2005-2009	2010-2014	2015-2019
Croplands	16.48	16.22	7.87	11.45	15.24	6.88
Grass Lands	15.36	18.43	8.99	6.60	9.18	8.53
High Vegetation	5.18	0.96	6.21	1.28	2.53	0.75
Barren Lands	0.83	1.30	0.69	1.35	1.14	0.52
Shrublands	11.99	14.20	14.30	18.21	10.60	15.01
Bare Soils	13.72	14.55	13.18	16.26	14.12	15.67
Urban Areas	0.89	1.80	1.48	1.17	1.51	2.67
Water Bodies	0.22	0.24	0.21	0.17	0.22	0.18

From 1995-1999, there's a slight decrease in PD for croplands, but an increase for grass lands and bare soils, suggesting a continued fragmentation of these classes. Interestingly, the PD for high vegetation drops significantly, possibly due to consolidation into larger forested areas or afforestation

efforts. In 2000-2004, croplands show a substantial decrease in PD, hinting at the consolidation of agricultural lands into larger patches, possibly due to changes in agricultural policies or practices. High vegetation sees an increase in PD, indicating more, but smaller, patches of forested areas, which could be due to natural regeneration or deliberate reforestation. By 2005-2009, croplands' PD rises again, which might reflect a subdivision of agricultural lands into smaller units. This period also shows a rise in PD for shrublands, suggesting an increase in fragmentation of these areas, which could be due to natural land cover changes or shifts in land use. In 2010-2014, the PD for croplands spiked, pointing to a significant fragmentation of agricultural land, which could be related to the diversification of crop types. However, by 2015-2019, the PD for croplands falls, suggesting a trend towards fewer, larger agricultural patches. Throughout these years, urban areas also show an increasing trend in PD, reflecting the ongoing development of urban spaces into many small, scattered units. The decrease in PD for high vegetation by 2015-2019 suggests possible deforestation or consolidation of forest patches.

3.3. LULC Prediction

Analyzing the Table 13 comparing predicted and actual areas of LULC for the Kerch Peninsula between 2000-2004, we observe interesting trends and discrepancies. The predicted areas for croplands were much larger than the actual, which could imply the model's drivers overestimated agricultural expansion or didn't fully account for factors limiting growth, like policy changes or market shifts. Grasslands, conversely, were underpredicted, possibly indicating a shift in land use to more extensive pastoral activities or natural reclamation of abandoned farmlands, which weren't captured by the model.

Table 13. Comparison of LULC class areas between actual and predicted scenarios in 2000-2004 period on Kerch Peninsula.

LULC Class	Predicted Area, km²	Actual Area, km²	Difference, km²
Croplands	681.48	315.25	+366.23
Grass Lands	945.19	1160.78	-215.59
High Vegetation	42.39	153.66	-111.27
Barren Lands	59.20	40.03	+19.17
Shrublands	651.65	640.07	+11.58
Bare Soils	383.33	488.43	-105.10
Urban Areas	27.12	20.82	+6.30
Water Bodies	108.88	81.70	+27.18

High vegetation areas also saw a shortfall in prediction, suggesting that conservation efforts or natural forest regeneration may have been more successful than anticipated, pointing to possible gaps in the model's environmental consideration. For barren lands, the slight overprediction hints at an expected degree of land degradation that may have been mitigated by effective land management or an underestimation of the landscape's natural resilience. Shrublands predictions were close to reality, showing that the model was well-tuned to the factors affecting this land cover type. Bare soils had a noticeable underprediction, which might reflect unanticipated erosion or other ecological processes exposing more soil than expected. Urban areas were slightly overestimated by the model, suggesting an optimistic forecast of urban sprawl or perhaps an overestimation of development pressures. Lastly, water bodies were overrepresented in the predictions, which could result from not fully considering changes in water management practices or the impacts of climatic variations on water levels.

The spatial analysis of the maps (Figure 12) reveals the nuanced interplay of land cover changes. Predicted croplands expansively blanket the landscape, suggesting an optimistic view of agricultural

spread that reality does not mirror. Actual grasslands, however, seem to creep into spaces where other land covers were expected, a testament to the land's natural succession or a shift in land management that eluded the model's foresight. Observing high vegetation, the actual map is more generous with its swathes of green than the predicted, hinting at successful forestation efforts or nature's unfettered reclamation of land. Conversely, the predicted map's barren lands paint a harsher picture than what came to pass, overestimating environmental degradation. Shrublands align closely between predicted and actual, indicating a well-calibrated aspect of the model. However, the reality of bare soils tells a story of greater exposure than expected, possibly reflecting activities that stripped the land more than the model could predict. Urban expansion is modestly overrepresented in the prediction, suggesting an anticipated growth that materialized less aggressively, possibly curbed by planning regulations or other societal factors. Finally, the predicted abundance of water bodies does not hold water when compared to the actual landscape, where such features are less prominent, likely altered by human intervention or ecological changes.
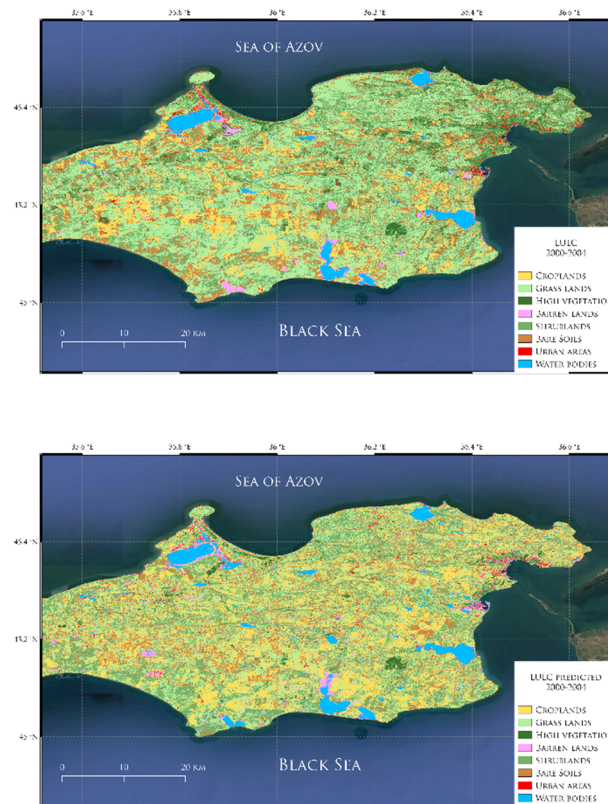
Figure 12. Actual and predicted LULC maps for 2000-2004 period.

The spatial distribution in the predicted of the predicted and actual LULC areas for the period 2005-2009 on the Kerch Peninsula map shows notable variations, implying that the underlying drivers used in the prediction may have failed to capture all the factors influencing land use changes.

Analyzing the table 14, the croplands were underpredicted by around 100 km², which may reflect an underestimation of agricultural development or failure to consider factors that would lead to its expansion during this period. In contrast, grass lands were overpredicted by approximately 175 km², possibly pointing to an overestimation of the transition from other LULC types to grass lands or an assumption about land management practices that did not occur as expected. High vegetation showed a surplus in the predicted area by about 57 km². This could mean the model predicted a larger scale of afforestation or natural vegetation growth than what was realized, or it could reflect the impacts of conservation efforts that were overestimated. The prediction for barren lands was less than the actual by around 32 km², which may indicate the model did not account for certain degradative processes affecting this LULC type. For shrublands, the model also underestimated the

area by nearly 69 km², which could suggest a prediction that didn't align with the natural progression or management of these lands. Bare soils showed a small underestimation in the model prediction by about 11 km², pointing to possible unexpected erosion or land clearing events. Urban areas were also underestimated by around 8 km², indicating the model may not have fully captured the extent of urban sprawl or development pressures. Water bodies were slightly overpredicted by the model by approximately 10 km², possibly due to not accounting for the impacts of climate variation or water management changes that could have reduced the extent of water bodies.

Table 14. Comparison of LULC class areas between actual and predicted scenarios in 2005-2009 period on Kerch Peninsula.

LULC Class	Predicted Area, km²	Actual Area, km²	Difference, km²
Croplands	189.13	289.52	-100.39
Grass Lands	1591.08	1416.27	+174.81
High Vegetation	126.65	69.55	+57.10
Barren Lands	38.89	71.08	-32.19
Shrublands	484.21	552.96	-68.75
Bare Soils	377.34	388.54	-11.20
Urban Areas	12.78	21.08	-8.30
Water Bodies	82.14	92.25	-10.11

The spatial analysis based on the maps for the 2005-2009 period (Figure 13) reveals a disparity in land cover distribution between the predicted and actual scenarios. In the predicted map, croplands do not extend as far as they do, pointing towards a conservative estimate of agricultural growth by the model. This suggests that actual developments in agricultural policy or market dynamics promoting cultivation were likely not fully integrated into the predictive framework. Grasslands, which sprawl over larger areas in the predicted map than in the actual one, suggest an overestimation of either the conversion of other lands into grasslands or the persistence of existing grasslands. It could reflect a misjudgment of either ecological succession or land management trends, perhaps assuming a continuity that was disrupted by changing land use practices. High vegetation is another category where prediction exceeds reality, indicating an expectation for more significant afforestation or natural regrowth than what transpired. This could hint at an overoptimistic view of conservation efforts or natural processes within the model's assumptions. Conversely, barren lands appear underrepresented in the prediction, revealing an oversight of factors leading to land degradation or a lack of accounting for the resilience of certain landscapes that remained productive or recovered to other land cover types. The prediction closely approaching actual urban areas might signify a nuanced understanding of urban development factors within the model, though it still slightly underestimates the actual urban sprawl, potentially missing some nuances of urbanization trends or regulatory changes that encouraged development. Water bodies display minor overestimation in the predicted map, which could mean that the model did not fully consider the hydrological alterations or changes in water management that could lead to a reduction in water bodies' extent.
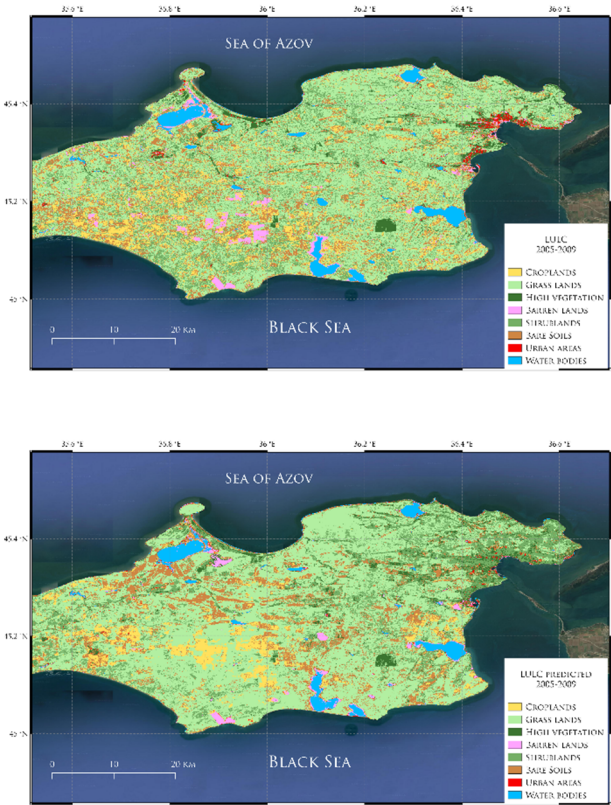
Figure 13. Actual and predicted LULC maps for 2005-2009 period.

The period from 2010 to 2014 for the Kerch Peninsula shows a distinct difference in land cover when comparing the predicted values with the actual ones. In the actual scenario, croplands expanded more than the model predicted, which may indicate economic or policy changes favoring agricultural development were not anticipated (Table 15).

Table 15. Comparison of LULC class areas between actual and predicted scenarios in 2010-2014 period on Kerch Peninsula.

LULC Class	Predicted Area, km²	Actual Area, km²	Difference, km²
Croplands	260.87	480.59	-219.72
Grass Lands	2040.24	1276.55	+763.69
High Vegetation	73.83	83.54	-9.71
Barren Lands	73.92	45.35	+28.57
Shrublands	225.41	415.80	-190.39
Bare Soils	112.08	501.71	-389.63
Urban Areas	19.50	18.40	+1.10
Water Bodies	95.40	78.04	+17.36

The model's overestimation of grassland areas could reflect an assumption about land conservation or usage that did not manifest as expected. The high vegetation predicted values were less than the actual, possibly pointing to effective environmental policies or natural regrowth that the model did not account for. Meanwhile, the predicted barren lands were over-calculated compared to the actual ones, perhaps due to an overestimation of land degradation or a lack of recovery factors included in the model. Shrublands and bare soils are both underrepresented in the prediction, hinting

at a transition between these land cover types that the model did not fully capture, or other influencing factors such as climate variability that were not considered. The predicted urban area is notably less than the actual, which may signify the model did not fully incorporate factors driving urban expansion. Finally, the model slightly overestimates water bodies, suggesting a misalignment with actual hydrological changes or water management strategies implemented during this period.

The actual LULC map for 2010-2014 reveals that croplands are more extensively distributed than predicted, suggesting that agricultural expansion has been more robust than anticipated (Figure 14). This could be a result of policy initiatives promoting agriculture or better-than-expected crop yields. The visual analysis suggests that these expansions are primarily at the expense of areas that were predicted to remain as grasslands and shrublands, indicating a possible shift in land-use priorities. Grasslands were predicted to occupy the largest area, but the actual map shows a lesser extent. This deviation could point to a conversion of grassland areas to agricultural uses, possibly due to increased demand for food production or changes in land management practices. The disparity might also reflect ecological transitions, such as grasslands developing into shrublands or other types due to climatic or human-induced changes not captured by the predictive model. The predicted map underestimates the actual spread of barren lands and bare soils. This may be due to unanticipated erosive processes, perhaps exacerbated by human activities like construction or natural factors such as less rainfall, leading to a loss of vegetative cover and an increase in barren lands. Urban areas show a slight increase in actual size compared to predictions. This could be attributed to an acceleration in urban development, potentially driven by economic growth that the model did not fully account for. This urban expansion often appears to encroach upon areas that were predicted to remain as agricultural or natural lands. Lastly, the predicted map overestimates the water bodies compared to the actual map, suggesting changes in water management or inaccuracies in modeling water dynamics. This could result from alterations in water usage, such as the construction of dams, or natural changes like drought conditions leading to a reduction in water bodies.
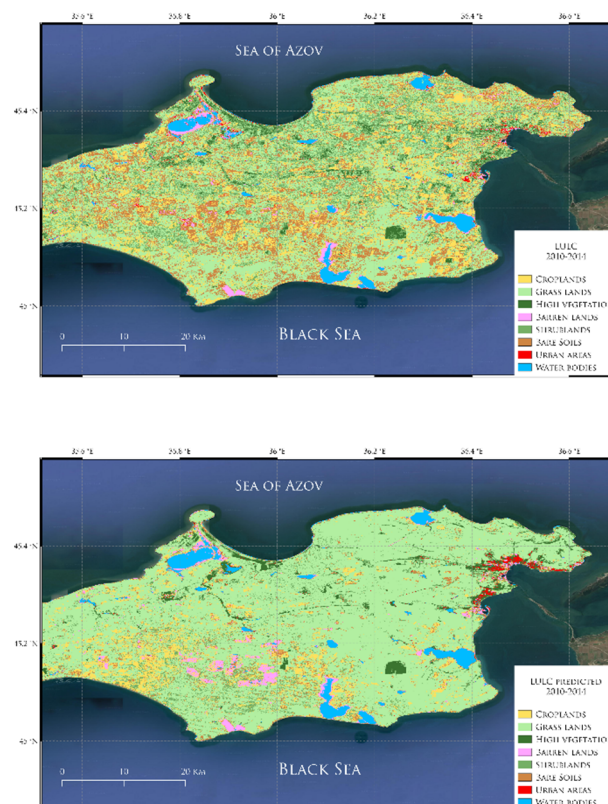
Figure 14. Actual and predicted LULC maps for 2010-2014 period.

For the 2015-2019 period, the analysis of the actual versus predicted LULC data in the Kerch Peninsula reveals that predicted area for croplands was lower than the actual area, indicating a considerable expansion of agricultural activity beyond what was anticipated. This might suggest that drivers such as policy changes favoring agriculture, shifts in crop demand, or even improved agricultural techniques could have played a role in accelerating cropland spread (Table 16).

Table 16. Comparison of LULC class areas between actual and predicted scenarios in 2015-2019 period on Kerch Peninsula.

LULC Class	Predicted Area, km²	Actual Area, km²	Difference, km²
Croplands	327.46	262.46	-65.00
Grass Lands	1665.97	1270.07	-395.90
High Vegetation	90.74	23.38	-67.36
Barren Lands	45.13	71.39	+26.26
Shrublands	203.78	486.42	+282.64
Bare Soils	481.21	702.91	+221.70
Urban Areas	12.99	50.63	+37.64
Water Bodies	75.46	32.70	-42.76

Grasslands were significantly overrepresented in the predictions, with the actual area being substantially smaller. This suggests that the model's assumptions about grassland persistence or expansion were not met, potentially due to these areas being repurposed for agriculture or urban development. High vegetation areas saw a dramatic reduction in the actual area compared to the predictions. This large discrepancy might be attributed to deforestation or conversion to other land uses, possibly for agricultural expansion or infrastructure development. The predicted areas for barren lands were less than the actual areas, suggesting that land degradation or natural processes like erosion could have been more prevalent than expected, leading to a loss of vegetative cover. For shrublands, the predicted areas were again smaller than the actual areas. This increase could be due to natural succession from grasslands to shrublands or from croplands lying fallow and returning to a more natural state. The model predicted more bare soils than were actually present, which could imply that soil conservation efforts have been more successful than anticipated, or that areas previously bare have been converted to other uses. Urban areas were predicted to be smaller than they actually were, pointing to a trend of urban expansion that exceeded model expectations. This could reflect economic growth leading to the expansion of existing urban areas or the creation of new ones. Lastly, the predicted area of water bodies was larger than the actual area, suggesting that water bodies have been reduced, possibly due to changes in climate patterns, increased water extraction for agricultural or urban use, or natural sedimentation processes filling in these areas.

Comparison of actual and predicted LULC maps for 2015-2019 period shows (Figure 15) that anticipated expansion of agricultural land, as posited by predictive models, was not fully realized in observed outcomes. This discrepancy may be attributed to an overestimation of agricultural proliferation, potentially neglecting the nuances of policy alterations or market forces that constrained agricultural growth. The variance underscores the challenge in accurately forecasting land use trends, emphasizing the necessity to integrate a comprehensive suite of variables into predictive algorithms. Contrastingly, grasslands were considerably overrepresented in predictions. This suggests a potential oversight in the model's calibration, perhaps underestimating the resilience of grasslands to conversion or misjudging the rate of land use transition. The actual diminution of grasslands, in favor of shrublands and urban development, indicates a dynamic ecological succession and urbanization process not fully captured by the model. Such transitions underscore the imperative of incorporating adaptive land management and policy frameworks into LULC change models. Moreover, the projected versus actual areas of high vegetation and barren lands highlight a critical

analysis point. The shortfall in high vegetation areas could point to a systemic underestimation of factors leading to deforestation or land repurposing for agriculture and urban expansion. Conversely, the actual increase in barren lands suggests a more pronounced environmental degradation or alteration in land management practices than the model anticipated, raising concerns about the peninsula's ecological resilience and the accuracy of degradation predictors within LULC models. Urban expansion, exceeding model predictions, reflects the formidable impact of economic development and demographic shifts, outpacing historical trends and model expectations. This expansion into areas designated for alternate uses in predictions illustrates the potent influence of socio-economic drivers on land cover change, underscoring the need for models to adaptively account for socio-economic trends. Lastly, the overestimation of water bodies in predictions versus actual observations may indicate an optimistic projection of hydrological stability or a misapprehension of water management practices' impact. This aspect accentuates the critical role of accurate hydrological modeling and the integration of water policy dynamics into comprehensive LULC predictive frameworks.
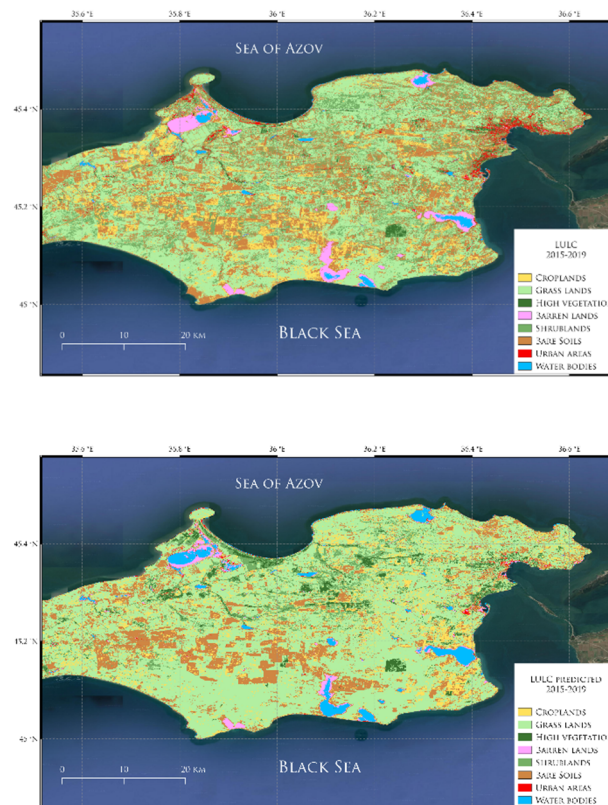
Figure 15. Actual and predicted LULC maps for 2015-2019 period.

Analyzing the LULC data for the Kerch Peninsula from the 2015-2019 period (Figure 16) and comparing it with the predicted data for 2020-2024 can be seen that croplands have seen a significant decrease from the actual 262.46 km² in the 2015-2019 period to a predicted 131.99 km² for 2020-2024. This drastic reduction could be attributed to a shift in agricultural policies, changes in land use priorities, or the influence of natural factors that render the land less suitable for farming. grass lands, conversely, show a substantial increase from an actual 1270.07 km² to a predicted 1553.43 km². This expansion may reflect a move towards conserving natural grasslands, a decline in intensive agricultural practices, or possibly the natural succession of grasses over previously cultivated lands. High vegetation sees a slight decrease from 23.38 km² to 21.95 km² in the predicted period. The marginal change suggests relative stability in forested or vegetated areas, possibly due to successful conservation efforts or a balance between deforestation and reforestation initiatives. barren lands have increased from 71.39 km² to 75.79 km². This suggests ongoing soil erosion issues or land

clearance activities that are striking the land off its vegetative cover, leading to more exposed terrain. Shrublands are predicted to decrease from 486.42 km² to 318.69 km², pointing to potential land conversion for other uses or changes in vegetation patterns due to climate or human influences. Bare soils, on the other hand, have shown a significant increase from 702.91 km² to an alarming 732.03 km². This could raise concerns regarding desertification processes, land mismanagement, or overexploitation of the land for construction and other developmental activities. Urban areas show a decrease in the predicted period, from 50.63 km² to 38.23 km². This unexpected decline could indicate a trend toward decentralization and rural development or successful policies limiting urban sprawl. Lastly, water bodies have reduced from 32.70 km² to 27.82 km². This reduction could be a result of water extraction for agricultural or urban needs, climate-induced changes affecting hydrological cycles, or the repurposing of water bodies for land development.
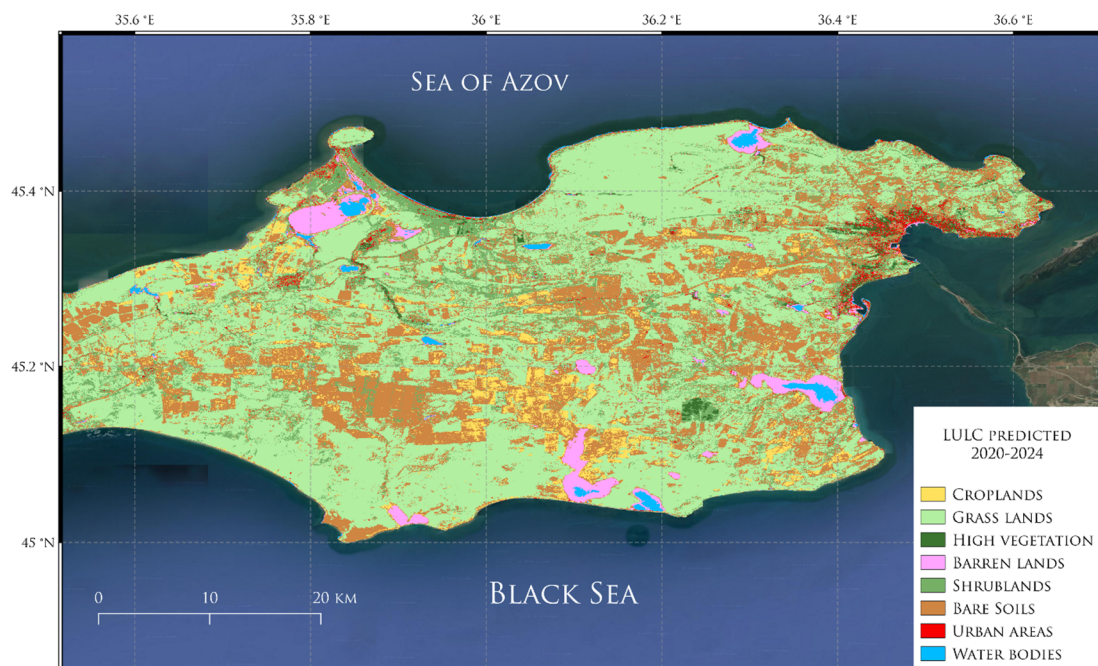
Figure 16. Predicted LULC map for the 2020-2024 period.

For the next predicted timestep the expansion of croplands to 145.45 km² in the 2025-2029 period (Figure 17) could reflect successful policy implementation or favorable agronomic conditions. Grasslands, the dominant landscape, have seen a noticeable increase in extent, suggesting either natural progression due to secondary succession in abandoned fields or perhaps the influence of land management policies prioritizing grassland conservation. This expansion from 1553.44 km² to 1760.74 km² could also be a response to changing grazing patterns or a shift in the ecological equilibrium favoring grassland biomes. Conversely, high vegetation areas have witnessed a contraction. This decline from 21.95 km² to 17.59 km² might hint at underlying stressors impacting forested regions and other densely vegetated habitats, including potential deforestation, urbanization, or changes in land management practices that favor other land uses over forest conservation. Similarly, barren lands have slightly diminished in area, suggesting either a restoration of vegetative cover in previously degraded lands or a change in land use where these areas are being converted for more productive purposes. The modest reduction to 74.68 km² in the 2025-2029 period could be indicative of either successful ecological restoration efforts or the exploitation of these lands for development purposes.
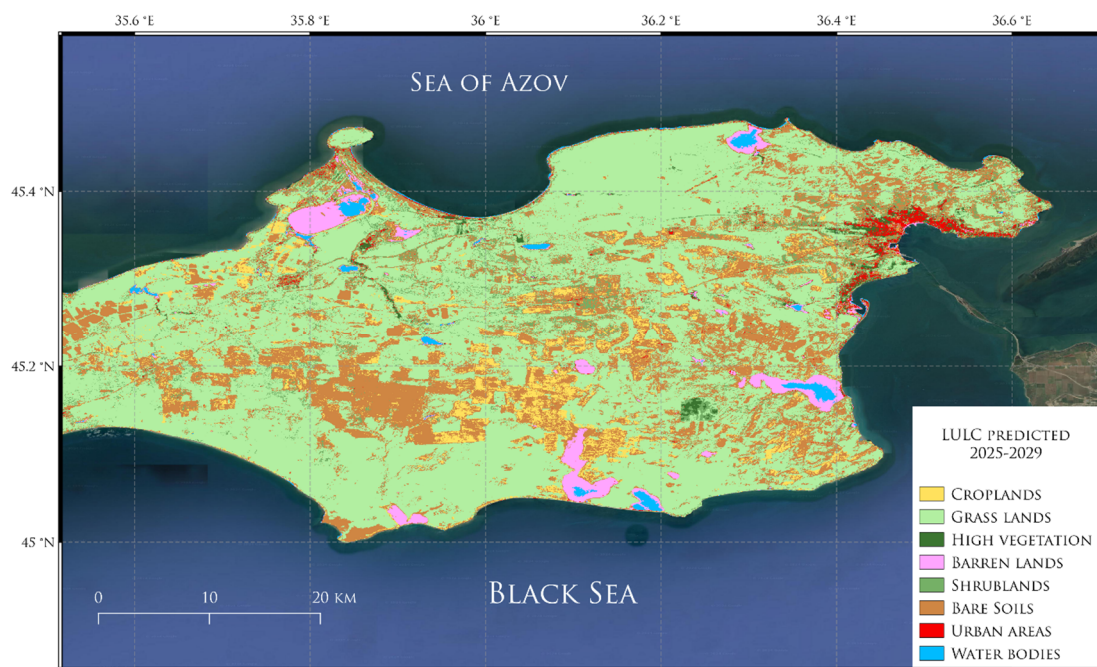
Figure 17. Predicted LULC map for the 2025-2029 period.

4. Discussion

Previous works [6,33,51] have laid a solid foundation in understanding the multifaceted nature of LULCC and its impacts on the Kerch Peninsula. However, the extant literature and our previous research have illuminated several areas where knowledge remains incomplete, underscoring the necessity for a continued and focused investigation.

Firstly, despite advancements in remote sensing and GIS technologies, the integration of socio-economic drivers with biophysical data remains underdeveloped in LULCC models. This gap hinders our ability to fully understand the complex interplay between human activities and land cover changes. Our current research endeavors to bridge this gap by employing a multidimensional approach that synthesizes socio-economic variables with spatial data, thereby offering a more nuanced understanding of LULCC drivers on the Kerch Peninsula.

Secondly, the dynamic and bidirectional relationship between LULCC and climate variability necessitates a more granular exploration. Previous studies have primarily focused on the unidirectional impacts of human activity on land cover or vice versa. This research aims to advance the discourse by examining how LULCC simultaneously influences and is influenced by climate variability, employing a holistic model that accounts for these bidirectional interactions.

Moreover, the specificity of LULCC impacts at the local and regional scales, particularly in areas undergoing rapid urbanization like the Kerch Peninsula, calls for a tailored analysis that can inform localized land management and conservation strategies. While our prior investigations have significantly contributed to the regional understanding of LULCC, the current study seeks to refine these insights by incorporating recent developments in LULCC modeling and analysis techniques. This includes the use of deep learning algorithms for enhanced classification accuracy and the analysis of long-term data sets for understanding temporal trends in land cover changes.

The results obtained in this paper on the Kerch Peninsula over recent decades, we've observed a complex interplay of socio-economic transformations, climatic variability, and the profound impacts of geopolitical shifts on LULCC. The intricate dynamics of these factors have shaped the peninsula's landscape, driving significant alterations in agricultural practices, urban development, and natural habitats. The disintegration of the Soviet Union and the subsequent economic transitions marked the initial phase of our study period, leading to industrial decline and a contraction in arable

land. This era was characterized by a noticeable de-intensification of agriculture, resulting in a reduction of cropland areas and a shift in land cover, as agricultural lands were abandoned or reverted to natural states. Concurrently, climatic shifts, notably warming trends, influenced LULCC, affecting vegetation phenology and agricultural yields. Our analysis highlighted a notable increase in urban areas, underscoring the ongoing urban sprawl and industrial development, which further intensified the anthropogenic pressures on the peninsula's natural ecosystems. This urban expansion, coupled with agricultural extension, necessitated a reevaluation of land management and conservation strategies to mitigate the adverse impacts on the region's ecological fabric. Throughout the subsequent years, the Kerch Peninsula experienced various land cover transformations, including the expansion of grasslands and high vegetation areas, possibly reflecting land abandonment or the transition towards more extensive farming practices. These changes suggest a landscape adapting to both the withdrawal of human agricultural activity and conservation policies aimed at preserving these ecosystems. The recent period, marked by Crimea's political transition and the shutdown of the North Crimean Canal, presented new challenges but also opportunities for adaptation and resilience within the region's LULC management. Despite the reduction in water resources, which significantly impacted agricultural practices, there was only a slight decrease in croplands, indicating an adaptation towards more water-efficient agricultural methods or a shift to less water-dependent crops. The resilience of grasslands, the slight expansion of barren lands, and the dynamic interplay between shrublands and bare soils, alongside modest urban development, reflect the peninsula's adaptive capacity to environmental and socio-economic challenges. The landscape's response to these challenges underscores the critical importance of sustainable land management practices in balancing development needs with environmental conservation.

Our research on the Kerch Peninsula elucidated the intricate dynamics between socio-economic factors, climatic variability, and the impacts of geopolitical shifts on LULCC. Like our findings, studies in other regions with comparable ecological and socio-economic conditions have documented substantial LULCC driven by a mix of natural and anthropogenic factors. For instance, the study by Paudel [52] on Nepal underscores the constant flux in LULC due to human activities and natural phenomena, mirroring our observations on the Kerch Peninsula regarding the dual influence of human and environmental drivers on LULCC.

Moreover, Chughtai [4] emphasize the role of remote sensing technology in monitoring LULCC, a methodological approach that was pivotal in our study as well. Their review highlights the utility of post-classification change detection methods, which aligns with our methodological rigor in detecting and analyzing LULCC over time. This common reliance on advanced technological tools underscores the growing importance of remote sensing in providing accurate, timely, and scalable insights into LULCC across diverse regions. Ren [2] synthesis of LULCC models further complements our analytical framework by delineating the spectrum of approaches from pattern- to process-based models. Their discussion on the challenges of integrating human-environment interactions into LULCC modeling resonates with our efforts to capture the complex interplay between anthropogenic activities and environmental changes on the Kerch Peninsula. This parallel underscore a broader challenge within LULCC research: the need for models that can holistically account for the multifaceted drivers and impacts of land cover changes. However, the comparative analysis also reveals divergences in LULCC dynamics across regions, attributed to specific local or regional drivers. For example, the study by Goldewijk [53] and Abdullah [54] on the advancements in LULCC classification techniques points to the significant role of agricultural expansion in driving LULCC, a factor that, while present, plays out differently on the Kerch Peninsula due to its unique socio-political and ecological context. This divergence highlights the importance of context-specific analyses in understanding LULCC processes.

In discussing the limitations and their potential impact on the results of the research on LULCC on the Kerch Peninsula, several key factors emerge that could influence the interpretation and applicability of the findings. These limitations also provide a pathway for future research directions, helping to refine and expand our understanding of LULCC dynamics.

One significant limitation is the reliance on available data, which might not cover all relevant aspects of LULCC comprehensively. For instance, historical land use data may be sparse, incomplete, or not available at a desirable spatial or temporal resolution. This limitation can lead to gaps in understanding the full extent of land use transitions over time, potentially underestimating, or overestimating the rate and nature of changes. Furthermore, inconsistencies in data collection methods across different time periods or regions can introduce biases, affecting the comparability and consistency of the results.

The methodological approaches employed to analyze LULCC, including remote sensing and GIS techniques, carry inherent limitations. For example, the accuracy of satellite imagery in identifying specific land cover types can vary, influenced by factors such as sensor resolution, atmospheric conditions, and the presence of similar spectral signatures between different land cover types. Additionally, the interpretation of remote sensing data requires substantial expertise, and the choice of classification algorithms can significantly impact the outcomes of the analysis. These methodological constraints might lead to misclassification errors, influencing the perceived extent and patterns of LULCC.

The research might not fully account for the complex interplay of socio-economic and political factors driving LULCC. While attempts can be made to incorporate these factors into the analysis, quantifying their direct and indirect impacts is challenging. Policies, economic shifts, and social dynamics can rapidly alter land use patterns, and without a nuanced understanding of these drivers, the research might not capture the complete picture of LULCC. Moreover, future changes in these areas could alter the trajectories of land use, posing a limitation to the predictability of current models.

The potential impacts of LULCC on ecosystems and climate are multifaceted and not entirely predictable. While the research can identify trends and correlations, the causal mechanisms underlying these relationships may be complex and not fully understood. Changes in land cover can affect biodiversity, ecosystem services, and climate feedback processes in ways that are difficult to quantify precisely.

Future studies could leverage advancements in remote sensing technologies, incorporating high-resolution satellite imagery and multi-source data, including drones and ground-based observations. This would enhance the accuracy of land cover classification and change detection, allowing for more detailed analyses of subtle changes and patterns. Additionally, integrating data from social media and mobile devices could offer insights into human activities related to land use, complementing traditional data sources.

Also, there is a need for the development and application of more sophisticated modeling techniques that can accurately simulate the complex interactions between socio-economic drivers, ecological processes, and LULCC. Agent-based models, coupled with system dynamics models and machine learning algorithms, could offer nuanced understandings of the causative factors behind land use transitions. These models can incorporate feedback loops, non-linear dynamics, and stochastic events, providing more realistic projections of future land use scenarios.

Future research should adopt a socio-ecological systems perspective, recognizing the intertwined nature of human and natural systems. This involves studying LULCC not just as a physical transformation of the landscape but as a component of broader socio-ecological dynamics. Research should explore the impacts of LULCC on ecosystem services, biodiversity, and human well-being, as well as how these changes feedback into land use decisions.

5. Conclusion

Our extensive study on the Kerch Peninsula over the past three decades has provided profound insights into the dynamics of LULCC, driven by a complex interplay of socio-political transformations, climatic variability, and environmental policies. Through the innovative use of convolutional neural networks and advanced satellite imagery, we have unveiled significant trends that highlight the profound impacts of urban growth, agricultural expansion, and the introduction of sustainable land management practices on the region's ecological fabric.

The transition matrices and LULC dynamics underscore the peninsula's adaptive responses to changing socio-economic conditions, policy shifts, and environmental challenges. Notably, the increase in urban areas and the concurrent decline in natural ecosystems like forests and wetlands illuminate the critical need for informed land management and conservation strategies to mitigate these effects.

Our findings emphasize the critical role of integrating change analysis and prediction in LULCC research. This approach allows for a nuanced understanding of the forces shaping the Kerch Peninsula's landscape, offering a foundation for developing effective land management policies and conservation strategies that can adapt to changing environmental conditions and socioeconomic needs.

The study has demonstrated the invaluable role of remote sensing and GIS technologies in enhancing LULCC analysis, enabling a deeper understanding of the intricate dynamics at play. These tools have not only facilitated a detailed examination of the region's landscape transformations but also underscored the potential for predictive modeling to forecast future land cover scenarios. Such insights are pivotal for planning sustainable development, implementing effective conservation strategies, and ensuring the ecological and socio-economic well-being of local communities.

Furthermore, our research has highlighted the importance of considering the broad spectrum of drivers influencing LULCC, from demographic shifts and economic trends to environmental factors and policy frameworks. Understanding these drivers is crucial for anticipating future land cover changes and devising strategies that balance development needs with environmental preservation.

Denis Krivoguz: Writing – review & editing, Writing – original draft, Visualization, Validation, Supervision, Software, Resources, Project administration, Methodology, Investigation, Funding acquisition, Formal analysis, Data curation, Conceptualization.

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References

1.   Bojer, A.K.; Ahmed, M.E.; Bekalo, D.J.; Debelee, T.G.; Al-Quraishi, A.M.F.; Deche, A. Analysis of Land Use/Land Cover Change (LULCC) and Debris Flow Risks in Adama District, Ethiopia, Aided by Numerical Simulation and Deep Learning-Based Remote Sensing. *Stoch Environ Res Risk Assess* **2023**, *37*, 4893–4910, doi:10.1007/s00477-023-02550-w.

2.   Ren, Y.; Lü, Y.; Comber, A.; Fu, B.; Harris, P.; Wu, L. Spatially Explicit Simulation of Land Use/Land Cover Changes: Current Coverage and Future Prospects. *Earth-Science Reviews* **2019**, doi:10.1016/J.EARSCIREV.2019.01.001.

3.   Gaur, S.; Singh, R. A Comprehensive Review on Land Use/Land Cover (LULC) Change Modeling for Urban Development: Current Status and Future Prospects. *Sustainability* **2023**, doi:10.3390/su15020903.

4.   Chughtai, A.H.; Abbasi, H.; Karas, I.R. A Review on Change Detection Method and Accuracy Assessment for Land Use Land Cover. *Remote Sensing Applications: Society and Environment* **2021**, *22*, doi:10.1016/J.RSASE.2021.100482.

5.   Prestele, R.; Arneth, A.; Bondeau, A.; Noblet-Ducoudré, N.; Pugh, T.; Sitch, S.; Stehfest, E.; Verburg, P. Current Challenges of Implementing Anthropogenic Land-Use and Land-Cover Change in Models Contributing to Climate Change Assessments. *Earth System Dynamics Discussions* **2017**, *8*, 369–386, doi:10.5194/ESD-8-369-2017.

6.   Krivoguz, D. Geo-Spatial Analysis of Urbanization and Environmental Changes with Deep Neural Networks: Insights from a Three-Decade Study in Kerch Peninsula. *Ecological Informatics* **2024**, *80*, 102513, doi:10.1016/j.ecoinf.2024.102513.

7.   Piekkoontod, T.; Pachana, B.; Hrimpeng, K.; Charoenjit, K. Assessments of Nipa Forest Using Landsat Imagery Enhanced with Unmanned Aerial Vehicle Photography. *Applied Environmental Research* **2020**, *42*, 49–59, doi:10.35762/AER.2020.42.3.5.

8.   Li, W.; Dong, R.; Fu, H.; Wang, J.; Yu, L.; Gong, P. Integrating Google Earth Imagery with Landsat Data to Improve 30-m Resolution Land Cover Mapping. *Remote Sensing of Environment* **2020**, *237*, 111563, doi:10.1016/j.rse.2019.111563.

9.   Moon, M.; Richardson, A.D.; Friedl, M.A. Multiscale Assessment of Land Surface Phenology from Harmonized Landsat 8 and Sentinel-2, PlanetScope, and PhenoCam Imagery. *Remote Sensing of Environment* **2021**, *266*, 112716, doi:10.1016/j.rse.2021.112716.

10.   Muhammad, R.; Zhang, W.; Abbas, Z.; Guo, F.; Gwiazdzinski, L. Spatiotemporal Change Analysis and Prediction of Future Land Use and Land Cover Changes Using QGIS MOLUSCE Plugin and Remote Sensing Big Data: A Case Study of Linyi, China. *Land* **2022**, *11*, 419, doi:10.3390/land11030419.

11.   Alshari, E.A.; Gawali, B.W. Modeling Land Use Change in Sana'a City of Yemen with MOLUSCE. *Journal of Sensors* **2022**, *2022*, e7419031, doi:10.1155/2022/7419031.

12.   Makinano-Santillan, M.; Bolanio, K.P.; Valeroso, J.D.; Yu, L.G. ANALYZING THE EFFECTS OF LAND USE / COVER CHANGE (LULCC) SIMULATION ON FLOODING: A CASE STUDY IN LAS NIEVES, AGUSAN DEL NORTE, MINDANAO, PHILIPPINES. *The International Archives of the Photogrammetry, Remote Sensing and Spatial Information Sciences* **2023**, *XLVIII-4-W6-2022*, 211–216, doi:10.5194/isprs-archives-XLVIII-4-W6-2022-211-2023.

13.   Amgoth, A.; Rani, H.P.; Jayakumar, K.V. Exploring LULC Changes in Pakhal Lake Area, Telangana, India Using QGIS MOLUSCE Plugin. *Spat. Inf. Res.* **2023**, *31*, 429–438, doi:10.1007/s41324-023-00509-1.

14.   Sajan, B.; Mishra, V.N.; Kanga, S.; Meraj, G.; Singh, S.K.; Kumar, P. Cellular Automata-Based Artificial Neural Network Model for Assessing Past, Present, and Future Land Use/Land Cover Dynamics. *Agronomy* **2022**, *12*, 2772, doi:10.3390/agronomy12112772.

15.   Nkolokosa, C.; Stothard, R.; Jones, C.M.; Stanton, M.; Chirombo, J.; Tangena, J.-A.A. Monitoring and Simulating Landscape Changes: How Do Long-Term Changes in Land Use and Long-Term Average Climate Affect Regional Biophysical Conditions in Southern Malawi? *Environ Monit Assess* **2023**, *195*, 1247, doi:10.1007/s10661-023-11783-9.

16.   Gondo, R.; Kolawole, O.D.; Mfundisi, K.B. Land Use and Land Cover Changes along the Boteti-Thamalakane River System in Ngamiland District, Botswana. *Scientific African* **2023**, *20*, e01595, doi:10.1016/j.sciaf.2023.e01595.

17.   Wang, J.; Bretz, M.; Dewan, M.A.A.; Delavar, M.A. Machine Learning in Modelling Land-Use and Land Cover-Change (LULCC): Current Status, Challenges and Prospects. *Science of The Total Environment* **2022**, *822*, 153559, doi:10.1016/j.scitotenv.2022.153559.

18.   Shimizu, K.; Murakami, W.; Furuichi, T.; Estoque, R.C. Mapping Land Use/Land Cover Changes and Forest Disturbances in Vietnam Using a Landsat Temporal Segmentation Algorithm. *Remote Sensing* **2023**, *15*, 851, doi:10.3390/rs15030851.

19.   Gidey, E.; Mhangara, P. An Application of Machine-Learning Model for Analyzing the Impact of Land-Use Change on Surface Water Resources in Gauteng Province, South Africa. *Remote Sensing* **2023**, *15*, 4092, doi:10.3390/rs15164092.

20.   Faqe Ibrahim, G. Urban Land Use Land Cover Changes and Their Effect on Land Surface Temperature: Case Study Using Dohuk City in the Kurdistan Region of Iraq. *Climate* **2017**, *5*, 13, doi:10.3390/cli5010013.

21.   Zhang, Y.; Odeh, I.O.A.; Han, C. Bi-Temporal Characterization of Land Surface Temperature in Relation to Impervious Surface Area, NDVI and NDBI, Using a Sub-Pixel Image Analysis. *International Journal of Applied Earth Observation and Geoinformation* **2009**, *11*, 256–264, doi:10.1016/j.jag.2009.03.001.

22.   Chen, X.-L.; Zhao, H.-M.; Li, P.-X.; Yin, Z.-Y. Remote Sensing Image-Based Analysis of the Relationship between Urban Heat Island and Land Use/Cover Changes. *Remote Sensing of Environment* **2006**, *104*, 133–146, doi:10.1016/j.rse.2005.11.016.

23.   Ronald, K.; Singh, K.J.; Lodhi, M.S. Land Use and Land Cover Change Study Using RS-GIS and Its Impact on Biodiversity: Study from Arunachal Pradesh, India. *Asian Journal of Environment & Ecology* **2023**, *22*, 120–132, doi:10.9734/ajee/2023/v22i4514.

24.   Sun, J.; He, J. Influence of Land Use and Land Cover Change on Land Surface Temperature. *E3S Web Conf.* **2021**, *283*, 01038, doi:10.1051/e3sconf/202128301038.

25.   Feng, Y.; Li, H.; Tong, X.; Chen, L.; Liu, Y. Projection of Land Surface Temperature Considering the Effects of Future Land Change in the Taihu Lake Basin of China. *Global and Planetary Change* **2018**, *167*, 24–34, doi:10.1016/j.gloplacha.2018.05.007.

26.   Krivoguz, D.O. Statistical Analysis Of Hydrological Drivers Of Landslide Inducing Factors Using Weights Of Evidence Approach On Kerch Peninsula. *International Journal of Environment and Geoinformatics* **2021**, *9*, 60–63.

27.   Kharytonov, M.M.; Mitsik, O.; Stankevich, S.; Titarenko, O. Geomining Site Ecological Assessment and Reclamation Along Coastal Line of the Kerch Peninsula. In Proceedings of the Black Sea Energy Resource Development and Hydrogen Energy Problems; Veziroğlu, A., Tsitskishvili, M., Eds.; Springer Netherlands: Dordrecht, 2013; pp. 325–336.

28.   Krivoguz, D.; Bespalova, L. Analysis of Kerch Peninsula's Climatic Parameters in Scope of Landslide Susceptibility. *Bulletin of KSMTU* **2018**, 5–11.

29.   Krivoguz, D.; Bespalova, L.; Zhilenkov, A.; Chernyi, S. A Deep Neural Network Method for Water Areas Extraction Using Remote Sensing Data. *JMSE* **2022**, *10*, 1392, doi:10.3390/jmse10101392.

30.   Krivoguz, D.; Bespalova, L. Landslide Susceptibility Analysis for the Kerch Peninsula Using Weights of Evidence Approach and GIS. *Russian Journal of Earth Sciences* **2020**, *20*, ES1003.

31. Krivoguz, D.; Zakharova, J. Geospatial Analysis of Ecology-Economic Grows of Kerch Peninsula Using GIS. *Geoinformatika* **2018**, 52–55.

32. Gao, C.; Cheng, D.; Iqbal, J.; Yao, S. Spatiotemporal Change Analysis and Prediction of Future Land Cover Changes Using QGIS MOLUSCE Plugin on a Large Regional Scale and the Relationship Analysis with Mountain Hazards: A Case Study of the Great Yellow River Region (GYRR). *China. Land*.

33. Krivoguz, D.; Chernyi, S.G.; Zinchenko, E.; Silkin, A.; Zinchenko, A. Using Landsat-5 for Accurate Historical LULC Classification: A Comparison of Machine Learning Models. *Data* **2023**, *8*, 138, doi:10.3390/data8090138.

34. Wu, Q. Geemap: A Python Package for Interactive Mapping with Google Earth Engine. *Journal of Open Source Software* **2020**, *5*, 2305, doi:10.21105/joss.02305.

35. Pettorelli, N.; Vik, J.O.; Mysterud, A.; Gaillard, J.-M.; Tucker, C.J.; Stenseth, N.C. Using the Satellite-Derived NDVI to Assess Ecological Responses to Environmental Change. *Trends in ecology & evolution* **2005**, *20*, 503–510.

36. Wang, J.; Price, K.P.; Rich, P.M. Spatial Patterns of NDVI in Response to Precipitation and Temperature in the Central Great Plains. *International Journal of Remote Sensing* **2001**, *22*, 3827–3844, doi:10.1080/01431160010007033.

37. Defries, R.S.; Townshend, J.R.G. NDVI-Derived Land Cover Classifications at a Global Scale. *International Journal of Remote Sensing* **1994**, *15*, 3567–3586, doi:10.1080/01431169408954345.

38. Gao, B.-C. NDWI—A Normalized Difference Water Index for Remote Sensing of Vegetation Liquid Water from Space. *Remote sensing of environment* **1996**, *58*, 257–266.

39. Gu, Y.; Brown, J.F.; Verdin, J.P.; Wardlow, B. A Five-year Analysis of MODIS NDVI and NDWI for Grassland Drought Assessment over the Central Great Plains of the United States. *Geophysical Research Letters* **2007**, *34*, 2006GL029127, doi:10.1029/2006GL029127.

40. Xu, H. Modification of Normalised Difference Water Index (NDWI) to Enhance Open Water Features in Remotely Sensed Imagery. *International Journal of Remote Sensing* **2006**, *27*, 3025–3033, doi:10.1080/01431160600589179.

41. Guha, S.; Govil, H.; Dey, A.; Gill, N. Analytical Study of Land Surface Temperature with NDVI and NDBI Using Landsat 8 OLI and TIRS Data in Florence and Naples City, Italy. *European Journal of Remote Sensing* **2018**, *51*, 667–678, doi:10.1080/22797254.2018.1474494.

42. Li, H.; Liu, Q. Comparison of NDBI and NDVI as Indicators of Surface Urban Heat Island Effect in MODIS Imagery. In Proceedings of the International conference on earth observation data processing and analysis (ICEODPA); SPIE, 2008; Vol. 7285, pp. 33–42.

43. Hunt, E.D.; Hubbard, K.G.; Wilhite, D.A.; Arkebauer, T.J.; Dutcher, A.L. The Development and Evaluation of a Soil Moisture Index. *Intl Journal of Climatology* **2009**, *29*, 747–759, doi:10.1002/joc.1749.

44. Quesney, A.; Le Hégarat-Mascle, S.; Taconet, O.; Vidal-Madjar, D.; Wigneron, J.-P.; Loumagne, C.; Normand, M. Estimation of Watershed Soil Moisture Index from ERS/SAR Data. *Remote sensing of environment* **2000**, *72*, 290–303.

45. Pesaresi, M.; Huadong, G.; Blaes, X.; Ehrlich, D.; Ferri, S.; Gueguen, L.; Halkia, M.; Kauffmann, M.; Kemper, T.; Lu, L. A Global Human Settlement Layer from Optical HR/VHR RS Data: Concept and First Results. *IEEE Journal of Selected Topics in Applied Earth Observations and Remote Sensing* **2013**, *6*, 2102–2131.

46. Pesaresi, M.; Ehrlich, D.; Ferri, S.; Florczyk, A.J.; Freire, S.; Halkia, M.; Julea, A.; Kemper, T.; Soille, P.; Syrris, V. *Operating Procedure for the Production of the Global Human Settlement Layer from Landsat Data of the Epochs 1975, 1990, 2000, and 2014*; Publications Office of the European Union Luxembourg, 2016;

47. Pesaresi, M.; Ehrlich, D.; Florczyk, A.J.; Freire, S.; Julea, A.; Kemper, T.; Syrris, V. The Global Human Settlement Layer from Landsat Imagery. In Proceedings of the 2016 IEEE international geoscience and remote sensing symposium (IGARSS); IEEE, 2016; pp. 7276–7279.

48. Beck, J.; Böller, M.; Erhardt, A.; Schwanghart, W. Spatial Bias in the GBIF Database and Its Effect on Modeling Species' Geographic Distributions. *Ecological Informatics* **2014**, *19*, 10–15.

49. Ivanova, N.V.; Shashkov, M.P. The Possibilities of GBIF Data Use in Ecological Research. *Russ J Ecol* **2021**, *52*, 1–8, doi:10.1134/S1067413621010069.

50. Hersbach, H.; Bell, B.; Berrisford, P.; Hirahara, S.; Horányi, A.; Muñoz Sabater, J.; Nicolas, J.; Peubey, C.; Radu, R.; Schepers, D.; et al. The ERA5 Global Reanalysis. *Quarterly Journal of the Royal Meteorological Society* **2020**, doi:10.1002/qj.3803.

51. Krivoguz, D.; Ioshpa, A.; Bespalova, L.; Zhilenkov, A.; Chernyi, S.; Goryachev, I.; Tsareva, P.; Voronova, A. Assessing Long-Term Lake Dynamics in Response to Climatic Variability: A Comprehensive Statistical Analysis. *Journal of Marine Science and Engineering* **2023**, *12*, 68.

52. Paudel, B.; Zhang, Y.; Li, S.; Liu, L.; Wu, X.; Khanal, N. Review of Studies on Land Use and Land Cover Change in Nepal. *Journal of Mountain Science* **2016**, *13*, 643–660, doi:10.1007/s11629-015-3604-9.

53. Klein Goldewijk, K.; Ramankutty, N. Land Cover Change over the Last Three Centuries Due to Human Activities: The Availability of New Global Data Sets. *GeoJournal* **2004**, *61*, 335–344, doi:10.1007/s10708-004-5050-z.

54. Abdullah, S.A.; Hezri, A.A. From Forest Landscape to Agricultural Landscape in the Developing Tropical Country of Malaysia: Pattern, Process, and Their Significance on Policy. *Environmental Management* **2008**, *42*, 907–917, doi:10.1007/s00267-008-9178-3.