

Fresh compost tea application does not change soil bacterial community structure, and has no effects on soybean growth or yield

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Abstract: Soil bacteria drive key ecosystem functions, including nutrient mobilization, soil aggregation and crop bioprotection against pathogens. Bacterial diversity is thus considered a key component of soil health. Conventional agriculture reduces bacterial diversity in many ways. Compost tea has been suggested as a bioinoculant that may restore bacterial community diversity and promote crop performance under conventional agriculture. Here, we conducted a field experiment to test this hypothesis in a soybean-maize rotation. Compost tea application had no influence on bacterial diversity or community structure. Plant growth and yield were also unresponsive to compost tea application. Combined, our results suggest that our compost tea bacteria did not thrive in the soil, and that the positive impacts of compost tea applications reported elsewhere may be caused by different microbial groups (e.g., fungi, protists, nematodes) or by abiotic effects on soil (e.g., contribution of nutrients and dissolved organic matter). Further investigations are needed to elucidate the mechanisms through which compost tea influences crop performance.

Keywords: Conventional agriculture, sustainable agriculture, compost tea, bacteria, biodiversity, Illumina MiSeq sequencing, plant growth, yield, soybean.

1. Introduction

Soil bacteria drive key ecosystem functions, including litter decomposition, nutrient mobilization, crop protection against pathogens, and soil aggregation [1-4]. Bacterial species are not functionally redundant, which translates into positive correlations between ecosystem functioning and bacterial diversity [5-7]. As a result, bacterial diversity is now recognized as a component of soil health and a central issue in the development of sustainable agriculture practices [2,8,9].

Conventional agriculture can negatively influence bacterial density and diversity in many ways. Tillage has been widely reported to negatively affect bacterial diversity in croplands [10-13]. The same is true for chemical fertilizers that have been shown to reduce soil functional diversity [14], and lead to community dominance by a few taxa [15-18]. Pesticides also reduce soil microbial diversity and enzymatic activity, which may compromise soil health and plant performance in the long run [19-21]. Various strategies have been suggested to alleviate the negative effects of these conventional practices. Among these, plant biostimulants such as microbial-based inoculants are deemed a promising solution for improvement of plant performance and ecosystem functioning [22-27].

Numerous microbial inoculants have been developed for organic and conventional farming [28-31]. Mycorrhizal fungi and nitrogen-fixing bacteria, for example, are widely used to promote plant growth and soil fertility in harsh conditions [32]. Likewise, various rhizobacteria have been found to promote plant growth and vigor [33]. However, the positive impacts of these microbial inoculants are likely to be context-dependent [34], and unlikely to restore microbial diversity in agricultural soils. Alternative solutions that better promote microbial diversity, and thus ecosystem

function [7], and resilience [35], should therefore be explored. Compost tea has been suggested as one such solution [36,37].

Compost tea is an inoculant prepared through aerobic, liquid-based incubation of compost amended with carbon sources. This promotes microbial proliferation [36-38]. Over a short incubation time (typically 48 hours), high cell densities can be achieved, allowing the application of a diluted suspension over large surfaces. Aerated compost tea is presumed to be an environmentally safe product that could enhance crop performance, in part, through the reintroduction of diverse soil bacteria contributing to nutrient cycling [37,38].

Positive yield responses to compost tea have been reported for a variety of crops [39-41]. Based on such findings, many authors have concluded that compost tea treatment could be used as a plant growth-promoting technique in organic cultivation of crops. However, we still lack basic information on the mechanisms through which compost tea influences indigenous microbial communities and crop yields. Specifically, we still do not know whether bacteria inoculated through compost tea can survive, successfully establish themselves and compete against indigenous bacterial communities in the soil. Further studies are required to trace and track changes in microbial communities following compost tea application, with appropriate experimental controls, in order to verify the potential impacts of the tea on crop growth and yield. We also need to better distinguish the biotic effects of applying compost tea to soil (i.e., its contribution of beneficial biotas) from its abiotic effects (i.e., its addition of nutrients to the soil in mineral and dissolved organic forms). In this study, we conducted a field experiment to evaluate the impact of

compost tea application on bacterial community structure and crop performance in a conventional soybean monoculture system. We hypothesized that compost tea application would promote bacterial diversity, and more specifically, Proteobacteria, which are commonly regarded as opportunistic copiotrophs [42], that should capitalize on simple sugars included as amendments during compost tea preparation. We anticipated that this shift in bacterial communities would, in turn, improve soybean growth and yield, given the wide range of known plant growth-promoting taxa among Proteobacteria.

2. Results

2.1. Bacterial community composition

From our total of 119 samples (108 soil samples and 11 compost tea samples), we retrieved 737 bacterial ASVs belonging to 13 phyla. Communities were dominated by Planctobacteria (63%), Verrucomicrobia (18%), Chloroflexi (7%), and Patescibacteria (6%) (Fig. 1.a).

When comparing plots treated with living vs sterilized compost tea, there was no significant effect of fresh compost tea application on bacterial communities, neither through shifts in α -diversity (ASV richness: $P = 0.64$ / Shannon's diversity: $P = 0.26$ / Inverse Simpson's diversity: $P = 0.56$), nor shifts in community structure (β -diversity; perMANOVA pseudo- $F = 1.17$, $d.f. = 2$, $P = 0.216$; Fig. 1.b,c).

Indicator species analysis revealed that: (1) there were many indicator bacterial ASVs of compost tea only (in fact, the majority of our indicator taxa belonged to this treatment category); (2) only one ASV (Planctobacteria) was an indicator of both compost tea samples and treated soil samples;

(3) several ASVs were indicators of both treated and control plots; and (4) no ASVs were indicators of only control plots or only treated plots. (Fig. 2).

2.2. Soybean growth and productivity

Compost tea application did not improve plant growth (shoot dry mass, $P = 0.36$) or grain yield (grain dry weight, $P = 0.14$; Fig. 3). Statistical power analyses indicated that compost tea application had small effect sizes (power = 23% and 30%, respectively, for growth and yield). We estimated that minimal sample size to detect an effect would have been 28 blocks for plant growth, and 20 blocks for plant yield, confirming the small effect size of our compost tea treatment.

3. Discussion

Surprisingly, both the control and treated soil samples were largely dominated by Planctobacteria (Fig. 1.a), a result contrasting with several studies identifying Proteobacteria as the dominant bacterial phylum in soils, followed by Chloroflexi, Bacteroidetes, Actinobacteria, and Acidobacteria [43-47]. Planctobacteria are a unique divergent phylum of aquatic bacteria [48-53], that can be isolated from nonaquatic environments such as soil [52,54]. These bacteria are assumed to prefer anaerobic soil micropores [55-57], as they can tolerate low O_2 pressures, which allows them to displace obligately aerobic taxa in low- O_2 microsites/horizons [55,56]. Here, we hypothesize that soil dominance by Planctobacteria could be explained by the recent installation of drainage infrastructures in the subsoil horizon of our study site. This caused the mixing of topsoil with deep subsoil (2 meters deep), which was presumably (1) less aerated and (2) less colonized by roots, which accordingly would account for the low abundance of copiotrophic rhizosphere specialists in the Proteobacteria [42]. This would be in line with Kepel et al. [58], who

recently found that the only soil in their dataset dominated by Planctobacteria was from a rice field, which are typically characterized by low soil O₂ pressures.

Because Planctobacteria were also abundant in our compost tea preparations (Fig. 1.a), we could also hypothesize that the Planctobacteria DNA retrieved in our soil samples (both from treated and control plots) belonged to dead bacterial cells, and this DNA had not fully degraded at the time of soil sampling. However, considering the total volume of liquid applied per surface in our treatments, we would find it surprising if the non-degraded portion of this dead DNA constituted the majority of the DNA we extracted afterwards from our soil samples. Moreover, by looking specifically at Planctobacteria communities in our soil and tea samples (i.e., by filtering our ASV table so that only Planctobacteria remain), we find that distinct Planctobacteria taxa dominated tea samples vs treated soil samples (Fig. S3).

Our principal component analysis (Fig. 1c) revealed a clustering of bacterial communities according to their plot origin rather than their treatment (i.e., living vs sterilized tea), which further shows that bacterial populations were spatially heterogeneous at our site, but not influenced by the treatment. This could be explained, in part, by contrasting soil properties across blocks (e.g., N availability; see Table 1).

The molecular analysis of bacterial community structure overall suggests a poor establishment of microbial taxa from the tea in the soil. This is supported by the fact that only 1 out of 737 ASVs was retrieved as an indicator taxon of both compost tea samples and treated soil samples (Fig. 2).

As our sensitivity analysis revealed that type I errors could represent around 3-4% of our dataset, we cannot rule out the possibility that the single indicator ASV for tea and for treated plots resulted from a type I error and thus was not truly an indicator for tea and treated plots. In fact, of our 737 ASVs, 322 were identified as indicator taxa (44%). This is well above the random expectation of 3-4%, but still, this means that roughly 10% of our indicator taxa may have arisen in the analysis by chance alone. However, this does not affect our conclusions, as most indicator taxa were indicators of either tea (probably taxa from tea that failed to thrive in the soil) or soil alone (resident soil taxa present prior to application). In both cases, this would suggest a poor establishment of tea bacterial taxa in our plots. Overall, this offers compelling evidence for the hypothesis that in our study, the tea bacteria failed to establish themselves in the soil, either because of low application density (and thus low initial population sizes) and/or because of a poor competitive ability against resident soil bacteria.

Compost tea application did not improve plant growth or yield in this experiment. Statistical power analyses confirmed the small effect size of our compost tea treatment, thus any impact of the compost tea on the living soil community (bacterial or not) would have been modest and would not have translated to large shifts in crop performance.

4. Materials and methods

4.1. Site Description

Our study was conducted in a field of approximately 3 hectares, located in Sainte-Christine, Quebec, Canada (see Fig. S1; 72.434353 W, 45.613667 N). This field has a several-decade history of conventional soybean-maize monocrop rotations and conventional agricultural practices. In spring 2018, installation of a drainage system in the field resulted in a severe soil disturbance in which the plow zone was mixed with the less biologically active, deeper horizons [12]. On June 6th, soybean was sown at a density of 382,850 seeds/ha.

4.2. Experimental design

We conducted our compost tea application using a randomized block design. We divided the field (344m x 82.5m) into 6 experimental blocks (172m x 27.5m). Each block was then divided into two plots, with one receiving the treatment (living compost tea) and the other receiving the control (compost tea sterilized by boiling). Before application, we characterized initial soil properties by collecting composite soil samples from each block (Table 1).

Table 1: Initial soil physio-chemical properties before seeding. Soil properties were measured on composite samples taken from each experimental block. Mehlich III – PO₄³⁻ = orthophosphates extractible with Mehlich-III solution; KCl-NH₄⁺ and KCl-NO₃⁻ = respectively ammonium and nitrates extractible using 2N KCl.

Block	pH	Organic matter content (%)	Gravimetric moisture (%)	Melich III - PO ₄ ³⁻ (mg/kg)	KCl-NH ₄ ⁺ (mg/kg)	KCl-NO ₃ ⁻ (mg/kg)
A	6.20	12.75	25.26	75.61	68.91	5.84
B	6.25	9.30	20.79	26.28	62.74	8.44
C	6.16	14.02	22.53	26.60	81.77	9.38

D	6.36	10.27	24.46	32.73	91.31	20.43
E	6.64	6.88	21.42	32.06	28.38	14.04
F	6.56	12.21	25.16	24.26	44.48	19.89
Mean	6.36	10.91	23.27	36.26	62.93	13.01

4.3. Compost tea preparation and application

Aerated compost tea was prepared in two phases: a pre-activation phase, aiming to increase microbial population densities in the compost, and a dilution phase, producing a liquid suspension from the compost (i.e., tea) for inoculation.

In the pre-activation phase, two different kinds of compost were mixed in equal quantities: the first, an especially carbon-rich vermicompost, consisting of up to 50% ramial wood chips and leaf litter and matured through the activity of earthworms; and the second, a thermal compost, consisting of 10% chicken manure, 15% horse manure, 30% fresh green plants, 20% ramial wood chips and 25% leaf litter, mixed at high temperatures (60-70° C) for 30 days. In a 75L container, we combined 20L of this compost mixture with 300 mL oatmeal, 150 mL alfalfa flour, 150 mL fish hydrolysate, 100mL seaweed flour, 30mL molasses, 5 ml humic acid solution and non-chlorinated water (to reach 50% humidity). This blend was incubated for 72 hours and mixed every 12 hours to maintain aerobic conditions. Compost tea was prepared by washing this aerated mixture at room temperature in 20L washing bags (mesh size = 400 µm), and then combining 10 L of the aerated mixture with 0.8 L water, 3 L oatmeal, 2.5 L fish hydrolysate, 1.5 L humic acid solution and 0.5 L soluble algae. Air was pumped into the mixture for two days to avoid anaerobic fermentation. Half of this compost tea preparation was then sterilized by heating at 95°C for 90

min, in order to be used as an experimental control (i.e., to distinguish the effects of the compost tea's living organisms from the abiotic effects of its minerals and dissolved organic nutrients).

The compost tea and sterilized control solution were prepared and applied to the field 4 times during the summer of 2018, on June 9th, June 22nd, July 5th and July 19th. Dilutions and dosages were adapted to weather conditions during the growing season, thus the compost tea dilution ratios for the specified dates were 1:1, 1:4, 1:4 and 1:3, respectively, with dosage densities of 121.57 l/ha, 486.26 l/ha, 486.26 l/ha and 364.7 l/ha, respectively. In addition, subsamples of each of the compost tea preparations (i.e., concentrated, applied, and sterilized) were kept at -20°C for molecular characterization of the bacterial communities.

4.4. Field samplings

Two field samplings were conducted. A first sampling campaign was done during the vegetative growing stage, on August 14th, in order to (1) measure the aboveground dry biomass as an indicator of vegetative plant growth, and (2) characterize the bacterial communities present in the soils. In each plot, nine individual soybean plants were excavated and their rhizospheric soil collected (by shaking the root system in a plastic bag) and kept frozen at -20°C for DNA extraction. Aboveground biomass was dried (at 65°C for 1wk) and weighed. The second field sampling was conducted a day before crop harvest, on October 3rd, when 30 individual soybean pods per plot were randomly collected and transferred to the laboratory to measure grain weight as an indicator of yield.

4.5. Molecular analyses

DNA was extracted from 250 mg of rhizospheric soil and compost tea samples using a Power Soil DNA kit (Qiagen Inc., Canada) according to the manufacturer's instructions. Double-stranded DNA was quantified using a Qubit® 2.0 Fluorometer (Thermo Fisher Scientific Inc., Canada). DNA extracts were PCR-amplified using 16S rDNA primers with CS1 and CS2 adapters show in italics (forward CS1-341F: 5' *ACACTGACGACATGGTTCTACACCTACGGGNGGCWGCAG*-3'; reverse CS2-806R: 5'-*TACGGTAGCAGAGACTTGGTCTGACTACHVGGGTATCTAATCC*-3'; [59]), targeting the hypervariable V3-V4 region of the 16 rRNA gene. PCR reactions were performed in a total volume of 50 µl containing 1X PCR buffer, 0.5 µM of each primer, 5.0 µL of dNTPs (10 mM), 0.4 µL of Taq DNA polymerase and 2 µL of template DNA. PCR conditions were as follows: 4 min denaturation at 94°C, followed by 35 cycles of denaturation (94°C for 30 s), annealing (55°C for 30 s), and extension (72°C for 60 s), and a final 10 min extension at 72°C. PCR reactions that gave a visible amplification band on agarose gel were sent for Illumina MiSeq sequencing (300 bp paired-end library) at the Génome Québec Innovation Center.

4.6. Bioinformatics

Analysis of the sequence data was coded in R (v4.0.1; R Development Core Team, 2014) using the *DADA2* R package (v1.1.2; [60]). Sequences were quality filtered and primers were removed. We removed sequences with less than 290 bp and 260 bp (forward and reverse, respectively), as the base quality of the sequences showed a clear drop below these thresholds. For this we used the *DADA2* command *filterAndTrim* with a maxEE score of 2 and trunQ score of 6. We then calculated the error rates using the *learnErrors* command and merged the forward and reverse

sequences. Next, chimeras were removed and the amplicon sequence variants (ASV) table was built, and finally taxonomy was assigned to the ASVs using the SILVA reference database [61].

A total of 2,171,433 raw reads were generated from 119 individual samples (108 soil samples and 11 compost tea samples). Sequences classified as chloroplasts or mitochondria were removed from the ASV table, as were any sequences classified as Eukarya or Archaea. Samples were then rarefied to 1749 reads per sample using the function *rrarefy* from the R package *vegan* [62]. To avoid focusing on potential sequencing artefacts or on especially rare bacterial taxa, we filtered our dataset to remove: (1) any occurrences with 5 reads; and (2) any ASVs that only appeared in 1 or 2 samples.

4.7. Statistical analysis

To determine the effect of compost tea and sterilized compost tea (control) treatments on plant growth and yield, we used linear mixed models (LMMs) as implemented in the R package *lme4*, including plot identity as a random effect [63]. We used the *pwr* R package to estimate the power of our analysis comparing the plant growth or yield of treated plots versus control plots [64].

We evaluated the impact of compost tea application on both bacterial α -diversity and community structure (β -diversity). Alpha diversity was assessed using ASV richness, the exponential form of Shannon diversity, and inverse Simpson diversity [65]. Alpha diversity was compared between treatments using Poisson regression (generalized LMM) for ASV richness and Gaussian LMMs for Shannon and Simpson diversities.

Shifts in community structure following treatments were assessed with permutational multivariate analysis of variance (perMANOVA; [66]) using the function *adonis* of the R package *vegan* [62].

The Hellinger distance [67] was used to evaluate pairwise β -diversity across samples, which was visualized using principal component analysis (PCA).

In order to identify bacterial ASVs that were specifically associated either with soil samples treated with compost tea or control samples treated with sterilized compost tea, we conducted an indicator species analysis (ISA) using the function *multipatt* of the R package *indicspecies* [68]. We used a threshold of $\alpha = 0.01$ because this analysis implied a high number of taxa in permutation-based statistical tests (i.e., 1 per bacterial taxon), which may inflate type I errors. However, traditional *P*-value correction methods (e.g., Bonferroni) would have resulted in overly conservative tests given the very high number of bacterial ASVs. This would have resulted in high type II error rates, which is why we decided to manually set α at 0.01. To evaluate how prone we were to detecting false positives (i.e., indicator taxa that would be associated with one treatment or another simply by chance), we conducted ISA on randomized metacommunities generated using the null model *vaznull* in the R package *bipartite* [69]. These simulations indicated that roughly 3-4% of indicator taxa may arise as false positives (Fig. S2).

5. Conclusions

Our results showed that aerated compost tea application had no influence on bacterial diversity or community structure. Accordingly, plant growth and yield were unresponsive to compost tea application. We note that our results do not undermine the potential role of compost tea in increasing crop yield or contributing to sustainable agriculture. Our design did not include plots where compost tea was not applied. Our study thus reveals that the positive effects of compost tea found in other studies could be due to: (1) the nutritional effects of compost tea (i.e., its

contribution of minerals to the soil through dissolved organic nutrients); or (2) its alteration of other physicochemical properties of the soil (e.g., increased cation exchange capacity due to dissolved organic matter in compost). Alternatively, the absence of effects in our study could be ascribed specifically to the compost we used or to the dose or frequency of tea application. Much remains to be studied regarding the mechanistic nature of the impact of compost tea on crop performance. Our only conclusion here is that in this field trial on a severely disturbed soybean monoculture field, living compost tea application did not influence bacterial communities or crop yield.

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Data Availability Statement: The datasets generated and analyzed during the current study are available in the Sequence Read Archive (SRA).

[<http://www.ncbi.nlm.nih.gov/bioproject/728448>].

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References:

1. Davison, J. (1988). Plant Beneficial Bacteria. *Nature Biotechnology*, 6(3), 282-286. <https://doi.org/10.1038/nbt0388-282>
2. Kennedy, A. C. (1999). Bacterial diversity in agroecosystems. In M. G. Paoletti (Ed.), *Invertebrate Biodiversity as Bioindicators of Sustainable Landscapes* (pp. 65-76). Elsevier. <https://doi.org/https://doi.org/10.1016/B978-0-444-50019-9.50007-8>
3. Van Der Heijden, M. G. A., Bardgett, R. D., & Van Straalen, N. M. (2008). The unseen majority: soil microbes as drivers of plant diversity and productivity in terrestrial ecosystems. *Ecology Letters*, 11(3), 296-310. <https://doi.org/10.1111/j.1461-0248.2007.01139.x>
4. Lladó, S., López-Mondéjar, R., & Baldrian, P. (2017). Forest Soil Bacteria: Diversity, Involvement in Ecosystem Processes, and Response to Global Change. *Microbiology and Molecular Biology Reviews*, 81(2), e00063-00016. <https://doi.org/10.1128/mmbr.00063-16>
5. Bonkowski, M., & Roy, J. (2005). Soil microbial diversity and soil functioning affect competition among grasses in experimental microcosms. *Oecologia*, 143(2), 232-240.
6. Wagg, C., Bender, S. F., Widmer, F., & Van Der Heijden, M. G. A. (2014). Soil biodiversity and soil community composition determine ecosystem multifunctionality. *Proceedings of the National Academy of Sciences*, 111(14), 5266-5270. <https://doi.org/10.1073/pnas.1320054111>
7. Delgado-Baquerizo, M., Maestre, F. T., Reich, P. B., Jeffries, T. C., Gaitan, J. J., Encinar, D., Berdugo, M., Campbell, C. D., & Singh, B. K. (2016). Microbial diversity drives multifunctionality in terrestrial ecosystems. *Nature Communications*, 7(1), 10541. <https://doi.org/10.1038/ncomms10541>
8. Hayat, R., Ali, S., Amara, U., Khalid, R., & Ahmed, I. (2010). Soil beneficial bacteria and their role in plant growth promotion: a review. *Annals of microbiology*, 60(4), 579-598.
9. Hayat, R., Ahmed, I., & Sheirdil, R. A. (2012). An Overview of Plant Growth Promoting Rhizobacteria (PGPR) for Sustainable Agriculture. In (pp. 557-579). Springer Netherlands. https://doi.org/10.1007/978-94-007-4116-4_22
10. Janušauskaite, D., Kadžienė, G., & Auškalnienė, O. (2013). The effect of tillage system on soil microbiota in relation to soil structure. *Polish Journal of Environmental Studies*, 22(5).

11. Silva, A., BABUJIA, L., Matsumoto, M., Guimarães, M., & Hungria, M. (2013). Bacterial diversity under different tillage and crop rotation systems in an oxisol of Southern Brazil. *Embrapa Soja-Artigo em periódico indexado (ALICE)*.
12. Sun, R., Li, W., Dong, W., Tian, Y., Hu, C., & Liu, B. (2018). Tillage changes vertical distribution of soil bacterial and fungal communities. *Frontiers in Microbiology*, 9, 699.
13. Liu, T., Li, S., Guo, L., Cao, C., Li, C., Zhai, Z., Zhou, J., Mei, Y., & Ke, H. (2020). Advantages of nitrogen fertilizer deep placement in greenhouse gas emissions and net ecosystem economic benefits from no-tillage paddy fields. *Journal of Cleaner Production*, 263, 121322.
14. Tsiafouli, M. A., Thébault, E., Sgardelis, S. P., De Ruiter, P. C., Van Der Putten, W. H., Birkhofer, K., Hemerik, L., De Vries, F. T., Bardgett, R. D., & Brady, M. V. (2015). Intensive agriculture reduces soil biodiversity across Europe. *Global change biology*, 21(2), 973-985.
15. Ji, L., Wu, Z., You, Z., Yi, X., Ni, K., Guo, S., & Ruan, J. (2018). Effects of organic substitution for synthetic N fertilizer on soil bacterial diversity and community composition: A 10-year field trial in a tea plantation. *Agriculture, Ecosystems & Environment*, 268, 124-132.
16. Ma, W., Abdulai, A., & Goetz, R. (2018). Agricultural cooperatives and investment in organic soil amendments and chemical fertilizer in China. *American Journal of Agricultural Economics*, 100(2), 502-520.
17. Bai, Y.-C., Chang, Y.-Y., Hussain, M., Lu, B., Zhang, J.-P., Song, X.-B., Lei, X.-S., & Pei, D. (2020). Soil chemical and microbiological properties are changed by long-term chemical fertilizers that limit ecosystem functioning. *Microorganisms*, 8(5), 694.
18. Liang, R., Hou, R., Li, J., Lyu, Y., Hang, S., Gong, H., & Ouyang, Z. (2020). Effects of Different Fertilizers on Rhizosphere Bacterial Communities of Winter Wheat in the North China Plain. *Agronomy*, 10(1), 93.
19. Hussain, S., Siddique, T., Saleem, M., Arshad, M., & Khalid, A. (2009). Impact of pesticides on soil microbial diversity, enzymes, and biochemical reactions. *Advances in agronomy*, 102, 159-200.
20. Lo, C.-C. (2010). Effect of pesticides on soil microbial community. *Journal of Environmental Science and Health Part B*, 45(5), 348-359.
21. Jacobsen, C. S., & Hjelmsø, M. H. (2014). Agricultural soils, pesticides and microbial diversity. *Current Opinion in Biotechnology*, 27, 15-20.
22. Du Jardin, P. (2015). Plant biostimulants: definition, concept, main categories and regulation. *Scientia Horticulturae*, 196, 3-14.

23. Bhardwaj, D., Ansari, M. W., Sahoo, R. K., & Tuteja, N. (2014). Biofertilizers function as key player in sustainable agriculture by improving soil fertility, plant tolerance and crop productivity. *Microbial cell factories*, 13(1), 1-10.
24. Souza, R. d., Ambrosini, A., & Passaglia, L. M. (2015). Plant growth-promoting bacteria as inoculants in agricultural soils. *Genetics and molecular biology*, 38(4), 401-419.
25. O'Callaghan, M. (2016). Microbial inoculation of seed for improved crop performance: issues and opportunities. *Applied microbiology and biotechnology*, 100(13), 5729-5746.
26. Swami, S. (2020). Soil Microbes for Securing the Future of Sustainable Farming. *Int. J. Curr. Microbiol. App. Sci*, 9(4), 2687-2706.
27. Raina, S. A., Bhat, R. A., Qadri, H., & Dutta, A. (2020). Values of Biofertilizers for Sustainable Management in Agricultural Industries. In *Bioremediation and Biotechnology, Vol 2* (pp. 121-137). Springer.
28. Khan, W., Rayirath, U. P., Subramanian, S., Jithesh, M. N., Rayorath, P., Hodges, D. M., Critchley, A. T., Craigie, J. S., Norrie, J., & Prithiviraj, B. (2009). Seaweed extracts as biostimulants of plant growth and development. *Journal of Plant Growth Regulation*, 28(4), 386-399.
29. Popko, M., Michalak, I., Wilk, R., Gramza, M., Chojnacka, K., & Górecki, H. (2018). Effect of the new plant growth biostimulants based on amino acids on yield and grain quality of winter wheat. *Molecules*, 23(2), 470.
30. Rafique, M., Sultan, T., Ortas, I., & Chaudhary, H. J. (2017). Enhancement of maize plant growth with inoculation of phosphate-solubilizing bacteria and biochar amendment in soil. *Soil science and plant nutrition*, 63(5), 460-469.
31. Hungria, M., Nogueira, M. A., & Araujo, R. S. (2015). Soybean seed co-inoculation with *Bradyrhizobium* spp. and *Azospirillum brasilense*: a new biotechnological tool to improve yield and sustainability. *Embrapa Soja-Artigo em periódico indexado (ALICE)*.
32. Requena, N., Perez-Solis, E., Azcón-Aguilar, C., Jeffries, P., & Barea, J.-M. (2001). Management of indigenous plant-microbe symbioses aids restoration of desertified ecosystems. *Applied and environmental microbiology*, 67(2), 495-498.
33. Gholami, A., Shahsavani, S., & Nezarat, S. (2009). The effect of plant growth promoting rhizobacteria (PGPR) on germination, seedling growth and yield of maize. *World Academy of Science, Engineering and Technology*, 49, 19-24.

34. Chaudhary, V. B., Akland, K., Johnson, N. C., & Bowker, M. A. (2020). Do soil inoculants accelerate dryland restoration? A simultaneous assessment of biocrusts and mycorrhizal fungi. *Restoration Ecology*, 28, S115-S126.
35. Girvan, M., Campbell, C., Killham, K., Prosser, J. I., & Glover, L. A. (2005). Bacterial diversity promotes community stability and functional resilience after perturbation. *Environmental microbiology*, 7(3), 301-313.
36. Naidu, Y., Meon, S., Kadir, J., & Siddiqui, Y. (2010). Microbial starter for the enhancement of biological activity of compost tea. *Int. J. Agric. Biol*, 12(1), 51-56.
37. Ingham, E. (2005). *The compost tea brewing manual* (Vol. 728). Soil Foodweb Incorporated Corvallis, OR, USA.
38. Kannangara, T., Forge, T., & Dang, B. (2006). Effects of aeration, molasses, kelp, compost type, and carrot juice on the growth of *Escherichia coli* in compost teas. *Compost science & utilization*, 14(1), 40-47.
39. Hargreaves, J. C., Adl, M. S., & Warman, P. R. (2009). Are compost teas an effective nutrient amendment in the cultivation of strawberries? Soil and plant tissue effects. *Journal of the Science of Food and Agriculture*, 89(3), 390-397.
40. Pant, A. P., Radovich, T. J., Hue, N. V., & Paull, R. E. (2012). Biochemical properties of compost tea associated with compost quality and effects on pak choi growth. *Scientia Horticulturae*, 148, 138-146.
41. Kim, M. J., Shim, C. K., Kim, Y. K., Hong, S. J., Park, J. H., Han, E. J., Kim, J. H., & Kim, S. C. (2015). Effect of aerated compost tea on the growth promotion of lettuce, soybean, and sweet corn in organic cultivation. *The plant pathology journal*, 31(3), 259.
42. Fierer, N., Bradford, M. A., & Jackson, R. B. (2007). Toward an ecological classification of soil bacteria. *Ecology*, 88(6), 1354-1364.
43. Li, W., Lv, X., Ruan, J., Yu, M., Song, Y.-B., Yu, J., & Dong, M. (2019). Variations in soil bacterial composition and diversity in newly formed coastal wetlands. *Frontiers in Microbiology*, 9, 3256.
44. Li, H., Xu, Z., Yang, S., Li, X., Top, E. M., Wang, R., Zhang, Y., Cai, J., Yao, F., & Han, X. (2016). Responses of soil bacterial communities to nitrogen deposition and precipitation increment are closely linked with aboveground community variation. *Microbial ecology*, 71(4), 974-989.
45. Rodrigues, J. L., Pellizari, V. H., Mueller, R., Baek, K., Jesus, E. d. C., Paula, F. S., Mirza, B., Hamaoui, G. S., Tsai, S. M., & Feigl, B. (2013). Conversion of the

- Amazon rainforest to agriculture results in biotic homogenization of soil bacterial communities. *Proceedings of the National Academy of Sciences*, 110(3), 988-993.
46. Navarrete, A. A., Cannavan, F. S., Taketani, R. G., & Tsai, S. M. (2010). A molecular survey of the diversity of microbial communities in different Amazonian agricultural model systems. *Diversity*, 2(5), 787-809.
47. da C Jesus, E., Marsh, T. L., Tiedje, J. M., & de S Moreira, F. M. (2009). Changes in land use alter the structure of bacterial communities in Western Amazon soils. *The ISME journal*, 3(9), 1004-1011.
48. Griepenburg, U., Ward-Rainey, N., Mohamed, S., Schlesner, H., Marxsen, H., Rainey, F. A., Stackebrandt, E., & Auling, G. (1999). Phylogenetic diversity, polyamine pattern and DNA base composition of members of the order Planctomycetales. *International Journal of Systematic and Evolutionary Microbiology*, 49(2), 689-696.
49. Schlesner, H. (1994). The development of media suitable for the microorganisms morphologically resembling Planctomyces spp., Pirellula spp., and other Planctomycetales from various aquatic habitats using dilute media. *Systematic and applied microbiology*, 17(1), 135-145.
50. Fuerst, J. A. (1995). The planctomycetes: emerging models for microbial ecology, evolution and cell biology. *Microbiology*, 141(7), 1493-1506.
51. Neef, A., Amann, R., Schlesner, H., & Schleifer, K.-H. (1998). Monitoring a widespread bacterial group: in situ detection of planctomycetes with 16S rRNA-targeted probes. *Microbiology*, 144(12), 3257-3266.
52. Wang, J., Jenkins, C., Webb, R. I., & Fuerst, J. A. (2002). Isolation of Gemmata-like and Isosphaera-like planctomycete bacteria from soil and freshwater. *Applied and environmental microbiology*, 68(1), 417-422.
53. Dedysh, S. N., & Ivanova, A. A. (2019). Planctomycetes in boreal and subarctic wetlands: diversity patterns and potential ecological functions. *FEMS microbiology ecology*, 95(2), fyy227.
54. Buckley, D. H., Huangyutitham, V., Nelson, T. A., Rumberger, A., & Thies, J. E. (2006). Diversity of Planctomycetes in soil in relation to soil history and environmental heterogeneity. *Applied and environmental microbiology*, 72(7), 4522-4531.
55. Derakshani, M., Lukow, T., & Liesack, W. (2001). Novel bacterial lineages at the (sub) division level as detected by signature nucleotide-targeted recovery of 16S rRNA genes from bulk soil and rice roots of flooded rice microcosms. *Applied and environmental microbiology*, 67(2), 623-631.

56. Elshahed, M. S., Youssef, N. H., Luo, Q., Najjar, F. Z., Roe, B. A., Sisk, T. M., Bühring, S. I., Hinrichs, K.-U., & Krumholz, L. R. (2007). Phylogenetic and metabolic diversity of Planctomycetes from anaerobic, sulfide- and sulfur-rich Zodletone Spring, Oklahoma. *Applied and environmental microbiology*, 73(15), 4707-4716.
57. Fuerst, J. A. (2017). Planctomycetes—new models for microbial cells and activities. In *Microbial Resources* (pp. 1-27). Elsevier.
58. Kepel, B. J., Gani, M. A., & Tallei, T. E. (2020). Comparison of bacterial community structure and diversity in traditional gold mining waste disposal site and rice field by using a metabarcoding approach. *International journal of microbiology*, 2020.
59. Takahashi, S., Tomita, J., Nishioka, K., Hisada, T., & Nishijima, M. (2014). Development of a prokaryotic universal primer for simultaneous analysis of Bacteria and Archaea using next-generation sequencing. *PloS one*, 9(8), e105592.
60. Callahan, B. J., McMurdie, P. J., Rosen, M. J., Han, A. W., Johnson, A. J. A., & Holmes, S. P. (2016). DADA2: high-resolution sample inference from Illumina amplicon data. *Nature methods*, 13(7), 581-583.
61. Pruesse, E., Quast, C., Knittel, K., Fuchs, B. M., Ludwig, W., Peplies, J., & Glöckner, F. O. (2007). SILVA: a comprehensive online resource for quality checked and aligned ribosomal RNA sequence data compatible with ARB. *Nucleic acids research*, 35(21), 7188-7196.
62. Oksanen, J., Blanchet, F. G., Kindt, R., Legendre, P., Minchin, P. R., O'hara, R., Simpson, G. L., Solymos, P., Stevens, M. H. H., & Wagner, H. (2013). Package 'vegan'. *Community ecology package, version*, 2(9), 1-295.
63. Zuur, A., Ieno, E. N., Walker, N., Saveliev, A. A., & Smith, G. M. (2009). *Mixed effects models and extensions in ecology with R*. Springer Science & Business Media.
64. Champely, S., & Champely, M. S. (2007). The pwr package. *UCB Lyon*, 1.
65. Hill, M. O. (1973). Diversity and evenness: a unifying notation and its consequences. *Ecology*, 54(2), 427-432.
66. Anderson, M. J. (2001). A new method for non-parametric multivariate analysis of variance. *Austral ecology*, 26(1), 32-46.
67. Legendre, P., Gallagher, E.D. Ecologically meaningful transformations for ordination of species data. *Oecologia* 129, 271–280 (2001). <https://doi.org/10.1007/s004420100716>.

68. De Caceres, M., Jansen, F., & De Caceres, M. M. (2016). Package ‘indicspecies’. *indicators*, 8, 1.
69. Dormann, Fründ, Blüthgen, & Gruber (2009) C.F. Dormann, J. Fründ, N. Blüthgen, B. Gruber Indices, graphs and null models: Analysing bipartite ecological networks *The Open Ecology Journal*, 2 (2009), pp. 7-24.

Figure legends:

Figure 1. a) Proportion of reads belong to the different bacterial phyla present in soils treated with living compost tea (“Treated”), with sterilized compost tea (“Control”), or from tea samples taken prior to application (“Tea sample”).samples, controlled soil samples (Control) and tea samples. b) Alpha-diversity of bacterial communities (respectively ASV richness, exponential Shannon and inverse Simpson diversity) c) Principal component analysis (PCA) of Hellinger-transformed bacterial relative abundances. Different symbols represent different treatments, and symbol colors indicate the experimental block to which sample belonged. Symbols represent the mean scores of samples from a given plot, and error bars represent 95% confidence intervals.

Figure 2. Ternary triangle presenting the relative distribution of reads from each ASV in treated plots, control plots or tea samples. Each symbol represents an ASV. Blue symbols are ASV indicator for tea samples; brown symbols are ASV indicator for control plots; the red symbol is the only ASV indicator for both tea and treated plots; grey symbols are those ASVs that are not indicator for any category.

Fig. 3: Boxplots showing plant growth (left) or yield (right) in control plots (red) or treated plots (cyan).