Article

Spatial and temporal variation of NO_2 vertical column densities (VCDs) over Poland: Comparison of the Sentinel-5P TROPOMI observations and the GEM-AQ model simulations

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Abstract: TRPOMI instrument aboard Sentinel-5P is a relatively new, high-resolution source of information about atmosphere composition. One of the primary atmospheric trace gases that we can observe through it is nitrogen dioxide. By now, we were using the chemical weather model (GEM-AQ) as a mean for estimating nitrogen dioxide concentration on a regional scale. Although well established in atmospheric science, the GEM-AQ simulations were always based on emission data, which in the case of the energy sector were reported by stack owners. In this paper, we attempted to compare the TROPOMI and GEM-AQ derived VCDs over Poland with a particular focus on large point emitters. We also checked how cloudy conditions influence TROPOMI results. Finally, we tried to link the NO_2 column number densities with surface concentration using boundary layer height as an additional explanatory variable

Keywords: air pollution; NO₂; Sentinel-5P; TROPOMI; GEM-AQ; Poland

1. Introduction

Nitrogen oxides ($NO_x = NO + NO_2$) play a significant role in tropospheric chemistry processes. As oxidisers precursors, they contribute to the tropospheric ozone formation process. Most of NO_x emissions are released as a form of NO molecules, but they quickly convert to NO_2 . Thus tropospheric NO_2 is commonly used as a more chemically stable proxy for NO_x emissions [1]. There are two major types of NO_x emissions - associated with traffic and industrial emissions, coming from high temperature combustion. The former is located at the earth surface and distributed proportionally to the road network, the latter at stacks located at bigger industrial incineration plants. For the most significant industrial NO_x sources in Poland, stack height is roughly within the range of 100-300 meters. This paper focuses on large point emitters since they are an issue of great concern, and their environmental impact exceeds the local scale.

There are several methods for obtaining gridded NO_2 estimates on larger than a local scale. To name the most significant: chemical transport models (CTMs) or online chemical weather models, spatial interpolation of station-based measurements, empirical models (like Landuse regression LUR or socioeconomic regression [2]), remote sensing (satellite or much less common on operational scale - aerial). Since each method has its intrinsic strengths and weaknesses, synergistic use of multiple sources of information and data-driven methods (also known as data assimilation or data fusion methods) is also gaining increasing attention [3,4].

Within the satellite remote sensing of atmospheric pollutants, significant progress has been made in recent decades. Starting from the first operational ultraviolet spectrometer, which was capable of delivering gridded data with a pixel size of $40 \times 320 \, km^2$ (Global Ozone Monitoring Instrument - GOME) in 1995[5], followed by SCIAMACHY aboard Envisat (res. XX in 2002 [6]) and GOME-2 [7].

An undeniable advantage of progress in satellite remote sensing of tropospheric NO_2 concentration is the growing archive record of past measurements on a global scale. This makes them a powerful tool for both spatial and temporal trend analysis[8–10] for environmental policy evaluation, industry and development assessment.

This paper aims to assess to what extent satellite-borne TROPOMI measurement can be used to evaluate the results of the operational chemical weather forecast model (GEM-AQ). We also check if TROPOMI results are valid under cloudy weather winter conditions within a temperate climate. Finally, we attempt to link a satellite-borne tropospheric column with the near surface concentrations using boundary layer depth as an additional regression variable.

2. Data and Methods

In this study we have used TROPOMI observations, the GEM-AQ model 24-h forecast from the operational run and the observations from national air quality monitoring network.

2.1. TROPOMI

TROPOMI, onboard Sentinel-5P satellite, is one of the most recently available instruments capable of monitoring NO_2 concentration in the atmospheric column. TROPOMI has a heritage to both the Ozone Monitoring Instrument (OMI) as well as to the SCanning Imaging Absorption spectroMeter for Atmospheric CartograpHY (SCIAMACHY). The Sentinel-5P is intended to extend the data records of these missions as well as be a preparatory mission for the Sentinel-5. Thus resolution as well as revisit time should be at least at the same level as for OMI and SCIAMACHY. Sentinel-5P performs on average one full and two partial scans over our area of interest per day.

The concentration retrieval algorithm (DOMINO, developed by KNMI) is based on the NO2 spectral properties in ultra-violet. It has previously been used for OMI [11] and with minor improvements, it has been adopted to TROPOMI data [12]. The retrieval algorithm uses several auxiliary atmospheric parameters within the processing, including atmospheric mass factor (AMF).

To provide the necessary meteorological data, the profile shape from the TM5-MP model is used (run at 1 x 1° resolution [13]). The surface albedo information is from a monthly OMI climatology (on a $0.5 \times 0.5^{\circ}$ resolution). Finally, a vertical column density (VCD) is provided by the algorithm in units mol/m^2 with a spatial resolution of approximately $7 \times 3.5 \ km^2$ (approx. $5.5 \times 3.5 \ km^2$ after 6 Aug 2019 [14]), aggregated as a tropospheric, stratospheric and total vertical column.

We used a level 2 product of TROPOMI (S5P_OFFL_L2_NO2) which is processed automatically by Copernicus Scientific data hub, up to 5 days after sensing. Data were downloaded from the data hub, using DHuSget 0.3.4 - an automatic sentinel data retrieving script. Within the level 2 product of TROPOMI, a quality assurance flag qa_value is provided for each pixel. This normalised flag is to be used as a threshold for discarding poor-quality retrievals from the useful ones. Most authors use the default threshold value of 0.75 [2,3,15,16]. However, this highly limits the number of retrievals in temperate climate due to intensive cloud cover, especially during winter months. According to TROPOMI ATBD [17] the value of 0.75 is recommended and should remove clouds, as well as scenes covered by snow, ice and other problematic retrievals. However, the value of 0.5 is also proposed as still good enough for model-comparison studies. Lower threshold (thus larger number of accepted retrievals) may be necessary if we still want to calculate monthly averages for the winter season. Exact the number will be given as examples in the results section. Therefore we decided to perform further processing using not only 0.75, but also 0.5 and 0.7 thresholds as a potential compromise.

Pixels which fulfil the above qa_value threshold requirement are also used to create a masking layers, which are later on used for calculating model-based monthly-average NO_2 column and model-based surface NO_2 concentration.

As the first processing step, TROPOMI data were regridded to GEM-AQ rectangular grid of size 300x470 and grid step 0.025°, using ESA Atmospheric Toolbox [18]. Secondly, regrided data were aggregated into monthly average raster. The term *monthly average*, although commonly used, maybe a

bit misleading in this context. Depending on location and time of the year - the monthly average may be an aggregate of 10 (in winter) to 40 (in the summer) cloud-free scans per pixel. TROPOMI NO2 column concentration is a scalar value, however it is produced with averaging kernel - an averaging vector, which describes how sensitive the instrument was to NO_2 at given time, altitude and location. In order to make GEM-AQ data comparable, the same averaging kernel was applied for tropospheric NO_2 column calculated from the model data.

2.2. The GEM-AQ model

The GEM-AQ is a semi-Lagrangian chemical weather model in which air quality processes (chemistry and aerosols), tropospheric chemistry are implemented online in the operational weather prediction model, the Global Environmental Multiscale (GEM) [19] model, which was developed at Environment Canada. The Gas-phase chemistry mechanism used in the GEM-AQ model is based on a modified version of the Acid Deposition and Oxidants Model (ADOM)[20] where additional reaction in the free troposphere was included [21].

The GEM-AQ model instance, run at the Institute of Environment Protection (Poland), is an ensemble member in CAMS50 and hence undergoes evaluation against satellite observation in the scope of CAMS84. However, the model output requested for column calculations reaches only 5km. For the sake of this paper, the entire troposphere was used.

Earlier study based on the comparison of the tropospheric NO2 column with satellite observations GEM-AQ with SCIAMACHY observations addressed the spatial correlation with total NOx emission fluxes [22]. Since the TROPOMI instrument provides significantly better resolution than Envisat SCIAMACHY, it is now feasible to focus on particular categories of emissions. We chose to focus on significant industrial NO_x sources because of the intensive contrast to the local NO_2 background.

Large emission sources within the model are driven by emission data from the national emission inventory. This data are based on annual reporting obligation, which is fulfilled by the facilities owners. Annual emission are transformed into monthly emission rates using weighting factor from annual emission profiles. Emission profiles are assigned to so-called SNAP categories [23]. In the case of NO_x emissions over Poland, the largest point emissions are assigned to SNAPs 1 (energy production from coal burning), 3 (non-energy manufacturing industry e.g. concrete or steel production) and 7 (road transport). Traffic emissions are considered to be uniform during the whole year, while SNAPs 1 and 3 are expected to follow a typical pattern high in winter, low in summer (fig. 1)

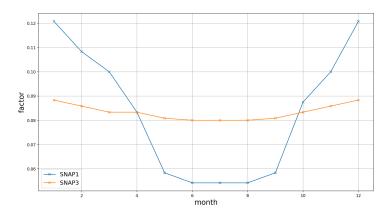


Figure 1. Annual emission profiles for energy production sector (SNAP 1) and non-energy manufacturing industry (SNAP 3)

The GEM-AQ model is set up to perform calculations using 28 vertical layers, out of which the lower 21 layers are classified as the troposphere. Troposphere averaging kernel is provided as an auxiliary variable of the TROPOMI level 2 NO_2 product. Averaging kernel values are provided at 35 levels of the TM5 model, which is the atmosphere model used within TROPOMI level 1 to level 2

processing [17]. TM5 averaging kernel is then linearly interpolated to GEM-AQ 28 levels (fig. 2). The NO_2 column number density is obtained for each layer using the following equation:

$$c_{NO_2,k} = f_k tn d_k \Delta z_k \tag{1}$$

where:

 $f_k[ppb]$ - molecular mixing ratio

 $tnd_k[molec/m^3]$ - total number density

 $\Delta z_k[m]$ - layer depth

The NO_2 column number density in the whole tropospheric column is then calculated using column number density from each GEM-AQ layer and averaging kernel derived from TROPOMI image:

$$C_{NO_2} = \frac{1}{\sum_k avk_k} \sum_{k=1}^{28} c_{NO_2,k} avk_k \tag{2}$$

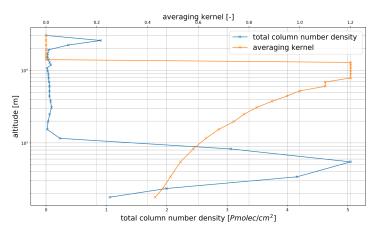


Figure 2. Example *NO*₂ vertical profile from GEM-AQ model and the TROPOMI-derived, troposphere averaging kernel, extracted at power plant stack location on 1st of April 2019

2.3. Boundary layer depth

The boundary layer is the lowest part of the troposphere, which is directly influenced by Earth surface and responds to these forcings in a short time scale [24]. Significant NO_x emissions occur within the boundary layer, while a satellite sensor observes the whole tropospheric column integrated. Therefore we expect boundary layer depth to be an additional variable that explains to what extent is the tropospheric column affected by concentrations from the boundary layer.

There are several ways of estimating boundary layer depth. Since GEM-AQ is an online chemical weather model model with the meteorological component, we decided to use Gradient Ricardson Number R_i with a critical value of $R_c = 0.025$. We assume that when $R_i < R_c$ we are within the boundary layer and turbulent mixing is the dominant form of transport [24]. The Gradient Ricardson Number is calculated as:

$$R_{i} = \frac{\frac{g}{\theta_{v}} \frac{\partial \theta_{v}}{\partial z}}{\left(\frac{\partial u}{\partial z}\right)^{2} + \left(\frac{\partial v}{\partial z}\right)^{2}} \tag{3}$$

where θ_v is a virtual potential temperature, u and v are horizontal components of the velocity vector, resulting from the meteorological component of the GEM-AQ model.

2.4. Surface observations

Observations of surface NO_2 concentrations during 2019 were obtained from the Chief Inspectorate of Environment Protection which is responsible for air quality monitoring in Poland. The dataset includes results from 112 automatic stations measuring with hourly time step.

3. Results

3.1. Overall Performance

Before detailed analysis, we performed a general linear regression analysis of TROPOMI NO_2 tropospheric column retrieval. We expect tropospheric columns retrieved using TROPOMI and the GEM-AQ model to be linearly correlated over the whole area of interest. Since we do not expect any additional bias, we assume that the noise is of Gaussian nature and the following regression equation is expected to be fulfilled:

$$N_{v,GEM}^{trop} = a * N_{v,TROPOMI}^{trop} + b (4)$$

where $N_{v,GEM}^{trop}$ is the monthly averaged GEM-AQ model-based tropospheric NO_2 column number density, $N_{v,TROPOMI}^{trop}$ is the monthly averaged TROPOMI-based tropospheric NO_2 column number density, a and b are regression parameters.

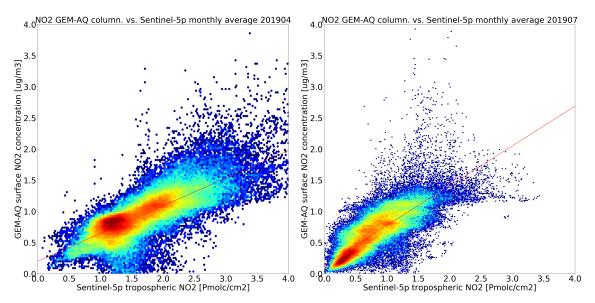


Figure 3. Scatter plots of monthly averaged tropospheric column retrieved by TROPOMI (horizontal axis) and GEM-AQ model (vertical axis) for **(A)**April and **(B)**July

Table 1. Parameters of equation 4, fitted to monthly averaged tropospheric column rasters; goodness of fit for each monthly average

| Month | a | b | R^2 | MSE |
|-----------|------|------|-------|------|
| January | 1.4 | 0.28 | 0.13 | 1.29 |
| February | 0.92 | 0.3 | 0.36 | 0.67 |
| March | 0.38 | 0.44 | 0.36 | 0.11 |
| April | 0.45 | 0.2 | 0.53 | 0.05 |
| May | 0.42 | 0.23 | 0.59 | 0.06 |
| June | 0.51 | 0.22 | 0.37 | 0.04 |
| July | 0.64 | 0.14 | 0.66 | 0.04 |
| August | 0.59 | 0.15 | 0.45 | 0.04 |
| September | 0.54 | 0.29 | 0.52 | 0.06 |
| October | 0.75 | 0.06 | 0.63 | 0.14 |
| November | 0.78 | 0.09 | 0.5 | 0.48 |
| December | 0.8 | 0.58 | 0.4 | 0.69 |

Table 1 summarises fitting results. The best (in terms of high R^2 and low MSE) linear regression was obtained for the July monthly average tropospheric column. In general, the MSE value follows the pattern - low during summer months, higher during winter months. R^2 does not seem to reveal any annual pattern. Thus it is either cloud cover or emission underestimation, making GEM-AQ and TROPOMI tropospheric column slightly different.

Both scatter plots (fig. 3) and regression parameters (a < 1) suggest that except for winter months, on a regional scale GEM-AQ model underestimates the NO_2 tropospheric column number densities. However at this stage, it is still questionable if it is overestimation caused by TROPOMI or underestimation by GEM-AQ.

3.2. The choice of qa_value

TROPOMI NO_2 OFFL product was processed by the DOMINO algorithm (version 1.2) on the ESA side. One of the auxiliary outputs of this algorithm is the quality assurance flag (qa_value). According to TROPOMI NO2 ATBD [17], the threshold of 0.75 should be used in order to remove clouds, pixels covered by snow and other problematic retrievals. Setting the threshold at 0.75 is sufficient for summer months, however in winter (November - February) only a few (less than 10) satellite images per month satisfy this condition (fig. 4 C,D). Reducing qa_value threshold to 0.7 or 0.5 may lead to some improvement (fig. 4 A,B). However, a lower threshold leads to underestimation in comparison to modelling results (fig. 8).

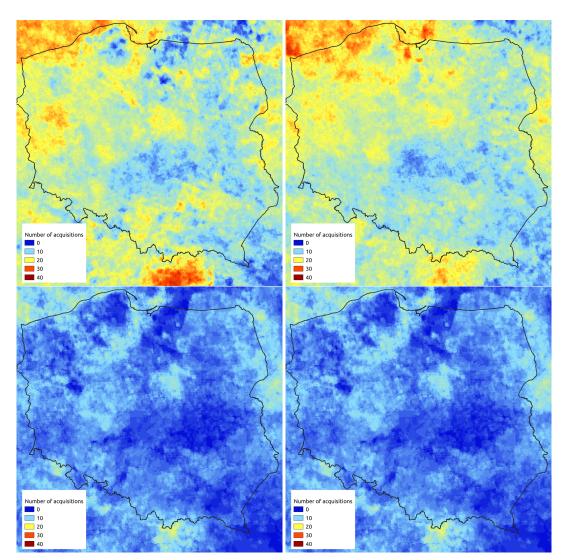


Figure 4. Number of pixels available for monthly averaging in January 2019 for given *qa_value* threshold **(a)**0.5 **(b)**0.7 **(c)**0.75 **(d)**0.8

3.3. Spatial Distribution

We investigated the spatial distribution of NO_2 . As figure 5 shows, the monthly averaged TROPOMI tropospheric NO_2 column reproduces the locations of large NO_x point sources. At the same time, this spatial pattern is not reproduced in the GEM-AQ surface layer. This confirms the fact, that the TROPOMI instrument at satellite level is not sensitive to surface layers. The only location where TROPOMI tropospheric column seems to be better correlated with surface concentration than with GEM-AQ tropospheric column is the coastal area, near the city of Gdańsk. Although there are no significant NO_x point source there - relatively higher values of column number density on the TROPOMI column (fig. 5 (A) as well as model-based surface concentration (fig. 6 B). This may be explained by harbour emission, which may be underestimated in the GEM-AQ model. Due to local see-breeze circulation the whole tropospheric column could be well mixed.

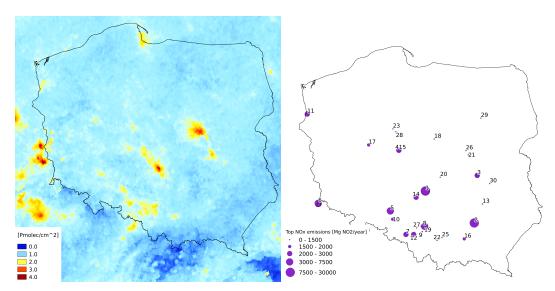


Figure 5. (**A**)Satellite distribution of NO_2 tropospheric column density (monthly average - June 2019) (B) Locations of top 30 NO_x emission point sources in Poland.

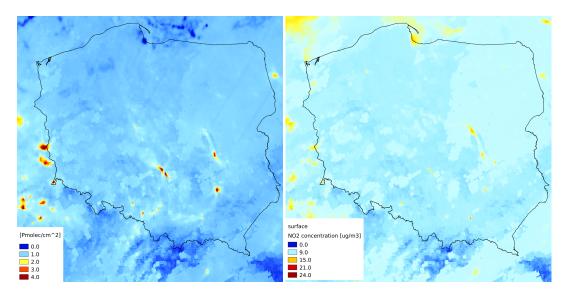
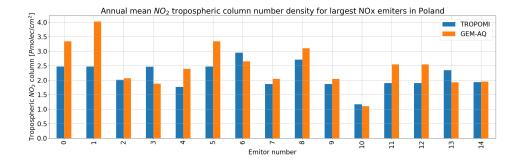


Figure 6. (A) GEM-AQ model-based NO_2 tropospheric column density (monthly average - June 2019) (B) GEM-AQ model-based NO_2 concentration at surface level (monthly average - June 2019)

The troposphere NO_2 column number density reveals not only the locations of major point sources but also the dominant wind direction. As the monthly average tropospheric NO_2 column is an average of non-cloudy days (mornings) also the resulting spatial distribution is highly dependent on accidental wind direction. Comparison of model-based and satellite-borne tropospheric NO_2 column over the whole domain (like fig.3 and eq.4) may be biased by a small error in modelled wind direction, which may lead to wrong concentration distribution.

In order to make the comparison less wind-depended we extracted the troposphere column number density value from pixels surrounding locations of the fifteen major point emitters (fig.5B) from both TROPOMI results and the GEM-AQ model. Yearly-averaged values for GEM-AQ and TROPOMI in most cases agree within the margin of 15%. Only in the case of tree emitters (out of fifteen) GEM-AQ model seems to underestimate tropospheric column number density, in other cases a small overestimation by the model is visible (fig. 7).



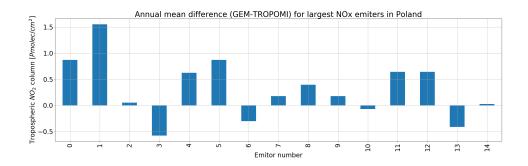


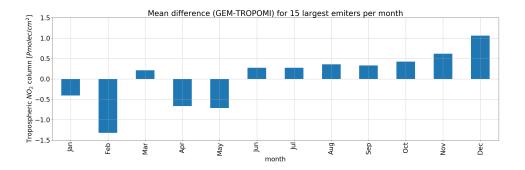
Figure 7. (A) Annual mean tropospheric NO_2 column number density extracted from pixels surrounding fifteen largest point emitters, from both TROPOMI acquisition and GEM-AQ model. **(B)** Annual mean difference between GEM-AQ and TROPOMI tropospheric column for fifteen largest emitters; Emitter numbers are the same as on figure 5

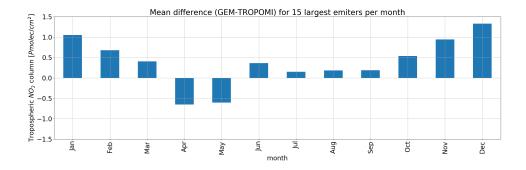
3.4. Temporal Comparison

Both NO_x emissions and NO_2 concentrations follow the same annual pattern - low in summer and high in winter. This fact is due to higher energy demand and low wind velocity episodes during the winter months. Moreover some of the largest coal-burning power plants in Poland are also a source of heat for city-wide heating systems. Thus they do burn more coal during low temperatures periods.

Because of cloud cover and non-point sources (road transport) we decided to analyse the temporal pattern only over the largest NO_x emitters. The difference between GEM-AQ tropospheric column and TROPOMI tropospheric column seems to be the smallest during the summer months (less than $0.5Pmolec/cm^2$, fig 8). In autumn the difference starts to grow and it exceeds $1.0Pmolec/cm^2$ in December.

The choice of the qa_value threshold seems to have a significant influence in January and February. For qa_value =0.5 TROPOMI returns higher values than the GEM-AQ model. This is probably due to partial cloud cover, which would be filtered out, when a higher qa_value is chosen. Regardless the qa_value April and May NO_2 column concentration seem to be underestimated in GEM-AQ model, which may be connected with overestimated ozone production during these months.





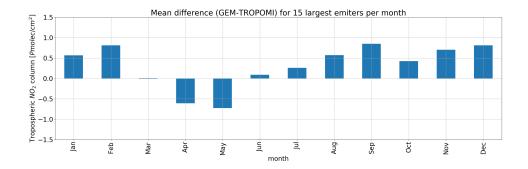


Figure 8. Mean difference in tropospheric column number density per month for different qa_value threshold: **(A)** $qa_value = 0.5$, **(B)** $qa_value = 0.7$, **(C)** $qa_value = 0.75$

3.5. Relation to near surface concentration

Although there are authors [2] who present linear relation between the near surface concentration and NO_2 tropospheric column, according to averaging kernel vertical distribution, the TROPOMI instrument is not very sensitive to NO_2 concentration surface level values(fig.2). It is probably hindered by the sensitivity at higher levels of the troposphere. Therefore a more complex relation linking NO_2 near surface concentration and the tropospheric column is needed.

A concept of explaining tropospheric NO_2 column density using nonlinear regression against surface concentration and boundary layer depth was introduced by Dieudonne et al.[25]. Further on it was applied to TROPOMI data over Paris by Lorrente [26], who showed that NO_2 surface

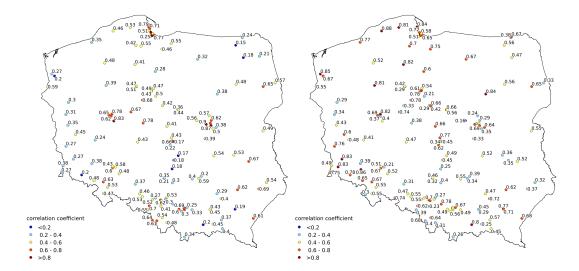


Figure 9. The correlation coefficient for nonlinear regression equation 6 in **(A)**April 2019 and **(B)** September 2019

concentration c_{surf} , tropospheric vertical column number density N_v^{trop} and boundary layer depth h the following empirical equation:

$$N_v^{trop} = K[0.244h(c_{surf} - 1.38) + 0.184(c_{surf} - 2.83)]$$
(5)

where *K* is a constant conversion factor $(1.31 \cdot 10^{15} molc/cm^2)$. We decided to introduce a more general nonlinear equation:

$$N_v^{trop} = [(a \cdot h + b) \cdot c_{surf} - c \cdot h + d]$$
 (6)

Parameters of equation 6 were fitted using Levenberg–Marquardt algorithm[27]. Fitting was performed separately for each measurement station in each month. Stations, where the number of TROPOMI tropospheric column values were lower than ten, were discarded. Therefore the results were highly dependent on cloud conditions. The best results were obtained for April and September 2019 (fig. 9). The spatial pattern of the correlation coefficient reveals that the equation 6 performs reasonably well within larger cities and densely populated area (fig. 9). This is probably caused by the larger contribution of road traffic NO_x emissions to the tropospheric NO_2 VCD.

4. Conclusions

Over the domain of the operational air quality forecast, performed routinely using the GEM-AQ model, we examined the results of the latest fine-scale satellite instrument (TROPOMI aboard Sentinel-5P) from the year 2019. The key findings from this study are the following:

- 1. In general, the GEM-AQ model tends to underestimate the NO_2 tropospheric column number density, which may be caused by either too intense mixing in the atmosphere, sink of NO_2 into further chemical processes (e.g. tropospheric ozone production) or too small background concentration
- 2. When looking at locations next to major NO_x emissions, the GEM-AQ model and TROPOMI converge reasonably well. Minor differences should be explained by individual emission examination
- 3. The annual temporal pattern is not properly reproduced by the TROPOMI instrument. It seems that cloud cover (thus qa_value threshold) and the number of satellite scenes averaged into monthly average play an important role. Lowering the qa_value during the summer months

- improve the convergence between TROPOMI and GEM-AQ, while during the winter months, it acts oppositely.
- 4. The relation between near surface concentration and troposphere column number density can be parametrised using boundary layer depth as an additional explanatory variable.

From the above findings, we conclude TROPOMI is a powerful and independent from ground measurements source of NO_2 distribution data. Although column number density is not to be used directly with surface concentration, it is still useful for validating modelling results, and after some additional processing, it can also be used for estimating surface concentrations in urban areas.

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