

ARIMA forecasting of PM_{2.5} and PM₁₀ trends: effects of continuing social distancing on air quality in a Southeast Asian urban area

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Abstract It has been hypothesized that social distancing as the prevention measures for COVID 19 can affect the air quality including PM_{2.5} and PM₁₀ in urban areas. According to this situation, this study aims to compare the PM_{2.5} and PM₁₀ before and after the implementation of social distancing. Likewise, this study also forecasts the benefits of social distancing on PM_{2.5} and PM₁₀ if social distancing period is continued and extended. To achieve these objectives, an Auto Regressive Integrated Moving Average (ARIMA) model to investigate the daily PM_{2.5} and PM₁₀ trends has been developed for social distancing periods (March–May 2020) and after May as well. The model confirms that if social distancing period is extended after May 2020 then the PM_{2.5} and PM₁₀ are estimated will be 4% and 9% lower. To confirm that the PM_{2.5} and PM₁₀ reductions are only due to social distancing effect, the study has investigated the possible effects of wind speed and rainfall on PM_{2.5} and PM₁₀. Nonetheless, the reductions do not correlate with those factors. To conclude social distancing should be considered as an option to control PM_{2.5} and PM₁₀ in urban areas.

Keywords: ARIMA, COVID 19, forecast, PM_{2.5}, PM₁₀, social distancing

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Presently, particulate matter (PM) is considered as one of the pollutants that can have impacts on urban air quality (Badura 2018). Volcanic dust or desert dust particles and sea spray aerosols are known as the origin of PM in natural settings. Nonetheless, in urban setting the PM is originated from anthropogenic activities. Those activities including transport, agriculture, industrial processes, and household fuel combustions.

The particulate matters which are also known as aerosols are divided into 2 types according to its coarseness. Those suspended particles in the atmosphere include fine particulate matter (PM_{2.5}) and coarse particulate matter (PM₁₀). The particle diameters of PM_{2.5} and PM₁₀ are <2.5 μm and <10 μm, and both measured as μg/m³. The PM_{2.5} and PM₁₀ have different sources and characteristics and able to interact with solar radiation, affect air quality, visibility, and the climate system including surface temperature and rain fall as well (Cruz and Pandis 1997, Sun and Ariya 2006, Zhang and Cao 2015).

Currently, the presence of fine particulate pollution especially in urban areas in the form of recorded PM_{2.5} and PM₁₀ has become a global issue due to its impact on human health, air quality, and the climate system as well. Correspondingly, routine air quality monitoring has been established in many countries for regular measurement of PM_{2.5} and PM₁₀, for epidemiological studies as well as for the management of air quality (Al-Saadi *et al.*

2005, Gomišček *et al.* 2004).

Regarding epidemiological studies, the presence of current social distancing as parts of current COVID 19 pandemic prevention has been considered has an effect on the air quality including PM_{2.5} and PM₁₀. A recent study by Xu *et al.* (2020) has provided a very comprehensive coverage regarding the impacts of the COVID 19 event on air quality in central China.

In urban areas mainly in Southeast Asian countries, there are several literatures that have discussed extensively the PM_{2.5} and PM₁₀ contents (Mohddin and Aminuddin 2014, Noor *et al.* 2015, Hossen and Hoque 2016). Nonetheless the COVID 19 event prevention in the form of social distancing related air quality studies are still limited. Meanwhile, several urban areas in Southeast Asian countries have experienced significant COVID 19 events. One of the populated urban areas is Jakarta. In here, social distancing was scheduled from mid-March through late May 2020. According that situation, this study aims to compare the PM_{2.5} and PM₁₀ before and after the implementation of social distancing. Likewise, an ARIMA model is used to investigate how far social distancing will affect the PM_{2.5} and PM₁₀ if social distancing period is extended.

Methodology

PM_{2.5} and PM₁₀ monitoring

The PM_{2.5} and PM₁₀ were monitored daily from March to May 2020. The March was representing the period before social distancing and April and May were representing the periods

after social distancing. The PM_{2.5} and PM₁₀ data were obtained from meteorology agency and covered Jakarta, the one of populated urban areas in Southeast Asian. All PM_{2.5} and PM₁₀ data were measured as µg/m³.

Wind and rainfall monitoring

It has been hypothesized that the meteorological parameters including wind speed and rainfall can affect the PM_{2.5} and PM₁₀ (Wang and Ogawa 2015, Zhang *et al.* 2017). In this study and to complement the PM_{2.5} and PM₁₀ analysis, the daily wind speed (km/h) and rainfall (mm/day) data were also collected from meteorology agency.

ARIMA forecasting

ARIMA is abbreviation for Auto Regressive Integrated Moving Average. This statistical approach is a versatile tool in forecasting an event in many fields (Alsharif *et al.* 2019, Tektas 2010). In ARIMA, autoregressive is stationarized series lags in the forecasting equation and moving average is the forecast error lags. The ARIMA forecasting equation is constructed as follows. First step to construct the equation is denoting the dth difference of Y by y below:

$$\text{If } d = 0: y_t = Y_t$$

$$\text{If } d = 1: y_t = Y_t - Y_{t-1}$$

$$\text{If } d = 2: y_t = (Y_t - Y_{t-1}) - (Y_{t-1} - Y_{t-2}) = Y_t - 2Y_{t-1} + Y_{t-2}.$$

Based on y terms, the ARIMA forecasting equation is:

$$\hat{y}_t = \mu + \phi_1 y_{t-1} + \dots + \phi_p y_{t-p} - \theta_1 e_{t-1} - \dots - \theta_q e_{t-q}$$

The ARIMA model is used to forecast the trends of $PM_{2.5}$ and PM_{10} after the end of May 2020.

Results

A quantile–quantile (QQ) plot of $PM_{2.5}$ and PM_{10} is shown in Figure 1 and 2. The QQ plots were presented as comparisons of before and after social distancing periods. Before social distancing period, there were high frequencies of $PM_{2.5}$ within the ranges of 150–200 $\mu\text{g}/\text{m}^3$. Nonetheless, after social distancing has been implemented, $PM_{2.5}$ within the ranges of 100–150 $\mu\text{g}/\text{m}^3$ were observed frequently. The similar patterns were also observed for PM_{10} for before and after social distancing periods (Figure 2). Even though the $PM_{2.5}$ and PM_{10} showed a fluctuation pattern, there were reductions observed in April and May after social distancing was implemented.

The Figure 3 and 4 shows how the $PM_{2.5}$ and PM_{10} were correlated with the external factors including the wind speed and rainfall. The frequent wind speed ranges were 1–15 km/h while rainfalls within the range of 1–2 mm/day were common. Nonetheless, according to the correlation plots there were no significant effects of wind speed and rainfall on the both $PM_{2.5}$ and PM_{10} .

The ARIMA forecasting confirms that the $PM_{2.5}$ average will be 4% lower and the trend remains stable in a period after social distancing (Figure 5, 6). The lower average value was also observed for forecasted PM_{10} . For the PM_{10} , it is

estimated that the average will be 9% lower and it also remains stable.

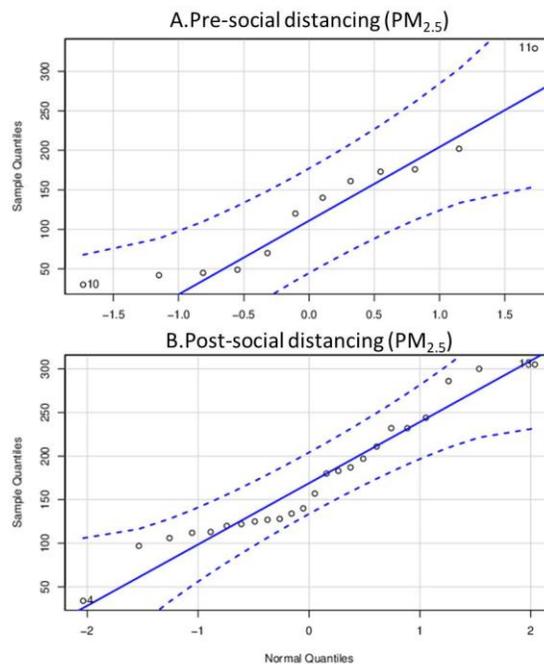


Figure 1. Normal probability plots (QQ-plot) of daily $PM_{2.5}$ before/pre/March (A) and after/post/April-May (B) social distancing.

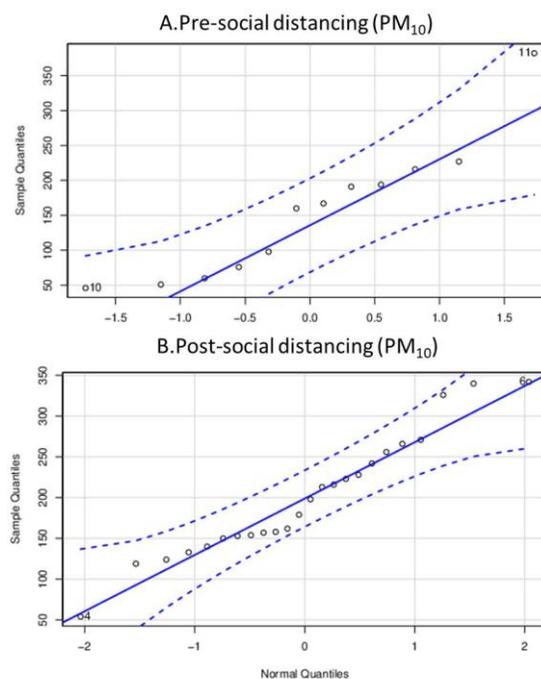


Figure 2. Normal probability plots (QQ-plot) of daily PM_{10} before/pre/March (A) and after/post/April-May (B) social distancing.

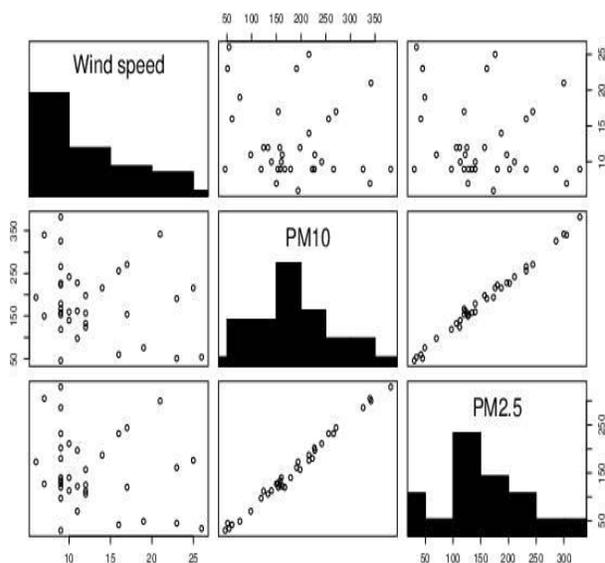


Figure 3. Correlation plots of daily $PM_{2.5}$ and PM_{10} with wind speed (km/h).

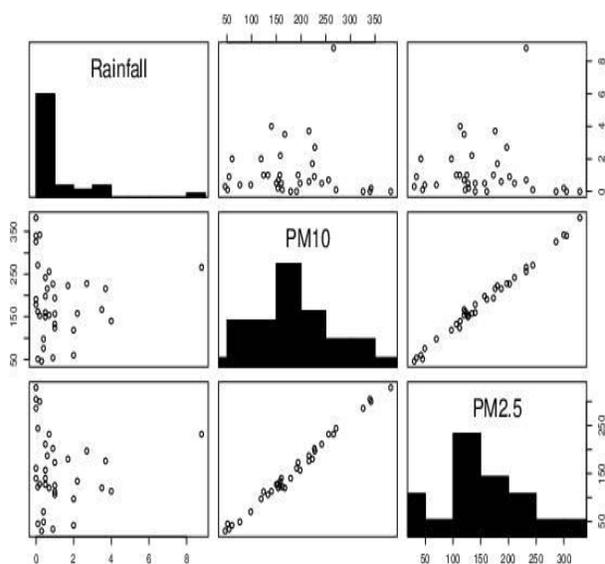


Figure 4. Correlation plots of daily $PM_{2.5}$ and PM_{10} with rainfall (mm/day).

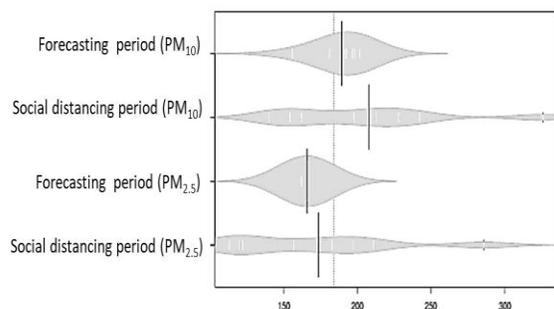


Figure 5. Violin plots of $PM_{2.5}$ and PM_{10} average for social distancing and forecasting periods.

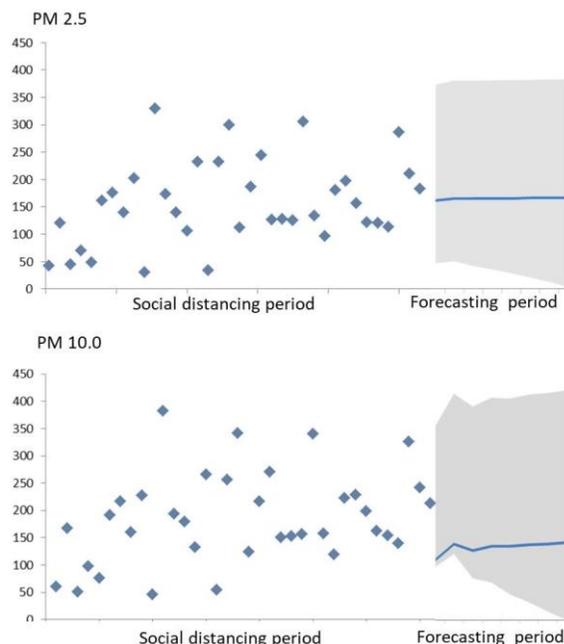


Figure 6. ARIMA model of daily $PM_{2.5}$ and PM_{10} for social distancing and forecasting periods.

Discussion

The $PM_{2.5}$ and PM_{10} measured in this study were also comparable to other urban areas. The data provided by Zhang *et al.* (2017) showed that in an urban industrialized area, the range of $PM_{2.5}$ was 6–209 $\mu\text{g}/\text{m}^3$ and PM_{10} can be higher with the range was 13–422 $\mu\text{g}/\text{m}^3$. Another study in 10 urban areas found the monthly average $PM_{2.5}$ ranges were >150–250 $\mu\text{g}/\text{m}^3$ and >250–300 $\mu\text{g}/\text{m}^3$ for PM_{10} (Li *et al.* 2020). In this study the ranges for $PM_{2.5}$ and PM_{10} are 32–329 $\mu\text{g}/\text{m}^3$ and 46–382 $\mu\text{g}/\text{m}^3$ respectively.

Both the PM_{10} and $PM_{2.5}$ were observed lower after social distancing than before social distancing was implemented. There are several literatures that have reported the impacts of social distancing on PM_{10} and $PM_{2.5}$. A decrease in air pollutant concentration as a consequence

of social distancing in the form of mobility reduction in São Paulo since March 22, 2020 has been recorded by Freitas *et al.* (2020). The $PM_{2.5}$ values originated from diesel vehicle exhaust, soil resuspension, and biomass burning were on average $3.7 \mu\text{g}/\text{m}^3$ lower than values without mobility reduction. Similarly, PM_{10} average concentrations were also $4.6 \mu\text{g}/\text{m}^3$ lower. The similar conditions were also reported in the Asian urban areas. In 3 cities in China after the epidemic prevention and control actions had been taken, the average values of $PM_{2.5}$ and PM_{10} were observed 30% and 40% respectively lower than the values without epidemic prevention (Xu *et al.* 2020).

A comprehensive study of social distancing impacts on air quality in particular Southeast Asian country has been reported by Abdullah *et al.* (2020). In their study, they reported that social distancing in Malaysia or widely known as Movement Control Order (MCO) has been found to reduce $PM_{2.5}$ levels. In several cities in Malaysia, the range variations of $PM_{2.5}$ reductions were 6–23%, 8–16%, 11–13%, 7–11%, 15–23%, and 9–16%.

The above aforementioned situations characterized by $PM_{2.5}$ and PM_{10} reductions are best explained by the anthropogenic activities. Social distancing was fundamental detrimental factors causing the reduction of daily routine activities commonly occurs in the urban, commercial, and industrial areas. Nonetheless, before social distancing was applied, there are significant activities lead to the peak of PM_{10}

and $PM_{2.5}$ Noor *et al.* (2015) reported that the PM_{10} peaks in residential areas were frequently observed at 9.00 am when people were leaving their home and reach the maximum peak at 10.00 pm when they were returning from school and work. During some particular times, the numbers of people performing their daily routine increase the motor vehicle numbers along with their emissions. The evening PM_{10} peak starts from 6.00 pm to 11.00 pm. The higher PM_{10} and $PM_{2.5}$ before social distancing period were evidence of the influence of traffic and other anthropogenic activities on PM_{10} and $PM_{2.5}$ which is still allowed at that period (Garcia *et al.* 2016).

Beside anthropogenic factors, the PM_{10} and $PM_{2.5}$ are related to other regional external factors. For instance, Noor *et al.* (2015) reported that the PM_{10} variations are influenced by wind, rainfall, and season. The sudden PM_{10} peak can happens due to the wind that contains and delivers particles from other places. Nevertheless, PM_{10} can be so low due to the rainfall (Sansuddin *et al.* 2010). A seasonal variation of rainfall can also affect the PM_{10} seasonal fluctuation (Juneng *et al.* 2009). It requires the increase of rainfall from 20 mm/day to 100 mm/day to reduce the $PM_{2.5}$ ranges from $20\text{--}35 \mu\text{g}/\text{m}^3$ to $5\text{--}7 \mu\text{g}/\text{m}^3$ (Li *et al.* 2017, Wang and Ogawa 2015).

Season was also responsible for the PM_{10} variations (Hamid *et al.* 2018). It was observed that during the Southwest Monsoon the annual average of PM_{10} concentration had exceeded the

thresholds while the PM_{10} concentration during the Northeast Monsoon was below the acceptable levels.

Despite the previous studies have found the fundamental correlations of the external factors including seasons, wind, and rainfall, nonetheless those factors are not significant in this study. Regarding the seasons, the location of this study is only having 2 seasons including the rain and dry seasons. This study was conducted during the homogenous rainy seasons with constants and stationary rainfalls. The recorded rainfalls during study period were dominated by the 0-2 mm/day ranges followed by the infrequent 3-8 mm/day ranges. While it takes a 20-100 mm/day rainfall range to cause variation and reduction of the $PM_{2.5}$ (Li *et al.* 2017).

Regarding the wind that assumed hypothetically can increase the PM_{10} and $PM_{2.5}$ due to the particles contained, this situation is not observed in this study. The study site was surrounded by settlement and suburban land uses rather than industrial, commercial, or emission generated land uses. Without the presence of industry in surrounding areas, then the prevailing winds will contain few particles and as consequence it will contribute less to the PM_{10} and $PM_{2.5}$.

Besides delivering the particles, the wind can also remove the particles and reduce the $PM_{2.5}$. Li *et al.* (2017) reported that the reductions of $PM_{2.5}$ from 35 to 15 $\mu\text{g}/\text{m}^3$ were caused by the increase of wind speed from 5 to 30 km/h.

While, in this study, the most frequent wind speed were ranging from 1 to 10 km/h and has less effects on $PM_{2.5}$.

Currently, there are several approaches can be used to forecast PM_{10} and $PM_{2.5}$, they include statistical models, chemical transport, and machine learning (Deters *et al.* 2017). Statistical models emphasized on single variable linear regression have generally shown a negative correlation between PM_{10} , $PM_{2.5}$ and different meteorological parameters (wind, rainfall, and temperature). The recent one is the ARIMA model. Unlike a pure statistical method, ARIMA model only requires the time series prior data to generalize the forecast. Likewise, this model can increase the forecast accuracy while keeping the number of parameters to a minimum. This is the reason why ARIMA model is increasingly used to predict air quality (Abhilash *et al.* 2018, Arafat and Amir 2012, Wang *et al.* 2009).

In this study, a social distancing related ARIMA forecasting of $PM_{2.5}$ and PM_{10} has been proposed. The finding confirms and forecasts that $PM_{2.5}$ and PM_{10} will decrease. A capability of ARIMA model to do the forecasting of $PM_{2.5}$ and PM_{10} has been discussed by Zhang *et al.* (2017). The model can also adopt the environmental factors that may affect the dynamic of $PM_{2.5}$ and PM_{10} . In their study, the forecasted PM_{10} was observed fluctuating according to the seasons. Most recent studies even though have investigated the forecasted $PM_{2.5}$ and PM_{10} , nonetheless none of them have discussed the forecasted $PM_{2.5}$ and PM_{10} resulted

from conditional factors in the form of policy in this case social distancing. However, as indicated by previous studies that environmental factors may influence the PM_{2.5} and PM₁₀ trends, then it is possible that conditional factors like tested in this study can affect the PM_{2.5} and PM₁₀ trends. As observed in this study, social distancing period factor if it is continued and extended can lower the PM_{2.5} and PM₁₀ in the future. Study by Zhang *et al.* (2018) has confirmed that policy factors can even contribute to reduce the PM_{2.5}. In their study, the forecasted PM_{2.5} ranges were reduced from 19–52 µg/m³ to 15–30 µg/m³. This reduction is assumed related to the local policy on air quality.

Conclusion

Social distancing period has caused reduction of PM_{2.5} and PM₁₀ ranges from 150–200 µg/m³ to 100–150 µg/m³. The wind speed and rainfall were having less contribution in affecting the PM_{2.5} and PM₁₀ in before and after social distancing periods. The ARIMA forecasting confirms that by continuing social distancing period, it is estimated that the PM_{2.5} and PM₁₀ will be 4% and 9% lower.

Recommendation

Social distancing period is forecasted can lower the PM_{2.5} and PM₁₀. For this reason, social distancing should be opted as a promising solution to control the PM_{2.5} and PM₁₀ mainly in the urban areas.

References

Abdullah S, Mansor AA, Liyana NN, Napi M,

Mansor WNW, Ahmed AN, Ismail M, Ramly ZTA. 2020. Air quality status during 2020 Malaysia Movement Control Order (MCO) due to 2019 novel coronavirus (2019-nCoV) pandemic. *Sci Total Environ.*

Abhilash M, Thakur A, Gupta D, Sreevidya B. 2018. Time Series Analysis of Air Pollution in Bengaluru Using ARIMA Model. *Ambient Communications and Computer Systems.* 413-426.

Al-Saadi J, Szykman J, Pierce RB, Kittaka C, Neil D, Chu DA, Remer L, Gumley L, Prins E, Weinstock L. 2005. Improving National Air Quality Forecasts with Satellite Aerosol Observations. *Bull. Am. Meteorol. Soc.* 86: 1249–1261.

Alsharif MH, Younes MK, Kim J. 2019. Time Series ARIMA Model for Prediction of Daily and Monthly Average Global Solar Radiation: The Case Study of Seoul, South Korea. *Symmetry.* 11(240)

Arafat RM, Amir HM. 2012. Time Series Analysis Model for Particulate Matter of Air Pollution Data in Dhaka City. *Asian Journal of Water, Environment and Pollution.* 9(4): 63-69.

Badura M. 2018. Evaluation of Low-Cost Sensors for Ambient PM_{2.5} Monitoring. *Journal of Sensors.*

Couvidat F, Bessagnet B, Vivanco MG, Real E, Menuet L, Colette A. 2018. Development of an inorganic and organic aerosol model (CHIMERE 2017 β v1.0): seasonal and spatial evaluation over Europe.

- Geoscientific Model Development*. 11: 165-194.
- Cruz CN, Pandis SN. 1997. A study of the ability of pure secondary organic aerosol to act a cloud condensation nuclei. *Atmos. Environ.* 31: 2205-2214.
- Deters JK, Zalakeviciute R, Gonzalez M, Yves Rybarczyk Y. 2017. Modeling PM_{2.5} Urban Pollution Using Machine Learning and Selected Meteorological Parameters. *Journal of Electrical and Computer Engineering*.
- Freitas ED, Ibarra-Espinosa SA, Gavidia-Calderón ME, Rehbein A, Abou Rafee SA, Martins JA, Martins LD, Santos UP, Ning MF, Andrade MF, Trindade RIF. 2020. Mobility Restrictions and Air Quality under COVID-19 Pandemic in São Paulo, Brazil. *Preprints*. 2020040515.
- García MÁ, Sánchez ML, de los Ríos A. 2019. Analysis of PM₁₀ and PM_{2.5} Concentrations in an Urban Atmosphere in Northern Spain. *Arch. Environ. Contam. Toxicol.* 76:331–345.
- Gomišček B, Hauck H, Stopper S, Preining O. 2004. Spatial and temporal variations of PM₁, PM_{2.5}, PM₁₀ and particle number concentration during the AUPHEP—Project. *Atmos. Environ.*, 38: 3917–3934.
- Hamid HA, Rahmat MH, Sapani SA. 2018. The classification of PM₁₀ concentrations in Johor Based on Seasonal Monsoons IOP Conference Series: *Earth and Environmental Science*. 140.
- Hossen MA and Hoque A. 2016,.A. Variation of Ambient Air Quality Scenario in Chittagong City: A Case Study of Air Pollution. *Preprints* 2016080174.
- Juneng L, Latif MT, Tangang FT, Mansor H. 2009.Spatio-temporal characteristics of PM₁₀ concentration across Malaysia. *Atmospheric Environment*. 43: 30.
- Li X, Feng Y, Liang H. 2017. The Impact of Meteorological Factors on PM_{2.5} Variations in Hong Kong. *IOP Conference Series: Earth and Environmental Science*. 78.
- Li X, Zhou X, Tong W. 2020. Spatiotemporal Analysis of Air Pollution and Its Application in Public Health. Elsevier.
- Mohddin S, Aminuddin N. 2014. The exposure assessment of airborne particulates matter (PM₁₀ & PM_{2.5}) towards building occupants: A case study at KL Sentral, Kuala Lumpur, Malaysia. *IOP Conference Series: Earth and Environmental Science*.
- Noor NM, Yahaya AS, Ramli N, Luca F, Abdullah M, Sandu AV. 2015. Variation of air pollutant (particulate matter - PM₁₀) in peninsular Malaysia: Study in the southwest coast of peninsular Malaysia. *Revista de Chimie -Bucharest- Original Edition-*. 66: 1443-1447.
- Sansuddin N, Yahaya AS, Ramli N, Yusof N, Ghazali NA, Al Madhoun W. 2010. Statistical analysis of PM₁₀ concentrations at different locations in Malaysia. *Environmental monitoring and assessment*. 180: 573-88.
- Sun J, Ariya P. 2006. Atmospheric organic and bio-aerosols as cloud condensation nuclei

- (CCN) :A review. *Atmos. Environ.*, 40: 795–820.
- Tektas M. 2010. Weather Forecasting Using ANFIS and ARIMA MODELS. *Environmental Research, Engineering and Management*. 1(51): 5 – 10.
- Xu K, Cui K, Young LH, Hsieh YK, Wang YF, Zhan J, Wan S. 2020. Impact of the COVID-19 Event on Air Quality in Central China. *Aerosol Air Qual. Res.* 20:915-929
- Zhang YL, Cao F. 2015. Fine particulate matter (PM_{2.5}) in China at a city level. *Sci. Rep.*5.
- Zhang B, Jiao L, Xu G, Zhao S, Tang X, Zhou Y, Chen G. 2017. Influences of wind and precipitation on different-sized particulate matter concentrations (PM_{2.5}, PM₁₀, PM_{2.5-10}). *Meteorology and Atmospheric Physics*.
- Zhang H, Zhang S, Wang P, Qinc Y, Wang H. 2017. Forecasting of particulate matter time series using wavelet analysis and wavelet-ARMA/ARIMA model in Taiyuan, China. *Journal Of The Air & Waste Management Association*. 67 (7): 776–788.
- Zhang L, Lin J, Qiu R, Hu X, Zhang H, Chen Q, Tan H, Lin D, Wang J. 2018. Trend analysis and forecast of PM_{2.5} in Fuzhou, China using the ARIMA model. *Ecological Indicators*. 95: 702-710.
- Wang J, Ogawa S. 2015. Effects of Meteorological Conditions on PM_{2.5} Concentrations in Nagasaki, Japan. *Int. J. Environ. Res. Public. Health*. 12(8): 9089–9101.
- Wang W, Guo Y. 2009. Air Pollution PM_{2.5} Data Analysis in Los Angeles Long Beach with Seasonal ARIMA Model. *International Conference on Energy and Environment Technology*, Guilin, Guangxi. 7-10.