1	Enemy impact on plant communities in a changing world					
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24 Highlights

- Predicting how global change affects our ecosystems remains a current challenge.
 Global change can directly affect plant communities, but it can also indirectly affect
 plant communities by altering species interactions.
- Plant enemies, such as herbivores or pathogens, can strongly affect plant productivity,
 diversity and composition. Their impact on plant communities is notoriously context dependent and may be strongly altered by global change drivers.
- We synthesize current knowledge on the context-dependency of plant-enemy
 interactions and illustrate how nitrogen enrichment, climate change, and elevated CO₂
 as well as a loss of plant diversity and insects, may alter enemy communities and their
 impact on plant communities.
- Understanding the indirect effects of global change on our ecosystems can
 fundamentally improve our ability to predict the consequences of global change.

37

38 Abstract

39 Indirect effects of global change via changing species interactions have been largely ignored 40 in studies predicting global change impacts on ecosystems. Antagonistic biotic interactions, 41 however, can strongly affect ecosystems and are likely to be affected by global change drivers 42 themselves. We synthesize current knowledge on the impact of invertebrate herbivores and 43 pathogens on plant productivity, diversity and community composition, and outline theory and 44 expectations on how important global change drivers – nitrogen enrichment, climate change and elevated CO², and plant and insect diversity loss, may affect enemy impact on plant 45 46 communities. We illustrate that our ability to predict global change impact requires a holistic 47 perspective, taking into account direct as well as indirect effects via the biotic component of 48 ecosystems.

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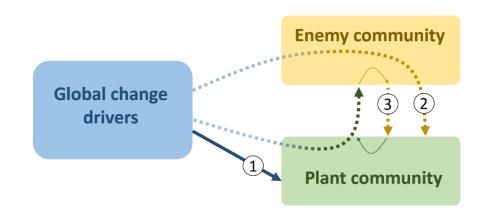
52 Global change can indirectly affect plant communities by changing species interactions

53 A major challenge is to predict how biodiversity and ecosystem functioning will respond to 54 environmental global change. Global change can directly affect the productivity, functional 55 composition and diversity of plant communities by changing environmental conditions [1–3], 56 however, it can also affect communities indirectly by altering species interactions [4, Fig. 1]. 57 Indirect effects have just recently started to be explored in studies on global change, mainly 58 mediated through changing competitive interactions [4,5] or plant mutualist interactions [6,7]. 59 However, we lack a general understanding of indirect effects mediated through altered 60 antagonistic trophic interactions, such as between plants and their herbivores or pathogens [8]. 61 Altered plant-enemy interactions following global change may profoundly affect, and even 62 reverse, ecosystem responses [9,10]. To predict ecosystem responses to environmental change, we need to also enhance our understanding of how global change drivers alter the interactions 63 64 of plant communities with their enemies.

Herbivores and pathogens can have strong impacts on plant communities, with large effects on primary production, species composition, plant diversity and ecosystem functioning [11–14]. The intensity of these interactions vary substantially in space [12,15,16], and enemy impact on ecosystems is likely to be altered by global change drivers. We still have a limited understanding of the key abiotic and biotic factors driving variation in enemy impact on plant communities [12,17], however developing general predictions is essential to understand how global change will affect our ecosystems.

In this review we summarize recent advances in our understanding of the impact of less well studied plant enemies, invertebrate herbivores and fungal pathogens, on three key aspects of plant communities: plant diversity, community composition and productivity, and highlight open questions and knowledge gaps. Then, we outline relevant ecological theory and empirical studies on how variation in environmental conditions will affect plant-enemy interactions, and 77 use this to develop predictions for how key global change drivers, nitrogen enrichment, climate 78 change, and increased CO₂, could indirectly affect plant communities via changing enemy 79 impact. We unpack these overall indirect effects and consider how global change might affect 80 enemy communities alone or how it might alter feedbacks between plant and enemy community structure. Finally, we consider how the current losses of plant and insect diversity 81 82 might alter enemy impact on plant communities. We do not discuss effects of global change 83 on the geographical distribution of plant species and their natural enemies (including biological 84 invasions), and effects on phenology.

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Fig. 1: Global change drivers can *directly* affect plant communities (solid line, (1)) but can also *indirectly* (dashed lines) affect plant communities through changing interactions of plant communities with their enemy communities. This can happen because global change drivers directly affect the enemy community ((2)) and hence their impact on plants, or because global change drivers directly affect the plant community which has consequences for the enemy community and hence their impact on plant communities ((3)).

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98 General impact of plant enemies on plant communities

99 Plant enemies can alter several aspects of plant communities, including productivity, diversity 100 and composition [12,13,18,19]. Enemies alter primary productivity by consuming or 101 preventing the accumulation of plant biomass. For herbivores, the strength of such top-down 102 control depends on whether they are limited by predation [20,21]. Herbivores and plant 103 pathogens can further be important drivers of plant coexistence and affect plant diversity 104 [13,22,23]. Even if enemies do not alter diversity, they frequently change competitive 105 interactions between plant species and alter the composition of plant communities [18]. We 106 discuss the impact of plant enemies on these three aspects of plant communities - productivity, 107 diversity, and composition.

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109 Impact of enemies on plant productivity

110 Plants are consumed by numerous organisms that feed in different ways and on different types 111 of plant tissue, above- or belowground. These enemies can be very restricted in their diet or can be highly generalist. Despite the high diversity of plant enemies, studies aiming to 112 113 understand their role in determining plant community dynamics have mainly focused on 114 vertebrate herbivores [24,25]. While highly debated in the past [26], today there is ample 115 evidence for their widespread, top-down effects on plant communities [11,12]. In contrast to 116 vertebrates, it has often been assumed that invertebrate herbivores and pathogens are too small, 117 specialised or strongly regulated by predators to have large effects on plant productivity [24]. 118 Several individual studies, however, have shown that invertebrate herbivores and fungal 119 pathogens *can* affect plant productivity [e.g. 13,27–29], but effects are often variable between 120 studies. While one meta-analysis on insecticide experiments in grassland found an increase in 121 primary production when insects where removed [30], a recent meta-analysis showed no 122 consistent top-down effect of aboveground invertebrate herbivores on plant biomass [12]. For

other groups, such as belowground invertebrates, molluscs or fungal pathogens, the impact on
plant communities is even less clear, although strong effects have been shown in some studies,
under certain conditions [28,29,31–33].

126

127 Impact of enemies on plant diversity

128 Enemies frequently alter competitive interactions between plants and can therefore alter the 129 *diversity* of plant species in a community [22,23,34]. Enemies can promote coexistence if they 130 more strongly attack the most abundant species within a community [34–37]. This can arise if 131 attack rate declines with decreasing host frequency (negative-frequency dependence). Enemies then drive niche differences between plant species, stabilizing coexistence (e.g. [23]). 132 133 Particularly the attack rate of specialist enemies has been suggested to decline non-linearly 134 with host abundance as rare hosts are difficult to locate and escape attack (Janzen-Connell 135 effect, e.g. [36,38]). Enemies can also promote coexistence and more strongly attack dominant 136 plant species through a correlation between palatability and dominance or a correlation 137 between sensitivity of plants to attack and dominance. This could be driven by, for instance, a 138 trade-off between the growth of plants in a resource rich environment and their defense against 139 enemies [39–42], or a trade-off between the competitive ability for belowground nutrients and 140 defense [43]. In this case, enemies promote coexistence by reducing the competitive ability 141 (fitness) of more vigorously growing plant species, equalizing fitness between dominant and 142 less dominant species.

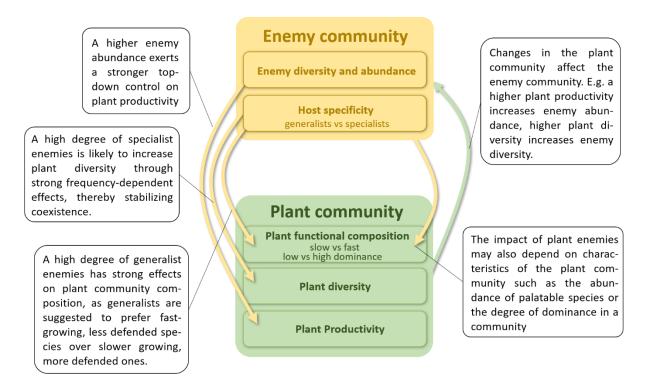
Empirical studies excluding plant enemies have found highly variable effects on plant diversity [24,25,28,29,44,45]. While large vertebrate herbivores were found to slightly promote grassland diversity (evenness) in a recent meta-analysis, effects of invertebrates were variable [12,32]. Evidence from the tropics [38,46] and also from plant-soil feedback experiments in temperate grasslands [36] suggest that soil-borne pathogens are able to maintain high levels of diversity through Janzen-Connell effects. However, pathogens or invertebrate herbivores may
also destabilize coexistence by creating positive feedbacks or by increasing fitness differences
among plant species [13]. Apparent competition through shared enemies could enhance fitness
differences, for instance through spillover of pathogens from a tolerant to a less tolerant host
[13,47].

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154 Effects of plant enemies on plant species composition

155 Plant enemies may also alter plant species composition without changing diversity. 156 Compositional changes are driven by the same basic mechanisms that also affect diversity. 157 Plant enemies can reduce the competitive ability of certain species, leading to competitive 158 release for others [18]. Plant enemies might generally change *plant functional composition* by 159 preferentially attacking faster growing, less defended plant species rather than slower growing, 160 and more defended ones [19,41,42]. Recent work on fungal pathogens in grasslands has shown 161 that fast-growing plant species were more likely to be infested than slow-growing species [42]. 162 Communities consisting of fast-growing species were thus more strongly impacted by 163 pathogens (they lost more biomass) than communities consisting of slower growing species, 164 leading to a shift towards more slow- growing species [42]. Similarly, invertebrates have been shown to affect plant community composition [e.g. 19,29], however the effects depended on 165 166 the group of enemies excluded, or herbivore preference, and were not linked to plant growth 167 strategy [16,45]. Pathogens and invertebrate herbivores might also change composition by 168 favouring later successional species [48]: Studies have shown that early successional plant 169 species experience stronger and more negative feedbacks with root-feeding invertebrates than 170 mid- or late-successional ones [49], and that species replacement during primary or secondary 171 succession can be driven by soil pathogens [50] and insect herbivores [51].

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Fig. 2: Illustration of the general effects of the enemy community on different aspects of theplant community (plant productivity, diversity, and functional composition).

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177 Context dependency in plant enemy impact: predicting indirect effects of global change

178 The outcome of biotic interactions are notoriously context-dependent, with enemy impact 179 depending on many abiotic and biotic factors [16,52,53]. In general, variation in impact can be 180 driven by changes in the enemy or the plant community. Changes in the abundance and per 181 capita feeding rate of enemies will alter their impact on plant productivity [54], while changes in enemy diversity and in their degree of specialisation will alter impact on plant diversity and 182 183 species composition (Fig. 2). At the same time, characteristics of the plant community may 184 determine the impact that enemies can have, e.g. enemies may have a larger impact in 185 communities dominated by a few plant species [55,56] or in communities dominated by 186 palatable, fast growing species [42, Fig.2]. Indirect effects of global change on plant 187 communities through altered biotic interactions can thus arise because global change alters the 188 enemy community alone (Fig. 1, (2)) or because it alters feedbacks between plant and enemy

189 community structure (Fig. 1, (3)). For example, increased temperature might directly alter 190 enemy per capita feeding rates (Fig. 1, pathway (2)). In addition, global change drivers might 191 alter enemy communities by changing the plant community: increases in plant productivity 192 might increase enemy abundance [57], increases in plant diversity are expected to increase 193 enemy diversity [58] and perhaps the proportion of specialists [59], while a shift to fast growing 194 plant species might increase the abundance of generalists (Fig. 1, pathway (3)). Understanding 195 how global change drivers alter the feedback between enemies and plants would fundamentally 196 enhance our ability to improve global change forecasts. Here, we provide a framework for how 197 important global change drivers, namely nitrogen enrichment, climate change, and an increase 198 in atmospheric carbon dioxide might modify the impact of plant enemies on plant productivity, 199 diversity and plant functional composition (Box 1). In addition, we discuss the consequences 200 of plant diversity loss and declines in insect abundance arising from other global changes.

201

202 Effects of nitrogen enrichment

203 Nitrogen enrichment is a key global change driver. It typically causes a loss of plant diversity 204 [60], increases plant biomass production [61] and tissue nitrogen concentrations [62] and leads 205 to a shift towards faster-growing plant species [63]. Nitrogen may alter enemy impact through 206 several of these mechanisms. Theory predicts that low-resource environments select for slow-207 growing, highly defended species whereas high-resource environments select for fast-growing, 208 less defended species [39]. Similarly, the nitrogen disease hypotheses predicts that plants 209 growing in fertile environments are more attractive to enemies due to higher tissue nitrogen 210 concentrations [64]. Consequently, fast-growing, acquisitive species, and species growing in 211 nitrogen rich conditions, should suffer higher levels of herbivory or pathogen infection than 212 species with a conservative strategy [40,65,66]. However, fast-growing species or species with 213 access to more soil resources might be more tolerant of enemy attack, meaning that the actual impact of enemies on them, in terms of plant biomass loss, could be lower despite higher infestation [67]. In addition, the exploitation ecosystem hypothesis [20,21] predicts that in resource rich environments with high primary production, natural enemies of herbivores (predators) should control herbivore populations and herbivores should only have large impacts at lower productivity. In this case nitrogen enrichment would not change enemy impact on plant productivity, if all the additional plant biomass is transferred to predators, however so far the pattern has been only shown for vertebrates and not for invertebrate herbivores [68].

221 The evidence for changes in herbivore and pathogen impact with productivity and fertility is 222 mixed. Herbivore biomass has been shown to increase with plant productivity [69,70], 223 however, herbivore impact (see Box 2) has sometimes been shown to be unaffected or even to 224 decrease with increasing productivity [30,69]. Results from a recent meta-analysis were 225 surprisingly uninformative for understanding patterns in the strength of herbivore impact 226 among systems differing in net primary production (indicator of soil fertility, [12]). For 227 pathogens, enhanced disease and impact with nitrogen enrichment has been shown in 228 agricultural studies, but findings from grasslands are contradictory, with some studies finding 229 support [71] and others not [42]. Whether the *impact* of plant enemies increases with increasing 230 soil fertility is therefore not fully resolved.

231

232 Effects of temperature and precipitation

Increases in temperature and changes in precipitation following climate change will not only affect plant growth directly but are likely to change plant communities indirectly by modifying their pathogen and herbivore communities. Warming is expected to increase the impact of herbivores and pathogens because higher temperatures lead to higher metabolic rates [72,73], accelerated development and reproduction [74,75]. However, this will only occur if precipitation remains high - warming in arid environments would be expected to reduce enemy 239 attack. Theory on how plant-enemy interactions change with temperature largely come from 240 latitudinal and altitudinal gradients. Biotic interactions have long been predicted to be strongest 241 in the warm, low latitudes [15,59]. The mechanisms driving this latitudinal gradient are likely 242 to emerge over long time scales and involve increased coevolution and speciation in more 243 predictable and benign climates [59,76]. Similarly, herbivore pressure is expected to be higher 244 at lower elevations, which are warmer and less variable climatically [77,78]. This should result 245 in decreased plant defense at high elevations and a higher proportion of generalist consumers. 246 Plant-fungal interactions are also likely to vary with elevation, as fungal communities are 247 sensitive to changes climatic conditions [79,80]. Altitudinal gradients occur over much smaller 248 scales and are less confounded by historical or macro-evolutionary processes [81]. They may 249 therefore be more promising than latitudinal gradients for predicting changes in enemy impact 250 with climate change [5,80].

251 Results from existing elevational gradients and warming experiments indicate that herbivore 252 pressure is indeed higher at lower elevations [77,78], however, recent studies questioned the 253 generality of previous findings [78,81]. Similarly, results from a global field survey suggests 254 that the relative abundance of soil-borne fungal pathogens increase with warming [82]. Most 255 warming experiments have shown an increase in pathogen and herbivore attack at higher 256 temperatures in the field [75,83–85] or in the lab [86,87]. However, other warming experiments 257 found no change [88] or even a decrease [89] in herbivory. Results from agriculture indicate 258 that temperature effects on pathogen impact are species-specific. While several cold-sensitive 259 pathogens would benefit from mild winters [90], others are impeded in their infection potential 260 by higher temperatures [91]. This suggests that species-specific changes in plant quality (leaf 261 nutrients, defensive compounds), herbivore and pathogen thermal limits [89] or changes in 262 herbivore natural enemy activity or abundance [92,93] also determine enemy attack and hence 263 their impact on plants. Rarely, studies assess *impact* of enemies on plant communities or 264 functioning along altitudinal gradients, i.e. by using exclusions [80], or combine warming 265 experiments with the exclusion of pathogens and herbivores (but see [94]. More studies are needed to reliable predict the impact of plant enemies on plant communities in a warmer world. 266 267 The effects of changes in precipitation or extreme whether events on enemy impact are more 268 difficult to predict, as water on the one hand affects habitat characteristics for enemies (soil 269 and air humidity), but also acts as a resource, affecting plants and enemies via changes in 270 productivity. Effects on habitat characteristics are likely to be species or group-specific. While 271 many pathogens, e.g. rusts, may depend on humidity during infection and may decrease during 272 drought, others may benefit [95,96]. Changes in soil moisture may affect habitat conditions 273 differently for herbivores or predators, with consequences for predator-prey interactions 274 [97,98]. Extreme weather events such as storms may facilitate natural dispersal for pathogens, 275 whereas other enemy groups may decrease [99,100]. When water acts as a resource, changes 276 in precipitation affect enemies indirectly via changes in plant quality and community 277 composition. E.g. periodic drought stress in plants may impede the ability of plants to 278 compensate for enemy attack, making plants more susceptible to disease. Continuing drought 279 may shift plant community composition towards more stress tolerant, well-defended species, 280 at the expense of fast-growing more competitive ones. A reduction in precipitation may further 281 increase mutualistic interactions, which can play an important role in plant resistance [101]. 282 This may decrease plant quality and the impact of enemies on plant communities, similar to 283 the opposite effects of increases in temperature and nitrogen. Because changes in precipitation 284 may result in complex responses of plants and enemies, drawing predictions on how such 285 changes affect the impact of enemies on plant communities is challenging.

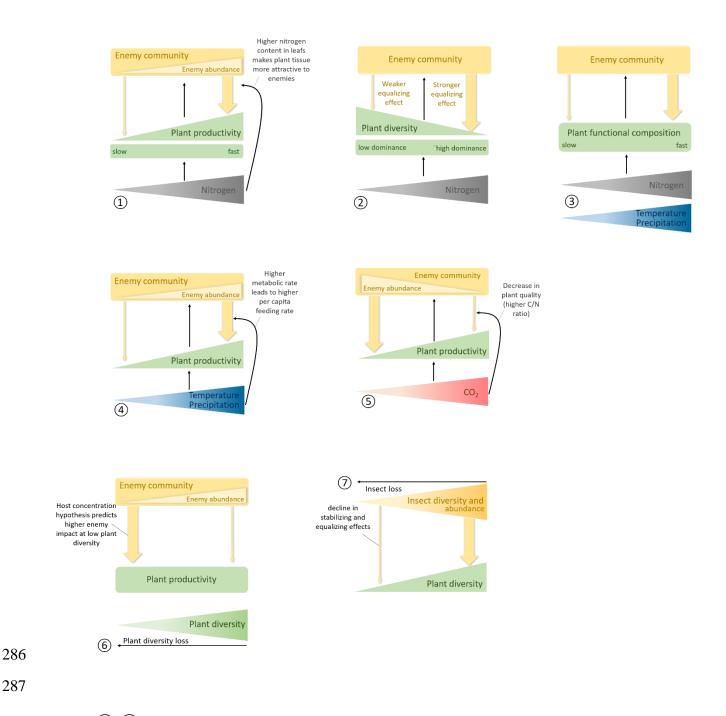


Fig. 3. (1)-(5) Illustrations indicate how the global change drivers nitrogen, temperature and elevated CO2 may affect the impact of plant enemies (*yellow arrows*) on plant productivity. Moreover, we depict how nitrogen enrichment may affect the impact of plant enemies on plant diversity and functional composition. (6) depicts the consequences of plant diversity loss on the impact of enemies on plant productivity, and (7) the effect of insect decline on plant diversity. Explanations see Box 1.

Box 1: Expectations of how global change drivers affect enemy impact

296 (1) Nitrogen enrichment affects impact of enemies on plant productivity. The nitrogen 297 disease and resource availability hypotheses would both suggest that the impact of herbivores 298 and pathogens on plant productivity should be higher following N enrichment, due to changes 299 in plant nutrient contents or functional composition. Moreover, a high soil fertility may reduce 300 the presence of beneficial soil microbes, which have been shown to provide plants with 301 resistance against enemies [128–130], further increasing enemy impact. While both plant and 302 herbivore biomass have been shown to increase with soil fertility, herbivore impact has not 303 [69]. This might suggest that fast-growing plant species are more tolerant and compensate for 304 leaf loss, or that top-down control of herbivores by predators is stronger at high soil fertility. 305 Predicting how the impact of enemies on plant productivity will change with nitrogen 306 enrichment will therefore require more information on variation in plant tolerance and predator 307 communities with nitrogen addition.

308 (2) Nitrogen enrichment increases impact of enemies on plant diversity. Nitrogen 309 enrichment usually reduces plant diversity and evenness, through an increase in light 310 competition, which is expected to alter enemy impact. We expect that the impact of enemies 311 on plant diversity is stronger at high soil fertility because there is more opportunity for plant 312 enemies to equalise fitness differences and promote diversity by reducing the dominance of the good light competitors [25]. Enemies might also be particularly likely to reduce the dominants 313 314 in resource rich environments because they are less defended (growth-defense trade-off), or 315 because they are tall and are therefore more likely to be removed by vertebrate grazers. Changes 316 in the feedbacks between enemies and plants may therefore offset some of the negative effects 317 of nitrogen enrichment on plant diversity.

318 (3) Nitrogen enrichment increases the impact of enemies on plant functional composition.
319 Nitrogen enrichment tends to shift plant functional composition towards dominance by fast

320 growing species. We expect that this will increase enemy impact on plant functional 321 composition because fast-growing plants are preferentially attacked by generalist enemies, 322 allowing slower growing but more defended species to persist. Excluding enemies from fertile 323 communities would result in a loss of slower growing species, shifting functional composition towards even faster growing species. However, excluding enemies from resource poor 324 325 environments would have minor effects on plant functional composition, as fast-growing 326 species will still be outcompeted by the slow growing ones and are therefore unlikely to be able 327 to increase.

328 (4) Temperature affects impact of enemies on plant productivity. A higher plant 329 productivity as well as an accelerated development and reproduction of plant enemies with 330 warmer temperatures and higher precipitations are both likely to increase enemy abundance 331 and consumption, and therefore enemy impact on *plant productivity*. However, this might be offset by changes in plant nutrients and defences, or - for herbivores - by altered top-down 332 333 control of predators. For example, it has been shown that spiders produce a second clutch in 334 arctic environments with climate warming [93]. Thus, whether the impact of invertebrate 335 herbivores and fungal pathogens on productivity increases with climate warming is not fully 336 resolved.

337 Predictions for how enemy impact on *plant diversity* and *composition* will change with
338 alterations in the climate are even less clear, as only few studies have interactively manipulated
339 temperature and enemy abundance [94].

5 Elevated CO₂ reduces impact of enemies on plant productivity. An increase in atmospheric carbon dioxide is expected to decrease plant quality via an increase in the C:N ratio and carbon-based secondary compounds. We therefore expect that, at least for herbivores, enemy abundance and hence impact will decrease. For pathogens, we need more studies to draw general conclusions on how increases in CO₂ affect pathogen impact on plants, particularly in natural systems [108]. We also expect that the effect of CO₂ on plant functional
composition will remain stable as herbivores will prevent an increase in N-fixing species with
elevated CO₂.

(6) Plant diversity loss affects impact of enemies on plant productivity. We would expect
that the impact of natural enemies on plant productivity increases with plant diversity loss (host
concentration hypothesis, diversity-disease hypothesis). However, the effect of plant diversity
loss on enemies will also depend on the species that are lost first from the community [131]: if
susceptible or palatable species are lost last, for instance if fast growing species remain
following nutrient enrichment, then diversity loss will enhance disease transmission. However,
the opposite will occur if resistant species are lost last.

355 (7) Insect decline alters effects on plant diversity. A loss of insect biomass, abundance and
 356 diversity is likely to weaken the stabilizing and equalizing effects of herbivores on plant
 357 coexistence. This might have far reaching consequences for plant diversity.

358

359 Effects of elevated atmospheric carbon dioxide

360 Another key element of anthropogenic climate change is an increase in atmospheric CO₂ levels. 361 The drastic rise in CO₂ is likely to not only affect plant growth but also plant interactions with 362 their enemies either by changing plant nutritional quality or functional composition and hence 363 the feedbacks between plant and enemy communities [102,103]. Experiments have shown that elevated CO₂ generally enhances plant growth and increases tissue C:N ratios [104,105]. 364 365 Moreover, it is likely that an increased C supply to plants leads to an increase in C-based secondary and structural compounds, and a dilution of nutrients [103], decreasing plant quality 366 367 [104,106]. As invertebrate herbivores and pathogens are N-limited, an increase in C:N ratios 368 is likely to lead to lower food quality.

Results from meta-analyses suggest that while chewing herbivores respond with increased food consumption in the short term to compensate for the lower nutritional quality, they also grow more slowly and have a longer development time, which increases exposure to enemies, all of which reduce herbivore abundance under elevated CO_2 [104,106]. Phloem feeding insects instead tend to increase with elevated CO_2 [104]. For pathogens, studies suggest that elevated CO_2 can have positive, negative or neutral effects on infection intensity, as different pathogens might respond differently [71,74,107,108].

376 In addition to changes in plant quality, changes in plant functional composition with elevated 377 CO_2 might alter enemy impact. While C_4 -plants are relatively unresponsive to elevated CO_2 , C₃-plants often increase their growth [109]. Even more positively affected are legumes thanks 378 379 to their close mutualistic interaction with nitrogen fixing bacteria in root nodules. Legumes 380 shunt excess carbon to their mutualists which stimulates their growth and subsequently the 381 growth of legumes [110]. However, the positive effect on legumes is likely to be offset by a 382 shift in preference of generalist plant enemies to the more nutritive N-fixing species, preventing 383 them from increasing.

384

385 Declines in plant and insect diversity – consequences for plant-enemy interactions

Other global change factors such as invasive species or pesticides may cause declines in plant 386 387 and insect diversity [1,111]. A decline in plant diversity is likely to alter the impact that 388 invertebrate herbivores and pathogens have on plant communities. The idea that plant 389 communities with higher densities of a few species should be more vulnerable to negative 390 effects of specialized enemiew, and that communities with high plant diversity harbour a larger 391 community of natural enemies, has a long history in ecology ([55], host concentration 392 hypothesis, [112]). Indeed, studies manipulating plant diversity of grasslands have shown that 393 pathogen attack is generally higher in species poor compared to species rich communities 394 [56,113–117]. For herbivores, previous results have been ambiguous, and abundance, diversity 395 and impact of herbivores have also been shown to increase be highest at high plant diversity 396 [32,33,118–121]. However, results for specialist insect pests are more consistent and resource 397 concentration effects are commonly observed [122]. Most of the evidence for resource 398 concentration effects comes from experiments in which diversity is manipulated and species 399 compositions are random [117]. Responses of herbivores and pathogens to real world gradients 400 of diversity are less well known [123] and will depend on how plant functional composition 401 changes at the same time.

402 Recent studies have also suggested that insect populations are declining in abundance 403 [111,124] and this loss of insect abundance and diversity is likely to have consequences for 404 plant communities. A loss of herbivore abundance and diversity is likely to result in a decline 405 in their impacts on plant communities. This might have far reaching consequences for plant 406 diversity by disrupting coexistence mechanisms and allowing competitive species to dominate.

407

408 **Concluding remarks and future avenues**

409 Our review highlights the multiple indirect effects that global change can have through altering 410 enemy impact. Global change is expected to profoundly alter feedbacks between plants and 411 enemies and to have strong indirect effects on biodiversity and ecosystem functioning. This 412 review also points out the gaps in our understanding of the context-dependency of plant-enemy 413 interactions [16]. The reason why we know so little about how enemy impact varies along 414 abiotic or biotic gradients is that large-scale patterns are usually pieced together from data 415 obtained by a variety of methods and protocols [125]. Moreover, while effects on enemy 416 communities are rather well studied, we often know little of whether such changes translate 417 into changes in *impact* of plant enemies, as studies often rely on comparing herbivore or 418 pathogen *attack* along gradients and rarely measure *impact* experimentally (see Box 2), i.e. by 419 using exclusions [80,126,127]. To predict the consequences of global change on ecosystems 420 we need to develop a predictive understanding of why and how antagonistic biotic interactions 421 vary in space. Exclusion of different plant enemies along environmental gradients is one way 422 to find general predictors of context-dependency. Moreover, experiments simultaneously manipulating various global change drivers while excluding enemy groups can inform us about 423 424 potential changes in plant-enemy impacts on plant communities under global change. 425 Enhancing our knowledge on how antagonistic biotic interactions alter the diversity and 426 functioning of plant communities, and how their impact depend on abiotic and biotic factors is 427 a pressing priority given accelerating global change.

428

429 Box 2: Assessing plant-enemy impact: impact does not equal damage!

430 The interactions of plants with their enemies are notoriously context dependent and the impact 431 of plant enemies therefore differs strongly in space. For example, ecological theory suggests 432 that enemy impact on plant communities is strongest at low latitudes, low elevation or in very 433 fertile habitats [15,39,77]. Most studies that test for such patterns collect data on herbivore or 434 pathogen attack [78,132]. However, damage by herbivores or pathogens is not necessarily 435 related to their impact, i.e. their effect on plant biomass, demography or other community 436 components. Some plant species might be more tolerant to enemy attack than others 437 (particularly species with an acquisitive, fast-growing strategy) and the actual *impact* of 438 enemies on them might be lower than one would expect based on enemy infestation. Similarly, 439 some plant species might be highly susceptible to already low levels of enemy damage and 440 decline in abundance, and these strong effects of enemies on plant community structure cannot 441 be assessed with damage assessments alone [133].

To best predict the impact of plant enemies on plant productivity, diversity or community composition a comparison of plant communities in the presence and absence (low densities) of enemies would be necessary, e.g. with addition/infection or exclusion experiments. Addition or infection experiments are only meaningful for determining the role of enemies in natural communities if the added enemy community composition and density mirrors the naturally occurring one, which is challenging and has so far not been done in terrestrial systems. Exclusion experiments have the advantage that they reduce natural levels of enemies, allowing 449 to draw realistic inferences about the role of enemies in plant communities. However, reducing 450 invertebrate herbivores and pathogens from plant communities often requires the use of 451 chemical biocide application. While this approach has sometimes been criticized as not being 452 effective and having non-target effects [133], as long as ecologists are careful about the 453 inferences and mindful of the limitation of this approach the chemical exclusion of plant 454 enemies offers a valuable opportunity to study the importance of enemy-plant interactions in 455 terrestrial ecosystems. Combined with investigations along environmental gradients [32] or 456 global change experiments (e.g. warming, [94]), the exclusion of enemies might reveal 457 important insights into how enemy impact in plant communities may change in the future. 458

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