

# Understanding the value and limits of nature-based solutions to climate change and other global challenges

Nathalie Seddon<sup>1</sup>, Alexandre Chausson<sup>1</sup>, Pam Berry<sup>2</sup>, Cécile AJ Girardin<sup>2</sup>, Alison Smith<sup>2</sup>, and Beth Turner<sup>1</sup>

<sup>1</sup>*Nature-based Solutions Initiative, Department of Zoology, University of Oxford*

<sup>2</sup>*Environmental Change Institute, School of Geography and Environment, University of Oxford*

**Abstract** There is growing awareness that “Nature-based Solutions” (NbS) can help to protect us from climate change impacts whilst slowing further warming, supporting biodiversity and securing ecosystem services. However, the potential of NbS to provide the intended benefits has not been rigorously assessed. There are concerns over their reliability and cost-effectiveness compared to engineered alternatives, and their resilience to climate change. Trade-offs can arise if climate mitigation policy encourages NbS with low biodiversity value, such as afforestation with non-native monocultures. This can result in maladaptation, especially in a rapidly changing world where biodiversity-based resilience and multifunctional landscapes are key.

Here we highlight the rise of NbS in climate policy—focussing on their potential for climate change adaptation as well as mitigation—and discuss barriers to their evidence-based implementation. We outline the major financial and governance challenges to implementing NbS at scale, highlighting avenues for further research. As climate policy turns increasingly towards greenhouse gas removal approaches such as afforestation, we stress the urgent need for natural and social scientists to engage with policy makers. They must ensure that NbS can achieve their potential to tackle both the climate and biodiversity crisis while also contributing to sustainable development. This will require systemic change in the way we conduct research and run our institutions.

**Keywords:** Adaptation, mitigation, climate change, ecosystems, nature, restoration

## 35 1. The rise of Nature-based Solutions

36 How do we meet three central challenges of the Anthropocene: mitigating and  
37 adapting to climate change, protecting biodiversity, and ensuring human wellbeing?  
38 A major part of the answer lies in addressing these interdependent challenges  
39 together; to do otherwise invites negative consequences and unintended feedbacks.  
40 Indeed, the ethos of the UN Sustainable Development Agenda is one of connectivity,  
41 inclusivity and partnership; it acknowledges interdependencies of the 17 social,  
42 environmental and economic goals and encourages actions that promote synergies  
43 among them [1]. Yet despite the importance of taking account of synergies and trade-  
44 offs between these goals [2-4], there is little evidence that this is happening in  
45 practice. As a direct result many goals are unlikely to be met by 2030. In particular,  
46 the failure to stabilise and adapt to climate change (SDG 13) [5] or protect  
47 biodiversity (SDGs 14 and 15) [6,7] has been exacerbated by the fact that these  
48 issues are being treated separately when in fact they are deeply interwoven and  
49 share many of the same drivers.

50 It is against this backdrop that nature-based solutions (NbS)—solutions to  
51 societal challenges that involve working with nature (Box 1)—are emerging as an  
52 integrated approach that can reduce trade-offs and promote synergies among the  
53 SDGs at relatively low cost [8-10]. In contrast to many engineered solutions, NbS  
54 have the potential to tackle both climate mitigation and adaptation challenges at  
55 relatively low-cost whilst delivering multiple additional benefits for people and nature.  
56 For example, restoring forests in upper catchments can help to protect communities  
57 downstream from flooding, at the same time as increasing carbon sequestration and  
58 protecting biodiversity. Planting trees and increasing green space in cities can help  
59 with urban cooling and flood abatement, while storing carbon, mitigating against air  
60 pollution, and providing recreation and health benefits (see Box 2 for examples).  
61 Consequently, NbS were endorsed in the IPBES Global Assessment [6], the Climate  
62 Change and Land Report of the Intergovernmental Panel on Climate Change (IPCC)  
63 [11] and the Global Adaptation Commission Report [148], and were highlighted as  
64 one of nine key action tracks at the 2019 UN Climate Action Summit [12]. Meanwhile,  
65 the World Economic Forum's (WEF) Global Risks Report 2019 specifically  
66 recognised the economic risks posed by biodiversity loss and ecosystem collapse  
67 [13] and the need for nature-positive business solutions. NbS are increasingly being  
68 viewed not only as a way to reconcile economic development with the stewardship of

69 ecosystems, but also as a means to diversify and transform business and enable  
70 sustainable development [14].

71         Here we take a critical look at the potential for NbS to deliver both climate  
72 change mitigation and adaptation whilst also supporting other ecosystem services.  
73 As the role of NbS for climate change mitigation rises up the policy agenda, we  
74 stress their vital role for climate change adaptation, and the importance of using  
75 evidence-based design to maximise synergies and minimise trade-offs. We focus on  
76 three key barriers: measuring the effectiveness of NbS; mobilising investment; and  
77 overcoming governance challenges. Finally, we identify the need for systemic  
78 institutional change to overcome these barriers, including a more holistic design and  
79 evaluation approach that fully incorporates the multiple benefits of NbS.

### Box 1 | Defining Nature-based Solutions

NbS involve working with and enhancing nature to help address societal challenges [15,16]. They encompass a wide range of actions, such as the protection and management of natural and semi-natural ecosystems, the incorporation of green and blue infrastructure in urban areas, and the application of ecosystem-based principles to agricultural systems. The concept is grounded in the knowledge that healthy natural and managed ecosystems produce a diverse range of services on which human wellbeing depends, from storing carbon, controlling floods and stabilizing shorelines and slopes to providing clean air and water, food, fuel, medicines and genetic resources [17]. NbS is an ‘umbrella concept’ for other established “nature-based” approaches such as ecosystem-based adaptation (EbA) and mitigation (EbM), eco-disaster risk reduction (eco-DRR), and Green Infrastructure (GI) [18]. More recently the term “natural climate solutions (NCS)” entered the lexicon [19]. NCS also falls under the umbrella of NbS, but refers explicitly to conservation and management actions that reduce greenhouse gas emissions from ecosystems and harness their potential to store carbon [19-21].

NbS vary in three important ways, which influence the range of benefits that they provide for people. (1) They **cover a spectrum of interventions** from protecting or restoring diverse natural ecosystems to creating new managed or hybrid “grey-green” approaches [22]. While healthy natural forests, grasslands and wetlands may store more carbon than their managed equivalents (e.g. due to greater soil depth, age and structural diversity [23]), managed and hybrid systems such as city parks or green roofs contribute to urban cooling, storm-water management, and bring mental and physical health benefits [24].

(2) NbS **vary in the extent to which they support biodiversity**, which in turn affects their resilience, i.e. their capacity to resist and recover from perturbation and maintain the flow of ecosystem services. NbS that protect and restore natural ecosystems and/or make use of diverse native species can play a key role in securing climate change mitigation and adaptation services, whilst also contributing to cultural ecosystem services such as inspiration and education. In contrast, NbS that do not harness ecological principles and support biodiversity (such as those involving non-native monocultures) are more vulnerable to environmental change in the long term and may also produce trade-offs among ecosystem services (e.g. carbon storage erosion control and water supply, as demonstrated in the Loess Plateau, Box 2).

(3) NbS **differ in how much they are designed and implemented by local communities** [25]. EbA places particular emphasis on this; it is a participatory community-based climate adaptation strategy which “may include sustainable management, conservation and restoration of ecosystems, as part of an overall adaptation strategy that takes into account the multiple social, economic and cultural benefits for local communities” [26].

By specifically aiming to address broad societal goals such as human wellbeing, including poverty alleviation and socioeconomic development, NbS differ from traditional biodiversity conservation and management approaches. However, to be resilient (and hence sustainable) NbS must be implemented in such a way as to support biodiversity and people [8, 27, 28].

## 2. Nature-based Solutions for climate change mitigation

Over the past ten years, there has been growing interest in the potential of NbS to help meet global goals for greenhouse gas (GHG) emissions reductions to mitigate climate change, reflecting the importance of natural ecosystems as sources and sinks for GHGs. The IPCC Climate Change and Land Report states that all scenarios that limit climate change to 1.5°C rely heavily on landuse change mitigation methods, as well as decarbonising the economy [11]. Agriculture, forestry and other landuse activities accounted for around 23% of total net anthropogenic emissions of GHGs during 2007-2016 ( $12.0 \pm 3.0$  Gt CO<sub>2</sub>e yr<sup>-1</sup>, includes CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O, [11]). Of this, net emissions of  $5.2 \pm 2.6$  Gt CO<sub>2</sub> yr<sup>-1</sup> were mostly due to deforestation, partly offset by afforestation/reforestation, and emissions and removals by other landuse activities [11].

Decreasing sources and increasing sinks of GHGs through terrestrial ecosystem stewardship and improvements in agriculture are widely cited as having the potential to provide around 30% of the CO<sub>2</sub> mitigation needed through to 2030 to keep warming to < 2 °C [19, 29, 30]. However, a more recent analysis focussing on tropical nations and involving tighter model constraints (e.g. on where ecosystem regeneration can take place) indicates that this figure is an overestimate, and emphasises the need to explore this potential on a national level [149]. Low GHG emissions and high forest cover in many tropical nations mean that NCS can mitigate over 50% of national emissions, mainly through avoided deforestation. Further, Griscom et al. [149]) highlight a particular set of countries with strong governance and intermediate financing capacity, where the focus on nature-based climate solutions would have the most potential for contributing to emissions mitigation (e.g. Indonesia, Brazil).

Some NbS may eventually reach a saturation point when the ecosystem is at equilibrium and sequestration is balanced by emissions. However, NbS have key advantages over other carbon dioxide removal (CDR) options. For example, Direct Air Capture is expensive, energy intensive and not yet deployable at scale; bioenergy with carbon capture and storage (BECCS) requires large areas of land for biofuel production; enhanced weathering entails quarrying, pulverizing and transporting rock on a large scale [31]; and all engineered approaches to CDR do not bring the suite of additional ecosystem services offered by well-implemented and managed NbS [150]. It is clear that engineered approaches to CDR are not an alternative to NbS but must work in tandem. Moreover they must only be deployed once we better understand

116 how to reduce trade-offs with biodiversity and ecosystem services [150].

117       The IPCC Climate Change and Land Report emphasizes that the mitigation  
118 potential from terrestrial ecosystems comes from restoration and management of  
119 forests and from curbing deforestation [11], especially in tropical and subtropical  
120 regions [32], where forests grow fast and there are no adverse effects from increased  
121 albedo (unlike boreal regions) [19,33-35]. The report [11] states a mitigation potential  
122 range of 0.4 to 5.8 GtCO<sub>2</sub> yr<sup>-1</sup> from avoided deforestation and land degradation, as  
123 well as a carbon sequestration potential of 0.5 to 10.1 GtCO<sub>2</sub> yr<sup>-1</sup> in vegetation and  
124 soils from afforestation/reforestation.

125       However, reliance on forests for GHG mitigation raises several practical and  
126 ethical concerns. First, if policy is not grounded in sound ecosystem and biodiversity  
127 science, parties risk investing in monocultures or low diversity plantations. For  
128 example, 45% of the 350 Mha currently pledged for reforestation is set to become  
129 commercial plantations, usually involving single species (i.e. monocultures) [28]. This  
130 is problematic for a number of reasons. Fast growing monocultures sequester  
131 carbon rapidly but they may not maximize carbon storage in the long-term as they  
132 are vulnerable to disease, pests and climate extremes [e.g. 39-41]. Moreover, when  
133 plantations are harvested, typically every 10-20 years in the tropics, much of the  
134 stored carbon is returned to the atmosphere [28]. In contrast, forests that regenerate  
135 naturally have high carbon sequestration rates [36], and older and more diverse  
136 forests store more carbon and are more resilient to climate extremes and disease  
137 [37,38]. When rotation times and GHG emissions from fertiliser application are taken  
138 into account, Lewis et al. [28] calculate that natural forest regeneration could store 40  
139 times more carbon than commercial plantations, and seven times more than  
140 agroforestry. They conclude that targets for climate stabilization cannot be achieved  
141 under current reforestation plans that comprise mainly plantations, even with the use  
142 of bioenergy with carbon capture and storage (BECCS). Another issue, is that  
143 plantations often involve fast growing non-native species which may become  
144 invasive, introduce new pests and diseases [42], or exacerbate water scarcity in arid  
145 or semi-arid regions [38, Box 2]. In environments where forests do not naturally  
146 thrive, such as savannahs prone to drought and fire risk, afforestation may reduce  
147 resilience to climate change and could compromise long-term carbon storage [43].  
148 More diverse ecosystems also tend to deliver a wider range of other regulating and  
149 cultural ecosystem services [38], increasing the cost-effectiveness of NbS.

150       Second, policies that offer financial incentives to scale up NbS for the

purpose of GHG mitigation risk compromising local land rights, leading to land grabs by governments and private investors. Whereas the Warsaw framework for REDD+ specifies conservation of biodiversity and the respect of indigenous people and local communities rights in the Cancun Safeguards [44], the guidance relating to other NbS action is too vague on both accounts.

Third, encroachment of tree plantations onto other ecosystems can have devastating impacts on biodiversity. It is particularly concerning, for example, that 9 Mha of ancient grassland is wrongly classified as degraded land suitable for afforestation [43]. It is clear that NbS needs to be grounded in robust understanding of the geographical distribution of the biomes of the world, the value of their biodiversity, and their ecological resilience.

Finally, and perhaps most critically, it is essential that enthusiasm for nature-based climate change mitigation does not curtail or distract from the urgent need to rapidly decarbonize our economy, including through radical systemic change [45].

Despite these caveats, well-designed NbS that incorporate diverse native species, avoid damaging biodiverse ecosystems and respect social safeguards, offer good opportunities for mitigation with key benefits for local people [45]. These options should include restoration of natural forests and wetlands (e.g. peatlands and mangroves), especially in tropical biodiversity hotspots [32], as well as agroforestry, and increasing carbon in agricultural soils [11]. It is urgent that we strengthen policy frameworks to ensure that NbS can provide multiple benefits for both climate mitigation and adaptation, and other vital ecosystem services secured by biodiversity [6].

### 3. Nature-based Solutions for climate change adaptation

The World Economic Forum Global Risks Report lists extreme weather events and natural disasters as the top two greatest risks to the global economy and human wellbeing, both in terms of severity of impact and likelihood of occurrence [13]. It also ranks the failure to mitigate and adapt to climate change—which exacerbates both extreme weather and natural disasters [46]—as one of the most impactful risks.

To date, the dominant approach to addressing the risks posed by extreme weather, natural disasters and climate change has involved, engineered interventions such as sea walls, levees or irrigation infrastructure [47]. For example, in Bangladesh—a country subject to some of the worse extremes of climate change



impacts and natural disasters—291 of 329 (88%) adaptation projects approved by the Bangladesh Climate Change Trust between 2009 and 2016 involved engineered (i.e. grey) interventions; only 38 involved nature-based (i.e. “green”) solutions [48]. This bias in investment towards engineered approaches is global (reasons why are discussed below). Yet there is growing evidence that NbS can in certain contexts provide a powerful complement (or alternative) to grey infrastructure [148].

A conceptual model for understanding nature’s role in supporting human adaptation to climate change is the vulnerability framework for social-ecological systems, formalised by the Intergovernmental Panel on Climate Change (IPCC) [49; Figure 1]. The framework explicitly integrates the vulnerability of ecosystems with the vulnerability of socioeconomic systems. It recognizes that, in each system, vulnerability to climate change has three dimensions. The first is *exposure*; that is, the extent to which a region, ecosystem, resource or community is impacted by climate change (dimension 1). The second is *sensitivity* to these impacts; that is, the degree to which a system is affected by, or responsive to, those effects (dimension 2). The third is the *adaptive capacity* of the system; that is, the ability to adjust or innovate in response to changing conditions (dimension 3). NbS act at the interface of the socioeconomic system and the ecosystem to reduce the vulnerability of the social-ecological system as a whole. In other words, through the protection, restoration and careful management of ecosystems (section 4), NbS can positively influence all three dimensions of socioeconomic vulnerability.

### ***NbS for reducing socioeconomic exposure (dimension 1)***

Most evidence for nature’s role in supporting human adaptation pertain to the first dimension of the vulnerability framework, i.e. reducing exposure to the immediate impacts of climate change (see Box 2 for examples). In particular, there is growing evidence that: (i) protecting, restoring or managing natural forests and wetlands in catchment areas (for example, in headwaters and along rivers) in many cases can secure and regulate water supplies [51], reduce flood risk [52] and/or reduce exposure to soil erosion and landslides [53]; (ii) restoring coastal ecosystems (i.e. mangroves, coral reefs, oyster beds, and salt marshes) protects communities from coastal flooding [54], reduces damages caused by storm surges [55], and limits coastal erosion [56,57,151]; (iii) nature-based agricultural practices such as agroforestry (planting trees among crops or crops among trees) can maintain and in some cases enhance yields in drier, more variable climates [58,59]; and (iv) creating



green roofs and walls, and/or planting trees and increasing green space in and around urban areas can moderate the impacts of heat waves [60-62] and regulate water flow [63; reviewed in 64].

### ***NbS for reducing socioeconomic sensitivity (dimension 2)***

Properly implemented and supported by biodiversity, NbS can reduce the sensitivity of individuals, communities and societies to climate change. They can secure or enhance the delivery of ecosystem services that sustain livelihoods and wellbeing, and provide diverse sources of income to help communities adapt to climatic or other environmental shocks (Box 2). For example, the rehabilitation of degraded semi-arid rangelands in Kenya cushions agro-pastoral communities against climatic shocks such as drought [65,66]. Communities using enclosures also reported having healthier, more productive livestock, more diverse sources of income (e.g. wood and grass cuttings, grass seeds, poultry products, fruits and honey) and an improved standard of living. Similarly, protecting forests in Zimbabwe ensures honey production during droughts, thereby providing a degree of food security when other crops fail [67]. Agroforestry also provides alternative income sources (fuelwood, fruit, timber) as well as reducing exposure to heat, drought, floods and erosion [68].

### ***NbS for supporting socioeconomic adaptive capacity (dimension 3)***

NbS can contribute to adaptive capacity in two main ways. First, NbS that are designed to support the genetic or species diversity will help to maintain a reservoir of wild species that can help us adapt to change, e.g. for breeding food and timber crop varieties that are resilient to climate change, pests and diseases, and as a source of knowledge for technical innovations based on biomimicry. Second, NbS can be implemented in a way that brings communities together to learn and experiment, for example through the process of ecosystem-based adaptation focusing on sustaining the supply of ecosystem services, including those that reduce exposure and sensitivity of vulnerable groups. For example, the implementation of community-based natural resource management in pastoral communities in Ethiopia is reported to have empowered local communities to develop systems for managing natural resources in the face of change, improved institutional governance and thereby potentially increased capacity to deal with future climate change [69]. Similar benefits of ecosystem-based approaches to adaptation have been reported across the globe, including in Togo [70], and Sri Lanka [71] (Box 2). However, NbS will only

deliver these benefits if they are specifically designed to do so. Many other factors influence adaptive capacity, including financial and human resources, as well as education and governance [72] and these factors also influence opportunities to implement NbS. The extent to which NbS contribute to adaptive capacity, however, is poorly understood and further monitoring and evaluation is needed [71].

#### 4. Effectiveness of NbS under climate change

The ability of ecosystems to act as a sink for CO<sub>2</sub> emissions (section 2) and reduce socioeconomic vulnerability to climate change (section 3) is directly and indirectly affected by the exposure, sensitivity and adaptive capacity of the ecosystems themselves (as illustrated in Figure 1). Sensitivity and adaptive capacity vary among ecosystems and can be strongly influenced by management approaches [73,74].

Natural ecosystems are usually well adapted to their natural disturbance regimes such as episodes of drought, flooding, storms or wildfires. Some ecosystems, such as grasslands, are able to recover normal ecosystem function after major droughts and fires [75]; others are more sensitive, as evidenced by dieback in forests across the globe [76]. Problems are arising because the increasing frequency and intensity of these disturbances under climate change, combined with other stressors such as land-use change and pollution, is causing disturbances to recur before the system has a chance to recover. This can result in a dramatic decline in the adaptive capacity of the ecosystem, leading to a transition to a new community of species or an entirely new ecosystem. For example, the increasing frequency and severity of fires in Yellowstone National Park is depleting the seed bank for forest regeneration [e.g. 77]. And while mangrove forests can keep pace with moderately high rates of sea-level rise (SLR) [78], saltmarshes cannot (and may be lost globally to SLR by the end of century without major intervention [79]). Exposure to such impacts can be reduced through active management such as tree thinning (shown to reduce fire frequency in *Eucalyptus* plantations [80]) or by maintaining or creating connectivity between ecosystems (which enables species to track preferred ecological niches across the landscape [27]).

Ecosystem sensitivity can be minimized by reducing the pressures affecting ecosystem function (pollution, invasive species, habitat loss and fragmentation, over-exploitation) and enhancing genetic, species and functional richness, which buffer the impacts of extreme weather [39, 81] and pests [82]. Greater diversity also safeguards the evolutionary potential of ecosystems, allowing for ecological

adaptation (often in the form of phenological changes), and reduces the likelihood for trade-offs among different ecosystem services. Diversity can be enhanced through active management (for example in multi-species crop or timber plantations), or through allowing degraded areas to regenerate naturally. Evidence is emerging that the latter can result in ecosystems with higher biodiversity that support a range of climate change adaptation services, with fewer trade-offs [27]. Areas of the Loess Plateau in China, for example, that were allowed to regenerate naturally into herbaceous cover and shrub land provide comparable levels of erosion control to those with afforestation, without compromising water supply or biodiversity (51, 83; Box 2).

With or without active management, many ecosystems have transitioned or are in the process of transitioning to alternative states under climate change [84]. Clearly, some of these new states cannot support human adaptation (e.g. algae-dominated reefs after mass coral mortality [85]). However sometimes, new communities will provide similar adaptation benefits to the pre-disturbance communities and/or provide additional novel adaptation services [86,87]. Further work is now urgently needed to model how the performance of NbS varies under climate change, drawing on knowledge of the eco-evolutionary mechanisms that underpin the ecosystem's capacity to resist and recover or adapt to major perturbations. Many physical models have been developed to forecast the effectiveness of hard infrastructure under different climate change scenarios; the equivalent ecological models now need to be developed for NbS.

## 5. Moving beyond pitching green solutions against grey

Over the last ten years, United Nations institutions (UN Environment, UN Development Programme and FAO) as well as international conservation organisations (e.g. IUCN, WWF, BirdLife International and Conservation International) have been implementing community-led nature-based approaches to adaptation (i.e. EbA) and/or ecosystem-based disaster risk reduction projects across the globe [e.g. 88,89]. Emerging evidence from these initiatives suggests that NbS, in certain contexts, provide low-cost solutions to many climate change-related impacts and offer key advantages over engineered solutions [90]. In particular, NbS are reported to deliver a wider range of ecosystem services especially to more vulnerable sectors of society, to protect us against multiple impacts and to be deliverable at

lower cost [90]. Many of these observations are increasingly backed up by research (Box 2), although there remains a lack of scientific synthesis and there are several knowledge gaps, in particular around how the cost-effectiveness of NbS compares to alternatives [91]. Here we argue that instead of framing NbS as an alternative to grey approaches, we suggest the focus should be on finding synergies among different solutions. We must identify sets of integrated actions that address a range of climatic impacts while providing additional ecosystem services, and that can be feasibly implemented and managed over landscapes and seascape in the long-run.

### (a) Difficulties in measuring effectiveness

A major difficulty comes in identifying appropriate indicators and metrics for the social-ecological effectiveness of nature-based interventions [92]. Effectiveness in delivering a specific climatic adaptation benefit—for example, reducing the impact of floods arising through increased precipitation—is influenced by many interacting, context-specific factors that fluctuate over time. These may be socioeconomic (e.g. institutional capacity to respond to an impact, including human and financial capital to design and implement an intervention); biophysical (e.g. frequency and intensity of natural hazards); and ecological (e.g. variation in the delivery of ecosystem services as a result of seasonal and spatial changes in biomass [93]). Also, what counts as effective depends on the perspectives and needs of those involved. Even if reasonable metrics could be identified, the dynamic and complex nature of social-ecological systems [94] including the potential for unexpected shifts in political support or ecosystem condition [95], make measuring and comparing the outcomes of interventions across scales extremely challenging [96]. As such, simple standardized metrics of NbS effectiveness that work across different scales, or that comprehensively capture the social-ecological dimensions of effectiveness, are unlikely to be found. Instead we must devise a suite of context-specific metrics (e.g. [97]). Such metrics will help increase our understanding of NbS effectiveness at the local level, and reduce the chance of unintended consequences or maladaptation.

### (b) How cost-effective are NbS?

The benefits of NbS have been found to outweigh the costs of implementation and maintenance in a range of contexts, including disaster (mainly flood) risk reduction along coasts [98-100] and in river catchments [101]. There is also growing evidence

that NbS can be more cost-effective than engineered alternatives, at least when it comes to less extreme hazard scenarios [102]. For example, across 52 coastal defence projects in the USA, NbS were estimated to be 2-5 times more cost-effective at lower wave heights and at increased water depths compared to engineered structures [103]. Natural flood management approaches in the UK (such as leaky dams and catchment woodland) significantly reduce hazards associated with small floods in small catchments, but do not appear to have a major effect on the most extreme events (though data from such events is lacking) [104,105].

The problem with current evidence for the cost-effectiveness of NbS is that appraisals in general do not use an appropriate framework, and as a result underestimate the economic benefits of working with nature, especially over the longterm. There are four major issues that need addressing. First, NbS are often highlighted as multi-functional, with the potential to deliver a wide range of benefits to both local and global communities. Yet benefits such as food and water security, carbon sequestration, and space for recreation, whether locally or beyond the immediate area of implementation [106], are rarely accounted for. This may be because they are difficult to monetise, or there is high uncertainty about non-market value [107,108].

Second, appraisals rarely factor in trade-offs among different interventions and ecosystem services, or between stakeholder groups, which may experience the costs and benefits of NbS differently (often reflecting differences in the extent of dependency on natural resources [109]).

Third, changes in the provision of ecosystem services over time, for example, under climate change and other stressors, are rarely considered, and there are major questions about how to balance future benefits with current costs [64, 110]. Engineered solutions can usually be implemented with relative certainty about the type and timescale of benefits, whereas NbS generally offer more flexible long-term solutions with benefits that might not be reaped when the costs are felt (or within standard political or electoral cycles).

Finally, perhaps the biggest challenge around estimating the cost-effectiveness of nature-based approaches relates to the variable levels of protection they offer (as discussed above, efficacy can vary with intensity and frequency of threats, the resilience of the ecosystem to withstand climate change impacts, and the vulnerabilities of the socioeconomic system). As a result, the response of ecosystems are much harder to predict and cost than engineered/grey infrastructure

[111] although recent modelling advances for predicting the efficacy of natural landforms in reducing hazards are helping to reduce this uncertainty [112].

In view of the complementary costs and benefits of NbS versus engineered approaches to dealing with the risks posed by climate change, there is growing consensus among ecologists, engineers and managers that a combination of green and grey may be the best solution in many contexts [113,114]. For example, the effectiveness saltmarshes for flood risk reduction can be increased by constructing breakwaters, or by artificially elevating salt-marsh foreshores [115]. Such a mix of interventions may also help address diverging stakeholder needs [109]. We urge researchers, policy makers and practitioners alike to focus on identifying integrated solutions that address a range of climatic impacts, provide additional ecosystem services, and can be feasibly implemented over the longterm.

## 6. Financing and governing NbS

To translate our understanding of the socioeconomic effectiveness of NbS into action on the ground, we need to consider the political processes that shape which interventions are adopted and why, and understand how to effectively finance, implement and govern those interventions.

### (a) Lack of investment in NbS

Despite broad recognition of the severe threats to the global economy posed by climate change [13], less than 5% of climate finance goes towards dealing with climate impacts, and less than 1% goes to coastal protection, infrastructure and disaster risk management including NbS [116]. This is despite growing evidence that natural habitats provide major economic benefits in the form of avoided losses from climate change related disasters [55,103], as well as supporting ecosystems services worth an estimated \$125 trillion annually [117]. For example, in their recent report, the Global Adaptation Commission highlights that the benefits of mangrove protection and restoration (i.e. fisheries, forestry, recreation and disaster risk reduction) are up to 10 times the costs [148]. However, NbS are “deplorably undercapitalized” [118], and this lack of finance is widely recognised as one of the main barriers to the implementation and monitoring of NbS across the globe [119-122].



Funding for NbS comes from public and private, bilateral and multilateral, national and international funds (e.g. Global Environmental Facility, Green Climate Fund, Adaptation Fund). Climate finance for forestry projects is mainly provided through payments for ecosystem services programs (PES, including carbon credits) under the UNFCCC compliance (GCF) or the voluntary market (private funding). However, there remains much uncertainty about the extent to which PES can deliver social and ecological benefits [123].

The availability of funding is often the trigger needed for action [124], especially when there are significant implementation costs [125,126] such as where infrastructure and people need to be relocated for planned retreat to create intertidal habitats for flood protection [125]. However, raising the necessary finance for such interventions is complex. Funding instruments can be difficult to apply for and/or require co-financing [64]. Moreover, the short-term nature of public and private sector decision-making hinders the longer-term planning and maintenance required for the emergence and sustained provisioning of NbS benefits [64].

A large part of the problem is that many of the benefits associated with NbS cannot be capitalized by any one party or organization. They create externalities that impact on many different groups, resulting in a problem of ownership. Financing for NbS requires the provision of appropriate risk sharing arrangements. In most cases investments are financed by debt, leaving those undertaking the projects to bear a substantial proportion of the risks. For example, bank lending and microfinance - the most widely used sources of external funding in developing countries - often impose risks on those least suited to bear them. Another problem with conventional finance is that it draws a distinction between providers (i.e. financial institutions and markets) and users of finance (i.e. business and individual borrowers), and traditional providers tend to lack understanding of and, most importantly, participation in the project.

Instead, what is emerging as critical to the provision of large scale, long-term investments in ecosystems is the creation of multilateral consortia of close partnerships between companies, communities, local governments, national governments, NGOs, local financial institutions, and national and international financial institutions. The consortia's willingness to provide various forms of capital reflects an understanding, influence and trust in the programme being undertaken [127,128]. Funding is then best provided in the form of equity to reflect mutual sharing and involving the measurement of less conventional forms of capital. In this



way, measurement and accounting are intimately related to the successful provision of finance.

Further work is urgently needed to test the effects of employing equity, risk sharing arrangements rather than debt finance for NbS, such as by conducting randomized control trials to examine the effects of moving from traditional to more innovative forms of financing. Finally, since the investments relate to human, social and natural capital, not just material and financial capital, there is also a need to greatly improve the measurement of these forms of capital. The failure to recognise expenditures on human, social and natural capital as assets, depreciated accordingly, partly explains the lack of investment in NbS projects.

Ultimately, the demands of growth-based economies, with entrenched policy and market conditions favouring industrialized and extractive land-uses present a serious barrier to upscaling sustainable landscape interventions [129]. The focus on economic-growth and short-term profits can reduce options considered by private or government sector actors which may not see NbS projects as bankable, particularly when faced with severe budget constraints [64].

## (b) Challenges to governing NbS

NbS often involve multiple actions taking place over broad landscapes and seascapes, crossing jurisdictional boundaries. For example, effective management of storm-water drainage across watersheds using nature-based approaches requires joint decision-making across different local, regional or even national governments and among multiple ministries (agriculture, forestry, and environment, finance, development, transport). Therefore, to be successful, governance of NbS requires (and indeed enables) active cooperation and coordinated action between stakeholders whose priorities, interest, or values may not align, or may even conflict [122]. A lack of policy coherence can lead to inaction when one agency sees “adaptation” as the responsibility of another [64]. It can also result in trade-offs, leading to conflicts. For example, landslide control through tree planting to protect infrastructure might come at the cost of agricultural productivity if reductions in ground water recharge is compromised, as shown in some NbS projects, including nature-based coastal management [122]. However, these trade-offs can be reduced if watersheds are restored with native trees where positive benefits of water supply materialize over time [130].

Unsupportive, or even conflicting incentives and regulations can also hinder the uptake of NbS [83, 120,131, 132]. For example, a lack of government incentives is a key barrier to scaling up green infrastructure in Hong Kong [83]. Existing regulatory frameworks, such as land-use rights or environmental and building permit schemes, plans, or codes, or sectoral policies can conflict with environmental management needs and hinder NbS uptake [120,122]. Examples include rural development payment schemes [121], post-disaster recovery policies [113], policies promoting intensive agriculture such as oil palm and subsidies for sheep farming [131].

Other institutional norms also limit uptake of NbS. Path dependency, whereby decision-makers implement solutions familiar to them, can be a formidable barrier to NbS uptake [132]. Decisions may be driven by power-relations, whereby choice of infrastructure is influenced by interests connected to property and appropriation regimes which do not support NbS [133]. Grey infrastructural approaches are deeply engrained in certain cultural contexts, and shape institutional practices. Such biases are compounded by cognitive factors such as a lack of awareness of ecosystem services provided by NbS, lack of perceived responsibility for action, or the discounting of climate risks [134-136] and similar issues that constrain innovation [87]. Overcoming these challenges requires strong institutions, and well-established planning structures, processes, and instruments to ensure benefits across landscapes and seascapes [124,131].

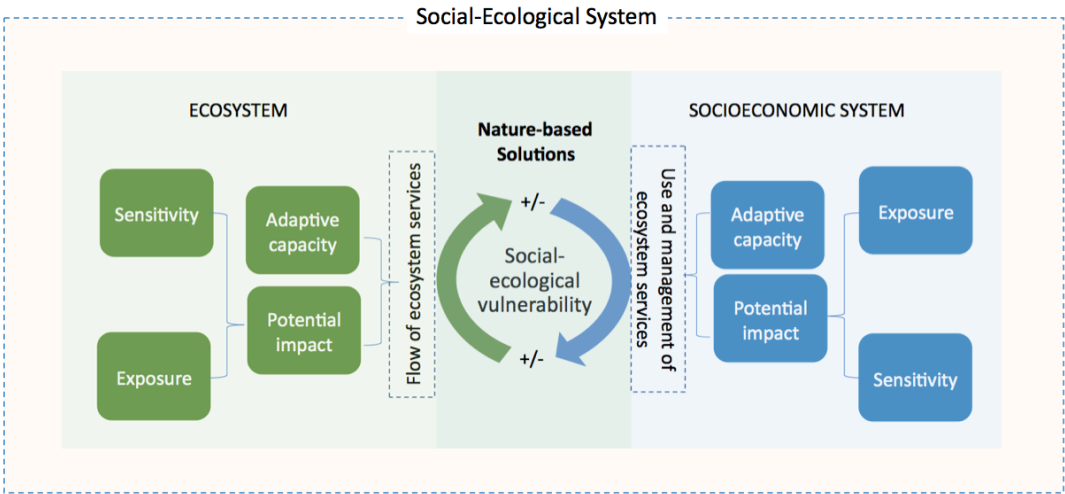
## 7. Conclusions

NbS are gaining traction in international policy and business discourse. They offer huge potential to address both causes and consequences of climate change whilst supporting biodiversity and thereby securing the flow of ecosystem services on which human wellbeing depends. Yet three barriers hinder the evidence-based integration of NbS into international, national and local climate and development policy and practice. First, challenges in measuring or predicting the effectiveness of NbS lead to high uncertainty about their cost-effectiveness compared to alternatives. Second, poor financial models and flawed approaches to economic appraisal lead to under-investment in NbS. And third, inflexible and highly sectoralised forms of governance hinder uptake of NbS, with grey, engineered interventions still being the default approach for many climate adaptation and mitigation challenges [137].

Overcoming these challenges requires major systemic change in how we conduct and communicate interdisciplinary research, and how we organise and run our institutions. More fundamentally, fully integrating NbS as solutions to both the climate and biodiversity crises requires a new approach in economic thinking, shifting from a focus on infinite economic growth to a recognition that the energy and material flows needed for human well-being must remain within safe biophysical limits [138,139]. NbS can play a key role in enabling sustainable development within planetary boundaries. However, their benefits will not be realised unless they are implemented within a systems-thinking framework that fully accounts for their potential to support multiple ecosystem services and the trade-offs among them. As nations revise their climate policies (Nationally Determined Contributions) in 2020, and climate policy increasingly turns towards greenhouse gas removal approaches to help achieve climate targets [140], further elucidation of this systematic framework should be an urgent priority for future research. The revision of CBD biodiversity targets and the launch of the Emergency Deal for Nature in 2020 should prompt scientists of all disciplines to fully engage with these issues, working together to find climate solutions that also address the biodiversity crisis and help to restore planetary health.

## **Acknowledgements**

This work was supported by a NERC Knowledge Exchange Fellowship awarded to NS.



**Figure 1: Integrating Nature-based Solutions to climate change impacts into the social-ecological vulnerability framework**

*Ecosystem exposure* is the extent to which systems are subject to pressures (floods, droughts, landslides, fires, etc). It is determined by the intensity, duration and frequency of events, geomorphology, and the extent of use and management of natural resources by human societies. *Ecosystem sensitivity* is the degree to which ecosystem structure and function alters as a result of perturbations. Ecosystem exposure combined with ecosystem sensitivity creates a *potential impact*. This is buffered by the *adaptive capacity* of the ecosystem. Both ecosystem sensitivity and adaptive capacity are determined by the diversity, heterogeneity and connectedness of the ecosystem and the characteristics and condition of its component species and habitats. Overall *ecosystem vulnerability* is shaped by the combination of potential impact and adaptive capacity. This ultimately affects the delivery of ecosystem goods and services upon which people and economies depend. In this way, ecosystem vulnerability affects *socioeconomic vulnerability* i.e., the degree to which the social system is adversely affected by change. *Socioeconomic sensitivity* is also influenced by a range of social, political and economic factors. For example, corruption or low levels of health, education or employment can increase socioeconomic sensitivity. Likewise, *socioeconomic adaptive capacity*, that can moderate the potential impact from social exposure and sensitivity, includes the ability to innovate (e.g. improving health, education, and finding alternatives sources of income). Nature-based solutions bring all these elements together and can, if implemented properly and equitably, decrease social-ecological vulnerability (see main text, and Box 2).

577 **Box 2.** Examples of Nature-based Solutions relevant for climate change adaptation organised with respect to dimension of socioeconomic  
578 vulnerability and type of climate change impact mitigated

Dimension 1: Reducing exposure	Protection from erosion and landslides
	<ul style="list-style-type: none"><li>• <b>China:</b> A combination of afforestation, reforestation, and conservation of existing natural forests over 25 years in the Poyang Lake basin halved heavy soil erosion while increasing net carbon sequestration five-fold and net income for local farmers six-fold [48]. Meanwhile, restoration of natural herbaceous and shrub-land vegetation on the Loess Plateau reduced soil erosion to a comparably or significantly greater extent than tree plantations, across a range of anti-soil erosion indices. Compared to afforested slopes, these naturally re-vegetated slopes also had 1.3-2 times higher soil water content [51].</li></ul>
	Protection from inland flooding
	<ul style="list-style-type: none"><li>• <b>Canada:</b> Reforestation in the headwaters of a river basin significantly reduced peak stream flows compared to an adjacent deforested basin, offering greater protection against flooding during spring snow melt [141].</li><li>• <b>USA:</b> Natural regeneration of mixed species hardwood watersheds following forest clearcutting reduced flood risk in lowland areas, reducing stream flows during periods of high precipitation by &gt;104 L/ha/day [142].</li><li>• <b>Europe:</b> Restoration of all but one of six rivers reduced flood damage to crops and forests, and was associated with increased agricultural production, carbon sequestration and recreation, with a net societal economic benefit over unrestored rivers of €1400 ± 600 [143].</li></ul>
	Protection from coastal hazards and sea level rise
	<ul style="list-style-type: none"><li>• <b>Global:</b> Natural coastal habitats significantly reduce wave heights, with coral reefs and salt-marshes being most effective, causing a reduction of 70%, followed by seagrass and kelp beds (36%), and mangroves (31%). Across 52 sites harnessing these habitats in coastal defence projects, nature-based solutions were 2-5 times more cost-effective at lower wave heights and at increased water depths compared to engineered structures [103].</li><li>• <b>Gulf of Mexico:</b> Construction of ‘living shorelines’ by aiding natural recruitment of oyster reefs can reduce vegetation retreat by 40% compared to unprotected sites, stabilizing the shoreline from the effects of waves and erosion, and increasing abundance and diversity of economically important species [57].</li></ul>
	Moderating urban heat waves and heat island effects
	<ul style="list-style-type: none"><li>• <b>USA:</b> Daytime air temperature is substantially reduced with greater canopy cover (≥40%) at the scale of a typical city block (60–90 m), especially on the hottest days [62].</li><li>• <b>Global:</b> Green spaces are on average 0.94 °C cooler in the day than urban spaces, with stronger effects the larger the green space, according to a meta-analysis of 47 studies comparing the cooling effects of green spaces in cities (parks, areas with trees) with those of purely urban areas [].</li></ul>
	Managing storm-water and flooding in urban areas

	<ul style="list-style-type: none"><li>• <b>Italy:</b> Establishment of wetlands and green recreational space has been effective in reducing flood risks, with a 10% higher reduction of downstream flooding and 7.5% higher reduction of peak flow compared to potential grey infrastructure alternatives. NbS also outperform grey infrastructure in terms of water purification and provide greater social-ecological benefits, such as recreation and habitat for biodiversity [144].</li></ul>
	<p><b>Sustaining natural resources in drier and more variable climates</b></p> <ul style="list-style-type: none"><li>• <b>Panama:</b> Agroforestry systems yield up to 21% higher economic return than farm mosaic approaches (i.e. where trees and crops are on separate parcels), including under a climate change scenario of more frequent droughts, in models that account for market and climate uncertainty [145].</li><li>• <b>Europe:</b> Agroforestry has reduced erosion and increased soil fertility, with greatest effects in hotter, drier regions such as the Mediterranean basin (which is suffering from soil damage through increasing aridity under climate change) [58].</li><li>• <b>Australia:</b> Forest management by tree thinning in <i>Eucalyptus delegatensis</i> forest significantly altered forest structure and canopy openness, decreasing fuel loads and hazards, and thereby achieving a 30% reduction in the intensity and 20% reduction in spread of fires compared to un-thinned forests under simulations of severe weather conditions. Such actions could be critical to ensure the resilience of commercially valuable <i>E. delegatensis</i> to wildfires, which are expected to increase in the future [80].</li></ul>
Dimension 2: Reducing sensitivity	<p><b>Buffering communities from climate shocks by enhancing and diversifying ecosystem services</b></p> <ul style="list-style-type: none"><li>• <b>Kenya:</b> Allowing rangelands in the Kenyan drylands to regenerate, through restoration within rangeland enclosures, diversifies income sources, which can cushion against climatic shocks [65,66]. Meanwhile, agroforestry in semi-arid regions provides alternative income sources including fuelwood, fruit, timber, as well as reducing exposure to heat, drought, floods and erosion [68].</li><li>• <b>Zimbabwe:</b> Protection of forested / wooded areas ensures honey production during droughts, thereby providing a degree of food security when other crops fail [67].</li></ul>
Dimension 3: Supporting adaptive capacity	<p><b>Governance reform, empowerment, and improving access to resources</b></p> <ul style="list-style-type: none"><li>• <b>Sri Lanka:</b> EbA empowered marginalized groups to respond to climate change impacts by supporting common-pool resource management institutions, and by supporting local adaptive strategies such as home gardening [71].</li><li>• <b>Ethiopia:</b> Community-based natural resource management in pastoral communities has improved institutional governance by transforming it towards a more flexible, inclusive, bottom-up approach, whereby community members become informed members of the decision-making process. This inclusivity in particular empowered women and the most vulnerable households. Altogether this has increased the capacity of these communities to deal with climate change [146].</li><li>• <b>Bangladesh:</b> Ecosystem-based adaptation has increased the adaptive capacity of coastal communities to extreme weather events and climate change by improving their access to institutional services and climate change information, as well as their access to natural resources to support diverse livelihood options [147].</li><li>• <b>Togo:</b> Ecosystem-based adaptation increased social inclusion and self-sufficiency of women and youth groups, leading also to increased crop yields for these savannah communities as a whole, whose food security is threatened by climate change. Community</li></ul>

members were also involved from the beginning, allowing them to learn how to design and implement such projects to be able to independently adapt to future changes as well [70].

579  
580



## References

1. United Nations. 2015 Transforming our world: the 2030 Agenda for Sustainable Development. Resolution A/RES/70/1, adopted by the General Assembly on 25 September 2015.
2. Moyer JD, Bohl DK. 2019 Alternative pathways to human development: Assessing trade-offs and synergies in achieving the Sustainable Development Goals. *Futures* **105**, 199-210.
3. Smith A. 2013 *The climate bonus: co-benefits of climate policy*. Routledge. 448 pp.
4. Gonzales-Zuñiga et al. 2018 SCAN (SDG & Climate Action Nexus) tool: Linking Climate Action and the Sustainable Development Goals. Key Findings Note. [http://ambitiontoaction.net/wp-content/uploads/2018/10/Key\\_findings\\_final.pdf](http://ambitiontoaction.net/wp-content/uploads/2018/10/Key_findings_final.pdf)
5. IPCC. 2018 Global warming of 1.5°C. An IPCC Special Report on the impacts of global warming of 1.5°C above pre-industrial levels and related global greenhouse gas emission pathways, in the context of strengthening the global response to the threat of climate change, sustainable development, and efforts to eradicate poverty. World Meteorological Organization, Geneva, Switzerland.
6. IPBES. 2019. Summary for policymakers of the global assessment report on biodiversity and ecosystem services of the Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services. S. Díaz, J. Settele, E. S. Brondizio E.S., H. T. Ngo, M. Guèze, J. Agard, A. Arneth, P. Balvanera, K. A. Brauman, S. H. M. Butchart, K. M. A. Chan, L. A. Garibaldi, K. Ichii, J. Liu, S. M. Subramanian, G. F. Midgley, P. Miloslavich, Z. Molnár, D. Obura, A. Pfaff, S. Polasky, A. Purvis, J. Razzaque, B. Reyers, R. Roy Chowdhury, Y. J. Shin, I. J. Visseren-Hamakers, K. J. Willis, and C. N. Zayas (eds.). IPBES secretariat, Bonn, Germany.
7. WWF. 2018 *Living Planet Report - 2018: Aiming Higher*. WWF, Gland, Switzerland.
8. Seddon N et al. 2019a Grounding nature-based climate solutions in sound biodiversity science. *Nature Climate Change* **9**, 84–87.
9. Stein et al. 2013 Preparing for and managing change: climate adaptation for biodiversity and ecosystems. *Frontiers of Ecosystems and Environment* **11**, 502–510.
10. Seddon N et al. 2019b Global recognition of the importance of nature-based solutions to climate change impacts *Global Sustainability* (in rev)
11. IPCC. 2019. Climate and Land: An IPCC Special Report on Climate Change, Desertification, Land Degradation, Sustainable Land Management, Food Security, and Greenhouse gas fluxes in Terrestrial Ecosystems. <https://www.ipcc.ch/report/srccl/>
12. <https://www.un.org/en/climatechange/climate-action-areas.shtml>
13. World Economic Forum. 2019 The Global Risks Report 2019, 14th Edition.
14. Calliari E et al. 2019 An assessment framework for climate-proof nature-based solutions. *Science of the Total Environment* **656**, 91-700.

- 626 15. Cohen-Shacham E et al. (eds.) 2016 Nature-based Solutions to address global  
627 societal challenges. Gland, Switzerland: IUCN. xiii + 97pp.
- 628 16. EC. 2015 Towards an EU research and innovation policy agenda for nature  
629 based solutions & re-naturing cities. Final report of the Horizon 2020 expert  
630 group on 'Nature-based solutions and re-naturing cities'.
- 631 17. Millennium Ecosystem Assessment. 2005 *Ecosystems and Human Well-being:  
632 Synthesis*. Island Press, Washington, DC.
- 633 18. Nature. 2017 Natural language: the latest attempt to brand green practices is  
634 better than it sounds. *Nature* 541, 133-134.
- 635 19. Griscom BW et al. 2017 Natural climate solutions. *Proc. National Acad.  
636 Sciences* 114, 11645–11650.
- 637 20. Griscom BW et al. 2019 We need both natural and energy solutions to stabilize  
638 our climate. *Glob. Change Biol.* doi:10.1111/gcb.14612
- 639 21. Fargione JE, Bassett S, Boucher T, Bridgham SD, Conant RT, Cook-Patton,  
640 SC. et al. 2018 Natural climate solutions for the United States. *Sci Adv.*  
641 4:eaat1869. doi: 10.1126/sciadv.aat1869
- 642 22. Sutton-Grier A. et al. 2018 Investing in natural and nature-based infrastructure:  
643 building better along our coasts. *Sustainability* 10, 523.
- 644 23. Watson JE et al. 2018 The exceptional value of intact forest ecosystems.  
645 *Nature Ecology and Evolution* 2, 599-610.
- 646 24. Keeler BL et al. 2019. Social-ecological and technological factors moderate the  
647 value of urban nature. *Nature Sustainability* 2, 29-38.
- 648 25. Reid, H, Bourne, A, Muller, H, Podvin, K, Scorgie, S & Orindi, V. 2018 A  
649 Framework for Assessing the Effectiveness of Ecosystem-Based Approaches  
650 to Adaptation. In *Resilience* (pp. 207-216, Elsevier).
- 651 26. Secretariat of the Convention on Biological Diversity. 2009 *Connecting  
652 Biodiversity and Climate Change Mitigation and Adaptation: Report of the  
653 Second Ad Hoc Technical Expert Group on Biodiversity and Climate Change*.
- 654 27. Brancalion PHS, Chazdon R. 2017 Beyond hectares: four principles to guide  
655 reforestation in the context of tropical forest and landscape restoration  
656 *Restoration Ecology* doi: 10.1111/rec.12519.
- 657 28. Lewis SL, Wheeler CE, Mitchard ETA, Kock A. 2019. Restoring natural forest is  
658 the best way to remove atmospheric carbon. *Nature* 568, 25-28.
- 659 29. Erb K-H et al. 2018 Unexpectedly large impact of forest management and  
660 grazing on global vegetation biomass *Nature* 553, 73–76.
- 661 30. Le Quéré C. et al. 2018 *Earth Syst. Sci. Data* 10, 405-448.
- 662 31. Royal Society and Royal Academy of Engineering. 2018. Greenhouse Gas  
663 Removal. Royal Society, UK. (royalsociety.org/greenhouse-gas-removal  
664 raeng.org.uk/greenhousegasremoval)
- 665 32. Brancalion PHS et al (2019) Global restoration opportunities in tropical  
666 rainforest landscapes. *Science Advances* 5 : eaav3223.
- 667 33. Canadell JG, Schulze ED. 2014 Global potential of biospheric carbon  
668 management for climate mitigation. *Nature Communications* 5, 1–12.
- 669 34. Grace J, Mitchard E, Gloor E. 2014 Perturbations in the carbon budget of the  
670 tropics. *Global Change Biology* 20, 3238–3255.

- 671 35. Houghton RA, Byers B, Nassikas AA. 2015 A role for tropical forests in  
672 stabilizing atmospheric CO<sub>2</sub>. *Nature Climate Change* **5**, 1022–1023.
- 673 36. Poorter L et al. 2016 Biomass resilience of Neotropical secondary forests  
674 *Nature* **530**, 211–214.
- 675 37. Liu X et al. 2018 Tree species richness increases ecosystem carbon storage in  
676 subtropical forests. *Proc. R. Soc. B* **285**, 20181240.
- 677 38. Smith A et al. 2017 How natural capital delivers ecosystem services: a typology  
678 derived from a systematic review. *Ecosystem Services* **26**, 111–126.
- 679 39. Hutchinson, C. et al. 2018 Effect of diversity on growth, mortality, and loss of  
680 resilience to extreme climate events in a tropical planted forest experiment  
681 *Scientific Reports* **8**, 15443.
- 682 40. Guyot V, Castagneyrol B, Vialatte A, Deconchat M, Jactel H. 2016 Tree  
683 diversity reduces pest damage in mature forests across Europe. *Biology*  
684 *Letters* **12**, 20151037.
- 685 41. Thompson I, Mackey B, McNulty S, Mosseler A. 2009 Forest Resilience,  
686 biodiversity, and climate change. A synthesis of the biodiversity/resilience/  
687 stability relationship in woodland ecosystems. Technical Series 43, Secretariat  
688 of the Convention on Biological Diversity, Montreal.
- 689 42. Ennos R, Cottrell J, Hall J, O'Brien D. 2019 Is the introduction of novel exotic  
690 forest tree species a rational response to rapid environmental change? – a  
691 British perspective. *For. Ecol. Manag.* **432**, 718–728
- 692 43. Veldman JW, Overbeck GE, Negreiros D, Mahy G, Stradic, Wilson Fernandes  
693 G, Durigan G, Buisson E, Putz FE, Bond WJ. 2015 Where tree planting and  
694 forest expansion are bad for biodiversity and ecosystem services, *BioScience*  
695 **65**, 1011–1018.
- 696 44. COP16, FCCC/CP/2010/7/Add.1
- 697 45. Anderson CM, DeFries RS, Litterman R et al. 2019 Natural climate solutions  
698 are not enough. *Science* **363**, 933–934.
- 699 46. <https://www.worldweatherattribution.org/>
- 700 47. Jones HP et al. 2012 Harnessing nature to help people adapt to climate  
701 change. *Nat. Clim. Chang.* **2**, 504–509.
- 702 48. Huq N, Bruns A, Ribbe L, Huq S. 2017 Mainstreaming Ecosystem Services  
703 Based Climate Change Adaptation (EbA) in Bangladesh: Status, Challenges  
704 and Opportunities. *Sustainability* **9**, 926. doi:10.3390/su9060926.
- 705 49. Marshall NA, Marshall PA, Tاملander J, Obura D, Malleret-King D, Cinner JE.  
706 2010. A framework for social adaptation to climate change sustaining tropical  
707 coastal communities and industries. IUCN, Gland, Switzerland.
- 708 50. Thiault L, Marshall P, Gelcich S, Collin A, Chlous F, Claudet J. 2017 Mapping  
709 social–ecological vulnerability to inform local decision-making. *Conservation*  
710 *Biology* **32**, 447–456
- 711 51. Jiao J. et al. 2012 Assessing the ecological success of restoration by  
712 afforestation on the Chinese Loess Plateau. *Rest. Ecol.* **20**, 240–249.
- 713 52. Bradshaw CJA, Sodhi NS, Pek SH, Brook BW. 2007 Global evidence that  
714 deforestation amplifies flood risk and severity in the developing world. *Global*  
715 *Change Biology* **13**, 2379–2395.

- 716 53. Huang L. et al. 2012 Forest restoration to achieve both ecological and  
717 economic progress, Poyang Lake basin, China. *Ecol. Eng.* **44**, 53–60.
- 718 54. Narayan S, Beck MW, Wilson C, Thomas CJ, Guerrero A, Shepard CC,  
719 Reguero BG, Franco G, Ingram JC, Trespalacios D. 2017 The Value of Coastal  
720 Wetlands for Flood Damage Reduction in the Northeastern USA. *Scientific*  
721 *Rep.* **7**, 9463.
- 722 55. Beck, MW, Losada, IJ, Menéndez, P, Reguero, BG, Díaz-Simal, P &  
723 Fernández, F. 2018 The global flood protection savings provided by coral reefs.  
724 *Nature communications* **9**, 2186.
- 725 56. Temmerman S, Meire P, Bouma TK et al. 2013 Ecosystem-based coastal  
726 defence in the face of global change. *Nature* **504**, 79-83
- 727 57. Scyphers SB et al. 2011 Oyster reefs as natural breakwaters mitigate shoreline  
728 loss and facilitate fisheries. *PLoS ONE* **6**, e22396.
- 729 58. Torralba M. et al. 2016 Do European agroforestry systems enhance  
730 biodiversity and ecosystem services? A meta-analysis. *Agriculture, Ecosystems*  
731 *and Environment* **230**, 150-161.
- 732 59. Tscharntke T, Clough Y, Bhagwat SA, Burchori D, Faust H, Hertel D, Scherber  
733 C 2011. Multifunctional shade-tree management in tropical agroforestry  
734 landscapes—a review. *Journal of Applied Ecology* **48**, 619-629.
- 735 60. Alexandri E, Jones P. 2008 Temperature decreases in an urban canyon due to  
736 green walls and green roofs in diverse climates. *Building and Environment* **43**,  
737 480-493;
- 738 61. Bowler DE et al. 2010 Urban greening to cool towns and cities: A systematic  
739 review of the empirical evidence. *Landscape and Urban planning* **97**, 147-155
- 740 62. Ziter C, Pederson EJ, Kucharik CJ, Turner MG. 2019. Scale-dependent  
741 interactions between tree canopy cover and impervious surfaces reduce  
742 daytime urban heat during summer. *Proc. Natl. Acad. Sciences* **116**, 7575-  
743 7580.
- 744 63. Liu W, Chen W, Peng C. 2014. Assessing the effectiveness of green  
745 infrastructures on urban flooding reduction: A community scale study. *Ecol.*  
746 *Modell.* **291**, 6–14.
- 747 64. Kabisch et al. 2016 Nature-based solutions to climate change mitigation and  
748 adaptation in urban areas: perspectives on indicators, knowledge gaps,  
749 barriers, and opportunities for action. *Ecology and Society* **21**, 26270403.
- 750 65. Mureithi SM, Verdoodt A, Njoka JT, Gachene CK, Van Ranst E. 2016 Benefits  
751 derived from rehabilitating a degraded semi-arid rangeland in communal  
752 enclosures, Kenya. *Land Degradation & Development* **27**, 1853-1862.
- 753 66. Wairore JN, Mureithi SM, Wasonga OV, Nyberg G. 2016. Benefits derived from  
754 rehabilitating a degraded semi-arid rangeland in private enclosures in West  
755 Pokot County, Kenya. *Land Degradation & Development* **27**, 532-541.
- 756 67. Lunga W, Musarurwa C. 2016 Exploiting indigenous knowledge commonwealth  
757 to mitigate disasters: from the archives of vulnerable communities in  
758 Zimbabwe. *Indian Journal of Traditional Knowledge* **15**, 22-29.

- 759 68. Quandt A , Neufeldt A, McCabe JT. 2017 The role of agroforestry in building  
760 livelihood resilience to floods and drought in semi-arid Kenya. *Ecology and*  
761 *Society* **22**, 10.
- 762 69. Reid H, Faulkner L, Weiser A. 2013. *The role of community-based natural*  
763 *resource management in climate change adaptation in Ethiopia* (No. 6).  
764 Climate Change Working Paper.
- 765 70. Munang R, Andrews J, Alverson K, Mebratu D. 2014. Harnessing ecosystem-  
766 based adaptation to address the social dimensions of climate  
767 change. *Environment: Science and Policy for Sustainable Development* **56**, 18-  
768 24.
- 769 71. Woroniecki S. 2019 Enabling Environments? Examining Social Co-Benefits of  
770 Ecosystem-Based Adaptation to Climate Change in Sri Lanka. *Sustainability*  
771 **11**, 772.
- 772 72. Abdul-Razak M, Kruse S. 2017 The adaptive capacity of smallholder farmers to  
773 climate change in the Northern Region of Ghana *Climate Risk Management* **17**,  
774 104-122.
- 775 73. Morris, RL, Konlechner, TM, Ghisalberti, M & Swearer, SE. 2018 From grey to  
776 green: Efficacy of eco-engineering solutions for nature-based coastal defence.  
777 *Global change biology* **24**, 1827-1842.
- 778 74. Standish, RJ, Hobbs, RJ, Mayfield, MM, Bestelmeyer, BT, Suding, KN,  
779 Battaglia, LL, Eviner, V, Hawkes, CV, Temperton, VM & Cramer, VA. 2014  
780 Resilience in ecology: Abstraction, distraction, or where the action is?  
781 *Biological Conservation* **177**, 43-51.
- 782 75. Dass P. et al. 2018 Grasslands may be more reliable carbon sinks than forests  
783 in California. *Environ. Res. Lett.* **13**, 074027.
- 784 76. Allen CD et al. 2010 A global overview of drought and heat-induced tree  
785 mortality reveals emerging climate change risks for forests. *Forest Ecology and*  
786 *Management* **259**, 660-684.
- 787 77. Turner MG, Braziunas KH Hansen WD, Harvey BJ. 2019 Short-interval fire  
788 erodes the resilience of subalpine lodgepole pine forests. *Proceedings of the*  
789 *National Academy of Sciences* (Minor revision was approved, awaiting editorial  
790 board decision).
- 791 78. Woodroffe K, Rogers KL, McKee CE, Lovelock IA, Mendelssohn, Saintilan N.  
792 2016 Mangrove Sedimentation and Response to Relative Sea-Level Rise. *Ann.*  
793 *Rev. Mar. Sci.* **8**, 243-266.
- 794 79. Valiela I, Lloreta J, Bowyer T, Minera S, Remsena D, Elmstrom E, Cogswell C,  
795 Thieler ER. 2018 Transient coastal landscapes: Rising sea level threatens salt  
796 marshes. *Science of The Total Environment* **640**, 1148-1156.
- 797 80. Volkova L. et al. 2017 Impact of mechanical thinning on forest carbon, fuel  
798 hazard and simulated fire behaviour in *Eucalyptus delegatensis* forest of south-  
799 eastern Australia. *Forest Ecology and Management* **405**, 92-100.
- 800 81. Isbell I. et al. 2017 Linking the influence and dependence of people on  
801 biodiversity across scales. *Nature* **546**, 65-72.
- 802 82. Jactel H. et al. 2017 *Curr. For. Rep.* **3**, 223-243.



- 803 83. Zhang X, Shen L, Tam VW, Lee WWY. 2012 Barriers to implement extensive  
804 green roof systems: a Hong Kong study. *Renewable and sustainable energy*  
805 *reviews* **16**, 314-319.
- 806 84. Scheffers BR, De Meester L, Bridge TCL, Hoffman AA, Pandolfi J, Corlett RT,  
807 Butchart SHM, Pearce-Kelly P, Kovac KM, Dudgeon D, Pacifici M, Rondinini C,  
808 Foden WB, Martin TG, Mora C, Bickford D, Watson JEM. 2016 The broad  
809 footprint of climate change from genes to biomes to people *Science* **354**,  
810 aaf7671, 10.1126/science.aaf7671.
- 811 85. Graham NAJ, S. Jennings, M. A. MacNeil, D. Mouillot, S. K. Wilson. 2015  
812 Predicting climate-driven regime shifts versus rebound potential in coral reefs.  
813 *Nature* **518**, 94–97.
- 814 86. Lavorel S. 2015 Ecological mechanisms underlying ecosystem-based  
815 adaptation. *Global Change Biology* **21**, 12-31
- 816 87. Lavorel S. et al. 2019 Mustering the power of ecosystems for adaptation to  
817 climate change. *Environmental Science and Policy* **92**, 87-97.
- 818 88. Rizvi AR. 2014 *Nature Based Solutions for Human Resilience A Mapping*  
819 *Analysis of IUCN's Ecosystem Based Adaptation Projects*. IUCN.
- 820 89. Osti M. et al. 2015 *UNEP's portfolio of Ecosystem-based Adaptation (EBA)*  
821 *projects around the world*. Internal Summary unpublished report.
- 822 90. Reid, H, Bourne, A, Muller, H, Podvin, K, Scorgie, S, Orindi, V. 2018 A  
823 Framework for Assessing the Effectiveness of Ecosystem-Based Approaches  
824 to Adaptation. In *Resilience* (pp. 207-216, Elsevier).
- 825 91. [www.naturebasedsolutionsevidence.info](http://www.naturebasedsolutionsevidence.info)
- 826 92. Christiansen L, Martinez, GS. 2018 Adaptation metrics: Perspectives on  
827 measuring, aggregating and comparing adaptation results. In *Adaptation*  
828 *metrics: Perspectives on measuring, aggregating and comparing adaptation*  
829 *results* (eds. L. Christiansen, G. S. Martinez & P. Naswa). Copenhagen, UNEP  
830 DTU Partnership.
- 831 93. Paul M, Amos CL 2011 Spatial and seasonal variation in wave attenuation over  
832 *Zostera noltii*. *J. Geophysical Res.* **116**, C08019.
- 833 94. Ostrom E. 2009 A general framework for analyzing sustainability of social-  
834 ecological systems. *Science* **325**, 419-422.
- 835 95. Nuno, A, Bunnefeld, N & Milner-Gulland, EJ. 2014 Managing social-ecological  
836 systems under uncertainty: implementation in the real world. *Ecology and*  
837 *Society* **19**.
- 838 96. van der Jagt, A, Dorst, H, Raven, R & UU, HR. 2017 The Nature of Innovation  
839 for Urban Sustainability. In *Naturvation* (Copernicus Institute for Sustainable  
840 Development Utrecht University).
- 841 97. Raymond CM et al. 2017 An Impact Evaluation Framework to Support Planning  
842 and Evaluation of Nature-based Solutions Projects. Report prepared by the  
843 EKLIPSE Expert Working Group on Nature-based Solutions to Promote  
844 Climate Resilience in Urban Areas. Centre for Ecology & Hydrology,  
845 Wallingford, United Kingdom.
- 846 98. Reguero BG, Beck MW, Bresch DN, Calil J, Meliane I. 2018 Comparing the  
847 cost effectiveness of nature-based and coastal adaptation: A case study from  
848 the Gulf Coast of the United States. *PLoS ONE* **13**: e0192132.  
849 <https://doi.org/10.1371/journal.pone.0192132>

- 850 99. Sutton-Grier, AE, Wowk, K & Bamford, H. 2015 Future of our coasts: The  
851 potential for natural and hybrid infrastructure to enhance the resilience of our  
852 coastal communities, economies and ecosystems. *Environmental Science &*  
853 *Policy* **51**, 137-148
- 854 100. Morris R, Konlechner T, Ghisalberti M, Swearer S. 2017 From grey to green:  
855 Efficacy of eco-engineering solutions for nature-based coastal defence. *Glob.*  
856 *Chang. Biol.* **24**, 1827–1842.
- 857 101. Daigneault A. et al. 2016 Dredging versus hedging: Comparing hard  
858 infrastructure to ecosystem-based adaptation to flooding. *Ecol. Econ.* **122**, 25–  
859 35.
- 860 102. Collentine D, Futter MN. 2016 Realising the potential of natural water retention  
861 measures in catchment flood management: trade-offs and matching interests.  
862 *J. Flood Risk Manag.* <http://dx.doi.org/10.1111/jfr3.12269>.
- 863 103. Narayan, S, Beck, MW, Reguero, BG, Losada, IJ, Van Wesenbeeck, B,  
864 Pontee, N, Sanchirico, JN, Ingram, JC, Lange, G-M & Burks-Copes, KA. 2016  
865 The effectiveness, costs and coastal protection benefits of natural and nature-  
866 based defences. *PloS one* **11**, e0154735.
- 867 104. Stratford C et al. 2017 Do trees in UK-relevant river catchments influence  
868 fluvial flood peaks? A systematic review. *CEH Research Report*. Centre for  
869 Ecology and Hydrology, Wallingford, UK.
- 870 105. Dadson SJ, Hall JW, Murgatroyd A et al. 2017A restatement of the natural  
871 science evidence concerning catchment-based ‘natural’ flood management in  
872 the UK. *Proceedings of the Royal Society A*, **473**: 20160706.
- 873 106. Wild T, Henneberry J, Gill L. 2017 Comprehending the multiple ‘values’ of  
874 green infrastructure—Valuing nature-based solutions for urban water  
875 management from multiple perspectives. *Environmental research* **158**, 179-  
876 187.
- 877 107. Mukherjee N, Sutherland WJ, Dicks L, Hugé J, Koedam N, Dahdouh-Guebas  
878 F. 2014 Ecosystem service valuations of mangrove ecosystems to inform  
879 decision making and future valuation exercises. *PloS one* **9**, e107706.
- 880 108. Czembrowski P, Kronenberg J & Czepkiewicz M. 2016 Integrating non-  
881 monetary and monetary valuation methods—SoftGIS and hedonic pricing.  
882 *Ecological Economics* **130**, 166-175.
- 883 109. Reddy SM, Guannel G, Griffin R, Faries J, Boucher T, Thompson M, Brenner J,  
884 Bernhardt J, Verutes G & Wood SA. 2016 Evaluating the role of coastal  
885 habitats and sea-level rise in hurricane risk mitigation: An ecological economic  
886 assessment method and application to a business decision. *Integrated*  
887 *environmental assessment and management* **12**, 328-344.
- 888 110. Brown P, Daigneault A, Gawith D, Aalbersberg W, Comley J, Fong P, Morgan  
889 F. 2014 Evaluating ecosystem-based adaptation for disaster risk reduction in  
890 Fiji. *Landcare Research* **161**.
- 891 111. Lacob O, Rowan JS, Brown I, Ellis C. 2014 Evaluating wider benefits of natural  
892 flood management strategies: an ecosystem-based adaptation perspective.  
893 *Hydrology Research* **45**, 774-787.



112. Moller I. 2019 Applying Uncertain Science to Nature-Based Coastal Protection: Lessons From Shallow Wetland-Dominated Shores *Front. Environ. Sci.* 24 April 2019 | <https://doi.org/10.3389/fenvs.2019.00049>
113. Sutton-Grier AE, Wowk K, Bamford H. 2015 Future of our coasts: The potential for natural and hybrid infrastructure to enhance the resilience of our coastal communities, economies and ecosystems. *Environmental Science & Policy* **51**, 137-148
114. Browder et al. 2019. World bank report - <https://www.wri.org/publication/integrating-green-gray>.
115. Vuik V, Borsje BW, Willemsen PWJM, Jonkman SN. 2019 Salt marshes for flood risk reduction: Quantifying long-term effectiveness and life-cycle costs. *Ocean & Coastal Management* **171**, 96 -110.
116. UN Environment 2019. Global Environment Outlook – GEO-6: Summary for Policymakers. Nairobi. DOI 10.1017/9781108639217.
117. Costanza R. et al. 2014 Changes in the global value of ecosystem services. *Global environmental change-human and policy dimensions* **26**, 152-158
118. Figueres C 2018. *What now? Next steps on climate change*. 29 October 2018, Oxford, UK. Accessed: 30 April 2019. Available from: <https://www.oxfordmartin.ox.ac.uk/videos/view/695>
119. Faivre, N, Fritz, M, Freitas, T, de Boissezon, B & Vandewoestijne, S. 2017 Nature-Based Solutions in the EU: Innovating with nature to address social, economic and environmental challenges. *Environmental research* **159**, 509-518.
120. Brink, E, Aalders, T, Ádám, D, Feller, R, Henselek, Y, Hoffmann, A, Ibe, K, Matthey-Doret, A, Meyer, M & Negrut, NL. 2016 Cascades of green: a review of ecosystem-based adaptation in urban areas. *Global Environmental Change* **36**, 111-123.
121. McVittie, A, Cole, L, Wreford, A, Sgobbi, A & Yordi, B. 2018 Ecosystem-based solutions for disaster risk reduction: Lessons from European applications of ecosystem-based adaptation measures. *International journal of disaster risk reduction* **32**, 42-54
122. Dale, P, Sporne, I, Knight, J, Sheaves, M, Eslami-Andergoli, L & Dwyer, P. 2019 A conceptual model to improve links between science, policy and practice in coastal management. *Marine Policy* **103**, 42-49.
123. Chan, KMA et al. 2017 Payments for Ecosystem Services: rife with problems and potential for transformation towards sustainability. *Ecological Economics* **140**, 110-122,
124. Wamsler C. 2015 Mainstreaming ecosystem-based adaptation: transformation toward sustainability in urban governance and planning. *Ecology and Society* **20**.
125. Harman, BP, Heyenga, S, Taylor, BM & Fletcher, CS. 2013 Global lessons for adapting coastal communities to protect against storm surge inundation. *Journal of Coastal Research* **31**, 790-801.
126. Guida RJ, Swanson TL, Remo JW, Kiss T. 2015 Strategic floodplain reconnection for the Lower Tisza River, Hungary: Opportunities for flood-height

- 939 reduction and floodplain-wetland reconnection. *Journal of Hydrology* **521**, 274-  
940 285.
- 941 127. Mayer C. 2013 *Unnatural Capital Accounting*. Natural Capital Committee,  
942 Discussion Paper
- 943 128. Barker, Mayer C. 2017 *How Should a 'Sustainable Corporation' Account for*  
944 *Natural Capital?* SSRN Discussion Paper 12.
- 945 129. Kremen C, Merenlender AM. 2018 Landscapes that work for biodiversity and  
946 people. *Science* **362**, 304.
- 947 130. Filoso S et al. 2017 Impacts of forest restoration on water yield: a systematic  
948 review *PLOS One* 0183210.
- 949 131. Harvey, CA, Chacón, M, Donatti, CI, Garen, E, Hannah, L, Andrade, A, Bede,  
950 L, Brown, D, Calle, A & Chara, J. 2014 Climate-smart landscapes:  
951 opportunities and challenges for integrating adaptation and mitigation in tropical  
952 agriculture. *Conservation Letters* **7**, 77-90.
- 953 132. Davies C, Laforcezza R. 2019 Transitional path to the adoption of nature-based  
954 solutions. *Land Use Policy* **80**, 406-409.
- 955 133. Young, R, Zanders, J, Lieberknecht, K & Fassman-Beck, E. 2014 A  
956 comprehensive typology for mainstreaming urban green infrastructure. *Journal*  
957 *of hydrology* **519**, 2571-2583.
- 958 134. Goldstein, A, Turner, WR, Gladstone, J & Hole, DG. 2018 The private sector's  
959 climate change risk and adaptation blind spots. *Nature Climate Change* **1**.
- 960 135. Brink, E, Aalders, T, Ádám, D, Feller, R, Henselek, Y, Hoffmann, A, Ibe, K,  
961 Matthey-Doret, A, Meyer, M & Negrut, NL. 2016 Cascades of green: a review  
962 of ecosystem-based adaptation in urban areas. *Global Environmental Change*  
963 **36**, 111-123.
- 964 136. Lange, W, Pirzer, C, Dünow, L & Schelchen, A. 2016 Risk perception for  
965 participatory ecosystem-based adaptation to climate change in the Mata  
966 Atlântica of Rio de Janeiro State, Brazil. In *Ecosystem-based disaster risk*  
967 *reduction and adaptation in practice* (pp. 483-506, Springer.
- 968 137. Finewood MH. 2016 Green Infrastructure, Grey Epistemologies, and the Urban  
969 Political Ecology of Pittsburgh's Water Governance. *Antipode* **48**, 1000– 1021.
- 970 138. Dietz R, O'Neill D. 2013 *Enough is enough: Building a sustainable economy in*  
971 *a world of finite resources*. Routledge.
- 972 139. Raworth, K. 2017 *Doughnut Economics: Seven Ways to Think Like a 21st-*  
973 *Century Economist*. 384pp. Random House, UK.
- 974 140. NASEM (National Academies of Sciences, Engineering, and Medicine) 2019.  
975 *Negative Emissions Technologies and Reliable Sequestration: A Research*  
976 *Agenda*. Washington, DC: The National Academies Press.  
977 <https://doi.org/10.17226/25259>.
- 978 141. Buttle JM. 2011 Streamflow response to headwater reforestation in the  
979 Ganaraska River basin, southern Ontario, Canada. *Hydrological Processes* **25**,  
980 3030-3041.
- 981 142. Kelly CN et al. 2016 Streamflow response to increasing precipitation extremes  
982 altered by forest management. *Geophysical Research Letters* **43**, 3727-3736.
- 983 143. Vermaat JE. et al. 2016 Assessing the societal benefits of river restoration  
984 using the ecosystem services approach. *Hydrobiologia* **769**, 121-135.

144. Liqueste C. et al. 2016 Integrated valuation of a nature-based solution for water pollution control. Highlighting hidden benefits. *Ecosystem Services* **22**, 392-401.
145. Paul C. et al. Agroforestry versus farm mosaic systems—Comparing land-use efficiency, economic returns and risks under climate change effects. *Science of the Total Environment* **587**, 22-35 (2017).
146. Reid H, Faulkner L, Weiser A. 2013 The role of community-based natural resource management in climate change adaptation in Ethiopia: Assessing participatory initiatives with pastoral communities. In *IIED Climate Change Working Paper No. 6* (eds. S. Fisher & H. Reid). London, Climate Change Working Paper.
147. Ahammad R, Nandy P, Husnain P. 2013 Unlocking ecosystem based adaptation opportunities in coastal Bangladesh. *Journal of coastal conservation* **17**, 833-840.
148. Global Commission on Adaptation (2019) Adapt now: a global call for leadership on climate resilience. <https://gca.org/global-commission-on-adaptation/report>
149. Griscom, B. et al. (in press) National potential for natural climate solutions in the tropics. *Phil. Trans. R. Soc. B*
150. Pete Smith, Justin Adams, David J. Beerling, Tim Beringer, Katherine V. Calvin, Sabine Fuss, Bronson Griscom, Nikolas Hagemann, Claudia Kammann, Florian Kraxner, Jan C. Minx, Alexander Popp, Phil Renforth, Jose Luis Vicente Vicente, Saskia Keesstra (2019) Land-Management Options for Greenhouse Gas Removal and Their Impacts on Ecosystem Services and the Sustainable Development Goals. *Annual Review of Environment and Resources* **44**, 255-286.
151. Chowdhury, M. S. N., Walles, B., Sharifuzzaman, S. M., Hossain, M. S., Ysebaert, T., & Smaal, A. C. (2019). Oyster breakwater reefs promote adjacent mudflat stability and salt marsh growth in a monsoon dominated subtropical coast. *Scientific reports* **9** 8549.