

1 *Article*

2 **Benefits and Trade-offs of Dairy System Changes Aimed at**  
3 **Reducing Nitrate Leaching**

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27 **Simple Summary:** Reducing inputs of nitrogen fertiliser and imported feed,  
28 with an associated reduction in stocking rate on pastoral dairy farms resulted in  
29 less nitrate leaching. A co-benefit was a reduction in greenhouse gas emissions.  
30 The exception was the implementation of a wintering barn where nitrate  
31 leaching was reduced, but greenhouse gas emissions remained unchanged due to  
32 greater manure storage and handling. Emission reductions in the lower-input  
33 systems came at an average loss of profit of approximately NZ\$100 per tonne  
34 CO<sub>2</sub>-equivalent.

35 **Abstract:** Between 2011 and 2016 small-scale farm trials were run across three  
36 dairy regions of New Zealand (Waikato, Canterbury, Otago) to compare the  
37 performance of typical regional farm systems with farm systems implementing  
38 a combination of mitigation options most suitable to the region. The trials ran  
39 for at least three consecutive years with detailed recording of milk production  
40 and input costs. Nitrate leaching per hectare of the milking platform (where  
41 lactating cows are kept) was estimated using either measurements (suction  
42 cups), models, or soil mineral nitrogen measurements. Post-trial, detailed farm  
43 information was used in the New Zealand greenhouse gas inventory  
44 methodology to calculate the emissions from all sources; dairy platform, dairy  
45 support land used for wintering non-lactating cows (where applicable) and  
46 replacement stock, and imported supplements. Nitrate leaching was also  
47 estimated for the support land and growing of supplements imported from off-  
48 farm using the same methods as for the platform. Operating profit  
49 (NZ\$/ha/year), nitrate leaching (kg N/ha/year), and greenhouse gas emissions (t  
50 CO<sub>2</sub>-e/ha/year) were all expressed per hectare of milking platform to enable  
51 comparisons across regions. Nitrate leaching mitigations adopted in lower-input  
52 (less imported feed and N fertiliser) farm systems reduced leaching by 22 to 30  
53 percent, and greenhouse gas emissions by between nine and 24 percent. The  
54 exception was the wintering barn system in Otago where nitrate leaching was  
55 reduced by 45 percent but greenhouse gas emissions were unchanged due to  
56 greater manure storage and handling. Important drivers of a lower  
57 environmental footprint are reducing nitrogen fertiliser and imported feed. Their  
58 effect is to reduce nitrogen surplus and feed flow through the herd and drive  
59 down both greenhouse gas emissions and nitrate leaching. Emission reductions

60 in the lower-input systems of Waikato and Canterbury came at an average loss  
61 of profit of approximately NZ\$100/t CO<sub>2</sub>-e (three to five percent of industry  
62 average profit per hectare).

63 **Keywords:** greenhouse gases; operating profit; mitigations; carbon price;  
64 environmental footprint

## 65 1. Introduction

66  
67 An important challenge facing global dairy industries is to develop farm  
68 systems that can maintain or increase production and profitability, while reducing  
69 environmental impacts, including on water and climate [1-3]. Water quality issues  
70 have been at the forefront of the environmental concerns in New Zealand (NZ) for a  
71 number of decades. More recently, the climate impacts from greenhouse gas (GHG)  
72 emissions from agriculture have gained increasing attention. Responding to the effects  
73 of anthropogenic GHG emissions on climate, NZ aims to transition to a low-emission  
74 economy to help meet the Paris Agreement target of limiting temperature increases to  
75 well below 2 °C above pre-industrial levels [4]. New Zealand's commitments under  
76 the Paris Agreement is to reduce GHG emissions by 30% below 2005 levels, by 2030  
77 [5]. In 2017, agriculture was the single biggest contributor (48%) to total GHG  
78 emissions in New Zealand, with the dairy sector contributing almost half (47%) of  
79 these emissions [6]. The largest sources of agricultural emissions are enteric methane  
80 (CH<sub>4</sub>) from ruminant animals and nitrous oxide (N<sub>2</sub>O) emissions from soils.

81 Although water quality was the focus of much of the environmental research in  
82 NZ in recent decades, many of the management practices to improve water quality  
83 were also expected to result in reductions in GHG emissions [7]. For example, the  
84 Pastoral 21 (P21) research programme [8] included farmlet (small farm) studies in  
85 three regions throughout NZ (Waikato, Canterbury, and Otago) that compared  
86 systems typical of that region ('Current') with 'Improved' systems, in which strategic  
87 changes were made to the Current system. The five key changes used to design the  
88 P21 Improved farmlets were using lower nitrogen (N) fertiliser inputs; fewer, but  
89 higher producing cows; lower herd replacement rate; greater use of high-energy/low-  
90 N feeds; and using off-paddock facilities to reduce the time cows spend on pasture (or  
91 on forage crops). In all regions, the Improved systems could reduce nutrient losses to  
92 water [8-11] while GHG emissions were estimated to be reduced in most of the

93 Improved Systems [11]. The total annual GHG emissions were strongly related to  
94 total feed eaten, and the lower feed supplies and associated lower stocking rates of the  
95 Improved systems were the key drivers of lower total GHG emissions in all three  
96 regions [11]. These findings align with international studies where the general trend  
97 was that increased farming intensity within a system (more input and more animals)  
98 may decrease the GHG intensity of milk (kg emissions/kg milk), but absolute  
99 emissions (kg emissions/ha) will increase [12-14]. Few studies have considered the  
100 wider issues of emissions to both air and water, impacts of mitigations on farm  
101 profitability, and the potential trade-offs from achieving these often-conflicting goals.  
102 The P21 farmlet studies utilised realistic grazing systems, and determined both N  
103 leaching and GHG emissions as well as systems' profitability. The aim of this study  
104 was, therefore, to analyse the results from these farmlet studies to assess the impact to  
105 environmental, production and economic outcomes of strategies applied to reduce N  
106 leaching.

107

## 108 **2. Materials and Methods**

109

### 110 *2.1. Regional farmlet trials*

111

112 The P21 programme ran small-scale farmlet studies (farmlets ranging from 13  
113 to 39 ha) that included 'Current' and 'Improved' systems in three regions in New  
114 Zealand (Waikato, Canterbury, and Otago; [10]). The 'Current' farmlets were  
115 designed to represent a system typical of the region in which it was located. The  
116 'Improved' farmlets were designed by applying a suite of strategic changes to the  
117 Current for each region to reduce N leaching (Table 1). Farm, animal and feed  
118 management practices for the farmlets in each region are described by Clark et al. [8]  
119 for Waikato, Chapman et al. [9] for Canterbury and Van der Weerden et al. [11] for  
120 Otago. A summary of the main features is given in Table 2. These farmlets were  
121 monitored for production, profitability and N leaching over the following years: 2011  
122 to 2016 - Waikato: 2011 to 2014 - Canterbury; 2012 to 2015 - Otago.

**Table 1.** System changes applied to typical regional dairy farms in developing nitrate leaching mitigated farms as part of the Pastoral 21 farmlet trials in the Waikato, Canterbury and Otago, New Zealand [11].

Region	Fewer, higher producing cows	Reduced N fertiliser inputs	Reduced herd replacement rate	Greater use of high energy/low N feeds	Off-paddock facilities
Waikato	✓	✓	✓	✓	✓
Canterbury	✓	✓		✓	
Otago		✓	✓		✓

123

124 **Table 2.** Key management features of control (Current) and improved systems

125 (Improved) in the Waikato, Canterbury and Otago; opt = optimised feeding; barn =

126 cows housed during winter and some wet days in autumn and spring. From [11].

127

Systems Features	Waikato		Canterbury		Otago		
	Current	Improved	Current	Improved	Current	Improved-opt	Improved-barn
Stocking rate (cows/ha)	3.2	2.6	5.0	3.5	3.0	2.8	2.8
Cow genetic merit (\$BW#)	90	170	133	140	109	105	104
N fertiliser (kg N/ha/year)	137	52	311	158	109	42	73
Replacement rate (%)	22	18	23	23	23	18	18
High energy/low N feed	N/A	0.24 (Grain t DM/cow/year)	N/A	40% diverse pasture	N/A	N/A	N/A
Stand-off/housing	No	Yes	No	No	No	No	Yes
Winter feed	On platform	On platform	Fodder beet + Pasture silage	Kale + Oat silage	Kale	Kale	N/A
N fertiliser for winter forage (kg N/ha/year)	N/A	N/A	200	307	200	200	N/A

128 N/A: not applicable; # Breeding worth, \$ (May 2011)

129

130 *2.2. Measuring production, nitrate leaching and greenhouse gases*

131

132 Individual milk yields (kg milk/cow) were measured for all cows at each  
 133 milking. Milk component concentrations (MS - milksolids = fat + protein) of both  
 134 morning and afternoon composite milk samples for each cow were determined weekly  
 135 for the Waikato farmlets [8] and fortnightly for Canterbury and Otago farmlets [9].  
 136 Nitrate leaching from the Waikato farmlets was determined from measurements of  
 137 nitrate-N concentration in the soil solution below plant rooting depth (collected in  
 138 porous ceramic cup samplers at a vertical depth of 60 cm). These measurements were

139 used in conjunction with drainage volume (from on-site lysimeters) to estimate  
140 leaching losses from the soil [1]. Off-farm sources of N leaching were estimated for  
141 fertiliser use for producing pasture for replacement stock, N-excreta deposited by  
142 replacement stock, and N fertiliser used for growing imported supplements using the  
143 New Zealand Agricultural Inventory methodology (NZAI; [16]). For the Canterbury  
144 farmlets nitrate leaching for the milking platform plus winter crop areas was estimated  
145 using the Overseer<sup>®</sup> nutrient budgeting tool ([9]; Overseer version 6.2.2 was operated  
146 using the standard industry operating protocol [15]). Nitrogen loss risk for the Otago  
147 farmlets was derived as average values weighted for the respective areas (“blocks”)  
148 required for the milking platform, winter and summer forage crops (if needed), young  
149 stock rearing and supplement provision. Estimates of N loss risk were assigned to  
150 each of the relevant blocks that made up an individual farmlet. This type of whole-  
151 system assessment was based on a combination of directly measured values, proxy  
152 values and literature values [10].

153 Annual average GHG emissions for each system were estimated for all the  
154 monitored years using calculations based on the NZAI methodology [16], but  
155 included key farmlet-specific activity data from the P21 farmlet systems as well as  
156 farmlet-specific emission factor values determined from targeted regional experiments  
157 (see [11] for more detail).

### 158 159 *2.3. Measuring system profitability*

160  
161 Operating profit (OP) was determined using a calculator developed specifically  
162 for research farmlet trials [17]. This involved scaling the farmlets up to more  
163 representative farm sizes for each region (Waikato: 100 ha; Canterbury: 160 ha;  
164 Otago: 220 ha), as many farm costs are related to farm and herd size (e.g. labour).  
165 Where physical outputs and inputs were known, these were used in the calculation.  
166 Where inputs could not be determined separately for each farmlet, average values  
167 were used based on regional information from DairyBase ([18], a DairyNZ database  
168 of farm financial and physical parameters used for benchmarking) and Glassey et al.  
169 [19]. A simplified economics model was applied to the biophysical data, using mean  
170 values for economically important variables, including supplementary feed prices and  
171 fertiliser prices, and cost data from DairyBase [20] to estimate the profitability of the  
172 farmlets. For all profitability calculations actual milk prices for the monitored years

173 were used. Average milk prices for the monitored years in NZ\$/kg MS were Waikato  
 174 6.59, Canterbury 7.40, and Otago 7.16. The economic calculations included the cost  
 175 of rearing replacement stock off-farm [8]. For the Waikato Improved farmlet, the base  
 176 depreciation rate for capital invested in the off-paddock infrastructure was \$350/ha,  
 177 with an additional \$61/ha for maintenance of the infrastructure [8].

178

### 179 3. Results and Discussion

#### 180 3.1. Waikato

181 System changes in the Waikato Improved farmlet resulted in a reduction in N  
 182 leaching on the milking platform of 23 kg N/ha (equivalent to a 43% reduction) [1].  
 183 However, when leaching losses accrued by grazing replacement stock, growing  
 184 imported supplements, and spreading loafing pad solids off-platform were accounted  
 185 to the milking platform, the reduction in N leaching was 16 kg N/ha (26% reduction)  
 186 (Table 3). The collateral benefit of the leaching reduction was a reduction in GHG  
 187 emissions of 2.2 t carbon dioxide equivalents per hectare (CO<sub>2</sub>-e/ha; 16% reduction).  
 188 However, the trade-off for the reduced environmental footprint of the Waikato  
 189 Improved farm was lower production (47 kg MS/ha; 4%) and lower profitability of  
 190 \$280/ha (13%) averaged over five farming seasons (Table 3).

191

192 **Table 3.** Average performance (production, profit and environmental losses) of three  
 193 regional farm system trials. All metrics are presented as per hectare of the milking  
 194 platform. Numbers in brackets indicate the range for all farming seasons available. In  
 195 the Canterbury region wintering of non-lactating cows can be either on kale followed  
 196 by an oats catch crop (Kale), or fodder beet (FB). Greenhouse gas data from [11].

Region	Farm system	Milk production (kg MS/ha)	Operating profit (\$/ha)	Nitrogen leaching (kg N/ha)	Greenhouse gas (kg CO <sub>2</sub> -e/ha)
Waikato	Current	1200 (1151 to 1232)	2086 (-244 to 3873)	62 (43 to 75)	13610
Waikato	Improved	1153 (1093 to 1207)	1807 (-834 to 3652)	46 (37 to 57)	11405
Canterbury	Current	2242 (1834 to 2428)	3893 (3596 to 4440)	Kale: 114 FB: 75	20615
Canterbury	Improved	1700 (1452 to 1808)	3535 (3283 to 3885)	Kale: 80 FB: 53	15582
Otago	Current	964 (915 to 1040)	715 (-1428 to 3226)	29 (24 to 38)	11827
Otago	Improved-barn	949 (913 to 983)	20 (-1980 to 2473)	16 (10 to 22)	11461

Otago	Improved- opt	931 (899 to 969)	777 (-1192 to 3040)	22 (15 to 31)	10792
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198           The substantial reduction in profit compared to the relatively small reduction in  
199 production can be explained by standing cows off pasture in the Waikato Improved  
200 system. Although this mitigation has been confirmed as highly effective for N  
201 leaching [1,21,22], the trade-offs are the increase in methane emissions from manure  
202 collected in effluent ponds [23,11], and the large costs of the capital investment,  
203 depreciation and maintenance of these facilities [24,25]. The cost of the standing cows  
204 off pasture is reflected in other working expenses and overheads and resulted in a  
205 10c/kg MS higher cost of milk production (Table 4). Production losses in the  
206 Improved system were minimised by using high genetic merit cows achieving high  
207 per-cow production, another important driver of efficiency and therefore footprint  
208 mitigation [26,27], although this target was negated to some extent by an exceptional  
209 run of dry years when the desired days in milk for the Improved system could not be  
210 achieved [8].



211 **Table 4.** Average financial results of the Pastoral 21 regional farm trials. All results are expressed per hectare of the milking platform. Numbers  
212 in brackets indicate the range for all farming seasons available.

Region	Waikato	Waikato	Canterbury	Canterbury	Otago	Otago	Otago
Farm system	Current	Improved	Current	Improved	Current	Improved-barn	Improved-opt
Dairy gross farm revenue (\$/ha)	7713 (5260 to 9702)	7363 (4670 to 9352)	15305 (15081 to 15510)	11445 (11357 to 11656)	6671 (4748 to 9216)	6430 (4710 to 8652)	6349 (4463 to 8565)
Total feed expenses (\$/ha)	965 (804 to 1179)	923 (719 to 1163)	4324 (3831 to 4657)	2208 (1995 to 2422)	1729 (1572 to 1950)	1458 (1269 to 1618)	1539 (1480 to 1629)
Total stock expenses (\$/ha)	745 (720 to 771)	632 (614 to 644)	1379 (1369 to 1387)	972 (970 to 978)	645 (606 to 666)	624 (617 to 638)	609 (600 to 617)
Total labour expenses (\$/ha)	1079 (1079 to 1079)	1026 (1026 to 1026)	1554 (1554 to 1554)	1554 (1554 to 1554)	1043 (1030 to 1052)	1036 (1034 to 1041)	1034 (1034 to 1034)
Total other working expenses (\$/ha)	1858 (1798 to 1884)	1924 (1803 to 2140)	3409 (3353 to 3446)	2479 (2468 to 2502)	1523 (1478 to 1602)	1901 (1852 to 1961)	1381 (1361 to 1412)
Total overheads (\$/ha)	980 (979 to 981)	1051 (1050 to 1052)	746 (742 to 750)	697 (695 to 699)	1014 (1002 to 1035)	1391 (1369 to 1432)	1010 (982 to 1024)
Dairy operating expenses (\$/ha)	5628 (5495 to 5829)	5556 (5457 to 5700)	11412 (10926 to 11775)	7910 (7682 to 8113)	5955 (5701 to 6175)	6411 (6179 to 6690)	5572 (5525 to 5655)
Operating expenses (\$/kg MS)	4.7 (4.5 to 4.8)	4.8 (4.6 to 5)	4.7 (4.5 to 4.9)	4.4 (4.3 to 4.6)	6.2 (5.8 to 6.6)	6.8 (6.3 to 7)	6 (5.7 to 6.3)
Dairy operating profit (\$/ha)	2086 (-244 to 3873)	1807 (-834 to 3652)	3893 (3596 to 4440)	3535 (3283 to 3885)	715 (-1428 to 3226)	20 (-1980 to 2473)	777 (-1192 to 3040)

213

214           Given the cost of installing and maintaining a stand-off pad in the Waikato, it is  
215 worthwhile exploring the potential impact of the multiple system changes where the  
216 stand-off approach is excluded. On average, using a stand-off pad would contribute to  
217 *ca* 60% of the N leaching reduction while the ‘low input’ strategies, including higher-  
218 producing cows, would contribute *ca* 40% [1]. The average reduction in N leaching in  
219 the Improved system excluding a stand-off pad can therefore be estimated as 6 kg N/ha  
220 (40% of 16 kg N/ha reduction). By excluding the *ca* \$400/ha cost associated with  
221 standing-off, farm profitability in the Improved would be greater than for the Current  
222 system. Similarly, by avoiding the increase in net GHG emissions due to the stand-off  
223 approach [11], total GHG emissions will be further reduced. This suggests farmers in  
224 the Waikato could increase profitability while reducing losses to air and water by  
225 implementing a subset of the ‘stacked’ mitigation strategies outlined in Table 1.

226           The cost of GHG mitigation in the Waikato trial amounted to c. \$127/t CO<sub>2</sub>-e at  
227 an average milk price of \$6.59/kg MS, which can be compared with the cost of \$103  
228 and \$114/t estimated by Adler et al. [25] for medium input (10-20% imported feed) and  
229 high input (20-40% imported feed) Waikato systems, respectively, using a milk price of  
230 \$5.50/kg MS. In another study focussing on three Waikato dairy systems (low, medium,  
231 high input) Adler et al. [28] estimated the marginal abatement cost for GHG of \$96/t  
232 CO<sub>2</sub>-e with a \$5.50 milk price. In a modelling study of a Waikato dairy system,  
233 Smeaton et al. [27] found a weak correlation ( $R^2 = 0.43$ ) between GHG emissions and  
234 profitability with an average abatement cost of c. \$250/t CO<sub>2</sub>-e. Carbon prices are rising  
235 [29], and about half of the global GHG emissions are now covered by carbon pricing  
236 initiatives priced at over US\$10/tCO<sub>2</sub>-e (~ NZ\$15), compared with one-quarter of  
237 emissions covered in 2017. It is clear that carbon prices will have to increase  
238 substantially more before it is economically worthwhile for dairy farmers to adjust their  
239 system instead of offsetting emissions by buying carbon credits (note: agriculture is  
240 currently not included in New Zealand Emissions Trading Scheme). However, the  
241 situation may change in the not too distant future if we consider that the High-Level  
242 Commission on Carbon Prices identified the carbon price to be in the range of US\$40–  
243 80/tCO<sub>2</sub>-e in 2020 and US\$50–100/tCO<sub>2</sub>-e by 2030, which will make it consistent with  
244 achieving the temperature goal of the Paris Agreement [29]. However, to shift  
245 investment at scale, carbon pricing coverage must expand, and prices must be stronger.  
246 Most initiatives saw increases in carbon prices in 2018 compared to price levels in

247 2017. But despite these, most initiatives in 2019 are still below the US\$40-\$80/tCO<sub>2</sub>-e  
248 needed in 2020 [29].

249 Compared with commercial farms in the Waikato region, the Current farm  
250 performed well above average in terms of production and profit (Tables 3 and 5), and it  
251 was clearly not an average or typical farm. The reasons could be the environmental  
252 conditions and/or measurement and managerial intensity applied at the research site. In  
253 the context of “average” commercial farms, the Waikato Improved system shows a lot  
254 of potential by maintaining production, trading a relatively small amount of profit, and  
255 leaving a modest environmental footprint. However, it should be considered that the  
256 gains made on the trial farms were made by running the farms with best-management  
257 practices, smaller reductions at higher profit trade-offs may be expected from most  
258 commercial farms.

259

260 **Table 5.** Average performances of typical commercial dairy farms in the same regions  
261 as the P21 farmlet trials. Extracted for the relevant years from DairyNZ Economic  
262 Survey data (<https://www.dairynz.co.nz/publications/dairy-industry/>).

	Waikato 2011-2016	Canterbury 2011-14	Otago 2012-15
Number of herds	56	23	28
Peak cows	343	751	587
Effective hectares	120	222	209
Stocking rate (cows/ha)	2.8	3.4	2.8
Milk production (kg MS/ha)	1025	1413	1120
Milk price (\$/kg MS)	6.59	7.40	7.16
Operating expenses (\$/kg MS)	4.80	4.96	4.95
Operating profit (\$/ha)	1949	3438	2505

263

264

### 265 3.2. Canterbury

266 In the Canterbury region the Improved system reduced N leaching from the  
267 milking platform by 14 kg N/ha (30% reduction) compared with the Current system [9].  
268 When including N leaching losses from the winter crop, the reductions in the Improved  
269 system were 22 kg N/ha (29%) with fodder beet, and 34 kg N/ha (30%) with kale (Table  
270 3). Leaching from both these winter crops were generally high (150-200 kg N/ha crop),  
271 but the larger area required for the lower-yielding kale crop resulted in higher N  
272 leaching losses per hectare of platform area, compared with fodder beet. The co-benefit  
273 of the lower leaching in the Improved system was a reduction in GHG emissions of 5 t

274 CO<sub>2</sub>-e/ha (24%) compared with the Current. However, trade-offs of the Improved  
275 system were reductions in production (*minus* 542 kg MS/ha, 24%) and profit (*minus*  
276 \$358/ha, 9%). The cost of GHG abatement was \$71/t CO<sub>2</sub>-e, which is much lower than  
277 for the Waikato, but still substantially higher than the current carbon price.

278 The operating profit for both Canterbury systems were higher than the average of  
279 surrounding commercial farms, mainly driven by higher production (Tables 3 and 5).  
280 Operating expenses for the trial farms (\$4.7 and \$4.4/kg MS for Current and Improved,  
281 respectively) were also lower than for the commercial farms (Table 4). The evidence  
282 from the Improved farmlet demonstrates that there are system options that Canterbury  
283 farmers could adopt to reduce their environmental footprint. Already, the Lincoln  
284 University Dairy Farm has successfully adopted the P21 Improved system at a whole  
285 farm scale [30]. There will be trade-offs compared to best-practice current systems but,  
286 with efficiency gains, both production and profit can be above the current averages for  
287 the region. Such efficiency gains will require improved management ability and  
288 processes on farm. This is important information for building farmer confidence in the  
289 face of regulatory change [9].

290

### 291 3.3. Otago

292 Two Improved systems were tested in Otago, one with duration-controlled  
293 grazing, where cows were housed in a barn for 12 hours/day on wet days in spring and  
294 autumn and 24 hours/day in winter from June to mid-August to reduce urinary N  
295 deposition onto wet soils (Improved-barn), and one attempting to optimise feed intake  
296 by changing calving date and type of home-grown feed (Improved-opt). Both Improved  
297 systems used less N fertiliser (Table 2). Leaching was reduced by 13 kg N/ha (45%) and  
298 7 kg N/ha (24%) in the Improved-barn and Improved-opt systems, respectively,  
299 compared with the Current system (Table 3). A collateral benefit was GHG reductions  
300 of 0.3 (3%) and 1.1 (9%) t CO<sub>2</sub>-e/ha from the barn and optimal-feeding systems,  
301 respectively. The small reduction in GHG emissions from the barn system was the  
302 result of an increase in the amount of manure that required active management with  
303 associated GHG emissions, which largely negated the gains made by reducing urinary N  
304 onto wet soils. Van der Weerden et al. [11] showed that off-paddock facilities can  
305 increase emissions per cow from manure management, with the magnitude of the  
306 increase depending on the extent of the facility's use. For the Otago situation, the use of

307 the barn for 24 h/day in winter and 12 h/day on wet days in autumn and spring  
308 corresponded to a 35% increase in GHG emissions per cow. For both Otago Improved  
309 systems trade-offs in production were small at -15 (2%) and -33 (3%) kg MS/ha for the  
310 barn and optimal-feeding systems. However, profitability of the barn system was  
311 significantly lower (-NZ\$700/ha). This was mainly due to extra depreciation on the  
312 capital required for the barn itself, the effluent spreader, and extra silage bunker space  
313 (Overheads, Table 4). Maintenance costs were also higher because of the need to deal  
314 with more captured effluent and the cost of replacing the woodchip bedding for the barn  
315 (other working expenses, Table 4). The average profit in the optimal-feeding system  
316 was moderately higher (NZ\$62/ha) compared with the Current system, mainly because  
317 of lower feed and fertiliser expenses (Table 4).

318 Compared with commercial farms in the Otago region (Table 5) the profitability  
319 of all systems was considerably lower (Table 3). The main contributor to the higher  
320 operating expenses/kg MS was the poor MS production/ha across all systems. Factors  
321 that contributed to low MS production/ha included the below average genetic merit of  
322 the herd, a third of the farm being a recent conversion from sheep farming without  
323 renovating the poor-quality sheep pastures and upgrading the water supply system,  
324 drainage issues on the low lying heavier soils and the geographical spread of the farm  
325 resulting in increased energy expenditure and lameness from long walks on undulating  
326 terrain. The complex management structure of the property meant the business was not  
327 as agile at responding to climatic challenges and market signals as commercial  
328 businesses in the region which impacted on the physical and financial performance of  
329 the farm.

330

### 331 *3.4. Insights across regions*

332 Greenhouse gas reductions from lower-input, lower-stocked systems in the  
333 Waikato and Canterbury regions came at an average loss of profit of approximately  
334 NZ\$100/t CO<sub>2</sub>-e. This mitigation cost needs to be viewed in the context of on-farm  
335 forestry that can achieve the largest emission reductions (3-96%), depending on the  
336 percentage of the land planted. However, this is an expensive option for dairy farms  
337 with an implied C cost in excess of NZ\$100-600/t CO<sub>2</sub>-e, mainly because of the large  
338 opportunity cost incurred when taking land out of dairy grazing. The most viable option  
339 for dairy farms would be forests planted only on marginal land and not for harvest,

340 which depend heavily on individual farm configurations and has a more limited  
341 mitigation potential of up to 10% of emissions [5].

342 Analysis of the Waikato Improved system without the loafing pad pointed to a  
343 profitable system that can achieve N leaching and GHG mitigations without requiring  
344 extra investment in infrastructure. The potential for environmental mitigation without  
345 infrastructure and without sacrificing profitability was further supported by the results  
346 from the Otago Improved-opt system. This is relevant to many farm systems that are  
347 starting from a low baseline where extra investment and/or lower profitability is simply  
348 out of the question. These systems can benefit from gradually improving the genetic  
349 merit of their herds over time.

350 The positive relationship between N leaching and GHG emissions observed in  
351 Waikato and Canterbury agrees with previous works [22,27,31], and confirms the  
352 potential positive by-product of N leaching regulation on GHG emissions. Two drivers  
353 of the lower environmental footprint of the Waikato and Canterbury Improved systems  
354 were lower N fertiliser use and lower stocking rate, which agree with the findings of  
355 several studies that these are key factors in pasture-based dairy systems determining the  
356 balance between production, profit and environmental footprint [25,26,27,28,32,33,34].

357

#### 358 **4. Conclusions**

359 Important drivers of a lower environmental footprint (GHG emissions and N  
360 leaching) are reducing nitrogen fertiliser and imported feed. This reduces nitrogen  
361 surplus and feed flow through the herd and drives down both GHG emissions and N  
362 leaching. Nitrate leaching mitigations in the P21 farmlet systems achieved leaching  
363 reductions of 24 to 30 percent. In addition, these lower-input (less imported feed and N  
364 fertiliser) systems also reduced GHG emissions by between 9 and 24 %. The exception  
365 was the Improved-barn system in Otago, where N leaching was reduced by 45 percent  
366 but GHG emissions were not reduced due to greater manure storage and handling.  
367 Greenhouse gas reductions in the lower input systems of Waikato and Canterbury came  
368 at an average loss of profit of approximately NZ\$100/t CO<sub>2</sub>-e (three to five percent of  
369 industry average profit per hectare). Economic impacts of Improved systems were  
370 highly regional specific and highlighted the need for future systems to perform better  
371 than current local systems, requiring strong management expertise, with consideration  
372 for investment in infrastructure. However, for system changes that do not include

373 infrastructure investment, profitability can increase while associated losses to air and  
374 water decrease.

375

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397

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399

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