

1 Article

2 Associations of Various Nighttime Noise Exposure 3 Indicators with Objective Sleep Efficiency and 4 Subjective Sleep Quality: A Field Study

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16 **Abstract:** It is unclear which noise exposure time window and noise characteristics during
17 nighttime are most detrimental for sleep quality in real life settings. We have conducted a field
18 study with 105 volunteers wearing a wrist actimeter to record their sleep during seven days,
19 together with concurrent outdoor noise measurements at their bedroom window. Actimetry
20 recorded sleep latency increased by 5.6 minutes (95% confidence interval: 1.6 to 9.6 minutes) per
21 10 dB(A) increase in noise exposure during the first hour after bedtime. Actimetry assessed sleep
22 efficiency was significantly reduced by 2-3 percent per 10 dB(A) increase in measured outdoor
23 noise ($L_{eq, 1h}$) for the last three hours of sleep. For subjectively reported sleepiness, noise exposure
24 during the last hour prior to wake up was most crucial with an increase in the sleepiness score of
25 0.31 units (95% CI: 0.08 to 0.54) per 10 dB(A) $L_{eq, 1h}$. Associations for estimated indoor noise were
26 not more pronounced than for outdoor noise. Considering noise events in addition to equivalent
27 sound pressure levels (L_{eq}) only marginally improved the statistical models. Our study provides
28 evidence that matching the nighttime noise exposure time window to the individual's diurnal
29 sleep-wake pattern results in a better estimate of detrimental nighttime noise effects on sleep. We
30 found that noise exposure at the beginning and the end of the sleep is most crucial for sleep
31 quality.

32 **Keywords:** sleep quality; road traffic noise; actimetry; indoor noise; noise measurements; noise
33 annoyance; noise sensitivity; time of day
34

35 1. Introduction

36 There is increasing epidemiological research demonstrating negative effects of transportation
37 noise exposure on various chronic diseases such as cardiovascular disease [1-4], metabolic
38 syndrome [5-9], depression [10-12], and cognitive functions [13-16]. Several mechanisms are
39 implicated in these negative noise effects like the activation of the hypothalamic-pituitary-adrenal
40 axis (HPA), which leads to increased cortisol and glucose levels as well as increased blood pressure,
41 with consequences for blood viscosity and blood coagulation [17]. Also, chronic sleep deprivation is
42 a stressor which contributes to allostatic load and has been found to be connected to all these
43 outcomes above. High allostatic load, characterized by repeated stress responses, affects the brain
44 regions involved in memory consolidation and affective processing with potential long term effects
45 on cognitive and mental health [18]. In large population studies, impaired sleep quantity and

46 quality has been associated with increased risk of developing coronary heart diseases [19]. There is
47 also solid evidence that short sleep duration, lack of slow wave sleep, and circadian
48 desynchronisation of sleep increases the sensitivity to food stimuli contributing to adverse
49 metabolic traits, in particular obesity and type 2 diabetes [20]. Further, sleep deprivation reduces
50 the motivation for physical activity and thus reduces the energy expenditure [21], which further
51 contributes to the risk for cardiometabolic syndromes.

52 Noise induced sleep disturbance as demonstrated in experimental human laboratory studies
53 [22-24], field trials [25] and observational epidemiological studies [26,27] are thus likely to be on the
54 pathway for detrimental effects on the cardiometabolic system and mental health. Strikingly,
55 evidence for noise effects on various sleep outcomes is only considered “very low” to “moderate”
56 in the recent systematic review of the World Health Organization Environmental Noise Guidelines
57 for the European Region [28]. Thus, many questions remain open.

58 Sound pressure level at the ear of the sleeper is the relevant entity for quantifying noise effects
59 on sleep. It is possible to quantify reactions of sleepers to noise in field experiments using contrived
60 exposure settings (i.e. reproducing noise in a controlled fashion with loudspeakers). This approach,
61 however, does not adequately consider long-term habituation to a noise source, which may be
62 relevant in a real life situation at home. Observational studies would be appropriate to account for
63 potential habituation effects, but are usually based on noise exposure modeling for the outdoor
64 façade thus rather imprecise regarding noise exposure at the ear.

65 Further, the average sound pressure level (L_{eq} and similar indicators) may not be the only
66 relevant factor; other noise exposure characteristics not captured in energy-based exposure
67 indicators may also be important. For instance, in experimental sleep studies on noise effects on
68 sleep, different effects have been observed for road, rail, and aircraft noise [22,23]. Thus, the effects
69 of noise on sleep might either be better predicted by the number of noise events [29], the maximum
70 sound pressure level [30], the sound pressure level slope [24,31] or by the order of events [31].
71 Confronted with the challenge of how to sum up and weight noise events we have developed an
72 acoustical metric, the intermittency ratio (IR), to characterize short-term temporal variations of
73 transportation noise exposure [32]. In a large cohort study, we found some evidence that IR may
74 have a modifying effect on the cardiovascular mortality risk [33].

75 Timing of noise exposure is also considered to be relevant for sleep effects. For instance, an
76 experimental sleep study observed that noise curfews at the end of the night are most beneficial for
77 sleep, because noise induced sleep disturbances at the beginning of the night were at least partly
78 compensated during the rest of the night [34].

79 In order to quantify relevant factors that affect sleep quality through noise exposure in a real
80 life situation, we conducted an observational field study with volunteers wearing wrist actimeters
81 to record their sleep-wake behavior during seven days with concurrent in- and outdoor noise
82 exposure measurements. The study explored i) the relevance of indoor noise compared to outdoor
83 noise, ii) the predictive contribution of IR to sleep effects in addition to equivalent continuous
84 sound pressure levels ($L_{eq,night}$), and iii) the effect of noise exposure at different times during the
85 night. We also tested potential effect modification by noise annoyance, noise sensitivity, and sex.

86 2. Materials and Methods

87 2.1. Study population and procedures

88 Study participants were recruited from a previous nationwide survey on transportation noise
89 annoyance and sleep disturbances among participants who expressed their willingness to be
90 contacted for further research [35]. Only adults with German language skills were included. People
91 with severe chronic disease or shift work were excluded. Calculated L_{den} for road traffic at the most
92 exposed façade had to be at least 50 dB to ensure sufficient outdoor noise to be detected inside. In
93 order to explore effects of road traffic noise exposure minimally contaminated by other noise
94 sources, modeled aircraft and railway noise had to be at least 10 dB lower than road traffic noise.

95 During recruitment we ensured no other relevant noise sources such as construction work or
96 tramways could disturb the sound level measurements.

97 After recruitment, study participants were visited at home between May and November 2016
98 and instructed about the procedures. All participants filled in a baseline questionnaire about
99 sociodemographic and other relevant characteristics such as noise annoyance (using the 11-point
100 ICBEN scale for road traffic noise in general [36]), noise sensitivity (using a 6-point Likert scale to
101 rate the statement "I am noise sensitive" from NoiSeQ) [37], and window opening habits during
102 night.

103 2.2. Sleep outcomes

104 Objective sleep behavior was evaluated using movement data collected by an actimeter device
105 (Daqtometer v2.4 by Daqtix GbR, Oetzen, Germany) continuously worn on the non-dominant wrist
106 during the seven day study period. Data were collected in 10-s to 60-s bins. Participants were
107 provided with diaries in which they logged sleep-relevant information two times during the day.
108 Prior to sleep, they gave information on consumption of coffee and alcohol, daytime naps, screen
109 time use, and medication intake during the preceding day. In the mornings, participants evaluated
110 the preceding sleep episode and noted bed and rise times, and whether or not they used an alarm
111 clock. Additionally, subjective sleep quality was rated on a verbally anchored visual analogue scale
112 from 0 (the worst sleep) to 100 (the best sleep). Subjective sleepiness was rated on a verbally
113 anchored Likert-type scale that ranged from 1 (extremely alert) to 9 (very sleepy–fighting sleep)
114 [38].

115 Actimetry data were analysed using the "Actiwatch Activity and Sleep Analysis" software
116 (Actiwatch software version 7, Cambridge Neurotechnology Ltd., Cambridge, United Kingdom).
117 Bed and rise times were entered manually based on actimetry records and sleep diary entries, while
118 sleep on- and offset were determined automatically by the algorithm to derive sleep efficiency
119 (proportion of actual sleep time per time in bed), sleep duration, sleep latency (time between
120 bedtime and sleep onset), and moving time (minutes moving per assumed sleep period as time
121 between sleep on- and offset).

122 2.3. Noise measurements and modelling

123 Noise exposure assessment included a seven-day outdoor measurement, a controlled short-
124 term measurement to derive the sound insulation of each bedroom for different window positions
125 and noise exposure modeling to identify eligible households. We modelled road traffic noise
126 exposure calculation at the most exposed façade of each study participant using the sonROAD
127 emission model [39] and the sound propagation model StL-86 as described in detail in Karipidis et
128 al, 2011 [40].

129 Outdoor noise at the study participant's bedroom window was measured with sound level
130 meters type Noise-Sentry RT, a class II measurement device with a measurement uncertainty of
131 about 1 dB(A) [41]. The sound level meters were flush mounted to the outer face of the closed
132 window and logged A-weighted 1-second- L_{eq} 's. To obtain free field estimates, a correction of -6 dB
133 was applied to the raw data. After a measurement period of seven days, the participants removed
134 the sound level meters and sent them back. From the measurement files $L_{eq,night}$ and IR_{night} (23:00-
135 7:00) were computed for each night and participant. Further, L_{eq} and IR for *a priori* specified time
136 windows during the night were derived: 19:00-23:00, 23:00-1:00, 1:00-5:00, 5:00-6:00, and 6:00-7:00.
137 We also calculated, for each individual, $L_{eq,1h}$ exposure for the first four hours of sleep as well as the
138 last four hours prior to wake up taking into account their individual sleep pattern on the
139 corresponding night.

140 Sound transmission of outdoor noise into the bedroom was measured during the home visit.
141 Outdoor and indoor (at the position of the pillow) levels were measured in parallel for three
142 window positions (open, tilted, closed) during three minutes in one-third octave bands from 50 Hz
143 to 10 kHz with a class I sound level meter (type NTI XL2, NTi Audio AG, Schaan, Liechtenstein)
144 and a free field microphone. Temporal resolution was set to one second. The indoor measurements

145 were compared with the concurrently conducted outdoor measurements and an algorithm was
 146 developed to derive the A-weighted attenuation factor in dB based on the correlation and the offset
 147 of these parallel measurements [42]. Finally, estimated indoor levels for analysis were obtained for
 148 each study participant by subtracting the individual attenuation factor from the outdoor
 149 measurements taking into account the preferred window position for the season when the data
 150 were collected, as indicated in the baseline questionnaire.

151 2.4. Data analysis

152 Primary endpoints were actimeter-derived sleep efficiency, sleep latency, sleep duration, and
 153 moving time. Secondary endpoints were subjective sleep quality and sleepiness as rated each
 154 morning in the sleep diary.

155 Association between sleep outcomes and noise exposure was analyzed using mixed regression
 156 models with auto-correlated residuals lag=1 and robust standard errors. To account for repeated
 157 nights from the same individual and multiple study participation per household a three level
 158 random intercept model was applied. Models were adjusted for confounding factors as indicated in
 159 the footnote of corresponding result tables (Table 3 and 4). Effect modification by noise annoyance
 160 and noise sensitivity was investigated by interaction analysis on dichotomized variables.

161 3. Results

162 For the study, we included 107 individuals from 96 households resulting in 720 nights with
 163 complete noise exposure data and recorded data on at least one sleep outcome (actimeter-derived
 164 or self-reported). Data from two individuals (14 nights) were excluded because their sleep was
 165 affected by their children. An additional 10 nights were excluded from the dataset because of
 166 sleeping out of home, two nights due to acute respiratory infection, and nine nights due to a
 167 recorded sleep duration of less than four hours. This left 694 nights from 105 individuals from 94
 168 households for further analyses, although number of observations were somewhat smaller for
 169 specific outcomes because for two individuals only actimetry data was available, and for two
 170 individuals only the subjective sleep quality data.

171 Mean age of the study participants was 52.1 years (SD=14.4 years, age range: 23 to 78 years).
 172 Fifty-three participants (51%) were female and 52 male. Twenty-nine individuals (28%) had a
 173 University degree, 34 (32%) a higher education, 39 (37%) an apprenticeship, and 3 (3%) compulsory
 174 education. Median noise annoyance was 6 with 23% being classified as highly annoyed (score ≥ 8).
 175 Forty-one subjects (39%) tended to agree (score 4-6) to the statement "I am noise sensitive".

176 Average sleep efficiency as recorded by actimetry was 88% and average sleep duration 7.0 h
 177 (Table 1). Subjective sleep quality scores ranged from 4 to 100 with a mean of 65. Average subjective
 178 sleepiness score was 4.1. Sleep efficiency was strongly negatively correlated with sleep latency (-
 179 0.79) and subjective sleep quality was negatively correlated with subjective sleepiness (-0.49).
 180 Correlations of all other outcomes were low (see Supplementary Table 1).

181 **Table 1.** Overview of the sleep outcomes (ACT=actimetry; SR=self-reported).

Outcome	N	Mean	SD	Min	Max
ACT Sleep efficiency [%]	634	88.4	7.9	46.3	98.2
ACT Sleep latency [min]	634	29.4	33	2	303
ACT Sleep duration [h]	634	7.0	1.2	4.1	11.7
ACT Moving time [%]	634	7.2	4.7	0.8	35.9
SR Sleep quality [0-100]	639	65	20	4	100
SR Sleepiness [1-9]	633	4.1	1.8	1	9

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183 Table 2 shows the summary estimates for the various noise exposure metrics. Mean measured
 184 nighttime noise level ($L_{eq,night}$) outside the window of the study participants' bedroom was 47.0
 185 dB(A) with a maximum of 62.7 dB(A). Estimated mean indoor nighttime noise level at the pillow of
 186 the study participants was 30.2 dB(A) (maximum: 55.3 dB(A)). Measured exposure was about 10
 187 dB(A) lower during 01:00 to 05:00 compared to the beginning and end of the night. Correlation
 188 between estimated indoor and measured outdoor nighttime exposure was 0.46 reflecting variation
 189 in noise attenuation between study participants. Correlations were ≥ 0.63 for measured L_{eq} between
 190 different time periods of the night (see Supplementary Table S3).

191 **Table 2.** Overview of the noise exposure data in dB(A) (for IR, see Supplementary Table S2).

Outcome	N	Mean	SD	Min	Max
Indoor $L_{eq,night}$	685	30.2	7.6	20.0 ¹	55.3
Outdoor $L_{eq,night}$	685	47.0	6.9	29.6	62.7
Outdoor $L_{eq,19-23}$	685	51.2	6.6	33.2	68.6
Outdoor $L_{eq,23-01}$	685	46.1	7.4	29.4	62.9
Outdoor $L_{eq,01-05}$	685	41.8	7.5	27.8	62.7
Outdoor $L_{eq,05-06}$	685	46.3	8.1	28.8	64.7
Outdoor $L_{eq,06-07}$	685	50.6	8.1	29.5	70.3

192 ¹ all estimated indoor values <20 dB(A) have been replaced by 20 dB(A)

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194 Measured nighttime noise ($L_{eq,night}$) tended to be negatively associated with sleep efficiency and
 195 positively associated with sleep latency, although not statistically significant (Table 3). For instance
 196 sleep efficiency decreased by 1.11% (95%-CI: -2.44% to 0.21%) and sleep latency increased by 5.67
 197 min (95% -CI: -1.00 to 12.34) per 10 dB(A) increase in $L_{eq,night}$. No indications of an association
 198 ($p \geq 0.27$) were seen for the other sleep outcomes including subjective sleep quality and sleepiness.

199 **Table 3.** Adjusted association between sleep outcomes and measured *outdoor* nighttime noise
 200 ($L_{eq,night}$) per 10 dB(A) increase; ACT=Actimetry; SR=Self reported,

Outcome	N	Coefficient	Confidence interval	p-value
ACT Sleep efficiency [%]	634	-1.11	-2.44 to 0.21	0.10
ACT Sleep latency [min]	634	5.67	-1.00 to 12.34	0.10
ACT Sleep duration [h]	634	0.01	-0.17 to 0.19	0.94
ACT Moving time [%]	634	-0.41	-1.13 to 0.31	0.27
SR Sleep quality [0-100]	639	-1.09	-4.96 to 2.78	0.58
SR Sleepiness [1-9]	633	0.02	-0.29 to 0.32	0.91

201 ¹ adjusted for age, sex, education, evening caffeine intake, evening alcohol consumption, evening screen time,
 202 day of the week, season and whether woken up by an alarm clock

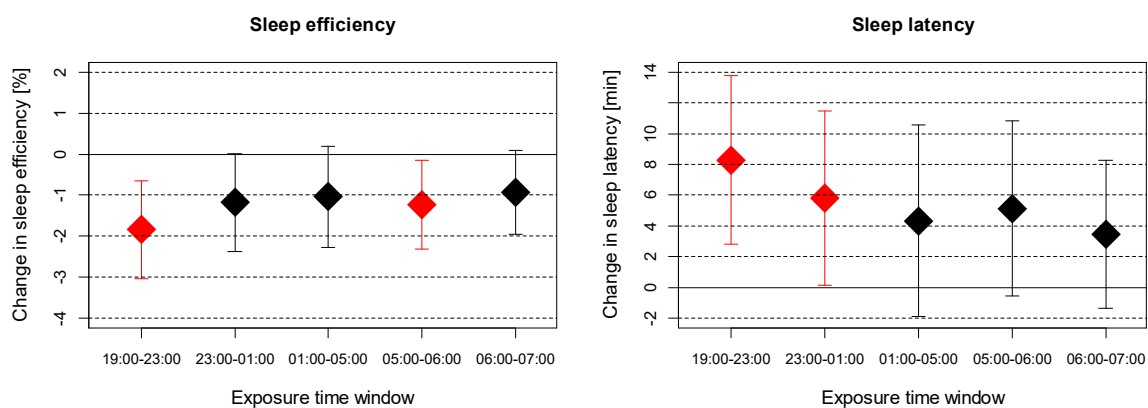
203 Using estimated indoor noise instead of measured outdoor noise yielded similar results in
 204 terms of regression coefficients, but none of the associations was even close to significance (Table 4).
 205 Similarly, we did not obtain any indications that IR was associated with any of the objectively
 206 recorded or subjectively reported sleep outcomes when also considering $L_{eq,night}$ in the same model
 207 (see Supplementary Table S4). A model with quartiles of IR_{night} instead of a continuous variable
 208 suggests a non-linear association for sleep efficiency and latency. For the second quartile of IR_{night}
 209 (50-63%), sleep efficiency was reduced by 1.24% (95% confidence interval: -2.73 to 0.25) and sleep
 210 latency was increased by 3.48 minutes (95% CI: -2.54 to 9.50) compared to the first quartile (4-50%).

211 **Table 4.** Adjusted association between sleep outcomes and estimated *indoor* nighttime noise ($L_{eq,night}$)
 212 per 10 dB(A) increase; ACT=Actimetry; SR=Self reported,

Outcome	N	Coefficient	Confidence interval	p-value
ACT Sleep efficiency [%]	634	-1.06	-2.86 to 0.74	0.25
ACT Sleep latency [min]	634	4.39	-5.54 to 14.32	0.39
ACT Sleep duration [h]	634	-0.06	-0.21 to 0.10	0.48
ACT Moving time [%]	634	-0.24	-0.90 to 0.42	0.47
SR Sleep quality [0-100]	639	0.21	-3.46 to 3.88	0.91
SR sleepiness [1-9]	633	-0.01	-0.28 to 0.26	0.95

213 ¹ adjusted for age, sex, education, evening caffeine intake, evening alcohol consumption, evening screen time,
 214 day of the week, season and whether woken up by an alarm clock

215 Since most indications for noise effects were found for sleep efficiency and sleep latency in
 216 relation to outdoor noise, we investigated the effect of noise exposure in potential critical time
 217 windows for these two outcomes in more detail (Figure 1). Noise exposure in the evening (19:00-
 218 23:00) and in the early morning hours (05:00-06:00) was significantly associated with sleep
 219 efficiency, whereas the other time windows reached only borderline significance. For sleep latency,
 220 noise exposure until 01:00 was most relevant.

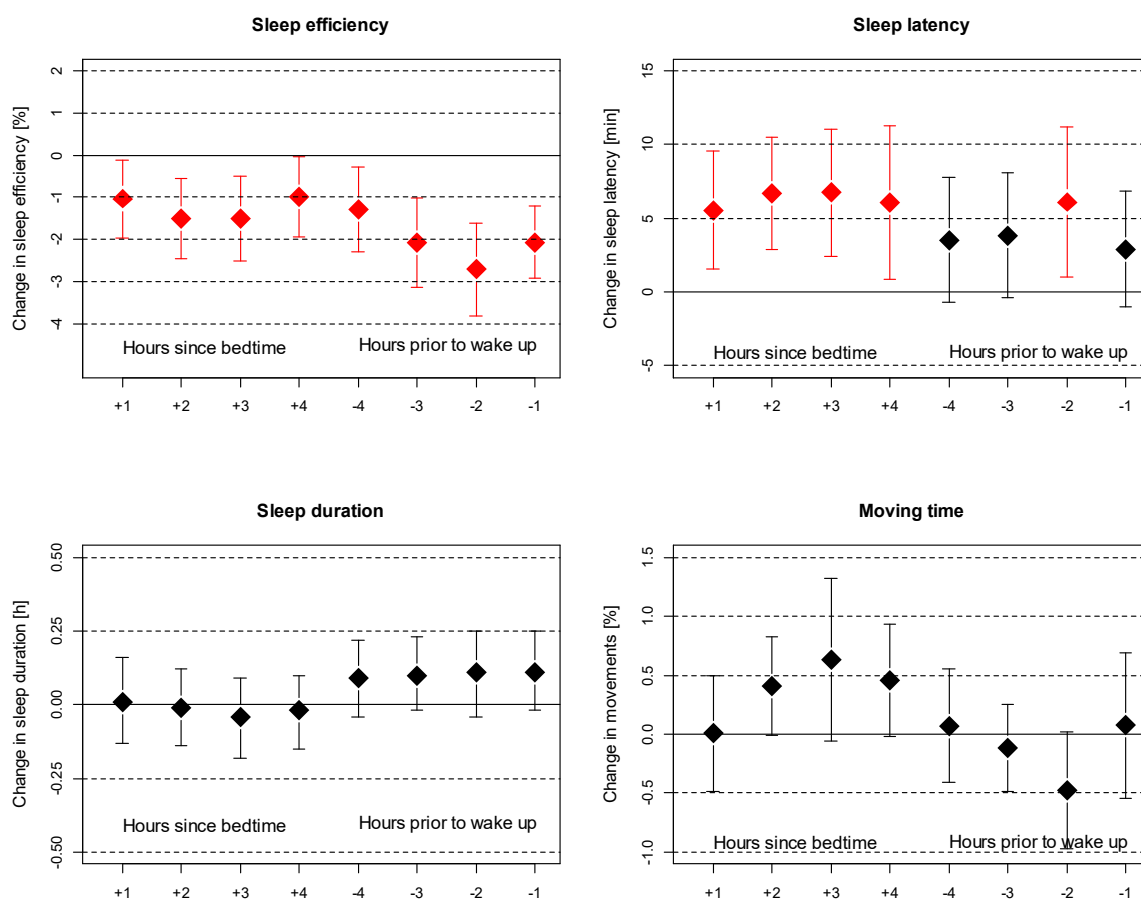


221 **Figure 1.** Association between measured outdoor noise exposure (L_{eq}) at different time windows
 222 during night and changes in all-night sleep efficiency and sleep latency per 10 dB(A) increase in
 223 noise exposure. Significant associations are highlighted in red. Analyses adjusted for age, sex,
 224 education, evening caffeine intake, evening alcohol consumption, evening screen time, day of the
 225 week, season, and whether woken up by an alarm clock.
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227 Exposure within fixed time periods (as shown above) may not match the individual time
 228 period a person is asleep or in bed. For instance, for somebody rising before 06:00, noise exposure
 229 between 06:00 and 07:00 is not relevant regarding sleep disturbance. We have thus calculated
 230 hourly noise exposure levels that matched the individual sleeping pattern of each night for the first
 231 four and the last four hours of sleep. The corresponding analyses with all actimetry-derived sleep
 232 outcomes are shown in Figure 2. This individualized noise exposure provides stronger associations
 233 between measured noise exposure and sleep efficiency. Sleep latency increased by five to seven
 234 minutes per 10 dB(A) increase in outdoor noise exposure during the first four hours after bedtime.
 235 For sleep efficiency, noise exposure during the last three hours prior to wake up was most critical
 236 with reductions of 2-3% per 10 dB(A) increase in measured outdoor noise exposure ($L_{eq,1h}$). For sleep
 237 duration and moving time, no significant associations were found; although for the latter, noise
 238 exposure two to four hours after bedtime tended to increase moving time during sleep and noise
 239 exposure two hours before wake-up tended to be negatively associated with moving time.

240 Subjective sleepiness in the morning was significantly associated with noise exposure in the
 241 last hour of sleep, whereas no associations were observed for noise exposure at the beginning of
 242 sleep (Figure 3). For subjective sleep quality, none of the noise exposure time windows were
 243 significantly associated, but noise exposure in the middle of the sleeping period (± 4 hours from
 244 bedtime and wake up) and at the end of the sleeping period showed the strongest trends for a
 245 relation. For IR matched to individualized sleep patterns, no significant effects on actimetry-derived
 246 and subjective sleep outcomes were observed.

247 Noise annoyance, noise sensitivity and sex were not significant effect modifiers for any of the
 248 outcomes and noise exposure time windows. A slight non-significant trend was seen for a stronger
 249 association of nighttime noise ($L_{eq,night}$) with sleep efficiency for males ($p=0.11$), people with a high
 250 (\geq median) annoyance score ($p=0.23$), and people not reporting to be noise sensitive ($p=0.25$).



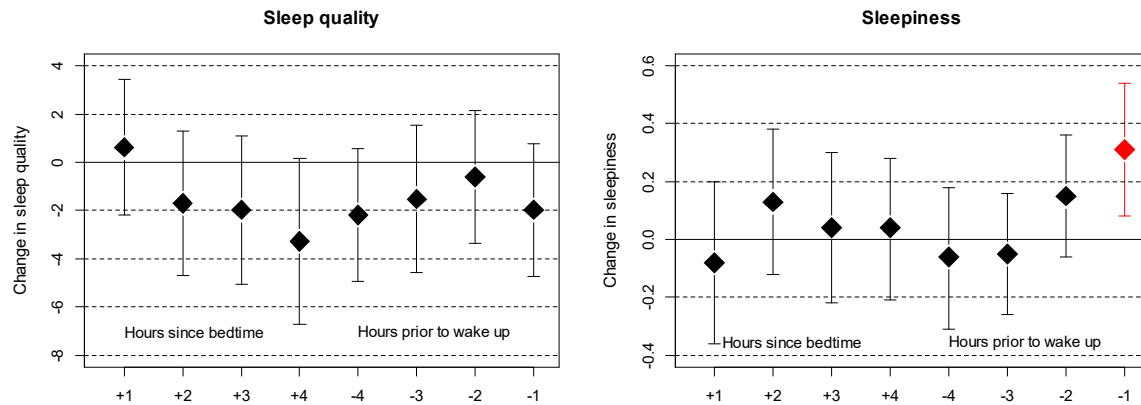
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Figure 2. Association between actimetry-derived outcomes and outdoor noise exposure for each hour after bedtime (+) or noise exposure for each hour prior to wake up (-). Changes refer to a 10 dB(A) increase in noise exposure in the respective hour. Same adjustments as indicated in Figure 1.



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Figure 3. Association between self-reported outcomes recorded each morning in the sleep diary and outdoor noise exposure for each hour after bedtime (+) or noise exposure for each hour prior to wake up (-). Changes refer to a 10 dB(A) increase in noise exposure in the respective hour. Same adjustments as indicated in Figure 1.

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4. Discussion

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Our study suggests that the timing of noise exposure within the night is a relevant factor for the deterioration of objective and subjective sleep quality. Using individual noise exposure time windows, matched to the individual bed- and rise times of each night, provided stronger associations compared to fixed time intervals (such as e.g. Leq between 19:00 and 23:00). Sleep latency, as expected, was most consistently associated with noise exposure at the beginning of the night, while for sleep efficiency and self-rated sleepiness noise exposure prior to wake up was most relevant. This suggests that noise exposure in the middle of the night may be less relevant for sleep quantity and quality, whereas noise exposure towards the end of the night, when sleep pressure is reduced and noise levels tend to be higher, is most disturbing for sleep.

This is in line with a field study using contrived aircraft noise exposure observing that aircraft noise events in the early morning elicited stronger reactions, as measured with high-resolution actigraphy, than events at the beginning of the sleep period [31]. Our findings are also in line with a large population based survey conducted 2009-2010 in Oslo with 13,019 participants. Road traffic nighttime noise was mostly associated with waking up too early and with difficulties falling asleep, but barely with awakenings during night [27]. In a Finnish study of 7,019 public sector employees, people exposed to >55 dB road traffic nighttime noise were more likely to report non-restorative sleep (OR=1.29, 95% CI: 1.01-1.65) and waking up too early (OR=1.24, 95% CI: 0.96-1.61) compared to people exposed to 45 dB or less. No associations were observed for frequently waking up during the night, short sleep duration, and difficulties falling asleep, the latter not in line with our results [43]. In a smaller Swiss survey of 1,375 adults, various self-reported sleep quality indices were not significantly related to road traffic nighttime noise [26]. However, most indications for an exposure-response trend were found for “waking up too early in the morning”, and least indications for “agitated sleep” and “waking phases during the night” in that same study [26].

Whether, or how, the observed pattern with stronger effects for evening and early morning noise translates into chronic health effects is unclear. Separating long-term noise effects regressed on average night noise exposure from time-specific effects in specific time periods is challenging for epidemiological studies, primarily because transportation noise in different time periods within the night is highly correlated, at least if no night curfews are in force. This is especially the case for modeled road traffic noise when traffic input data are based on traffic count samples, which are then extrapolated [2,3]. In reality, as demonstrated in our study, diurnal and day-to-day variation in traffic flows leads to a lower correlation between environmental noise exposures at different time intervals than one would observe between such computed standard metrics [44]. For large-scale

293 epidemiological studies on long-term risks, individual noise measurements as done here are not
294 feasible. A Swiss cohort study on cardiovascular mortality has evaluated the effects of diurnal noise
295 variation in their analysis by combining road, rail and aircraft noise. Although such a combination
296 introduces additional diurnal variability due to different pattern and night curfews for aircraft
297 noise, the correlation between different exposure time windows remained high (≥ 0.94), precluding
298 any firm conclusion. There was a trend that, for all cardiovascular causes combined, exposure
299 during 01:00 to 05:00 was most relevant. For hypertensive related causes of death early morning
300 noise (05:00-06:00) was most relevant; and for ischemic stroke and heart failure early evening and
301 early morning noise were more detrimental than the rest of the night, although strongest
302 association were seen with daytime noise [45].

303 We hypothesized that we would find stronger associations for estimated indoor noise than
304 measured outdoor noise but could not confirm this with our data. This is in line with results from a
305 survey on self-reported sleep disturbance in the same special issue of this journal [46]. However,
306 our findings do not match the results of an epidemiological study done on indoor noise, which
307 found stronger associations for hypertension with indoor noise compared to modeled noise at the
308 most exposed façade [47].

309 We did not measure indoor noise, but rather applied an indirect procedure to estimate indoor
310 noise levels from measured outdoor noise because our main interest was outdoor noise penetrating
311 into the building. Direct indoor noise measurements would be heavily affected by the behavior of
312 the participants. For instance, a sleepless person may produce some sound, which would yield a
313 biased correlation between sleeplessness and noise exposure. There is no obvious explanation why
314 indoor noise was less good a predictor than outdoor noise in our study. One may speculate that
315 people who feel disturbed by outdoor noise close their bedroom window and thus have a lower
316 indoor noise estimates, as this is the strongest predictor of indoor noise. This would mean that we
317 deal with reverse causation in the sense that the sleep quality affects the noise exposure and not the
318 other way round. Alternatively, estimated indoor noise levels may be subject to higher exposure
319 misclassification as some levels were low. Estimated indoor noise levels below 20 dB(A) were
320 censored and replaced with 20 dB(A) for the analysis; this was the case for 79 nights in 21
321 individuals. Finally, we obtained window opening habits for each season from the baseline
322 questionnaire but did not specifically ask about the window position for each night, which may also
323 add to exposure misclassification. Exposure misclassification may have resulted in reduced
324 statistical power, which would explain the observed similar regression coefficients for indoor and
325 outdoor noise but higher p-values for indoor noise.

326 We also hypothesized that sleep effects depend on the exposure characteristics. In particular,
327 exposure situations with individual noise events clearly standing out from average (background)
328 noise, as quantified with the Intermittency Ratio (IR) [32], were considered to be more detrimental
329 for sleep than exposure to a steady sound level. However, we could not confirm that sleep
330 disturbances are increasing with increasing IR. There was a non-significant non-linear pattern with
331 lowest sleep efficiency for moderate (50-63%) levels of IR, which is in line with findings of the effect
332 of IR in a cohort study on cardiovascular mortality [33]. However, in a cross-sectional survey on
333 arterial stiffness, rather number of events during night was relevant [48]; and in a cohort study on
334 diabetes IR was not associated with the diabetes risk [7]. Thus, it remains open whether IR is
335 contributing to Leq-based metrics for predicting long term health effects of noise.

336 The strengths of this study include the prospective and detailed data collection with
337 acquisition of objective sleep data (actimetry) and measurement of noise exposure. By measuring
338 instead of calculating outdoor noise exposure, we could adequately record the diurnal variability of
339 noise acting upon study participants. Our measurements allowed us to estimate outdoor noise
340 passing into the sleeping rooms for individual bedroom characteristics. We could also match
341 exposure time windows to the individual bed and rise times. The relatively small sample is a
342 limitation and thus the power of the study rather limited. P-values of the analyses are not adjusted
343 for multiple comparisons because we were interested in the pattern of the effect estimates rather
344 than hypothesis testing. Thus some significant coefficients may in fact be chance findings. Note also

345 that some of the exposure and outcome measures are correlated and thus analysis should not be
346 considered as mutually independent. This explains, for instance, the paradox result that noise
347 exposure two hours prior to wake up is significant associated with sleep latency.

348 5. Conclusions

349 Our study provides evidence that matching the nighttime noise exposure time window to the
350 individual's diurnal sleep-wake pattern results in a better estimate of detrimental nighttime noise
351 effects on sleep. The study suggests that noise exposure in the early morning hours is probably
352 most crucial for a negative impact on objective sleep efficiency and subjective sleep quality.
353 However, evening noise exposure was also associated with longer sleep latency. We could not
354 confirm that noise induced sleep effects are better explained by indoor noise compared to outdoor
355 noise, which might be due to reverse causality. This needs to be confirmed in larger studies.

356 **Supplementary Materials:** The following are available online at www.mdpi.com/xxx/s1, Table S1: Pearson
357 correlation matrix of sleep outcomes. Table S2: Summary of IR exposure data, Table S3: Pearson correlation
358 matrix for measured nighttime noise exposure metrics, Table S4: Association between all outcomes and
359 IRnight

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361 RP, BF, DV, JMW; formal analysis, BF, MRÖ; resources, RP, JMW; data curation, BF, FR, RP; writing—original
362 draft preparation, MRÖ; writing—review and editing, all; visualization, MRÖ; funding acquisition, MRÖ, CC,
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