

1 Article

2 **You can't see the woods for the trees: Invasive *Acer***
3 ***negundo* L. in urban riparian forests harms**
4 **biodiversity and limits recreation activity**

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22 **Abstract:** Public access to high quality green environments has become a key issue for city
23 managers and a matter of environmental justice. Remnants of natural ecosystems allow citizens a
24 direct contact with nature, but conversely the presence of people contributes further to the existing
25 disturbances. Urban pressures on ecosystem remnants may act to favour the expansion of some
26 invasive species in cities. Whilst the negative impacts of invasive species on ecosystem function is
27 well documented little is known how invasive species influence the use of green spaces by people.
28 Here, we examined one of the few remnants of urban riparian forests in Europe, the Vistula river
29 valley in Warsaw which has recently become an attractive recreation site. Despite their high
30 ecological value, the poplar and willow forests have been increasingly taken over by the invasive tree
31 species *Acer negundo*. We examined the status of the invasion process and the relationship between
32 recreational ecosystem services and the characteristics of the tree stands – tree species, tree density
33 and age and NDVI values. We found the willow forest to be more susceptible to invasion by *A.*
34 *negundo* than the poplar forest, which was revealed in significantly higher share of the maple
35 individuals and their greater volume per unit area. Presence of *A. negundo* affected biodiversity
36 resulting in decreased undergrowth density and number of species. The use intensity by the public,
37 assessed on the basis of trampling intensity and the density of existing informal tracks, were
38 negatively correlated to the presence of *A. negundo*. This study highlights the need to integrate
39 invasive species management into green infrastructure planning and management.

40 **Keywords:** blue-green infrastructure, nature-based solutions, urban green spaces, invasive trees,
41 trampling, forest remnants

42

43 1. Introduction

44 The increasing urban population of cities worldwide combined with anxiety over the life quality of
45 residents has resulted in an increasing interest in the benefits to be derived from city green areas [1,
46 2, 3, 4, 5, 6; 7]. Many cities were established on the banks of rivers and frequently the remnants of
47 former riparian forests constitute an important part of the green systems of the city [8, 9]. Despite
48 remaining under strong anthropogenic pressure, including land-use change, pollution, lack of cyclic
49 flooding, human trampling or the introduction of non-indigenous organisms [10]. Those ecosystems
50 generate multiple important services for the city residents [11, 12]. Natural riparian forests are
51 recognized for their positive contribution to nutrient removal, carbon capture, air purification,
52 pollination, noise buffering and water cleaning [13, 14, 15, 16, 17]. The few examples of riparian
53 forests which were preserved in cities play an additional important role by contributing recreation
54 opportunities to citizens and allowing them to interact with nature. The role of recreation in city green
55 spaces in improving citizen health and well-being has been widely recognized and is regarded as one
56 of the most important ecosystem services urban green spaces offer [18, 19, 20, 21, 22, 23, 24]. More
57 and more studies also reveal an increasing demand from city residents for “less ordered” green areas
58 providing the possibility to experience nature more directly than the traditional urban green spaces
59 such as parks can offer [25, 26].

60 Physical disturbances to nature in cities, such as pollution, drought or drainage lead to
61 irreversible changes in the existing ecosystems. Preserving high biodiversity in urban environments
62 faces major difficulties, and river valleys are additionally highly susceptible to other negative changes,
63 including a high risk of biological invasions [27]. Invasive plant species (IAS) not only severely alter
64 the biodiversity of the areas they colonize but can also alter the ecosystem services provisioning by
65 those areas [28]. The negative effects of various IAS on ecosystem services provisioning were
66 recognized for multiple services [29], as affecting pollination [30], water supply [32], carbon
67 sequestration or erosion control [32]. Yet little is known about the impact of IAS on the recreational
68 potential of an area, and still less about ecosystem services in riparian forests due to their scarcity and
69 difficulties in RES assessment [33].

70 Nature preservation in cities is strongly driven by social approval and a participatory
71 approach to spatial urban planning has become good practise [34]. Consequently, it is crucial to
72 recognize if IAS in urban areas, riparian forests in particular, affect the perception of urban green
73 areas and result in a decline of ecosystem service provisioning [29]. Invasive tree species are common
74 in urban forests, and likely to remain so due to low effectiveness of removal actions [29], yet little is
75 known concerning their perception as a component of natural vegetation remnants or regenerating
76 forms. The presence of invasive tree species affects the ecosystem function [35, 36, 37, 38, 39], but also
77 changes in their visual appearance which can affect how people use the space [39, 40]. Invasive plant
78 species can be therefore perceived negatively by the public and be associated with their
79 environmental impact, but still a significant share of society might have no negative attitude towards
80 the presence of invasive species [41]. Many plants, now declared as invasive species, were
81 deliberately introduced to urban recreational areas due to their aesthetic values, hence the public may
82 be positive towards IAS. Despite any ecosystem services lost, other services such as use in medicine
83 or biofuels may occur [29]. The public may perceive IAS similarly to native trees, and IAS
84 management thus receives low priority [42, 43]. The dispersal process of invasive plant species and

85 their negative effects on the ecosystems still require further examination [36], but it is also predicted
86 that climate change and human actions will escalate of the process [55, 56].

87 Ash-leaved maple (*Acer negundo*), a native of North America was commonly planted in Europe due
88 to high tolerance to heat and water stress and has become one of the most invasive plant species
89 occurring in riparian forests in Europe. Due to intensive seed production and easy dispersal by wind
90 and water it has spread throughout Poland, and other parts of Europe, being most successful in
91 urbanized regions along rivers [44, 45]. In this study we investigated the process of invasion in
92 riparian forests by *A. negundo* and evaluate how it affects the structure of the riparian forests
93 preserved in Vistula river in Warsaw and the effect of *A. negundo* on recreational ecosystem services
94 of this area. We address the following questions: 1) What are the forest stand characteristics of
95 riparian forest invaded to various extent by *A. negundo* - habitat type, tree density and age; 2) What
96 is the effect of *A. negundo* on the biodiversity and NDVI, as indirect measure of biomass and quality
97 of the riparian forest stands; 3) How are the recreational ecosystem services, expressed in intensity of
98 penetration by visitors, linked to the presence of *A. negundo* and the forest characteristics.

99 2. Materials and Methods

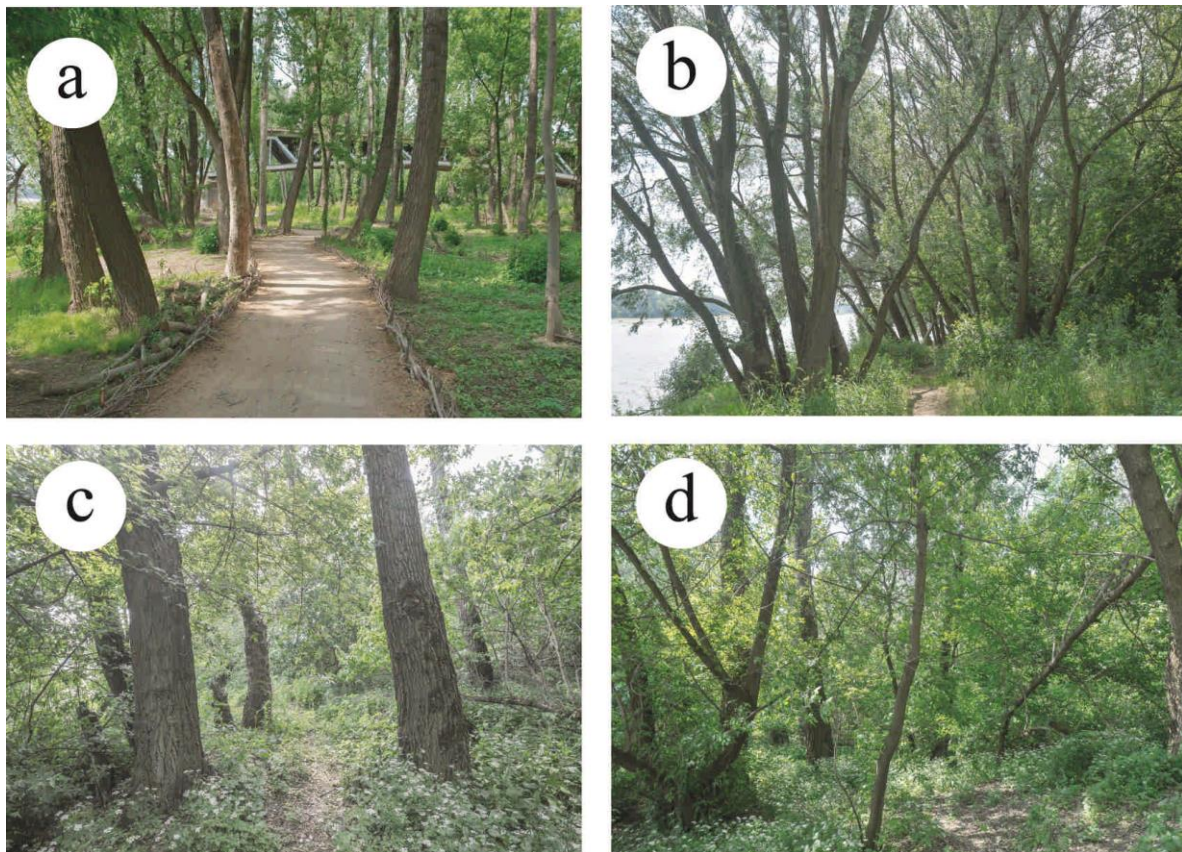
100 2.1. Study area

101 The investigation was conducted in Warsaw, Poland's capital city, which covers an area of 517 square
102 kilometres and is inhabited by over 1.76 million citizens (GUS, 2018). The city is characterized by high
103 greenness, with the green infrastructure, including agricultural areas, accounting for nearly 50% of
104 the city's area, along with over 14% forested lands [46]. The urban green space of Warsaw consists of
105 201 parks and forests, and 12 nature reserves. The unique element of Warsaw's green system is the
106 strip of natural riparian forest which developed along the Vistula river (Fig 1) in between the
107 embankments. The case of Warsaw is a unique example of a natural riparian forest formed within a
108 narrow strip between the river and the embankments which is also located in the strict city centre.
109 Today the area is covered by the 50-year old riparian forests *Populetum albae* and *Salicion albae* – 91E0
110 (Interpretation Manual - EUR28 2013), accompanied by the oxbow lakes represented by the *Potamion*
111 and *Nymphaeion* plant communities (code 3150-2) and muddy banks with *Chenopodion rubri* p.p
112 *Bidention* p.p. vegetation (code 3270), all being Natura 2000 habitat types, temporarily appearing
113 between the groyenes. Despite the high ecological value, revealed in number of species the area is
114 subjected to various pressures and the negative changes can be revealed in the presence of the
115 invasive tree species such as *A. negundo* which has been noted in substantial numbers in the forest
116 stand, both in the canopy and undergrowth [47].

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122 **Figure 1.** Main formal route along the riparian forest (formal trail) (a), informal trails in poplar riparian forest123 *Populetum albae* (b), willow riparian forest with *Populus x canescens* (c) and willow riparian forest heavily124 invaded by *A. negundo* (d)

125

126 Since the 70's the area of riparian forests received little attention from the city's managers, due to the

127 cyclic floods and associated lack of installed permanent infrastructure and remained unmanaged. In

128 2007 the Municipality of Warsaw made the area accessible to the public by creating a "nature track"

129 along the river in the forest. The track was an approximately 20 km section of a sandy route parallel

130 to the river course designated for walking and cycling, which remains periodically flooded along

131 with the temporary infrastructure. Since that time the number of visitors has grown substantially. In

132 years 2014-2018 the yearly number of cyclists recorded has grown to over 140 000 (source:

133 <http://rowery.um.warszawa.pl>). The track along the natural shore of the river, accompanied by

134 cultural and educational events remains one of the most popular places in Warsaw [47] offering the

135 citizens a wide range of ecosystem services and the possibility to experience nature in the middle of

136 a major city. Apart from the main formal trail along the forest a growing interest of the citizens in

137 nature-seeking has resulted in increased activity of the visitors off the main track and led to multiple

138 off-trail routes throughout the forests.

139

140 **2.2. Canopy tree inventory**

141 In the years 2015-2017 a detailed tree canopy inventory was held in the 20 km riparian forest strip

142 (Figure 1). We identified all tree individuals and recorded the species, the area was divided into

143 homogenous forest patches in terms of tree age, species (Figure 2) and habitat. A detailed vector map
144 of the forest stand patches was created based on field maps and orthophotomap from year 2017. The
145 patches varied from 0.01 to 1.63 ha. Within each distinguished patch every single tree trunk of breast
146 height diameter larger than 4 cm was recorded. All tree species were identified, including the hybrids
147 between *Populus* sp. and *Salix* sp. We also estimated the age of the forest stand patches on the basis
148 of rectified RGB orthophotomaps from years (1945, 1975, 1977, 1982, 1987, 1994, 2001, 2005, 2008, 2010,
149 2011 and 2012 retrieved from the Warsaw Municipality Office).

150 The forest stand patches were categorized based on the dominant tree species in the canopy and
151 vegetation composition as two habitat types - poplar or willow riparian forests and further analysed
152 separately, due to the differences in habitat characteristics. The willow forests patches were located
153 closer to the river and were of approximately 50 m width, the undergrowth was dominated by the
154 species characteristic of the *Salicetum albo-fragilis* type the content of sand in soil was >50% and cyclic
155 flooding took place every 2-10 years [48]. The poplar forest stand patches were located further from
156 the water course and stretched further to the embankment. The habitat was less frequently flooded
157 and the vegetation composition was characteristic for the *Populetum albae* type [49].
158



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160

161 **Figure 2.** Photograph presenting the investigated riparian habitats – typical willow and poplar forests used for
162 recreation in Warsaw

163

164 2.3. The effect of *A. negundo* on biodiversity

165 The distribution dynamics and effects of invasive *A. negundo* presence on the riparian forests
166 structure was investigated in detail. Vegetation was recorded in 83 representative 20×20 m plots in
167 homogenous areas to a various level invaded by *A. negundo*. In each plot the percentage of *A. negundo*,
168 number of species in the canopy and undergrowth, and percentage cover of each of the layers - trees,
169 shrubs, and undergrowth cover was recorded.

170 Vegetation quality assessment was performed by calculating average Normalized Difference
171 Vegetation Index (NDVI) for each of the previously identified forest patches from Sentinel 2

172 multispectral images. NDVI is frequently used to quantitatively assess urban vegetation, and values
173 vary from -1 to 1, where maximum values correspond to highest coverage of vegetation. We used a
174 cloudless scene obtained during the vegetation season (6.08.2017) to calculate NDVI using a software
175 addendum (Quality Assurance Tools) (van Leeuwen, TBRS, Tucson, Arizona) applied to ENVI
176 software. Normalized difference vegetation index was calculated using red and NIR reflectance as:
177 $NDVI = [(NIR - Red)/(NIR + Red)]$ [50].
178

179 **2.4. Off trail activity as an indicator of recreational ecosystem services of riparian forests**

180 We assumed the intensity of visits to be an indirect measure of user preference and willingness to
181 spend time in this area. We inventoried the “traces” of user activity beyond the main walking track
182 by mapping informal trails in the riparian forest, identified as linear damages in vegetation and soil
183 being the effect of off-trail activity [51]. We used two indicators: 1) off trail density per patch, which
184 better indicates the tendency of visitors to spatially penetrate the area and cause damage to the
185 vegetation and 2) off trails soil compaction, which more describes the intensity of visits and can be
186 related to the number of visitors who had used the trail. Mapping of informal trails was performed
187 in years 2016 and 2017 in May-August in the field using a GPS device along the whole 20 km river
188 section in the willow and forest stands. Linear tracks with visible vegetation losses and bare ground
189 were inventoried. The density of tracks was presented as density in m/ha. In the central part of each
190 inventoried trail section in the representative part soil compaction was measured using a
191 penetrometer Eijkelkamp 06.01.SA and presented in kN/cm². The measurement was performed in all
192 patches at least 5 repetitions and the score averaged. Previous studies revealed that the main walking
193 trail was visited by 136 to 687 people per day over the weekend and from 69 to 343 during weekdays
194 [47]. A small proportion of the visitors, 1.1-11.5% walk off the main trail which equates to around
195 three thousand people/year who spend time beyond the main track within the natural riparian forest,
196 most walk off the track only for a short period of time [47].

197 One-way ANOVA/MANOVA were used, followed by Tukey to test differences in forest type and
198 vegetation parameters – tree stand composition and NDVI values. Pearson correlation was used to
199 assess the relationship between the presence of *A. negundo* and the total vegetation cover at various
200 heights and recreational ecosystem services expressed in the length and density of informal tracks.
201 Statically significant differences were at $p < 0.05$. All statistical analyses were performed in SPSS 13
202 software.
203

trees									
[years]									
willow	165.0	17.8	162.3	47.8	245.9	877.6	4649.8	4.3	0.55
poplar	96.8	163.9	28.3	33.5	157.1	6907.9	836.3	2.0	0.60
p-value	0.09*	0.00*	0.00*		0.01*	0.00*	0.00*		0.00*
willow									
<10	633.1a	3.9a	223.6	73.6	454.5a	79.0a	2606.3a	14.5	0.50a
10-20	150.1b	5.3ab	149.3	49.3	209.9b	164.7a	3898.6ab	4.9	0.52ab
20-30	77.4b	25.3b	160.2	29.4	194.9b	949.9ab	3838.0ab	3.9	0.55ab
>30	102.9b	22.1ab	152.7	37.1	239.4b	1361.6b	6082.1b	3.1	0.57b
p-value	0.00*	0.01*	0.71		0.00*	0.00*	0.01*		0.01*
poplar									
<10	240.5a	59.2	30.1a	72.9	309.5a	3402.7	1030.4	6.5	0.59
10-20	200.4ab	125.1	25.1a	57.2	229.0ab	6341.2	524.5	3.2	0.58
20-30	81.4bc	175.3	50.5b	26.5	101.2b	7260.5	1247.7	1.2	0.60
>30	61.8c	177.2	15.7c	24.3	158.0ab	7132.9	664.4	2.0	0.61
p-value	0.00*	0.31	0.00*		0.03*	0.44	0.40		0.72

229

230 The structure of the forests stands strongly depends on their age. Young stands are significantly more
 231 highly invaded (73.6 %) than younger stands (37.1%), the number of individuals being 14.5% and 3.1%
 232 in terms of volume, respectively. Whereas the native tree species show both increased numbers and
 233 volume in older stands (Table 1). In both willow and poplar forest stands at the age up to 20 years
 234 the maple is always the dominating tree species, while the older stands from 20 to 40 years showed
 235 only a small fraction of this species in terms of the basal area of trunks, which is at least 5 times lower
 236 than of the native trees. In willow stands a significant increase in numbers of poplars is noted in older
 237 stands when compared to younger ones, while both, poplar and willows increase in volume. In
 238 poplar forests, the increase in numbers is not significant, while the willows decrease in numbers, the
 239 increase in volume is not homogenous (table 1).

240

241 3.2. The effect of *A. negundo* on biodiversity

242 We found that the presence of *A. negundo* strongly modifies the structure of the forest stand at many
 243 levels. *A. negundo* is a main component of the shrub layer (2-5 m height) along with *Sambucus nigra*
 244 and *Populus nigra* and *Salix alba* seedlings, hence the high correlation obtained ($r^2=0.775$; Table 2). The
 245 presence of *A. negundo* contributes to lower density of the canopy ($r^2=-0.338$) and high density of *A.*
 246 *negundo* is associated with a decline of the plant density of the undergrowth ($r^2=-0.584$). The
 247 development of the species in the canopy is also followed by a continuous decline of total number of
 248 plant species in undergrowth ($r^2=-0.378$).

249

250 **Table 2.** The relationship between the presence of *A. negundo* and the total vegetation cover at various heights.
 251 Pearson correlation coefficients. significant values shown in bold at $p < 0.05$

	Tree layer	Shrubs	Undergrowth	Number of species
Share of <i>A. negundo</i> in layer (p-value)	-0.338*	0.775*	-0.584*	-0.378*

252

253 3.3. Effect of *A. negundo* on Recreation Ecosystem Services

254 By linking Recreational Ecosystem Services (RES) to indicators describing visitor activity throughout
 255 the area them walking off the main track we found that the relation between the maple density and
 256 off-trail activity strongly depends on the habitat type. In the willow forest the relation between the
 257 maple coverage and visitor activity is statistically significant and as the coverage of the maple
 258 increases the soil compaction decreases ($p = -0.212$; table 3). The soil compaction decreases which
 259 allows us to conclude that the presence of *A. negundo* results in reduced off-trail activity of the visitors.
 260 This relationship with the density of *A. negundo* was not observed in the poplar forest.
 261 A clear positive relationship between the abundance of maple, expressed in the average tree basal
 262 area were found with the penetration intensity by visitors expressed in path soil compaction and path
 263 density ($p = 0.276$ and $p = 0.652$; table 3). While the number of trees *per se* (highest in youngest stands,
 264 table 1) was not related to the visitor activity or was negatively correlated (Table 3). In the case of the
 265 poplar forest a negative relation between the presence of *A. negundo* and the trail density was found
 266 ($p = -0.345$; table 3). Other indicators and tree species cover were unequivocally related to off-trail
 267 activity.

268

269 **Table 3.** Pearson's correlation coefficients for the recreational ecosystem services indicators (off trail density
 270 and off-trail soil compaction) relation with the tree canopy structure in willow and poplar riparian forests.

271 Significant differences were shown in bold

272

	Poplar forest							
	maple		poplar		willow		all trees	
	no	b.area	no	b.area	no	b.area	no	b.area
Off trail	-	-0,179*	0,192	0,173*	0,048	0,128	0,133	0,652*
Off trail	-	-0,406*	0,098	0,081	-	-0,017	-	0,276*

273

274

	Poplar forest							
	maple		poplar		willow		all trees	
	num	basal	num	basal	num	basal	num	basal
	ber	area	ber	area	ber	area	ber	area
Off trail	-	-0,060	0,096	0,099	-	-0,093	-	0,073
Off trail	0,052	-0,039	-	-0,078	0,033	0,022	0,005	-0,071

275

no – number, b.area – basal area

276 4. Discussion

277 This study revealed significant alterations of the structure and biodiversity of riparian forests when
278 invaded by *A. negundo* from the herbaceous layer up to the canopy. The maple always prevailed as
279 the dominating species in younger stands, both in terms of number of individuals and the volume in
280 both willow and poplar tree stands, while in tree stands of 30-40 years the indigenous tree species
281 were present in higher densities. The presence of maple was associated with a selective negative
282 impact on the patterns of how people move across the area, revealed in informal track density and
283 soil compaction. We found the willow forests where the share of maple was higher to be less
284 frequently visited while in poplar forests this relationship was less visible. The study showed the
285 major effect of the invasive maple on both the biodiversity and recreation, indicating at accurate
286 management implications for this species to optimize nature conservation and recreation in such
287 valuable areas.

288 With an increased demand from city residents for direct contact with nature [52], natural ecosystems
289 such as riparian forests, will become more frequently visited and gain more importance as places for
290 recreation. Here, we found that willow stands invaded by *A. negundo* maple were less visited by
291 people, while a reduction in visitation was not evident for poplar tree stands with a high maple
292 density. *A. negundo* is perceived no differently to other trees by the public and is treated as an
293 accepted component of the green space, more attractive than the view of the built-up areas [53]. Yet
294 we found reduced usage of willow stands with a high maple content. The effect was more visible in
295 the willow forests located closer to the water where the density of the trails and their compaction
296 were negatively correlated to both number and volume of maple. The use of poplar stands was less
297 affected, the track density was only found to be smaller in sites where the number of individual
298 maples was high. In the willow stands with high *A. negundo* the activity of visitors on informal tracks
299 was reduced by over one third. These areas are of a high importance to the visitors as the presence of
300 water in the recreational area has a great influence on the aesthetical judgment [54]. For many visitors
301 the presence or absence of invasive species may have little impact on recreation, a previous study
302 found that many users accept spontaneous vegetation on grasslands providing it remains green [26]
303 and many citizens may not recognize a species as invasive. Our measurements were based on the
304 "traces" of the visitors and may be affected by the soil difference between the two habitats or by
305 proximity to water and user preferences for areas close to water. Also, differences in the "visibility"
306 and easiness of penetration and herbaceous vegetation density which might have affected the
307 distribution of the informal tracks over the area.

308 In this study, *A. negundo* proved to be a widespread component of the riparian forest flora of Warsaw.
309 The riparian forests in cities are reported to have 10-40% of invasive tree species [36, 55] and over the
310 last 100 years an increase of IAS in urban floras has been noted [57]. The invasion alters the
311 composition of the forests, affecting for instance mineralization processes leading to replacement of
312 the riparian native species [58]. In extreme cases in the riparian zones the invasion can lead to the
313 development of homogenous communities of novel ecosystems little resembling the former riparian
314 forests [59]. Our study showed widespread invasion of *A. negundo* in riparian forests but restricted to
315 the young forests stands. We also found major differences between the willow forests situated close
316 to the river course and the poplar forests located further from the river. The willow forest may be
317 more susceptible to invasion by *A. negundo*, where the density of shoots of the native willow was
318 larger than analogically poplars in the neighbouring habitat (Table 2). The river catchments are in

319 general at risk of being subject to biological invasions [60], but the sites which are being regularly
320 flooded are also more frequently subject to establishment of the seedlings from the seeds carried by
321 water [61]. The correlation between the risk of invasion with the distance to the river course is
322 previously recognized [62]. The willow forests compared to poplar forests, despite higher invasion
323 rate, were characterized by lower NDVI, indicating at higher biomass values in contrast to a study
324 from Bulgaria where it was the more invaded sites in the riparian zones which had the highest canopy
325 cover [63].

326

327 None of the patches distinguished was bereft of *A. negundo*, and the oldest stands did not reveal
328 trends of this species disappearing. Young tree stands up to the age of 20 years were characterized
329 by numerous maple shoots of small basal area. In the older forest patches of age 20-40 years the share
330 of maple is lower while the indigenous trees prevail both in terms of number of shoots and their basal
331 area (Table 2). This development stage of the regenerating riparian forest is for *A. negundo* a moment
332 of entering the survival phase, as its shoots grow in volume and develop under the canopy of other
333 trees but do not increase in numbers. *A. negundo* encounters biotic resistance in an intermediate
334 successional niche when the indigenous species grow more rapidly [64]. The maple in older stands
335 indicate *A. negundo* does stabilize at a low level showing little need to undertake removal actions for
336 this species. Removal of trees results in creating gaps which would be quickly filled by the maple as
337 it reproduces effectively from both shoots and rhizomes [65]. The seedlings of maple are
338 characterized by high tolerance and quick growth rate in the gaps [66].

339

340 The presence of *A. negundo* strongly affects the structure of the tree stand and the herbaceous
341 vegetation were noted. Woodlands with *A. negundo* had less dense canopy and more developed shrub
342 layer with floristically poor herb layer which relates to loss of ecological value [37, 56, 67]. The
343 examined sites which were partially artificially created by river regulation and further developed
344 due to natural processes (accumulation, succession). Increased penetration intensity by invasive
345 species and degradation of indigenous plant communities is expected to be intensified by increasing
346 pressure from recreational activities [68], and the future may be bleak for the remnants of natural
347 forest in urban environments.

348

349 High demand for green areas and outdoor recreation in cities results in riparian areas being
350 incorporated into the city's green system, both the ones of high ecological value and the degraded
351 ones [69]. In Warsaw, similarly to other European cities, the residents seek direct contact with nature,
352 it especially to the most valuable components of nature [7, 70]. Areas along rivers, overgrown with
353 dense vegetation of riparian forests and plentiful shore vegetation are highly attractive to visitors [71],
354 and the characters destroyed patches are repaired [65, 72]. Such natural ecosystems are exposed to
355 high numbers of visitors and serve as recreation areas, but they are also subject to increased trampling
356 and biodiversity loss [69, 73, 74]. The preservation of biodiversity may then contradict the ability of
357 residents to freely and actively use the space. It is not known how to manage degraded areas still
358 having traces of naturalness but with high share of invasive species. Should they be protected as
359 remnants of natural vegetation or a rather more flexible approach should be used, eg. to apply no
360 removal of invasive species, despite them posing a threat to biodiversity and to make it available to
361 the public because of their contribution to recreation. Invasive species impede biodiversity, but also

362 offer a multitude of other services to the residents [29, 75], therefore their removal in cities, where
363 they were introduced in the first place as ornamental plants requires understanding of both, the
364 ecological process and taking into account the preferences of the public.
365 This study showed that recreation can be affected by the presence of the invasive *A. negundo*.
366 However, if removal actions were to be undertaken it is more desirable in the areas which are more
367 intensively visited, therefore the willow forests close to the water. Removal of high numbers of the
368 maple could result in a rapid decline of aesthetics and social approval of the more valuable areas by
369 the water, as the public perceives the maple better than areas bereft of vegetation [41]. Recreational
370 activity in such areas should rather be supported by creating additional infrastructure, as in protected
371 areas, which could limit the pressure and boost indigenous riparian vegetation regeneration. From
372 the point of the biodiversity preservation the maples should be removed [72] but taking into account
373 the costs and constantly occurring disturbances in the area it is expected that the maple will become
374 an inseparable component of the city's green spaces. The maple can be treated more gently, the risk
375 of invasion expansion should be taken into account but so should its role in contribution to recreation
376 and ecosystem services provisioning [29]. If removal of *A. negundo* is necessary it should be carefully
377 considered and if occurring should be spread over a longer period (over 30 years) allowing the
378 indigenous plant communities to regenerate and eliminate the maple through natural abiotic
379 pressure [66].

380 5. Conclusions

- 381 1. *A. negundo* is a permanent and abundant component of the urban riparian forests in Vistula
382 river valley in Warsaw, it was found to be more abundant in willow forests stands than in
383 poplar forests.
- 384 2. The abundance of *A. negundo* was found to be significantly higher in younger stands than in
385 older ones, the differences were manifested in both number of stems and their volume.
- 386 3. Occurrence of *A. negundo* in riparian forests negatively affects biodiversity, shrub layer and
387 herbaceous vegetation, the more invaded stands were poorer in species.
- 388 4. Increased share of *A. negundo* were found to be related to decreased activity of visitors in the
389 forest, but the effect was stronger in the willow forests, the poplar forests were less affected
390 Presence of *A. negundo* can play an important role in providing recreation possibilities for the
391 city dwellers.

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398 References

- 399 1. Tzoulas, K.; Korpela, K.; Venn, S.; Yli-Pelkonen, V.; Kazmierczak, A.; Niemela, J.; James, P. Promoting
400 ecosystem and human health in urban areas using green infrastructure: a literature review. *Landsc.*
401 *Urban Plan.* **2007**, *81*, 167–178.
- 402 2. Breuste, J.; Niemela, J.; Snep, R.P.H. Applying landscape ecological principles in urban environments.
403 *Landsc. Ecol.* **2008**, *23*, 1139–1142.

- 404 3. Vandermeulen, V.; Verspecht, A.; Vermeire, B.; Van Huylenbroeck, G.; Gellynck X. The use of economic
405 valuation to create public support for green infrastructure investments in urban areas. *Landsc. Urban*
406 *Plan.* **2011**, *103*, 198–206.
- 407 4. Gómez-Baggethun, E.; Barton, D.N. Classifying and valuing ecosystem services for urban planning.
408 *Ecol. Econ.* **2013**, *86*, 235–245.
- 409 5. Schäffler, A.; Swilling, M. Valuing green infrastructure in an urban environment under pressure — The
410 Johannesburg Case. *Ecol. Econ.* **2013**, *86*, 246–257.
- 411 6. Breuste, J.; Rahimi, A. 2015, Many public urban parks, but who profits from them? The example of
412 Tabriz, Iran. *Ecol. Process.* **2015**, *4*, 1–15.
- 413 7. Richards, D.R.; Warren, P.H.; Moggridge, H.L.; Maltby, L. Spatial variation in the impact of dragonflies
414 and debris on recreational ecosystem services in a floodplain wetland. *Ecosystem Serv.* **2015**, *15*, 113–121.
- 415 8. Shanahan, D.F.; Lin, B.B.; Gaston, K.J.; Bush, R.; Fuller A.. What is the role of trees and remnant
416 vegetation in attracting people to urban parks? *Landsc. Ecol.* **2015**, *30*, 153–165.
- 417 9. Liquete, C.; Kleeschulte, S.; Dige, G.; Maes, J.; Grizzetti, B.; Olah, B.; Zulian, G. Mapping green
418 infrastructure based on ecosystem services and ecological networks: A Pan-European case study.
419 *Environ. Sci. Policy* **2016**, *54*, 268–280.
- 420 10. Pennington, R.T.; Lavin, M.; Oliveira-Filho, A. Woody plant diversity, evolution, and ecology in the
421 tropics: perspectives from seasonally dry tropical forests. *Ann. Rev. Ecol., Evol. Syst.* **2009**, *40*, 437–457.
- 422 11. Haase, D. Holocene floodplains and their distribution in urban areas—functionality indicators for their
423 retention potentials. *Landscape Urban Plann.* **2003**, *66*, 5–18.
- 424 12. Juutinen, A.; Koseniusec, A.K.; Ovaskainen, V. Estimating the benefits of recreation-oriented
425 management in state-owned commercial forests in Finland: A choice experiment. *J. Forest Econ.* **2014**,
426 *20*, 396–412.
- 427 13. Bayley, P.,B. Understanding large river: floodplain ecosystems, *BioScience* **1995**, *45*, 153–158.
- 428 14. Barbier, E.B.; Thompson, J.R. The value of water: floodplain versus large-scale irrigation benefits in
429 northern Nigeria. *Ambio* **1998**, *27*, 434–440.
- 430 15. Tockner, K.; Stanford, J.A. Riverine flood plains: present state and future trends. *Environ. Conserv.* **2002**,
431 *29*, 308–330.
- 432 16. Grygoruk, M.; Mirosław-Świątek, D.; Chrzanowska, W.; Ignar, S. How much for water? Economic
433 assessment and mapping of floodplain water storage as a catchment-scale ecosystem service of
434 wetlands. *Water* **2013**, *5*, 1760–1779.
- 435 17. Looy van K.; Tormos, T.; Souchon, Y.; Gilvear, D. Analyzing riparian zone ecosystem services bundles
436 to instruct river management, *Ecosyst. People* **2017**, *13*, 330–341.
- 437 18. Fitter, A.; Elmqvist, T.; Haines-Young, R.; Potschin, M.; Rinaldo, A.; Setälä, H.; Stoll-Kleemann, S.;
438 Zobel, M.; Murlis, J. An assessment of ecosystem services and biodiversity in Europe. *Environ. Sci.*
439 *Technol.* **2010**, *30*, 1–28.
- 440 19. Suding, K.N. Toward and area of restoration in ecology: successes, failures and opportunities ahead.
441 *Annu. Rev. Ecol. Evol. Syst.* **2011**, *42*, 465–487.
- 442 20. Lachowycz, K.; Jones, A.P. Towards a better understanding of the relationship between greenspace and
443 health: development of a theoretical framework. *Landsc. Urban Plan.* **2013**, *118*, 62–69.
- 444 21. White, M.P.; Alcock, I.; Wheeler, B.W.; Depledge, M.H. Would you be happier living in a greener urban
445 area? A Fixed-effects Analysis of Panel Data. *Psychol. Sci.* **2013**, *24*, 920–928.
- 446 22. Alcock, I.; White, M.P.; Wheeler, B.W.; Fleming, L.E.; Depledge, M.H. Longitudinal effects on mental
447 health of moving to greener and less green urban areas. *Environ. Sci. Technol.* **2014**, *48*, 1247–1255.
- 448 23. Gascon, M.; Triguero-Mas, M.; Martínez, D.; Dadvand, P.; Fors, J.; Plasencia, A. Mental health benefits
449 of long-term exposure to residential green and blue spaces: a systematic review. *Int. J. Environ. Res.*
450 *Public Health* **2015**, *12*, 4354–4379.
- 451 24. Gascon, M.; Triguero-Mas, M.; Martínez, D. Residential green spaces and mortality: a systematic review.
452 *Environ. Intern.* **2016**, *86*, 60–67.
- 453 25. Kim, M.; Rupprecht C.D.D.; Furuya K. Residents' Perception of informal green space – a case study of
454 Ichikawa City, Japan. *Land* **2018**, *7*, 1–20.
- 455 26. Sikorski, P.; Wińska-Krysiak, M.; Chormański, J.; Sikorska, D. Low-maintenance green tram tracks as a
456 socially acceptable solution to greening a city. *Urban For. Urban Gree.* **2018**, *35*, 148–164.

- 457 27. Charkes, H.; Dukes, J.S. Impacts of invasive species on ecosystem Services. [in:] *Biological Invasions*;
458 Nentwig, W., Eds.; Springer, Berlin Heidelberg, Germany, **2008**; pp. 217–237.
- 459 28. Vilà, M.; Hulme, P.E. Non-native species, ecosystem services, and human well-being. [in:] *Impact of*
460 *Biological Invasions on Ecosystem Services*; Vilà, M.; Hulme, P.E., Eds.; Cham, Switzerland, 2017; pp. 1-14.
- 461 29. Sladonja, B.; Poljuha, D.; Uzelac, M. Non-native invasive species as ecosystem service providers. [in:]
462 *Ecosystem services and global ecology*; Hufnagel, L., Eds.; IntechOpen, London, GB, **2018**; pp. 39-59.
- 463 30. Nienhuis, C.; Stout, J. Effectiveness of native bumblebees as pollinators of the alien invasive plant
464 *Impatiens glandulifera* (Balsaminaceae) in Ireland. *J. Pollin. Ecol.* **2009**, *1*, 1-11.
- 465 31. Preston, I.R.; Le Maitre, D.C.; Blignaut, J.N.; Louw, L.; Palmer, C.G. Impact of invasive alien plants on
466 water provision in selected catchments. *Water SA* **2018**, *44*, 719–729.
- 467 32. Greenwood, P.; Baumann, P.; Pulley, S.; Kuhn, N.J. The invasive alien plant, *Impatiens glandulifera*
468 (Himalayan Balsam), and increased soil erosion: causation or association? Case studies from a river
469 system in Switzerland and the UK. *J. Soils Sediment* **2018**, *18*, 3463–3477.
- 470 33. Kapitza, K.; Zimmermann, H.; Martín-López, B.; von Wehrden, H. Research on the social perception of
471 invasive species: a systematic literature review. *NeoBiota* **2019**, *43*, 47–68.
- 472 34. Newig, J.; Fritsch, O. Environmental governance: Participatory, multi-level - And effective? *Environ.*
473 *Policy and Gov.* **2009**, *19*, 197–214.
- 474 35. Schnitzler, J.; Benzler, J.; Altmann, D.; Mucke, I.; Krause, G. Survey on the population's needs and the
475 public health response during floods in Germany 2002. *J. Public Health Manag. Pract.* **2007**, *13*, 461–464.
- 476 36. Höfle, R.; Dullinger, S.; Essl, F. Different factors affect the local distribution, persistence and spread of
477 alien tree species in floodplain forests. *Basic Appl. Ecol.* **2014**, *15*, 426–434.
- 478 37. Dyderski, M.K.; Gdula, A.K.; Jagodzinski, A.M. The rich get richer" concept in riparian woody species
479 – A case study of the Warta River Valley (Poznan, Poland). *Urban For. Urban Gree.* **2015**, *14*, 107–114.
- 480 38. Rieger, I.; Kowarik, I.; Cherubini, P.; Cierjacks, A. A novel dendrochronological approach reveals
481 drivers of carbon sequestration in tree species of riparian forests across spatiotemporal scales. *Sci. Total*
482 *Environm.* **2017**, *574*, 1261–1275.
- 483 39. Foster, J.; Sandberg, L. Friends or foe? Invasive species and public green space in Toronto. *Geograph.*
484 *Rev.* **2004**, *94*, 178–198.
- 485 40. Shackleton, R.T.; Richardson, D.M.; Shackleton, C.M.; Bennett, B.; Crowley S.L.; Dehnen-Schmutz, K.;
486 Estevez, R.A.; Fischer, A.; Kueffer, C.; Kull, C.A.; Marchante, E.; Novoa, A.; Potgieter, L.J.; Vaas, J.; Vaz,
487 A.S.; Larson, B.M.H. Explaining people's perceptions of invasive alien species: A conceptual framework.
488 *J. Environ. Manage.* **2018**, *229*, 1-17.
- 489 41. Potgieter, L.J.; Gaertner, M.; O'Farrell, P.J.; Richardson, D.M. Perceptions of impact: Invasive alien
490 plants in the urban environment. *J. Environ. Manage.* **2018**, *229*, 76-87.
- 491 42. Hoyle, H.; Hitchmough, J.; Jorgensen, A. Attractive, climate-adapted and sustainable? Public perception
492 of non-native planting in the designed urban landscape. *Landscape Urban Plann.* **2017**, *164*, 49–63
- 493 43. Sikorska, D.; Sikorski, P.; Richard, H.J. High biodiversity of green infrastructure does not contribute to
494 recreational ecosystem services. *Sustainability* **2016**, *8*, 334-347.
- 495 44. Pyšek, P.; Chytrý, M.; Pergl J.; Sádlo, J.; Wild, J. Plant invasions in the Czech Republic: current state,
496 introduction dynamics, invasive species and invaded habitats, *Preslia* **2012**, *84*, 575–629.
- 497 45. Tokarska-Guzik, B.; Dajdok, Z.; Zając, M.; Zając, A.; Urbisz, A.; Danielewicz, W.; Hołdyński, C. Plants
498 of foreign origin in Poland with particular emphasis on invasive species (in Polish), GDOŚ, Warszawa,
499 Poland, **2012**; pp. 1-197.
- 500 46. Degórska, B., Degórski, M. Green infrastructure as a very important quality factor in urban areas –
501 Warsaw case study. *Europa XXI* **2017**, *32*, pp. 51–70.
- 502 47. Sikorski, P.; Parafjańczuk, S.; Wierzba, M.; Sikorska, D.; Borowski, J.; Vitasović Kosić, I. The
503 phenomenon of illegal dispersion in riparian forests under high tourist pressure (in Polish). In *Problems*
504 *of water management in forest, urban and non-urbanized areas*. Kałuża T., Strzeliński P. Eds.; Wydawnictwo
505 Naukowe Bogucki: Poznań, Poland, **2014**; pp. 131-144.
- 506 48. Matuszkiewicz, W.; Sikorski, P.; Szwed, W.; Wierzba, M.; Danielewicz, W.; Wysocki, C.; Kiciński, P.
507 Vegetation of Poland. Illustrated guide: forests and shrubs (in Polish), PWN, Warszawa, Poland, **2012**,
508 pp. 189-207.
- 509 49. Leuschner, C.; Ellenberg, H. *Vegetation ecology of Central Europe*, 6th ed. Springer International
510 Publishing: Cham, Switzerland, 2017, pp. 652–688.

- 511 50. Carlson, T.; Ripley, D. On the relationship between NDVI, fractional vegetation cover, and leaf area
512 index. *Remote Sens. Environ.* **1997**, *62*, 241–252.
- 513 51. Coppes, J.; Braunisch, V. Managing visitors in nature areas: where do they leave the trails? A spatial
514 model. *Wildlife Biol.* **2013**, *19*, 1–11.
- 515 52. Abbott, A. City living marks the brain, *Nature* **2011**, *22*, 429.
- 516 53. Arnberger, A.; Schneider, I.E.; Ebenberger, M.; Eder, R.; Venette, R.C.; Snyder, S.A.; Gobster, P.H.; Choi,
517 A.; Cottrell, S. Emerald ash borer impacts on visual preferences for urban forest recreation settings.
518 *Urban For. Urban Gree.* **2017**, *27*, 235–245.
- 519 54. Kaplan, R.; Kaplan, S. The experience of nature: A psychological perspective. Cambridge University
520 Press, New York, NY, **1989**, pp. 1–368.
- 521 55. Schnitzler, A.; Hale, B.W.; Alsum, E.M. Examining native and exotic species diversity in European
522 riparian forests. *Biol. Conserv.* **2007**, *138*, 146–156.
- 523 56. Pennington, D.N.; Hansel, J.R.; Gorchov, D.L. Urbanization and riparian forest woody communities:
524 Diversity, composition, and structure within a metropolitan landscape. *Biol. Conserv.* **2010**, *143*, 182–194
- 525 57. Kowarik, I. Novel urban ecosystems, biodiversity and conservation. *Environ. Pollut.* **2011**, *159*, 1974–
526 1983.
- 527 58. Krevš, A.; Kučinskiene, A. Influence of invasive *Acer negundo* leaf litter on benthic microbial
528 abundance and activity in the littoral zone of a temperate river in Lithuania. *Knowl. Manag. Aquat.*
529 *Ecosyst.* **2017**, *418*, 1–10.
- 530 59. Marozas, V.; Cekstere, G.; Laivins, M.; Straigyte, L. Comparison of neophyte communities of *Robinia*
531 *pseudoacacia* L. and *Acer negundo* L. in the eastern Baltic Sea region cities of Riga and Kaunas. *Urban*
532 *For. Urban Gree.* **2015**, *14*, 826–834.
- 533 60. Sunga C.Y.; Li M-H.; Rogers G.O.; Volder A.; Wang Z. Investigating alien plant invasion in urban
534 riparian forests in a hot and semi-arid region. *Landsc. Urban Plan.* **2011**, *100*, 278–286.
- 535 61. Pyšek, P.; Jarošík, V.; Hulme, P.E., Pergl, J.; Hejda, M.; Schaffner, U.; Vilà, M. A global assessment of
536 invasive plant impacts on resident species, communities and ecosystems: the interaction of impact
537 measures, invading species' traits and environment. *Glob. Chang. Biol.* **2012**, *18*(5), 1725–1737.
- 538 62. Höfle, R.; Dullinger, S.; Essl, F. Different factors affect the local distribution, persistence and spread of
539 alien tree species in floodplain forests. *Basic Appl. Ecol.* **2014**, *15*(5), 426–434.
- 540 63. Dyakov, N.; Zhelev, P. Alien species invasion and diversity of riparian forest according to
541 environmental gradients and disturbance regime. *Appl. Ecol. Env. Res.* **2013**, *11*, 249–272.
- 542 64. Saccone, P.; Girel, J.; Pages, J.P.; Brun, J.J.; Michalet, R. Ecological resistance to *Acer negundo* invasion
543 in a European riparian forest: relative importance of environmental and biotic drivers. *Applied Veg. Sci.*
544 **2013**, *16*, 184–192.
- 545 65. Merceron, N.R.; Lamarque, L.J.; Delzon, S.; Porté, A.J. Killing it softly: Girdling as an efficient eco-
546 friendly method to locally remove invasive *Acer negundo*. *Ecol. Rest.* **2016**, *34*, 297–305.
- 547 66. Saccone, P.; Pagès, J.P.; Girel, J.; Brun, J.J.; Michalet, R. *Acer negundo* invasion along a successional
548 gradient: early direct facilitation by native pioneers and late indirect facilitation by conspecifics. *New*
549 *Phytologist* **2010**, *187*, 831–842.
- 550 67. Straigytė, L.; Cekstere, G.; Laivins, M.; Marozas, V. The spread, intensity and invasiveness of the *Acer*
551 *negundo* in Riga and Kaunas. *Dendrobiology* **2015**, *74*, 157–168.
- 552 68. Vakhlamova, T.; Rusterholz, H.P.; Kamkin, V.; Baur, B. Recreational use of urban and suburban forests
553 affects plant diversity in a Western Siberian city. *Urban For. Urban Gree.* **2016**, *17*, 92–103.
- 554 69. Sikorski P. 2013. Influence of the urban park nature on the floristic diversity of undergrowth and park
555 lawns (in Polish). Wyd. Wieś Jutra, Warsaw, Poland, 1–108.
- 556 70. Mihalič, T. Performance of environmental resources of a tourist destination: concept and application. *J.*
557 *Travel Res.* **2013**, *52*, 614–630.
- 558 71. Kenwick, R.A.; Shammin, M.R.; Sullivanc, W.C. Preferences for riparian buffers. *Landsc. Urban Plan.*
559 **2009**, *91*, 88–96
- 560 72. Gonzalez, E.; Martínez-Fernandez, V.; Shafroth, P.B.; Sher, A.A.; Henry, A.L.; Garofano-Gomez, V.;
561 Corenblit, D. Regeneration of Salicaceae riparian forests in the Northern Hemisphere: A new
562 framework and management tool. *J. Environ. Manage.* **2018**, *218*, 374–387.
- 563 73. Cole D.N.; Marion J.L. Recreation impacts in some riparian forests of the Eastern United States. *Environ.*
564 *Manage.* **1988**, *12*, 99–107.

- 565 74. Bötsch, Y.; Tablado, Z.; Scherl, D.; Kéry, M.; Graf, R.F.; Jenni, L. Effect of recreational trails on forest
566 birds: human Presence matters. *Front. Ecol. Evol.* **2018**, *6*, 1–10.
- 567 75. Sharp, R.L.; Larson, L.R.; Green, G.T. Factors influencing public preferences for invasive alien species
568 management. *Biol. Conserv.* **2011**, *144*, 2097–2104.