

Article

## Human health risk assessment of nitrate and trace metals *via* groundwater in central Bangladesh

Md. Mostafizur Rahman<sup>1\*</sup>, Md. Bodrud-Doza<sup>2</sup>, Md. Iftakharul Muhib<sup>3</sup>, Kaniz Fatima Binte Hossain<sup>4</sup>, Md. Sabbir Hossain<sup>1</sup>, Sharmin Akter<sup>1</sup>, Md. Tajuddin Sikder<sup>5</sup>, Mashura Shammi<sup>1</sup>, M Khabir Uddin<sup>1</sup>

<sup>1</sup>Department of Environmental Sciences, Jahangirnagar University, Dhaka, Bangladesh. [rahmanmm@juniv.edu](mailto:rahmanmm@juniv.edu)

<sup>2</sup>International Centre for Climate Change and Development, Bangladesh

<sup>3</sup>Department of Textile Engineering, City University, Bangladesh

<sup>4</sup>Graduate School of Environmental Science, Hokkaido University, Japan

<sup>5</sup>Department of Public Health and Informatics, Jahangirnagar University, Dhaka, Bangladesh

\*Correspondence: [rahmanmm@juniv.edu](mailto:rahmanmm@juniv.edu); Tel.: +8801717514833

**Abstract:** Groundwater plays a pivotal role as the largest potable water sources in Bangladesh. However, the quality of the groundwater faces challenges due to xenobiotic compounds in it. As agriculture is widely practiced in Bangladesh, potential nitrate ( $\text{NO}_3^-$ ) pollution may appear in near future. Besides, excess amount of arsenic (As) has already been found in groundwater in many parts of Bangladesh including the present study area. Thus, this study was conducted to assess the water quality and associated human health risk in central Bangladesh. A total of 99 groundwater samples from the central part of Bangladesh were analyzed to assess human health risk due to high level of  $\text{NO}_3^-$  and other trace elements i.e. arsenic (As), iron (Fe), and manganese (Mn). Concentration of  $\text{NO}_3^-$  was determined using column chromatography and inductively coupled plasma optical emission spectrometer (ICPOES) was used to measure As, Fe and Mn concentrations. It was found that  $\text{NO}_3^-$  concentration (253.17 mg/L) in the groundwater samples exceeds the recommended guideline value by the WHO (50 mg/L). Moreover, this study area also characterized with elevated concentration of As (19.44  $\mu\text{g/L}$ ), Fe (811.35  $\mu\text{g/L}$ ), and Mn (455.18  $\mu\text{g/L}$ ) in the groundwater. Non-carcinogenic human health risk was calculated by justifying HQ (Hazard Quotient) and HI (Hazard Index) and attributed potential conjunctive human health risks due to  $\text{NO}_3^-$ , As, Fe and Mn in the study area. Child (9.941) is more vulnerable than adult (7.810) considering non-carcinogenic human health risk. Moreover, high carcinogenic risk was found due to As contamination in the groundwater samples and children ( $1.94 \times 10^{-3}$ ) are more susceptible to carcinogenic risk compared to adults ( $9.2 \times 10^{-4}$ ).

**Keywords:** nitrate; arsenic; health risk assessment; hazard quotient; hazard index.

### 1. Introduction

Groundwater is one of the major drinking water sources in Bangladesh; however its quality has been degraded by human activities over the time. The presence of trace elements such as arsenic (As), iron (Fe), manganese (Mn) and nitrate ( $\text{NO}_3^-$ ) in ground water is predominantly influenced by natural and anthropogenic sources [1, 2]. A line of evidence already suggested the possible sources of As in the groundwater of Bangladesh. For instance, geogenic processes, discharge of wastewater, smelters and mines, agricultural runoff, leakage from heavily contaminated areas and geothermal waters into ground water reservoir, could be the possible sources of metals in groundwater [2-5]. We have demonstrated the status of As contamination in Singair Upazila, Manikganj District and geogenic factors were found to be the major source of As in the groundwater in that area [2]. The quality of

groundwater, such as concentration of trace elements is predominantly related to human health. Consumption of water contaminated with significant amount of trace elements *via* oral ingestion may cause serious health effects varying from shortness of breath to several types of cancers in human [6]. The consumption of As contaminated ground water is the main concern of As toxicity, skin lesion, gangrene in leg, cancer in lung, bladder, liver and skin etc., these diseases could be found among humans due to excessive consumption of As contaminated drinking water *via* oral ingestion pathway [7]. Fe and Mn also could accumulate into human body through drinking pathway [8]. Exposure to Mn could cause neurotoxicity [9] while gastrointestinal distress and dysfunction of multiple organs could occur due to iron consumption *via* drinking water [10-11]. Long term consumption of excessive  $\text{NO}_3^-$  can create cancer risk on human body due to formation of nitrosamines [12]. It has also been known to create spontaneous absorptions, birth defects, respiratory tract infections and changes in immune system [13-18]. Thus health risk assessment of these contaminants for predicting the health hazard is very important. According to United States Environment Protection Agency (USEPA), human health risk assessment is the process of estimating the nature and probability of adverse health effects in humans who may be susceptible to chemicals in contaminated environmental media, now or in the near future [19]. Mostly four steps must be followed to fulfill the health risk assessment e.g. identification of hazard, dose-response relationship, exposure assessment as well as risk characterization [20]. Thus, human health risk assessment is an effective way to justify health risk levels posed by various contaminants [21]. Oral ingestion is considered to be the route of exposure to chemical contaminants like  $\text{NO}_3^-$ , As, Fe, Mn in groundwater.

Considering the above facts, we hypothesized that the level of contaminants ( $\text{NO}_3^-$  and trace metals) might be higher than acceptable level in central Bangladesh and could be major risk factor for human health. Therefore, the aim of this study was to determine the level of  $\text{NO}_3^-$  along with As, Fe and Mn in groundwater and to evaluate the health risk associated with exposure to these contaminants *via* oral ingestion. This research may be the first attempt for assessing the health risk associated with  $\text{NO}_3^-$  and trace metals in the groundwater of central Bangladesh.

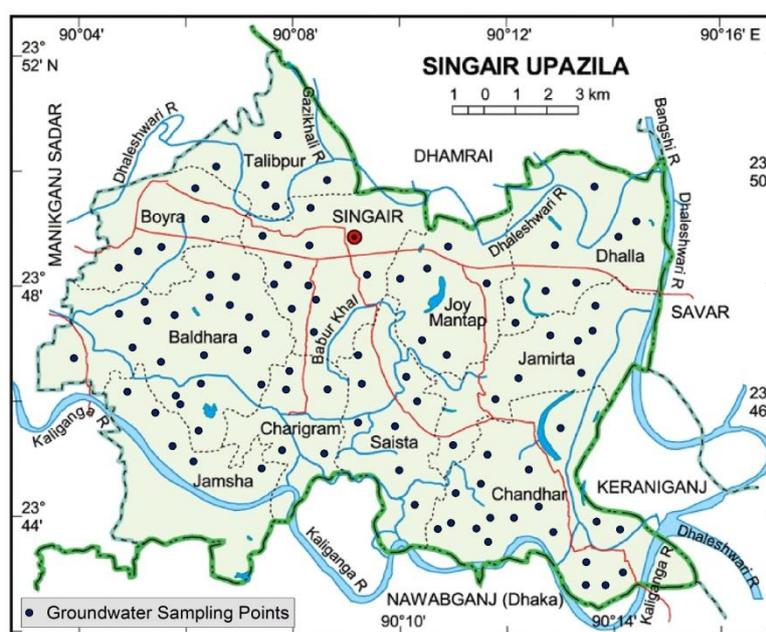


Figure 1. Map of the study area.

## 2. Materials and Methods

Total 99 groundwater samples were collected from different depth depending on the tube-well bore depth (m) in the central Bangladesh (Figure 1). The study area is located in Manikganj district Bangladesh. Sampling time is August 2011 and analysis was performed within 1 month from the sampling time. Before taking the sample the tube-well was pumped for around 10 minutes. All of the collected samples were stored in 250 ml polypropylene plastic bottles with (for trace metal)/without ( $\text{NO}_3^-$ ) acidification by 2 drops of concentrated  $\text{HNO}_3$  to prevent any precipitations. The inductively coupled plasma (ICP-OES) optical emission spectrometer was used (SII NanoTechnology inc.) for the determination of As, Fe, and Mn as depicted else where [2]. Nitrate ( $\text{NO}_3^-$ ) was determined by chromatography method (DX-120, Dionex, USA) using the columns "Ion Pac AS12A" (Dionex, Abbreviation of state, USA). Samples were filtered with glass fiber filters with 0.45  $\mu\text{m}$  pore. All the samples were analyzed triplicate for ensuring reproducibility and statistical validity. Statistical analysis was done by MS-excel 2007 and Principal Component Analysis (PCA) was done by using Origin 9.0 software (OriginLab, USA).

### 2.1. Risk Analysis

Risk assessment is the process of estimating the occurrence probability of any given magnitude of adverse health effects over a specified time period [22]. The health risk assessment of each element is based on evaluation of the risk level and is expressed in terms of a carcinogenic or non-carcinogenic health risk [23]. Slope factor (SF) for carcinogen risk characterization and the reference dose (RfD) for non-carcinogen risk characterization were also evaluated [24]. Oral ingestion pathway was considered for the study [25]. Below equation was adopted to calculate the chronic daily intake for oral and dermal adsorption pathways [26-29].

$$CDI_{Oral} = \frac{(CW \times IR \times EF \times ED)}{(BW \times AT)} \quad (1)$$

Here, CDI = Chronic daily intake ( $\mu\text{g}/\text{Kg}/\text{day}$ ) CW = Concentration of trace metal in water ( $\mu\text{g}/\text{L}$ ), IR = Ingestion rate (L/day; 2.2 for adult 1 for Child), EF = Exposure frequency (Days/year, 365), ED = Exposure duration [Year; for oral = 70 for Adult, 10 for Child], BW = Body weight (Kg; 70 kg for Adult, 25 kg for Child), AT = Average time (Days; 25,550 for Adult, 3650 for Child) [22, 28-33].

The health risk due to groundwater consumption was justified on the basis of chronic (noncarcinogenic) and carcinogenic effects. The noncarcinogenic risk was calculated as Hazard quotient (HQ) by the following equation-

$$HQ = \frac{CDI}{RfD} \quad (3)$$

Where HQ is hazard quotient (unitless) and RfD ( $\mu\text{g}/\text{Kg}/\text{day}$ ) originates from risk-based concentration table [34].

For elemental risk assessment, the individual HQs are combined to form Hazard Index (HI). If the value of HQ and HI exceeds 1, there could be potential noncarcinogenic effects on health while HI less than 1 indicates the no risk of health effects [35-36].

$$HI = HQ_1 + HQ_2 + \dots + HQ_n \quad (4)$$

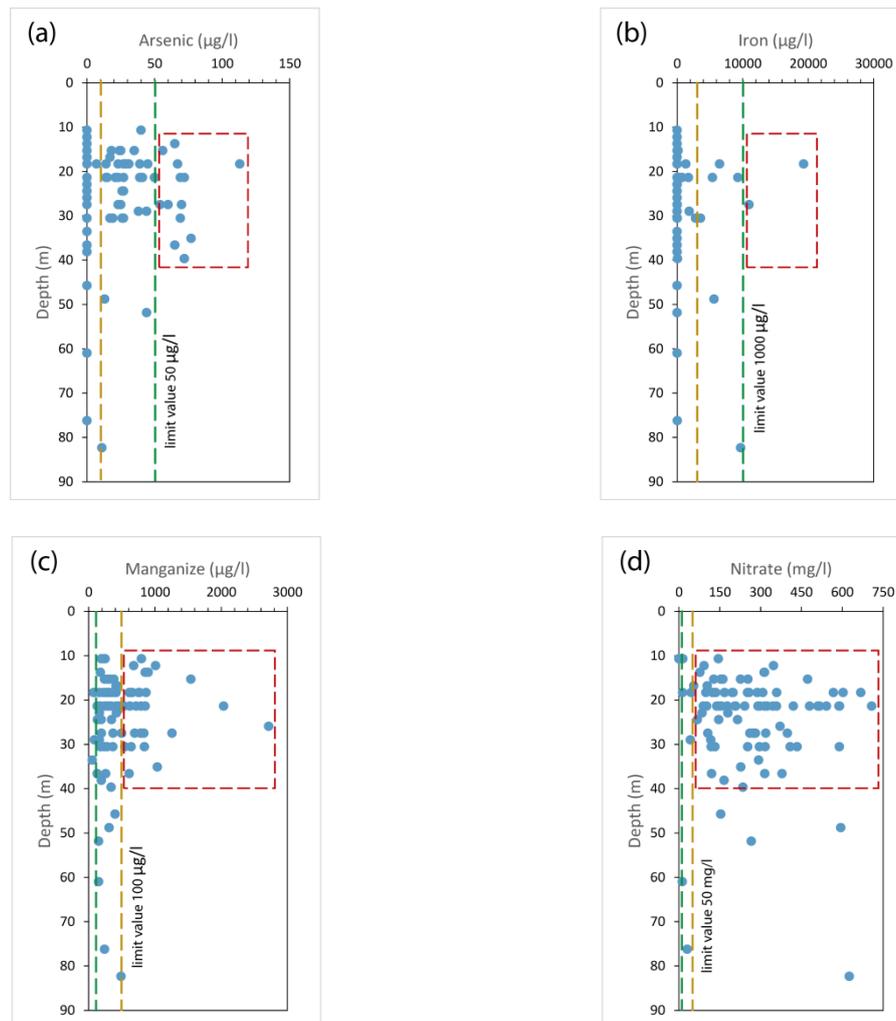
The carcinogenic risk was measured from the calculation of  $CDI_{oral}$  (mg/kg/day) and Slope Factor (SF) (mg/kg/day)<sup>-1</sup> and a range of characterization has been provided in **Table 1** chronic and cancer risk assessment [37].

**Table 1.** Scales for chronic and carcinogenic risk assessment [31, 21]

Risk level	HQ or HI	Chronic risk	Calculated cases of cancer occurrence	Cancer risk
1	< 0.1	Negligible	< 1 per 1000,000 inhabitants (10 <sup>-6</sup> )	Very low
2	≥ 0.1 < 1	Low	> 1 per 1000,000 inhabitants (10 <sup>-6</sup> ) < 1 per 100,000 inhabitants(10 <sup>-5</sup> )	Low
3	≥ 1 < 4	Medium	> 1 per 100,000 inhabitants(10 <sup>-5</sup> ) < 1 per 10,000 inhabitants(10 <sup>-4</sup> )	Medium
4	≥4	High	> 1 per 10,000 inhabitants(10 <sup>-4</sup> ) < 1 per 1000 inhabitants(10 <sup>-3</sup> )	High
5			> 1 per 1000 inhabitants(10 <sup>-3</sup> )	Very high

**Table 2.** Descriptive Statistical Analysis ) (detailed data were provided in supplementary material 1).

Parameters	Minimum	Maximum	Mean	Std. Deviation	Bangladesh Standard (1997)	Indian standards (2012)		WHO (2011)
						Acceptable Limit	Permissible Limit	
Well Depth (m)	10.67	82.3	24.322	11.82475				
As (µg/L)	BDL	113	19.444	24.75604	50	10	50	10
Fe (µg/L)	BDL	19296	811.353	2747.66231	300-1000	300	No relaxation	No proposed value
Mn (µg/L)	54	2716	455.181	394.11977	100	100	300	500
NO <sub>3</sub> <sup>-</sup> (mg/L)	BDL	708.11	253.175	168.80272	10	45	No relaxation	50



**Figure 2.** Depth wise water analysis. Dotted line representing recommended/limit value and the red dotted box refers the clusters of sampling points exceeding limit value for each of the parameters limited to 40 m depth.

### 3. Results

#### 3.1. Statistical Description

The average depth wise results of 99 ground water samples for As, Fe, Mn and  $\text{NO}_3^-$  are provided in **Table 1**. The mean concentrations of As, Fe, Mn and  $\text{NO}_3^-$  were also compared in **Table 2** with the BDWS (1997), Indian Standards (2012) and WHO (2011). Statistical analysis from **Table 2** represented that mean concentration of  $\text{NO}_3^-$  (253.175  $\text{mg/L}$ ) exceeded the standard concentration of Bangladesh Drinking Water Standard (BDWS 1997), India (2012) and WHO (2011). Again the mean concentration of Mn (455.181  $\mu\text{g/L}$ ) was found to be greater than BDWS (1997) and Indian Standards (2012) but lesser than WHO (2011). Though the mean concentration for Fe (811.253  $\mu\text{g/L}$ ) was within the range of BDWS (1997), it crossed the parameters standardized by India (2012) and WHO (2011). The mean

concentration of As (19.444  $\mu\text{g/L}$ ) did not exceed the limit for BDWS (1997) and permissible limit of Indian Standards (2012) but break the concentration limit given by acceptable limit of Indian Standards (2012) and WHO (2011).

Elemental depth wise distribution was also illustrated in terms of depth (m) vs. elemental concentrations ( $\mu\text{g/L}$ ). [Figure 2(a-d)]. Figure 2d suggested that approximately 87.89% of groundwater samples exceeded all the standard limit values of  $\text{NO}_3^-$  in the depth range between 10m-40m. Figure 2c, 2a and 2b also showed that approximately 21.21% ground water sample for Mn, 13.13% for As and 2.02% for Fe crossed the standard limit values in the same depth range.

**Table 3.** Principle component Analysis

	PC1	PC2
As	0.56	-0.536
Fe	0.83	-0.229
Mn	0.249	0.752
$\text{NO}_3^-$	0.691	0.439
Eigenvalues	1.542	1.099
% of Variance	38.562	27.479
Cumulative %	38.562	66.04

### 3.2. Principal Component Analysis (PCA)

To assess inter-elemental correlation results Principal Component Analysis (PCA) of groundwater samples were analyzed and have been shown in Table 3. PCA was calculated to interpret the elemental data [38]. This powerful method helps to identify and correlate different group of elements that can be identified as possessing similar behavior and common origin [39]. The eigenvalue was considered higher than 1 on the basis of Kaiser Criterion [35]. The principal component analysis of collected ground water samples elucidated that the variables were co-related to 2 principal components in which 66.04% total variance was identified. The source components (factors) were extracted on the basis of varimax rotation [38].

**Table 4.** Summary of HQ and HI of As, Fe, Mn and  $\text{NO}_3$  and Carcinogenic risk of Arsenic for oral ingestion in groundwater samples (average of 99 groundwater samples) (detailed data were provided in supplementary material 1)

Health Risk	Inhabitants	HQ for As	HQ for Fe	HQ for Mn	HQ for $\text{NO}_3$	Hazard index (HI)	Non-Carcinogenic Risk
Non-carcinogenic risk	Adult	2.037	0.085	0.715	4.973	7.810	High
	Child	2.593	0.108	0.910	6.329	9.941	High
		Carcinogenic risk (CR) of As					Carcinogenic Risk

<b>Carcinogenic risk</b>	Adult	$9.2 \times 10^{-4}$	High
	Child	$1.17 \times 10^{-3}$	Very High

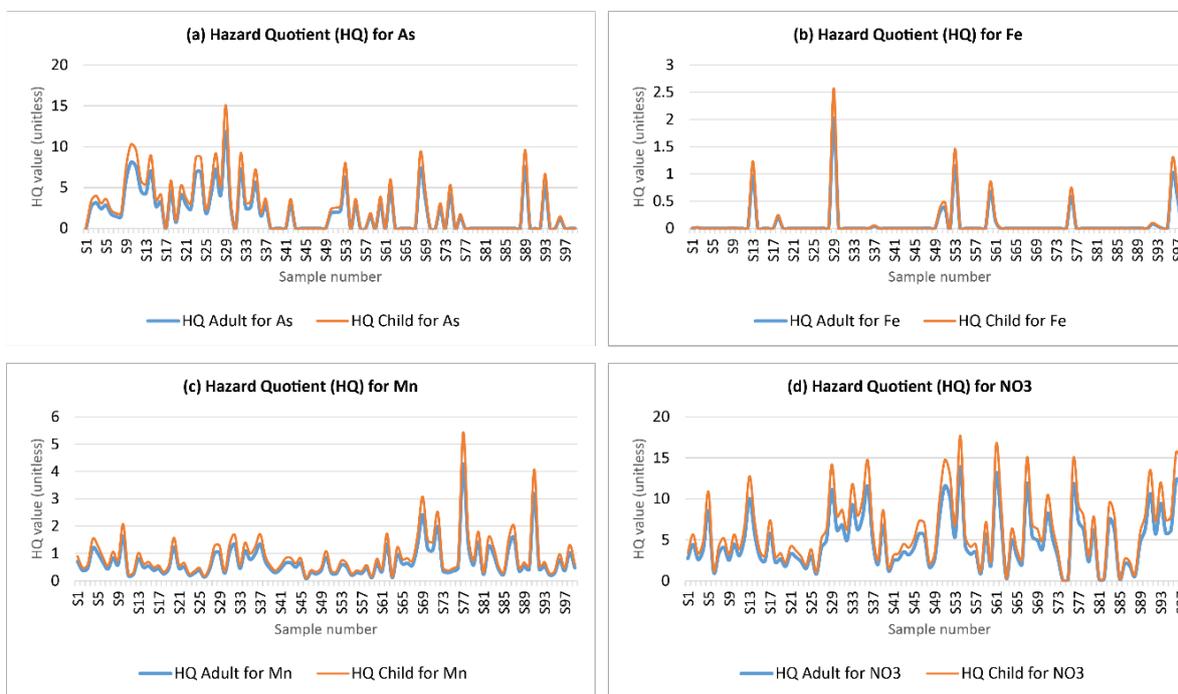
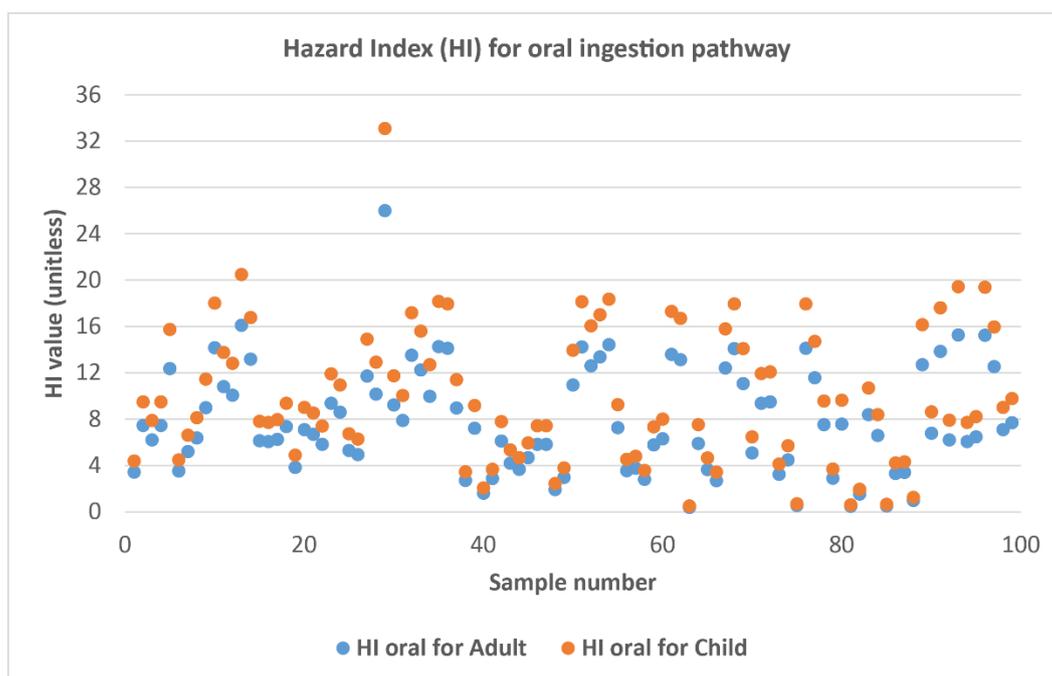


Figure 3: Hazard Quotient (HQ) oral ingestion pathway for As, Fe, Mn and NO<sub>3</sub><sup>-</sup>. Water sample identification number and HQ are plotted in X and Y axis respectively (Sampling points with depth provided in supplementary material 2).

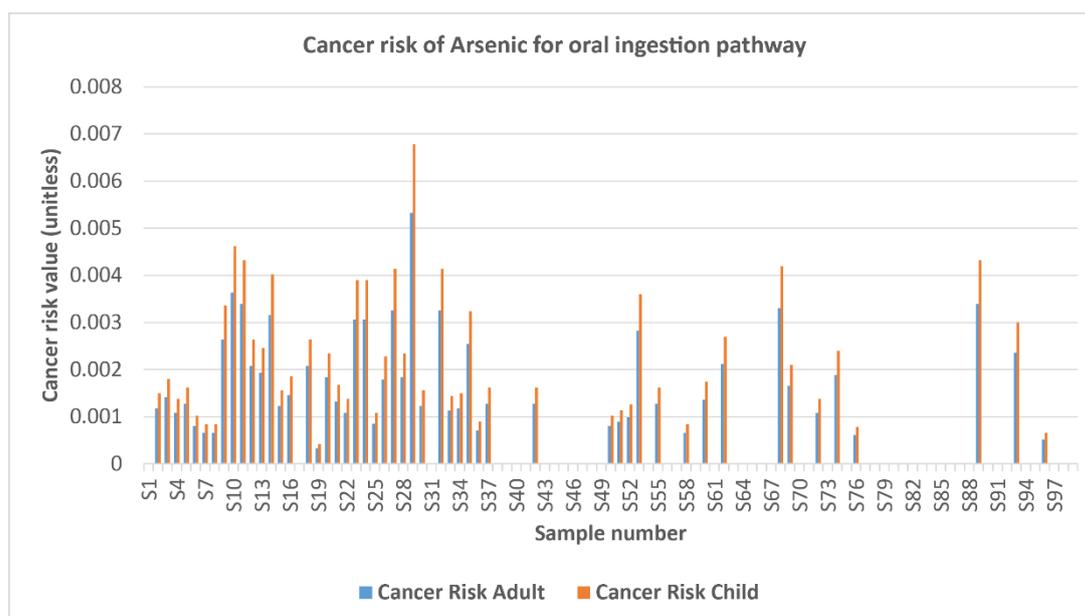


**Figure 4.** Hazard Index (HI) for oral ingestion pathway (Sampling points with depth provided in supplementary material 2).

### 3.3. Health Risk Assessment

#### 3.3.1 Hazard Quotient (HQ) and Hazard Index (HI) measurement

Above described table 2 and table 3 already have shown that the groundwater of this area is polluted with nitrates and trace metals. To evaluate the impact on human health related to this pollution, establishing Hazard Quotient (HQ) with related Hazard index (HI) is an important tool to calculate non-carcinogenic health risk. In this study, HQ and HI of As, Fe, Mn and  $\text{NO}_3$  for 99 ground water samples are illustrated in **Figure 2 and 3**. Considering average of these elements, both oral as well as dermal adsorption risks were identified in terms of non-carcinogenic risk measurement for both child and adult generation (**Table 4**). Oral ingestion pathway for adult and child showed high HQ level in As, Fe, Mn and  $\text{NO}_3$ . Considering oral ingestion of As, Fe, Mn and  $\text{NO}_3$ , HQ values for adults were 2.037, 0.085, 0.715 and 4.973; whereas, HQ values for child were 4.321, 0.180, 1.517 and 10.549, respectively. The following order was found in respect of HQ for adults in oral ingestion  $\text{NO}_3 > \text{As} > \text{Mn} > \text{Fe}$ . For child the order was  $\text{NO}_3 > \text{As} > \text{Mn} > \text{Fe}$  in oral ingestion. HI calculation also done for adult and child. For adult, about 75.76% samples showed high non-carcinogenic risk while 19.19% samples showed medium non-carcinogenic risk. For child, about 83.84% samples showed high non-carcinogenic risk while 11.11% samples showed medium non-carcinogenic risk. Low non-carcinogenic health effect (5.05% samples) was shown in both adult and child. Average hazard index result for adult and child also placed in **Table 4**. Results showed that both adult (HI=7.810) and child (HI=9.941) showed high non-carcinogenic risk effect.



**Figure 5.** Cancer Risk (HI) for oral ingestion pathway (Sampling points with depth provided in supplementary material 2).

### 3.3.2. Cancer Risk Assessment

Cancer risk on adult and child due to As for both oral ingestion pathway was illustrated in **Figure 4 and 5** and the calculated result placed in **Table 4**. Considering oral ingestion pathway, high ( $9.2 \times 10^{-4}$ ) and very high ( $1.94 \times 10^{-3}$ ) carcinogenic risk was justified for adult and child respectively. Out of 99 samples for adult, 39.39% samples exhibited very high carcinogenic risk and 12.12% samples showed high carcinogenic risk. Moreover, for children, 42.42% samples possessed very high carcinogenic risk while 9.09% samples showed high carcinogenic risk. For both adult and children about 48.48% samples seemed to be not harmful.

## 4. Discussion

Water quality parameters reflect the level of contamination in water resources. In this study, the quality of groundwater in terms of nitrate ( $\text{NO}_3^-$ ), arsenic (As), iron (Fe) and manganese (Mn) varied drastically among different sampling sites in central Bangladesh. The results has shown that the mean value of  $\text{NO}_3^-$  (253.175  $\mu\text{g/L}$ ) exceeded Bangladesh standard (BDWS, 1997), Indian Standards (2012) and WHO standard (2011) (Table 2). Both the geogenic and anthropogenic sources can be the contributor of elevated level of  $\text{NO}_3^-$  in groundwater of the study area e.g. application of nitrogenous fertilizer or manure slurries in agriculture and aquaculture, poor soil profile, and irrigation mechanisms. Relatively little of the  $\text{NO}_3^-$  found in natural waters is of mineral origin, most coming from organic and inorganic sources, the former including waste discharges and the latter comprising chiefly artificial fertilizers. Alongside, anthropogenic activities like sewage effluents, unsewered sanitation, and unplanned disposal of solid waste in densely populated areas may also leached  $\text{NO}_3^-$  into groundwater [43]. The higher occurrence of  $\text{NO}_3^-$  in shallow depth groundwater was likely due to anthropogenic activities e.g. excessive use of agrochemicals and unmanaged irrigation system that could cause microbial mineralization in the ground water [1]. This excess  $\text{NO}_3^-$  contamination in drinking water can cause increased cancer risk [38].

Nitrogenous fertilizers rapidly convert into  $\text{NO}_3^-$  in soils, are highly soluble and hence easily leachable to deep soil layers and ultimately enter into shallow aquifers [44]. For instance, sandy or sandy clay soil with high coarse texture have high water filtration rate and possibly contributing in nitrate leaching to underground waters [45]. Moreover, extensive agricultural practices in the study area may contribute in  $\text{NO}_3^-$  leaching in aquifers of this area [46]. Similar  $\text{NO}_3^-$  concentration increase over the acceptable limit was reported by Majumdar *et al.* (2008) in shallow groundwater of central-west region of Bangladesh [44].

Interest is centered on  $\text{NO}_3^-$  concentrations in groundwater is mainly due to its potential deleterious effects on human health. Hemoglobin carries oxygen into the blood but high level of  $\text{NO}_3^-$  intake reduces the capacity of oxygen binding with hemoglobin which creates a condition referred to as methemoglobinemia or "blue baby syndrome," which may cause mortality by asphyxiation especially in newly born infants [48]. Although,  $\text{NO}_3^-$  itself is not a direct toxicant but is a health hazard because of its conversion to nitrite. According to Walker (1990) when nitrate reduces to nitrite, it can undergo nitrosation reactions in the stomach with amines and amides to form a variety of *N*-nitroso compounds (NOC) [49]. These *N*-nitroso compounds (NOC) are mainly carcinogens [54]. The results of this study thus indicates that groundwater from the 10 m to 40 m depth in the study area is severely polluted with  $\text{NO}_3^-$ . Furthermore, the HQ and HI value of  $\text{NO}_3^-$  implicated for potential non-carcinogenic health risk in the study area both for adult and for children (table 4 and figure 3). Dellavalle *et al.* (2014) suggested that high dietary nitrate intake expected to have higher exposure to

endogenously formed NOCs and increases risk of colorectal cancer [50]. Moreover, our result has been shown that children health risk due to increased NO<sub>3</sub><sup>-</sup> concentration in drinking water is higher than that of adults. Similar results were reported by Zhai *et al.* (2017) in China [51].

The study area is also characterized with elevated levels of trace metals (As, Fe, and Mn) in the groundwater. The mean concentration of all three metals exceeds the limit of WHO standards (2011). The major potential sources of these trace metals is geogenic [2, 52] and Mn might be originated from both geogenic and anthropogenic sources as depicted by the PCA analysis (Table 3). The first component (PC1) was considered to be associated with the earth's crust and the geological basin formation [31]. This component showing 38.562% variance comprised of As, Fe, Mn and NO<sub>3</sub><sup>-</sup> with high loadings. The second component (PC2) was also attributed to the geogenic source [31] and explaining 27.479% of variance (Table 3). The mean value of arsenic (19.44 µg/L) lies within the range of BDWS (1997) standard and permissible limit of Indian Standards (2012) but exceeded the concentration limit given by acceptable limit of WHO standard (2011). This occurrence of As, Mn and Fe in the shallow depth water might be attributed to the geogenic formation of the Bengal Basin [2] as well as from anthropogenic activities [39]. Arsenic in groundwater is associated with skin damage, increased risk of cancer, and problems with circulatory system [53]. This study shows that carcinogenic risk of arsenic for oral ingestion from groundwater sample is high for child than the adults. Similar high carcinogenic risk for adult and child were also found by Rahman *et al.* (2017) in ground water samples of Gopalganj of Bangladesh [50], which is located southern side of our sampling area. Moreover, Chen *et al.* (2011) reported that exposure to arsenic in drinking water is adversely associated with mortality from heart disease [51]. Skin lesions caused by As contaminated drinking water have already been reported among adults and children in Bangladesh [52]. However, Sohel *et al.* (2009) stated that increased risks at low level exposure of As (50–149 µg/L) were observed for death due to cancers, cardiovascular disease, and infectious diseases in Bangladesh [53]. However, the HQ for Fe (children; 0.108) and Mn (children; 0.910) are relatively very lower compared to NO<sub>3</sub><sup>-</sup> and As in the study area. Though the mean concentration of Fe and Mn is quite high but the calculated HQ is within the safe limit for both adult and children in the study area. Thus the study shows that the shallow aquifer groundwater of central is polluted by NO<sub>3</sub><sup>-</sup> and As with substantial human health risk.

## 5. Conclusions

The outcome of the study depicts the human health risk of NO<sub>3</sub><sup>-</sup> and As by analyzing 99 ground water samples from the Central Bangladesh. The elevated occurrence of NO<sub>3</sub><sup>-</sup> and As is predominant within the 10 m to 40 m depth zone. Furthermore, HQ values of NO<sub>3</sub><sup>-</sup> and As are the dominant contributors in HI for the non-carcinogenic health risk for both adult and children. Additionally, carcinogenic risk due to As is also high in the study area. Finally, the shallow tube-well is not safe for drinking water collection in the study area. However, further study warrants considering wider area with spatial and temporal variability to predict health risk more precisely including the human level direct detection of the NO<sub>3</sub><sup>-</sup> and As.

**Author Contributions:** “Conceptualization, M.M.R. and M.B.D.; Methodology M.M.R. M.B.D. and M.I.M.; Software, M.M.R. M.B.D. and M.I.M; Validation, M.M.R. M.B.D. and M.K.U.; Formal Analysis, M.M.R. M.B.D. and M.I.M; Investigation, M.M.R. K.F.B.H. M.K.U. and M.I.M; Data Curation, M.M.R. M.S. and M.B.D.; Writing-

Original Draft Preparation, M.M.R. M.S.H. and K.F.B.H; Writing-Review & Editing, M.M.R. M.T.S. M.S.H. and S.A.; Visualization, M.M.R. S.A. and M.S.; Supervision, M.M.R. M.T.S. and M.K.U.

**Conflicts of Interest:** The authors declare no conflict of interest.

## References

- [1]. Sharma, C.; Mahajan, A.; Garg, U.K. Fluoride and nitrated in ground water of south-western Punjab, India- occurrence, distribution and statistical analysis. *Desalination Water Treat.* **2014**, *57*, 1-12.
- [2]. Rahman, M.M.; Sultana, R.; Shammi, M.; Bikash, J.; Ahmed, T.; Maruo, M.; Kurasaki, M.; Uddin, M.K. Assessment of the status of groundwater arsenic at SingairUpazila, Manikganj, Bangladesh; Exploring the correlation with the other metals and ions. *Expo. Health* **2016**, *8*, 217–225.
- [3]. Anawar, H.M.; Akai, J.; Mostofa, K.M.G.; Safiullah, S., Tareq, S.M. Arsenic poisoning in groundwater Health risk and geochemical sources in Bangladesh. *Environ. Int.* **2002**, *27*, 597– 604.
- [4]. Ritter, L.; Solomon, K.; Sibley, P.; Hall, K.; Keen, P.; Mattu, G.; Linton, B. Sources, pathways and relative risks of contaminants in surface water and ground water: a perspective prepared for the Walkerton Inquiry. *J. Toxicol. Environ. Health* **2002**, *65*, 1-142.
- [5]. Buchet, J.P.; Lison, D. Clues and Uncertainties in the risk assessment of arsenic in drinking water. *Food Chem. Toxicol.* **2000**, *38*, S81-S85.
- [6]. Kavcar, P.; Sofuoglu, A.; Sofuohlu, S.C. A health risk assessment for exposure to trace metals via drinking water ingestion pathway. *Int. J. Hyg. Environ. Health* **2009**, *212*, 216-227.
- [7]. Muhib, M.I.; Chowdhury, M.A.J.; Easha, N.J., Rahman, M.M, Shammi, M.; Fardous, Z.; Bari, M.L.; Uddin, M.K.; Kurasaki, M.; Alam, M.K. Investigation of heavy metal contents in Cow milk samples from area of Dhaka, Bangladesh. *Int. J. Food Contam.* **2016**, *3*, 16.
- [8]. Wasserman A.G.;Liu, X.; Parvez, F.; Ahsan, H.; Levy, D.; Factor-Lltvak, P.; Kline, J.; Geen, A.V.; Slavkovich, V.;Lolacono, N.G.; Cheng, Z.; Zheng, Y.; Graziano, J.H. Water Manganese Exposure and Children's Intellectual Function In Araihasar, Bangladesh. *Environ. Earth Perspec.* **2006**, *114*, 124-129.
- [9]. Merrill, R.D.; Shamim, A.A.; Ali, H.; Jahan, N.; Labrique, A.B.; Schulze, K.; Christian, P.; West Jr, K.P. Iron status of women is associated with the iron concentration of potable groundwater in rural Bangladesh. *J. Nutr.* **2011**, *141*, 944–9.
- [10]. Heming, N.; Montravers, P.; Lasocki, S. Iron deficiency in critically ill patients:highlighting the role of hepcidin. *Crit. Care.* **2011**, *15*, 210.
- [11]. El-Midaoui, A.; Elhannouni, F.; Taky, M.; Chay, L.; Sahli, M.A.M.; Echihabi, L.; Hafsi, M. Optimization of nitrate removal operation from ground water by electro dialysis. *Sep. Purif. Technol.* **2002**, *29*, 235– 244.
- [12]. Pannala, A.S.; Mani, A.R.; Spencer, J.P.E.; Skinner, V.; Bruckdorfer, K.R.; Moore, K.P.; Rice-Evans, C.A. The effect of dietary nitrate on salivary, plasma, and urinary nitrate metabolism in humans. *Free Radic. Biol. Med.* **2003**, *34*, 576–584.
- [13]. Lohumi, N.; Gosain, S.; Jain, A.; Gupta, V.K.; Verma, K.K. Determination of nitrate in environmental water samples by conversion into nitrophenols and solid phase extraction-spectrophotometry, liquid chromatography or gas chromatography–mass spectrometry, *Anal. Chim. Acta* **2004**, *505*, 231–237.
- [14]. Fewtrell, L. Drinking-water nitrate, methemoglobinemia, and global burden of disease: A discussion. *Environ. Health Perspect.* **2004**, *112*, 1371–1374.
- [15]. Greer, F.R.; Shannon, M. Infant methemoglobinemia: The role of dietary nitrate in food and water. *Paediatrics* **2005**, *116*, 784–786.
- [16]. Ward, M.H.; DeKok, T.M.; Levallois, P.; Brender, J.; Gulis, G.; Nolan, B.T.; Derslice, J.V. Workgroup report: Drinking-water nitrate and health—Recent findings and research needs, *Environ. Health Perspect.* **2005**, *113*, 1607–1614.
- [17]. Rachid, A.; Christophe, M.; Marc, B.; Laure, O.; Sylvie, T.; Paul, P. Methemoglobinemia by cerium nitrate poisoning, *Burns* **2006**, *32*, 1060–1061.
- [18]. Ma, H.W.; Hung, M.L.; Chen, P.C. A systemic health risk assessment for chromium cycle in Taiwan. *Environ. Int.* **2007**, *33*, 206-218.
- [19]. Momot, O.; Synzynyns, B. Toxic aluminium and heavy metals in ground wate of middle Russis: health risk assessment. *Int. J. Environ. Res. Pub. Health* **2005**, *2*, 214-218.
- [20]. Wu, B.; Zhang, Y.; Zhang, X.; Cheng, S. Health risk from exposure of organic pollutants through drinking water consumption in Nanjing, China. *Bull. Environ. Contam. Toxicol.* **2010**, *84*, 46-50.

- [21]. Bortey-Sam, N.; Nakayama, S.M.; Ikenaka, Y.; Akoto, O.; Baidoo, E.; Mizukawa, H.; Ishizuka, M. Health risk assessment of heavy metals and metalloid in drinking water from communities near gold mines in Tarkwa, Ghana. *Environ. Monit. Assess.* **2015**, *187*, 397. DOI 10.1007/s10661-015-4630-3
- [22]. USEPA (US Environmental Protection Agency). National primary/secondary and drinking water regulations **2009**. Washington, D.C.
- [23]. Lim, H.S.; Lee, J.S.; Chon, H.T.; Sager, M. Heavy metal contamination and health risk assessment in the vicinity of the abandoned Songcheon Au–Ag mine in Korea. *J. Geochem. Explor.* **2008**, *96*, 223–230.
- [24]. Ning, L.; Ni, T.; Xia, J.; Dai, M.; He, C.; Lu, G. Non-carcinogenic risks induced by metals in drinking source water of Jiangsu Province, China. *Environ. Monit. Assess.* **2011**, *177*, 449–456. DOI 10.1007/s10661-010-1646-6.
- [25]. USEPA (US Environmental Protection Agency). Risk Assessment Guidance for Superfund Volume I Human Health Evaluation Manual (Part A). United States Environmental Protection Agency **1989**, Washington, D.C.
- [26]. Giri, S.; Singh, A.K. Human health risk assessment via drinking water pathway due to metal contamination in the ground water of Subarnarekha river basin, India. *Environ. Monit. Assess.* **2015**, *187*, 63.
- [27]. Wongsasuluk, P.; Chotpantarat, S.; Siritwong, W.; Robson, M. Heavy metal contamination and human health risk assessment in drinking water from shallow groundwater wells in an agricultural area in Ubon Ratchathani province, Thailand. *Environ. Geochem. Health* **2013**, *36*, 169–182. DOI.10.1007/s10653-013-9537-8
- [28]. USEPA (US Environmental Protection Agency). Risk assessment guidance for superfund (RAGS), volume I: human health evaluation manual (part E) interim. United States Environmental Protection Agency **1993**, Washington, D.C.
- [29]. ECETOC (European Center for Ecotoxicology of Chemical). Aquatic Toxicity of Mixtures. Technical Report. **2001**, 80, Brussels.
- [30]. USEPA (US Environmental Protection Agency). Baseline human health risk assessment, **2001**. Vasquez Boulevard and 1-70 superfund site, Denver CO.
- [31]. USEPA (US Environmental Protection Agency). A risk assessment-multiway exposure spreadsheet calculation tool. United States Environmental Protection Agency, **1999**. Washington, D.C.
- [32]. Weyer, P.J.; Cerhan, J.R.; Kross, B.C.; Hallberg, G.R.; Kantamneni, J.; Breuer, G.; Jones, M.P.; Zheng, W.; Lynch, C.F. Municipal drinking water nitrate level and cancer risk in older women: the Iowa women’s health study. *Epidemiology* **2001**, *11*, 3.
- [33]. Kavcara, P.; Sofuoglu, A.; Sofuoglu, S.C. A health risk assessment for exposure to trace metals via drinking water ingestion pathway. *Int. J. Hyg. Environ. Health* **2009**, *212*, 216–227.
- [34]. Tahri, M.; Beny’Ich, F.; Bounakhla, M.; Bilal, E.; Gruffat, J.J.; Moutte, J. Multivariate analysis of heavy metal contents in soil, sediment and water in the region of Meknes (Central Morocco). *Environ. Monit. Assess.* **2005**, *102*, 405–417.
- [35]. Kaiser, H.F. The application of electronic computers to factor analysis. *Educ. Psychol. Meas.* **1960**, *20*, 141–151.
- [36]. Sikder, M.T.; Kihara, Y.; Yasuda, M.; Yustiawati,; Mihara, Y.; Tanaka, S.; Odgerel, D.; Mijiddorj, B.; Syawal, S.M.; Hosokawa, T.; Saito, T.; Kurasaki, M. River Water Pollution in Developed and Developing Countries: Judge and Assessment of Physicochemical Characteristics and Selected Dissolved Metal Concentration, Clean- soil air and water **2013**, *41*, 60–68.
- [37]. Kundu, M.C.; Mandal, B.; Sarkar, D. Assessment of the potential hazardous of nitrate contamination in surface and groundwater in a heavily fertilized and intensively cultivated district of India. *Environ. Monit. Assess.* **2008**, *146*, 183–189
- [38]. Suthara, S.; Bishnoib, P.; Singh, S.; Mutiyara, P.K.; Nema, A.K.; Patil, N.S. Nitrate contamination in groundwater of some rural areas of Rajasthan, India. *J. Hazard. Mater.* **2009**, *171*, 189–199
- [39]. Kundu, M.C.; Mandal, B. Agriculture activities influence nitrate and fluoride contamination in drinking groundwater of an intensively cultivated district in India, *Water air Soil Pollut.* **2008**. doi:10.1007/s1270-008-9842-5
- [40]. Wakida, F.D.; Lerner, D.N. Non-agriculture sources of groundwater nitrogen: a review and case study. *Water Res.* **2005**, *39*, 3–16.
- [41]. Reddy, A.G.S.; Kumar, K.N.; Rao, D.S.; Rao, S.S. Assessment of nitrate contamination due to groundwater pollution in north eastern part of Anantapur District, A.P., India, *Environ. Monit. Assess.* **2009**, *148*, 463–476.

- [42]. Walker, R. Nitrates nitrites and N-nitroso compounds: a review of the occurrence in food and diet and the toxicological implications. *Food Addit. Contam.* **1990**, *7*, 717–768.
- [43]. Tricker, A.R.; Preussmann, R. Carcinogenic N-nitrosamines in the diet: occurrence, formation, mechanisms and carcinogenic potential. *Mutat. Res.* **1991**, *259*, 277–289.
- [44]. Majumder, R.K.; Hasnat, M.A.; Hossain, S.; Ikeue, K.; Machida, M. An exploration of nitrate concentrations in groundwater aquifers of central-west region of Bangladesh. *J. Hazard. Mater.* **2008**, *159*, 536–543
- [45]. Dellavalle, C.T.; Xiao, Q.; Yang, G.; Shu, X.O.; Aschebrook-Kilfoy, B.; Zheng, W.; Lan, L.H.; Ji, B.T.; Rothman, N.; Chow, W.H.; Gao, Y.T.; Ward, M.H. Dietary nitrate and nitrite intake and risk of colorectal cancer in the Shanghai Women's Health Study. *Int J Cancer.* **2014**, *134*(12), 2917–2926. Doi: 10.1002/ijc.28612.
- [46]. Zhai, Y.; Zhao, X.; Teng, Y.; Li, X.; Zhang, J.; Wu, J.; Zuo, R. Groundwater nitrate pollution and human health risk assessment by using HHRA model in an agricultural area, NE China. *Ecotoxicol. Environ. Saf.* **2017**, *137*, 130–142
- [47]. Li, P.; Wu, J.; Qian, H.; Lyu, X.; Liu, H. Origin and assessment of groundwater pollution and associated health risk: a case study in an industrial park, northwest China. *Environ. Geochem. Health*, **2013**.
- [48]. Smith, L. A.; Means, J. L.; Chen A.; et al., *Remedial Options for Metals-Contaminated Sites*, Lewis Publishers, Boca Raton, Fla, USA,, **1995**.
- [49]. Bodek, I.; Lyman, W. J.; Reehl, W. F.; Rosenblatt, D. H. *Environmental Inorganic Chemistry: Properties, Processes and Estimation Methods*, Pergamon Press, Elmsford, NY, USA, **1988**
- [50]. Rahman, M.M.; Islam, M.A.; Bodrud-Doza, M.; Muhib, M.I.; Zahid, A.; Shammi, M.; Tareq, S.M.; Kurasaki, M. Spatio-temporal assessment of groundwater quality and human health risk: a case study in Gopalganj, Bangladesh. *Expo. Health* **2017**, DOI.10.1007/s12403-017-0253-y
- [51]. Chen, Y.; Graziano, J.H.; Parvez, F.; Liu, M.; Slavkovich, V.; Kalra, T.; Argos, M.; Islam, T.; Ahmed, A.; Rakibuz-Zaman, M.; Hasan, R.; Sarwar, G.; Levy, D.; Geen, A.V.; Ahsan, H. Arsenic exposure from drinking water and mortality from cardiovascular disease in Bangladesh: prospective cohort study. *BMJ* **2011**, *342*, d2431
- [52]. Das, B.; Rahman, M.M.; Nayak, B.; Pal, A.; Chowdhury, U.K.; Mukherjee, S.C.; Saha, K.C.; Pati, S.; Quamruzzaman, Q.; Chakraborti, D. Groundwater arsenic contamination, its health effects and approach for mitigation in West Bengal, India and Bangladesh. *Water Qual. Expo. Health* **2009**, *1*, 5–21. DOI.10.1007/s12403-008-0002-3
- [53]. Sohel, N.; Persson, L.A.; Rahman, M.; Streatfield, P.K.; Yunus, M.; Ekstroöm, E.C.; Vahter, M. Arsenic in Drinking Water and Adult Mortality A Population-based Cohort Study in Rural Bangladesh. *Epidemiology* **2009**, *20*, 824–830