

1 Article

# 2 Phosphorus Budget for a Forested-Agricultural 3 Watershed Influenced by East-Asian Summer 4 Monsoon

5 Arif Reza<sup>1,2,3,#</sup>, Jaesung Eum<sup>1,#</sup>, Sungmin Jung<sup>1</sup>, Youngsoon Choi<sup>1</sup>, Changwon Jang<sup>1</sup>, Kiyong Kim<sup>1</sup>,  
6 Jeffrey S. Owen<sup>4</sup> and Bomchul Kim<sup>1,\*</sup>

7 <sup>1</sup> Department of Environmental Science, Kangwon National University, Chuncheon 24341, South Korea;  
8 bd.rezaarif@gmail.com (A.R.); jseum@kangwon.ac.kr (J.E.); jungsm@kangwon.ac.kr (S.J.);  
9 youngsoonchoi@kangwon.ac.kr (Y.C.); jcw0110@kangwon.ac.kr (C.J.); kiyong@kangwon.ac.kr (K.K)

10 <sup>2</sup> Department of Environmental Science, College of Agricultural Sciences, International University of  
11 Business Agriculture and Technology, Dhaka 1230, Bangladesh;

12 <sup>3</sup> Department of Animal Industry Convergence, Kangwon National University, Chuncheon 24341, South  
13 Korea;

14 <sup>4</sup> Department of Environmental Science, Hankuk University of Foreign Studies, Yongin 17035, South Korea;  
15 jeffreyscottowen@gmail.com (J.S.O.)

16 \* Correspondence: bkim@kangwon.ac.kr; Tel.: +82-33-252-4443

17 # These authors contributed equally to this work.

18

19 **Abstract:** Despite increased awareness of and attention to the need for sustainable agriculture,  
20 fertilizers and compost application in excess of crop requirements remain common agricultural  
21 practices in South Korea, causing eutrophication of freshwater and coastal ecosystems. In this study,  
22 a phosphorus (P) budget was developed to quantify P inputs, outputs, and retention in a forested-  
23 agricultural watershed. The P budget showed that chemical fertilizers and organic compost were  
24 the largest source of P (97.6% of the total) followed by atmospheric deposition (2.1% of the total P),  
25 whereas forest export (0.2% of the total) and sewage treatment plants (STPs) (0.1% of the total) were  
26 negligible. The dominant P outputs were crop harvesting and hydrologic export to surface water.  
27 The P balance showed a significant accumulation of P in the watershed; approximately 87% of the  
28 total P input was retained in the soils within the watershed. However, P concentrations in drainage  
29 water were still high enough to cause eutrophication of downstream reservoirs. The results provide  
30 useful information on the proportion of P export and retention in soils and will help support efforts  
31 to improve water quality and design better management of agricultural non-point source pollution.

32 **Keywords:** Phosphorus; nutrient budget; non-point source pollution; Korea; monsoon climate

33

## 34 1. Introduction

35 Anthropogenic impacts on the global phosphorus (P) cycle have been a major research topic in  
36 the field of P biogeochemistry, especially in agricultural ecosystems. Because of the increase in P  
37 fertilizer use during the last few decades, global P inputs from terrestrial to aquatic ecosystems have  
38 doubled and eutrophication remains a global issue [1,2].

39 Eutrophication is a serious environmental challenge in Korea and P export from agricultural  
40 fields has been identified as major contributor to non-point source (NPS) nutrient pollution [3]. The  
41 characteristic intense rainfall during the East-Asian summer monsoon dramatically influences the  
42 export of NPS derived nutrients from highland agricultural croplands in Korea [4]. Some studies have  
43 shown that P export from agricultural fields in Korea is higher than in other regions [3,4]. To ensure  
44 short-term productivity as well as long-term sustainability, minimizing P losses and achieving a  
45 balance between P inputs and P outputs in agricultural systems is important. P budget studies can

46 help to identify the sources and fates of P, which are important for regional water quality issues.  
 47 Improved knowledge of watershed P budgets can also help evaluate the impacts of agricultural  
 48 practices on P accumulation and losses.

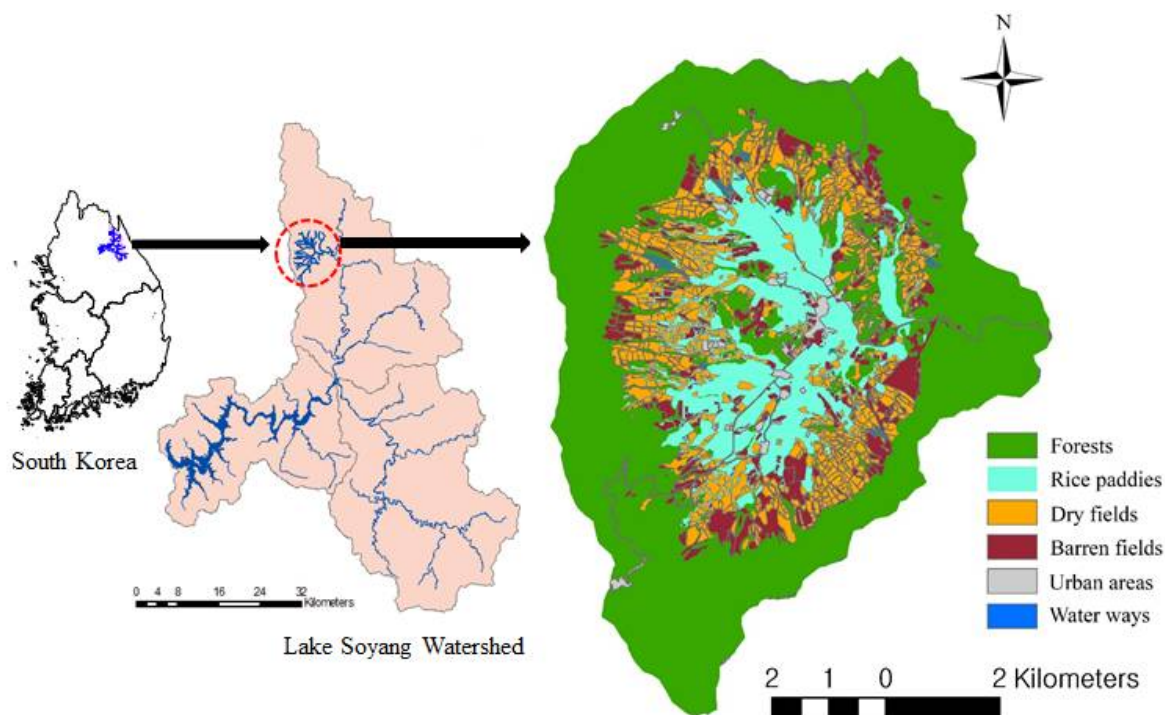
49 Using a mass balance approach to calculate P budgets is a commonly used approach in  
 50 agricultural systems [5-8]. Continued accumulation of P in agricultural land beyond crop  
 51 requirements may well become a long-term concern in Korea. Although there are many watersheds  
 52 with mixed (forested and agricultural) land use distributed throughout most areas of Korea [9], no  
 53 research has focused on constructing P budgets for these mixed land use watersheds. The current  
 54 study was conducted in the Haeon watershed, a sub-watershed located at the upper reaches of Lake  
 55 Soyang (the largest reservoir in South Korea). The study area is an advantageous location for  
 56 studying the watershed P budget because it has only a single outlet (Mandae Stream). Previous  
 57 studies on NPS of sediment and nutrient have conducted in the Haeon watershed and the watershed  
 58 is known as a hot spot for soil erosion and export of NPS pollution to Lake Soyang [4].

59 The aim of this study was to construct a P budget for the 2013 growing season to quantify P  
 60 inputs, outputs, and retention for the Haeon watershed.

## 61 2. Materials and Methods

### 62 2.1. Study site and land use

63 The Haeon watershed (longitude 128°50' to 128°11'E and latitude 38°13' to 38°20'N) is located in  
 64 the Gangwon Province in the eastern region of Korea (Figure 1). The physical characteristics of the  
 65 study watershed are shown in Table 1. The hydrologic characteristics of the watershed are strongly  
 66 influenced by the East-Asian summer monsoon climate. Nearly 70% of the annual precipitation  
 67 occurs in July and August [10]. Approximately 90% of the annual rainfall occurs during the April to  
 68 October cropping season [11]. The 10 year (2004-2013) mean annual temperature in the Haeon  
 69 watershed is 10.9°C with a mean annual precipitation of 1506 mm.



70

71 **Figure 1.** Location of the Haeon watershed in eastern Korea and map showing the land use in the  
 72 mixed-use watershed.

73

**Table 1.** Characteristics of the study watershed.

Watershed	Area (ha)	Population density (person/ha)	Elevation (m)	Slope (°)	% Coverage				
					Forest	Crop	Urban	Other <sup>1</sup>	Total
Haean	6,174	0.23	340-1320	5-20	61.1	29.6	0.62	8.68	100

74

<sup>1</sup>Barren fields and water ways

75 The soils in the watershed are classified as Cambisols but most agricultural soils are Anthrosols  
 76 due to the long-term use of soil amendments with sandy soil as a top dressing [12]. The study area is  
 77 one of the largest highland agricultural areas in the Lake Soyang watershed [10]. Generally, in  
 78 highland agricultural fields high value horticultural row crops such as potato, radish, cabbage and  
 79 soybean are preferred over other crops [9]. In the study area, rice fields account for more than 30% of  
 80 the agricultural land, while dryland crops (cabbage, potato, radish, soybean) account for  
 81 approximately 52% of the agricultural land (Table 2) [13].

82

**Table 2.** Agricultural land use in the Haean watershed (2012) [13].

Croplands	Area (ha)	Percent of the total (%)
<i>Annual</i>		
Rice paddy	569	31.1
Cabbage	142	7.77
Potato	225	12.3
Radish	213	11.7
Soybean	163	8.92
Maize	52	2.84
Pumpkin	34	1.86
<i>Perennial</i>		
Orchard	85	4.65
Ginseng	82	4.49
Codonopsis	28	1.53
Peach	19	1.04
Grape	18	0.98
Others	198	10.8
Total	1828	100

83

A summary of local agricultural practices in the Haean watershed is shown in Table 3 [14].  
 84 Conventional tillage and seedbed preparation without surface cover are common agricultural  
 85 practices in the Haean watershed. Some crops are planted with plastic sheets used as a mulch.  
 86

87 **Table 3.** Local agricultural practices for the five most common crops accounting for about 72% of the total  
 88 agricultural area in the Haean watershed and comparison between recommended fertilizer application rates by  
 89 RDA (Rural Development Administration) and fertilizer applied by farmers [14].

Crop	Cultivation period	Number of applications	Fertilizer application rates (kg P/ha/yr)		
			Types	RDA Recommended	Applied
<i>Rice paddy</i>	Mid May – End October	1-2	Chemical	20	78
			Compost	-	30
<i>Cabbage</i>	End July – End October	1-2	Chemical	26	64
			Compost	19	83
<i>Potato</i>	End April – Mid August	1	Chemical	14	55
			Compost	13	103
<i>Radish</i>	End July – End September	1-3	Chemical	13	37
			Compost	19	91
<i>Soybean</i>	End May – End October	1	Chemical	13	51
			Compost	19	92

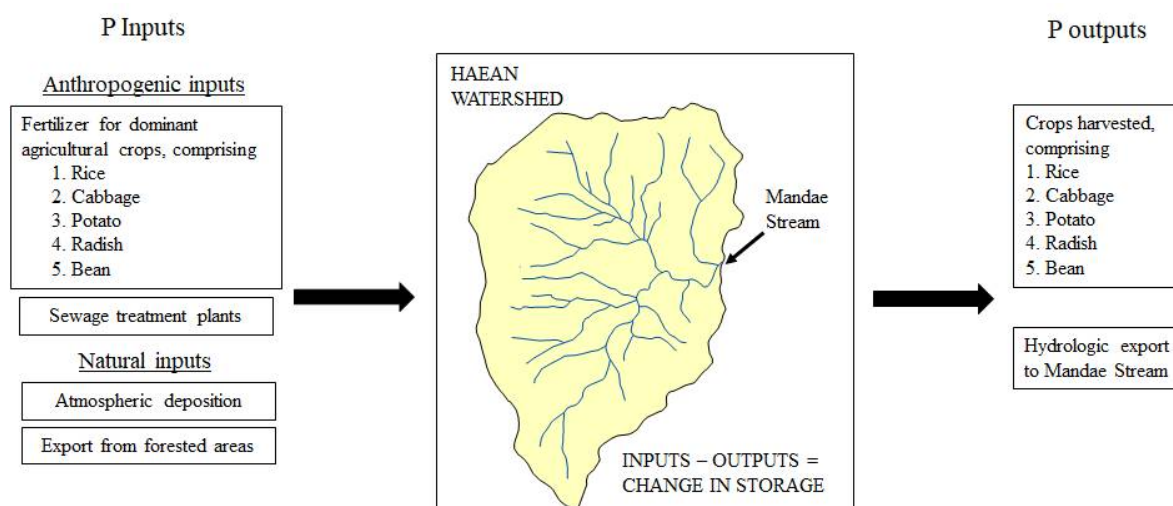
90

## 91 2.2. Sampling and chemical analysis

92 Crop (5-8 replicates from 1m<sup>2</sup> area for each crop) and soil (3-5 samples from 0-20 cm depth)  
 93 samples were collected from randomly selected crop fields within the watershed at the time of final  
 94 harvest. Immediately after separating the plant parts, the fresh weight (FW) of leaves, stems, roots,  
 95 and grains were measured. Dry matter (DM) of plant parts was measured after oven drying at 80°C  
 96 for 72 hours. A finely ground aliquot of each sample was analyzed for total phosphorus (TP)  
 97 concentration [15]. Soil P concentration was measured using the modified Lancaster method [16].  
 98 Rainfall data were collected from a weather station maintained by the Korean Meteorological  
 99 Administration (KMA). Twelve rainfall samples were collected using tipping buckets near a long-  
 100 term monitoring site in a forested sub-watershed during the study period. TP concentration in rain  
 101 water samples was measured using the ascorbic acid method [17].

## 102 2.3. Budget methodology

103 A mass balance approach was used to compare P inputs and P outputs and estimate annual P  
 104 storage in the watershed in 2013 (Figure 2). To develop a more realistic budget, information on local  
 105 agricultural practices and crop production were collected using a farm census. A total of 52 farmers  
 106 involved in cultivating the major crops in the study watershed were interviewed during the study  
 107 period. Data from the results of the farm census together with data from field and literature values  
 108 were used to develop the P budget. Conceptually, P inputs to the Haean watershed were separated  
 109 into two groups considered as anthropogenic (fertilizer application, sewage treatment plants) and  
 110 background sources (atmospheric deposition, forest loading). P outputs from the Haean watershed  
 111 were calculated using data on P concentrations from the crop harvest and hydrologic export to  
 112 surface water.



113  
114 **Figure 2.** Schematic diagram of P inputs and outputs used to calculate the phosphorus budget for the Haeon  
115 watershed.

### 116 2.3.1. P inputs to the Haeon Watershed

- 117 • Inputs from fertilizer application

118 Interviews with farmers were conducted to collect information on fertilizer (chemical and organic  
119 compost) application rates. Fertilizer P content was obtained from data provided by the RDA and  
120 from previous studies [14, 18]. The total amount of P in fertilizer used in the Haeon watershed was  
121 calculated by multiplying the rate of fertilizer application by the fertilizer P content and the  
122 agricultural land area for each crop and expressed as:

$$123 \quad P_{\text{Fert}} = FA \times C_{\text{Fert}} \times A \quad (1)$$

124 where  $P_{\text{Fert}}$  is the P input from the fertilizer (kg/yr); FA is the rate of fertilizer application (kg/ha/yr),  
125  $C_{\text{Fert}}$  is the P content in fertilizer (chemical and organic compost) (%) and A is the agricultural land  
126 area for each crop (ha).

- 127 • Inputs from sewage treatment plants

128 Sewage treatment plants (STPs) are an obvious point source of P inputs. Though most of the  
129 study area is rural, three small STPs discharge treated wastewater in the Haeon watershed. The  
130 outflow and TP concentrations were obtained from the managing authority [13]. The amount of P  
131 from STPs was calculated using the following equation:

$$132 \quad P_{\text{STP}} = C_{\text{STP}} \times FR \quad (2)$$

133 where  $P_{\text{STP}}$  is the P input from the STPs (kg/yr);  $C_{\text{STP}}$  is the average P concentration (mg/L) and FR is  
134 the average annual flowrate of the effluent ( $\text{m}^3/\text{yr}$ ).

- 135 • Inputs from atmospheric deposition

136 In regions with sufficient rainfall, P inputs in through wet deposition is more important than  
137 dry deposition of P [19]. Therefore, P input from dry deposition was not considered in this study;  
138 input in wet deposition was estimated by the following approach [6]:

$$139 \quad P_{\text{Rain}} = A \sum C_{\text{Rain}} \times V_{\text{Rain}} \quad (3)$$

140 where  $P_{\text{rain}}$  is P input from wet deposition (kg/yr),  $A$  is the watershed area (ha),  $C_{\text{rain}}$  is the average P  
141 concentration in wet deposition during the study period (mg/L) and  $V_{\text{rain}}$  is the 10 year (2004-2013)  
142 mean annual rainfall (mm/yr).

- 143 • Inputs from forested areas

144 The P input from forested land was considered as a background source. Due to the widely  
145 different sampling frequency for rain event and dry period samples, inputs from forested areas were  
146 estimated by summing the inputs during sampled and non-sampled rain events and dry period [20].

### 147 2.3.2. P outputs from the Haean Watershed

- 148 • Removal through crop harvesting

149 Crop production data were obtained from a local farm census. P removal from the Haean  
150 watershed was calculated by multiplying the measured P content in crop samples and the crop  
151 harvest estimated for each crop and expressed as:

$$152 \quad P_{\text{CH}} = CP \times C_{\text{CH}} \times A \quad (4)$$

153 where  $P_{\text{CH}}$  is the P output through crop harvesting (kg/yr);  $CP$  is the crop production (kg/ha/yr),  $C_{\text{CH}}$   
154 is the P content (%) and  $A$  is the agricultural land area for each crop (ha).

- 155 • Hydrologic export to Mandae Stream

156 To obtain an estimate of P export in surface water to Mandae Stream (the watershed outlet), the  
157 hydrologic export from an earlier intensive monitoring study conducted by the authors (2009-2013)  
158 was used [4].

## 159 3. Results

### 160 3.1. P inputs

#### 161 3.1.1. Fertilizer application

162 The calculated annual P input from chemical fertilizer was 81,779 kg P/yr (47.8% of total P  
163 inputs). Organic compost supplied 86,640 kg P/yr (50.7% of total P inputs) to the Haean watershed.  
164 Among the five most common crops, fertilizer used in rice fields was found to be the largest  
165 contributor of P in the watershed followed by potato, radish, soybean and cabbage (Table 4). The P  
166 input from fertilizer use for rice, potato, radish, soybean, and cabbage was 61,447, 35,461, 27,297,  
167 23,347, and 20,867 kg P/yr, respectively.

168

169

170

171

172

173

174

175

176

177 **Table 4.** P inputs from fertilizer (chemical and organic matter) for the five most common crops in the Haean  
178 watershed.

Crop	Fertilizer input (kg P/yr)		P content (%) <sup>1</sup>		P input (kg P/yr)		
	Chemical	Compost	Chemical	Compost	Chemical	Compost	Total
Rice	441,981	1,437,443	10		44,198	17,249	61,447
Cabbage	303,188	980,915	3		9,096	11,771	20,867
Potato	308,306	1,927,391	4	1.20	12,332	23,129	35,461
Radish	261,072	1,622,053	3		7,832	19,465	27,297
Soybean	64,005	1,252,168	13		8,321	15,026	23,347
		Total			81,779	86,640	168,418

179 <sup>1</sup>P content was calculated from values in published sources [14, 18]. Pig manure is the main raw material of  
180 compost used in the study watershed; rice hulls, crushed wood and saw dust were used as bedding materials.

### 181 3.1.2. Fertilizer application

182 P input from STPs accounted for a small fraction of the total P input to the Haean watershed.  
183 With the limited human population and little industrial development in the study area, STPs  
184 accounted for only 196 kg P/yr (0.1% of total P inputs) in the Haean watershed (Table 5).

185 Table 5. Summary of phosphorus inputs in the Haean watershed.

Phosphorus inputs	Amount (kg/yr)	Percent of the total (%)
<i>Anthropogenic</i>		
Chemical Fertilizer	81,779	47.4
Organic matter	86,640	50.2
Sewage treatment plants	196	0.1
<i>Natural</i>		
Atmospheric deposition	3,719	2.1
Export from forested areas	274	0.2
Total	172,607	100

### 186 3.1.3. Atmospheric deposition

187 The mean P concentration of the wet deposition samples was 0.04 mg/L. In 2013, the annual  
188 rainfall was 713 mm, less than half of the 10 years (2004–2013) average annual rainfall (1,506 mm) in  
189 the study area. Therefore, the annual P input from wet deposition was only 3,719 kg P/yr (Table 5).  
190 Atmospheric inputs accounted for about 2.1% of total P inputs.

### 191 3.1.4. Export from forested areas

192 Though forest land covers more than half of the Haean watershed, the P input from surface  
193 runoff was small. The calculated amount of P from surface runoff from forest areas was 274 kg P/yr  
194 (0.2% of total input; Table 5) [20].

### 195 3.1.5. Summary of P inputs

196 Our inventory of the P inputs showed considerable variation in the P content among the five  
197 most common crops in the watershed (Table 5). The total P inputs to the watershed were  
198 approximately 172,607 kg/yr for the 2013 growing season. Agricultural P inputs in the form of organic

199 compost were the largest source of P, accounting for 50.2% of total P inputs. Chemical fertilizers were  
 200 the second largest P source, accounting for 47.4% of the total P inputs. P inputs from STPs added only  
 201 0.1% of all the P entering the watershed. Background P inputs such as wet atmospheric deposition  
 202 and export from forested area accounted for only 2.3% of the total P inputs. Overall, anthropogenic  
 203 P inputs (chemical fertilizer and organic compost applications) were the largest P inputs in the Haeon  
 204 watershed (Table 5).

### 205 3.2. P outputs

#### 206 3.2.1. P removal through crop harvesting

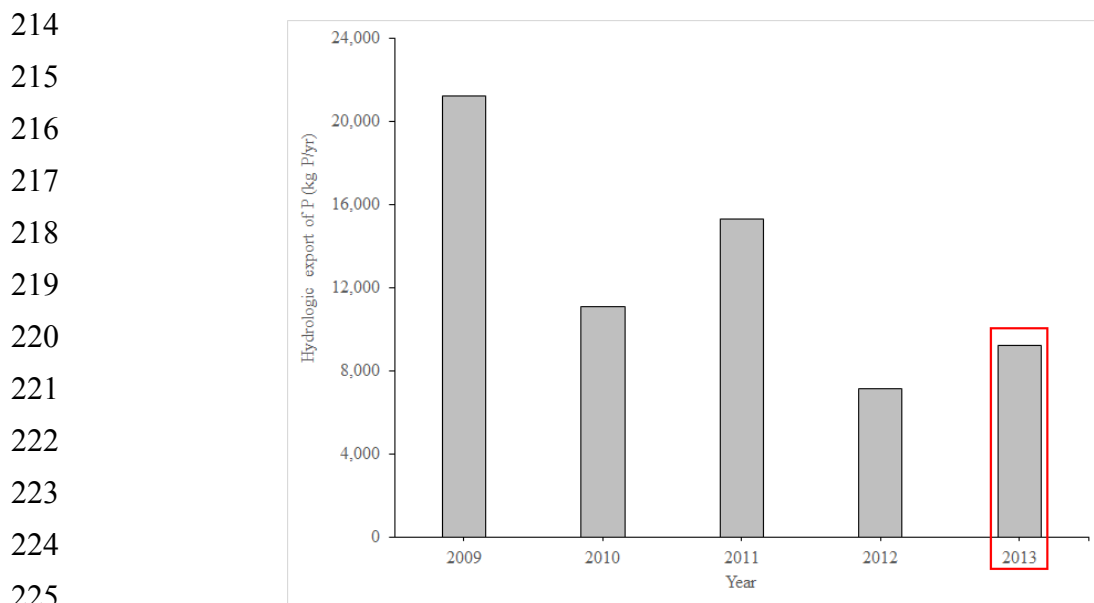
207 Crop production in the Haeon watershed is solely commercial. We assumed that all the harvested  
 208 crops left the watershed and none of the crops were consumed within the watershed. P output as  
 209 harvested crops was 13,216 kg P/yr, accounting for 58.9% of total P outputs (Table 6).

210 **Table 6.** Phosphorus outputs from the watershed through crop harvesting.

Crop	Yield (kg/yr)	Water content (%)	Dry weight (kg/yr)	P content (%)	Harvested P (kg P/yr)
Rice paddy	3,218,833	11.3	2,855,105	0.22	6,281
Cabbage	6,882,456	91.8	564,361	0.35	1,975
Potato	3,578,850	77.2	815,978	0.23	1,877
Radish	6,184,242	94.0	371,055	0.40	1,484
Soybean	261,126	10.0	235,013	0.68	1,598
Total					13,216

#### 211 3.2.1. Hydrologic export to surface water

212 Hydrologic export of P to the Mandae stream was estimated to be 9,215 kg/yr in 2013 (Figure 3;  
 213 Table 7) [4]. This P export in surface water accounted for 41.1% of total P outputs.



226 **Figure 3.** P output from the watershed in hydrologic export to the Mandae Stream (2009-2013)

227

228



229 **Table 7.** Summary of phosphorus outputs in the Haean watershed

Phosphorus outputs	Amount (kg/yr)	Fraction of total output (%)
Crop harvesting	13,216	58.9
Hydrologic export	9,215	41.1
Total	22,431	100

## 230 3.2.3. Summary of phosphorus outputs

231 We calculated a net export of P in harvested crops of 58.9 % of the total P output compared to  
 232 export in surface water of 41.1 % of the total P output (Table 7). Thus, P export through crop  
 233 harvesting was identified as the main pathway for P outputs.

## 234 3.3. Summary P budget

235 We found that fertilizer use (including chemical fertilizer and organic compost application) is  
 236 the largest P input to the Haean watershed. Secondary P inputs are wet deposition and surface runoff  
 237 from forested land. P input from STPs accounted for only a small portion of P inputs. Hydrologic  
 238 export in the watershed outlet and crop harvesting were the most important P outputs. Based on the  
 239 current study, a net positive P balance was indicated for the Haean watershed (Table 8). We found  
 240 that approximately 87% of the total P inputs were retained in the watershed (Table 8).

241 **Table 8.** Summary of the annual phosphorus budget of the Haean watershed

Phosphorus	Amount (kg/yr)
Inputs (I)	172,607
Outputs (O)	22,431
Net (I - O)	150,177
% Retained	87.0

242 **4. Discussion**

## 243 4.1. Fertilizer application and fate of P in soils

244 We have identified fertilizer use as the major P input in the Haean study watershed. Specifically,  
 245 organic compost was the biggest contributor to the total P input for most of the common crops except  
 246 rice. For the case of rice, chemical fertilizer was found to be the largest P input. Although the RDA  
 247 has not recommended any organic compost application in rice fields, most farmers in the Haean  
 248 watershed do apply organic compost in their rice fields. The estimated fertilizer consumption in  
 249 Korea is 3.6 times higher than the global average [21]. Furthermore, application rates in the study  
 250 watershed for chemical fertilizers and organic compost application rates have been 2.5 to 4 and 4.5 to  
 251 8 times higher than the recommended application rates (kg/ha/yr) by the RDA, respectively.

252 As part of this study, our interviews with local farmers showed that fertilizer application rates  
 253 for rice, cabbage, potato, radish, and soybean surpassed the RDA recommended P fertilizer  
 254 application rates by 5.7, 3.4, 6, 4 and 4.6 times, respectively. Farmers in the Haean watershed generally  
 255 applied all the P fertilizer in a single dose at the beginning of the growing season and did not use  
 256 additional applications of P fertilizer later in the year. At the beginning of the growing season, plant  
 257 growth is thought to be slow with relatively low P uptake. Thus, excess P fertilizer application early  
 258 in the growing season combined with substantial rainfall could result in enhanced loss of P from  
 259 soils. Therefore, splitting P fertilizer applications into two or three separate fertilizer applications  
 260 based on crop P requirement could be a useful measure to reduce harmful P losses. While annual  
 261 chemical fertilizer consumption in the Haean watershed has been decreasing in recent years [13],

262 there are still considerable opportunities to reduce P fertilizer inputs and eventually decrease P  
263 surpluses in the watershed.

264 The detailed inventory of P inputs and outputs in this study has shown that P inputs are higher  
265 than surpass the outputs (Table 8). Approximately 150,177 kg P/yr (Table 8) is retained in soils in the  
266 Haean watershed, signifying the potential for contributing to eutrophication and leaching into  
267 ground water. This finding is further supported by the increase in soil P concentration in the  
268 watershed. An increase in average soil test P concentrations in the dry fields from 76.3 mg/kg  
269 (estimated in 2009) [22] to 95.9 mg/kg (this study) suggests that the excess P is being held in the soil.  
270 Based on the calculation method used by Bennet et al [1], if farmers in the watershed immediately  
271 stop applying excess P fertilizer and apply the same amount of P that is exported from the watershed  
272 (assuming that agricultural production and P export will not decline with reduced fertilizer  
273 application), more than 22 years would be required for P in the soil to reach equilibrium P  
274 concentrations. Therefore, achieving reductions in soil P concentrations in the Haean watershed is a  
275 goal with long-term benefits.

276 P sequestered in soil is often considered a 'chemical time bomb' [23]. Excessive P stored in soils  
277 can result in increased water-soluble P concentrations in soils, thus amplifying the potential for  
278 hydrologic export [24]. The general soil texture of soils in the Haean watershed is sandy loam; these  
279 soils are susceptible to erosion during the summer monsoon season [25]. Related to this, observations  
280 from other monitoring studies show that most (>83%) of the P exported from the Haean watershed  
281 in surface water is in particulate form [26]. Even if P fertilizer applications are reduced, soil P will  
282 continue to be available for hydrologic export to Lake Soyang. Soil P can be desorbed by changes in  
283 soil chemistry which might be related to changing inputs of acidic precipitation [5]. With continued  
284 climate change, the number of tropical and subtropical cyclones is likely to increase in summer  
285 monsoon regions [27,28]. Future efforts focused on the design and implementation of revised P  
286 management programs should consider the importance of P retained in highland agricultural soils  
287 in the study area.

288 In this study, we measured P inputs in wet deposition of 3,719 kg P/yr. This estimate is similar  
289 to P inputs in wet deposition in other studies [29, 30]. Although forested land covers more than 61%  
290 of the total watershed area, only a small amount of P was lost from forests in the Haean watershed  
291 (274 kg P/year). This estimate is within the range of measurements of P lost in surface runoff from  
292 forested land in other regions [31, 32]. Forested-agricultural watersheds can be found throughout  
293 Korea and future research directed at understanding P retention in mixed forested-agricultural  
294 systems would provide insight on how to better manage P losses at long-term scales.

295 Our study was not focused on identifying the biogeochemical processes controlling specific  
296 pathways of P sequestration or export. Generally, P losses from agricultural land are related to soil  
297 erosion [33, 34]. In the Haean watershed, hydrologic export of P is influenced by meteorological  
298 factors including rainfall amount, rainfall intensity, and the number of antecedent dry days.

#### 299 4.1. P input and output budgets from other studies

300 Studies on the global P cycle confirm that the average P accumulation rate in the world's soil has  
301 increased along with the dramatic increase in P consumption observed since 1960 [35, 36]. P  
302 accumulation in agricultural soils has been documented in most areas of Korea [37, 38]. The amount  
303 of P retained in the Haean watershed, 24.4 kg P/ha/yr, is similar to P retention in countries with  
304 intensive agricultural practices such as the Netherlands [39]. To understand the magnitude of P  
305 retention in a regional context, we compared P retention in the Haean watershed with P retention in  
306 watersheds in other locations (Table 9). The Haean watershed shows high P retention and has a larger  
307 area of forest land use compared to the other watersheds. Changing crop types in the watershed from  
308 annual crops to perennial crops with lower recommended fertilizer P applications is a possible  
309 management strategy that would lower the amount of P sequestered in the watershed. However,  
310 perennial crops cannot replace the economic value of annual crops in Korea, and this strategy might  
311 be limited to only certain fields. In addition to P input and output budgets, more studies on the  
312 biogeochemical processes controlling P dynamics are needed. Developing more efficient and

313 sustainable fertilization strategies will be useful to minimize long-term P sequestration in the Haean  
314 watershed.

315 **Table 9.** Comparison of the P budget in this study with P budgets for other watersheds

Watershed	Area (ha)	Land use (%)			Phosphorus flux (kg/ha/yr)			Retention (%)	Ref.
		Forest	Agri.	Others	Input	Output	Balance		
Haean, Korea	6,174	61.1	29.6	5.45	28.0	3.63	24.4	87.0	This study
EAA, South Florida	191,198	-	93.3	7.70	3.32	0.82	2.50	75.2	[8]
Lake Okeechobe*, Florida	1,392,987	30.3	55.2	14.5	5.99	0.88	5.11	85.3	[40]
Lake Okeechobe**, Florida	1,392,874	29.5	56.2	14.3	4.59	0.97	3.62	78.9	[41]
Bui SB, Vietnam	2,751	13.6	45.8	40.6	5.7	0.90	4.80	84.1	[42]
Cuyahoga, Ohio	182,300	41.6	31.0	27.4	12.1	2.43	9.47	79.9	[43]
Grand, Ohio	182,500	44.2	38.8	17.0	7.42	1.81	5.61	75.6	[43]
Guayaquil, Ecuador	136,000	6.40	39.8	53.8	4.36	0.93	3.43	78.7	[7]
Liuchahe, China	732	17.9	70.1	12.0	53.4	5.46	47.9	89.8	[6]
Roanoke, North Carolina	2,194,756	60.4	33.3	6.30	2.55	0.46	2.09	81.9	[44]

## 316 5. Conclusions

317 The P budget approach used in this study represents the first comprehensive attempt to identify  
318 the origin and fate of P in highland forested-agricultural watersheds in Korea. The P budget in the  
319 Haean watershed is highly influenced by agricultural activities. P inputs to the watershed are  
320 dominated by fertilizer inputs, while crop harvesting and hydrologic export are almost equally  
321 important in the P outputs. There is a net accumulation of P in the system even though nutrient loss  
322 and soil erosion occurs during the summer monsoon season in areas with intensive agriculture.  
323 Comparing the balance between P inputs and outputs of the present study with a number of diverse  
324 watersheds throughout the world shows common patterns with the Haean watershed, despite the  
325 significant differences in site specific soil characteristics, agricultural practices, geologic, and  
326 meteorological factors among the watersheds. Comparing the P retention in the Haean watershed in  
327 the present study with other studies shows that amount of P retained in the Haean watershed is much  
328 higher than for many watersheds. This study also confirmed that even if all the P inputs in the  
329 watershed were eliminated, soil P concentrations may not drop immediately because of the high P  
330 accumulation in agricultural soils. The P budget in this study provides a useful analysis tool to design  
331 better management practices to mitigate NPS-derived P pollution problems in mixed land use  
332 watersheds in Korea.

333 **Author Contributions:** Conceptualization, A.R., J.E. and B.K.; methodology, A.R. and J.E.; formal analysis, A.R.  
334 and J.E.; investigation, A.R., J.E., S.J., Y.C., C.J. and K.K.; data curation, J.E., S.J., Y.C., C.J. and K.K.; writing—  
335 original draft preparation, A.R.; writing—review and editing, J.S.O. and B.K.; supervision, B.K.; funding  
336 acquisition, J.S.O. and B.K.

337 **Funding:** This research was funded by the Korean Ministry of Environment, Kangwon National University and  
338 Hankuk University of Foreign Studies.

339 **Conflicts of Interest:** The authors declare no conflict of interest. The funders had no role in the design of the  
 340 study; in the collection, analyses, or interpretation of data; in the writing of the manuscript, or in the decision to  
 341 publish the results”.

## 342 References

- 343 1. Bennett, E.M.; Carpenter, S.R.; Caraco, N.F. Human impact on erodable phosphorus and eutrophication: a  
 344 global perspective. *Bioscience*. **2001**, *51*, 227-34.
- 345 2. Elser, J.J.; Bracken, M.E.S.; Cleland, E.E.; Gruner, D.S.; Harpole, W.S.; Hillebrand, H.; Ngai, J.T.; Seabloom,  
 346 E.W.; Shurin, J.B.; Smith, J.E. Global analysis of nitrogen and phosphorus limitation of primary producers  
 347 in freshwater, marine and terrestrial ecosystems. *Ecol. Lett.* **2007**, *10*, 1135-42.
- 348 3. Kim, B.; Park, J.H.; Hwang, G.; Jun, M.S.; Choi, K. Eutrophication of reservoirs in South Korea. *Limnology*.  
 349 **2001**, *2*, 223-9.
- 350 4. Reza, A.; Eum, J.; Jung, S.; Choi, Y.; Owen, J.S.; Kim, B. Export of non-point source suspended sediment,  
 351 nitrogen, and phosphorus from sloping highland agricultural fields in the East Asian monsoon region.  
 352 *Environ. Monit. Assess.* **2016**, 188:692.
- 353 5. Bennett, E.M.; Reed-Andersen, T.; Houser, J.N.; Gabriel, J.R.; Carpenter, S.R. A phosphorus budget for the  
 354 Lake Mendota watershed. *Ecosystems*. **1999**, *2*, 69-75.
- 355 6. Yan, W.; Yin, C.; Zhang, S. Nutrient budgets and biogeochemistry in an experimental agricultural  
 356 watershed in Southeastern China. *Biogeochemistry*. **1999**, *45*, 1-9.
- 357 7. Borbor-Cordova, M.J.; Boyer, E.W.; McDowell, W.H.; Hall, C.A. Nitrogen and phosphorus budgets for a  
 358 tropical watershed impacted by agricultural land use: Guayas, Ecuador. *Biogeochemistry*. **2006**, *79*, 135-61.
- 359 8. Faridmarandi, S.; Naja, G.M. Phosphorus and water budgets in an agricultural basin. *Environ. Sci. Technol.*  
 360 **2014**, *48*, 8481-90.
- 361 9. Kim, G.; Chung, S.; Lee, C. Water quality of runoff from agricultural-forestry watersheds in the Geum  
 362 River Basin, Korea. *Environ. Monit. Assess.* **2007**, 134:441.
- 363 10. Kim, B.; Choi, K.; Kim, C.; Lee, U.H.; Kim, Y.H. Effects of the summer monsoon on the distribution and  
 364 loading of organic carbon in a deep reservoir, Lake Soyang, Korea. *Water Res.* **2000**, *34*, 3495-3504.
- 365 11. Kettering, J.; Park, J.H.; Lindner, S.; Lee, B.; Tenhunen, J.; Kuzyakov, Y. N fluxes in an agricultural  
 366 catchment under monsoon climate: a budget approach at different scales. *Agric. Ecosys. Environ.* **2012**,  
 367 *161*, 101-11.
- 368 12. IUSS Working Group WR. World reference base for soil resources. In *World Soil Resources Report No.103*.  
 369 FAO: Rome, Italy, 2007.
- 370 13. Yanggu Statistical Yearbook 2011. Yanggu County Office: Gangwon Province, South Korea, 2012.
- 371 14. The Standard Rate of Chemical Fertilizer for Crops. Rural Development Administration: Suwon, South  
 372 Korea, 2010.
- 373 15. Andersen, J.M. An ignition method for determination of total phosphorus in lake sediments. *Water Res.*  
 374 **1976**, *10*, 329-331.
- 375 16. Methods of soil and plant analysis. National Institute of Agricultural Science and Technology: Suwon,  
 376 South Korea, 2000.
- 377 17. Murphy, J.A.; Riley, J.P. A modified single solution method for the determination of phosphate in natural  
 378 waters. *Anal. Chim. Acta.* **1962**, *27*, 31-36.
- 379 18. Lee, S.B.; Kim, J.G.; Lee, D.B.; Lee, K.B.; Han, S.S.; Kim, J.D.; Baek, S.H. Changes of physico-chemical  
 380 properties and microflora of pig manure due to composting with turning time and depth. *Korean J. Soil  
 381 Sci. Fertil.* **2001**, *35*, 127-135.
- 382 19. Loye-Pilot, M.D.; Martin, J.M. Saharan Dust Input to the Western Mediterranean: An Eleven Years Record  
 383 in Corsica. In *The Impact of Desert Dust Across the Mediterranean*. Guerzoni, S., Chester, S., Eds.; Springer:  
 384 Dordrecht, The Netherlands, 1996; pp 191-199. ISBN 978-94-017-3354-0.
- 385 20. Reza, A.; Eum, J.; Jung, S.; Kim, B. Influence of East Asian summer monsoon on suspended sediment and  
 386 nutrient export from a pristine forested watershed. In *Proceedings of the Annual Conference on  
 387 Environmental Research*, Kangwon National University, South Korea, 2016.
- 388 21. Fertilizer consumption (kilograms per hectare of arable land) 2012. Available online:  
 389 <https://data.worldbank.org/indicator/AG.CON.FERT.ZS> (accessed on 26 June 2016)
- 390 22. Yoo, D. Evaluation of sediment yield using area-weighted measured slope and slope length at Haean  
 391 Myeon watershed. M.Sc. Thesis, Kangwon National University, Chuncheon, South Korea, 2011.

- 392 23. Stigliani, W.M.; Doelman, P.; Salomons, W.; Schulin, R.; Smidt, G.R.; Van der Zee, S.E. Chemical time  
393 bombs: predicting the unpredictable. *Environ.: Sci. Policy Sustain. Dev.* **1991**, *33*, 4-30.
- 394 24. Stout, W.L.; Sharpley, A.N.; Gburek, W.J.; Pionke, H.B. Reducing phosphorus export from croplands with  
395 FBC fly ash and FGD gypsum. *Fuel.* **1999**, *78*, 175-178.
- 396 25. Choi, J.D.; Lee, H.J.; Park, S.Y.; Choi, Y.H.; Lim, K.J.; Gilkes, R.J. Sediment control practices in sloping  
397 highland fields in Korea. In *Proceedings of the 19th World Congress of Soil Science: Soil solutions for a*  
398 *changing world*, Brisbane, Australia, 2010.
- 399 26. Eum, J. Runoff characteristics of NPS from agricultural watershed and water quality simulations with crop  
400 change. Ph.D. Thesis, Kangwon National University, Chuncheon, South Korea, 2015.
- 401 27. Webster, P.J.; Holland, G.J.; Curry, J.A.; Chang HR. Changes in tropical cyclone number, duration, and  
402 intensity in a warming environment. *Science.* **2005**, *309*, 1844-1846.
- 403 28. Kang, N.Y.; Elsner J.B. Climate mechanism for stronger typhoons in a warmer world. *J. Clim.* **2016**, *29*,  
404 1051-1057.
- 405 29. Gibson, C.E.; Wu, Y.; Pinkerton, D. Substance budgets of an upland catchment: the significance of  
406 atmospheric phosphorus inputs. *Freshwater Biol.* **1995**, *33*, 385-392.
- 407 30. Haygarth, P.M.; Chapman, P.J.; Jarvis, S.C.; Smith, R.V. Phosphorus budgets for two contrasting grassland  
408 farming systems in the UK. *Soil Use Manage.* **1998**, *14*, 160-167.
- 409 31. Lin, J.P. Review of Published Export Coefficient and Event Mean Concentration (EMC) Data. Mississippi:  
410 U.S.Army Corps of Engineers, Engineer Research and Development Center; 2004.
- 411 32. White, M.; Harmel, D.; Yen, H.; Arnold, J.; Gambone, M.; Haney, R. Development of sediment and nutrient  
412 export coefficients for US ecoregions. *J. Am. Water Resour. Assoc.* **2015**, *51*, 758-775.
- 413 33. Sharpley, A.N.; Daniel, T.C.; Edwards, D.R. Phosphorus movement in the landscape. *J. Prod. Agri.* **1993**, *6*,  
414 492-500.
- 415 34. Haygarth, P.M.; Jarvis, S.C. Transfer of phosphorus from agricultural soils. *Adv. Agron.* 1999, *66*, 196-249.
- 416 35. Carpenter, S.R.; Caraco, N.F.; Correll, D.L.; Howarth, R.W.; Sharpley, A.N.; Smith, V.H. Nonpoint pollution  
417 of surface waters with phosphorus and nitrogen. *Ecol. Appl.* **1998**, *8*, 559-568.
- 418 36. Lu, C., Tian, H. Global nitrogen and phosphorus fertilizer use for agriculture production in the past half  
419 century: shifted hot spots and nutrient imbalance. *Earth Syst. Sci. Data.* **2017**, *9*, 181.
- 420 37. Lee, C.H.; Park, C.Y.; Do Park, K.; Jeon, W.T.; Kim, P.J. Long-term effects of fertilization on the forms and  
421 availability of soil phosphorus in rice paddy. *Chemosphere.* **2004**, *56*, 299-304.
- 422 38. Park, M.; Singvilay, O.; Shin, W.; Kim, E.; Chung, J.; Sa, T. Effects of long-term compost and fertilizer  
423 application on soil phosphorus status under paddy cropping system. *Comm. Soil Sci. Plant Anal.* **2004**, *35*,  
424 1635-1644.
- 425 39. Chardon, W.J.; Koopmans, G.F. Critical evaluation of options for reducing phosphorus loss from  
426 agriculture. In: *Proceedings of the 4th International Phosphorus Workshop*, Wageningen, The  
427 Netherlands, 2004.
- 428 40. Hiscock, J.G.; Thourot, C.S.; Zhang, J. Phosphorus budget—land use relationships for the northern Lake  
429 Okeechobee watershed, Florida. *Ecol. Eng.* **2003**, *21*, 63-74.
- 430 41. He, Z.; Hiscock, J.G.; Merlin, A.; Hornung, L.; Liu, Y.; Zhang, J. Phosphorus budget and land use  
431 relationships for the Lake Okeechobee Watershed, Florida. *Ecol. Eng.* **2014**, *64*, 325-336.
- 432 42. Luu, T.N.M.; Garnier, J.; Billen, G.; Le, T.P.Q.; Nemery, J.; Orange, D.; Le, L.A. N, P, Si budgets for the Red  
433 River Delta (northern Vietnam): how the delta affects river nutrient delivery to the sea. *Biogeochemistry.*  
434 **2012**, *107*, 241-259.
- 435 43. Han, H.; Bosch, N.; Allan, J.D. Spatial and temporal variation in phosphorus budgets for 24 watersheds in  
436 the Lake Erie and Lake Michigan basins. *Biogeochemistry.* **2011**, *102*, 45-58.
- 437 44. McMahan, G.; Woodside M.D. Nutrient mass balance for the Albemarle-Pamlico drainage basin, North  
438 Carolina and Virginia, 1990. *J. Am. Water Resour. Assoc.* **1997**, *33*, 573-589.