

1 Article

2 Experimental Analysis of Heavy Metal Uptake in Soil 3 and Citrus Plants from Irrigation Water

4 Ailin Zhang¹, Veronica Cortes², Bradley Phelps², Hal van Ryswyk³, Tanja Srebotnjak^{4*}

5 ¹ Harvey Mudd College, Chemistry Department, 301 Platt Blvd, Claremont, CA 91711, USA;

6 azhang@g.hmc.edu

7 ² Harvey Mudd College, Engineering Department, 301 Platt Blvd, Claremont, CA 91711, USA;

8 vcortes@g.hmc.edu

9 ³ Harvey Mudd College, Chemistry Department, 301 Platt Blvd, Claremont, CA 91711, USA;

10 vanryswyk@g.hmc.edu

11 ⁴ Harvey Mudd College, Hixon Center for Sustainable Environmental Design, 301 Platt Blvd, Claremont, CA
12 91711, USA; tsrebotnjak@g.hmc.edu

13 * Correspondence: tsrebotnjak@g.hmc.edu; Tel.: +1-909-621-8751

14

15 **Abstract:** Oilfield produced water (OPW) is used to boost freshwater sources for crop irrigation in
16 California's agriculturally important Central Valley. OPW is known to contain salts, metals,
17 hydrocarbons, alkylphenols, naturally radioactive materials, biocides, and other compounds from
18 drilling and production processes. Less is known about the potential uptake and accumulation of
19 these compounds in crops and soil irrigated with OPW. In this study 23 potted mandarin orange
20 plants were irrigated 2-3 times weekly (depending on season) with water containing three different
21 concentrations of the known OPW heavy metals barium, chromium, lead, and silver. Seven sets of
22 samples of soil and leaves and all fruits were collected and processed using microwave-assisted
23 digestion (EPA Method 3051A). Processed samples were analyzed using ICP-OES. ANOVA,
24 ANCOVA, and Tukey's honest significant difference test were used to examine the effects of metal
25 concentrations in the irrigation water, sample number, and number of watering days on the metal
26 concentrations in the soil, leaf, and fruit samples. Accumulation of barium in soil and leaves was
27 strongly positively associated with sample and number of watering days, increasing nearly
28 2,000-fold. Lead also showed an upward trend, increasing up to 560-fold over baseline level.
29 Chromium showed an increase in the soil that tapered off, but less consistent results in the leaves
30 and fruit. The silver results were more volatile, but also indicated at least some level of
31 accumulation in the tested media. The smallest absolute accumulation was observed for chromium.
32 Concentrations in the fruit were highest in the peel, followed by pith and juice. Accumulation of all
33 heavy metals was generally highest in the soil and plants that received the highest irrigation water
34 concentration. Considering the potential for adverse human health effects associated with
35 ingesting soluble barium contained in food and drinking water, and to a lesser extent chromium
36 and lead, the study signals that it is important to conduct further research into whether OPW
37 contaminants can enter the food chain and pose risks to consumers.

38 **Keywords:** oilfield produced water; wastewater; heavy metals; irrigation; bioaccumulation; soil
39 health; California

40

41 1. Introduction

42 California is still the third-largest oil-producing state in the United States.[1] During the oil
43 production process, substantial volumes of oilfield produced water (OPW, also referred to as oilfield
44 brine, connate water, or formation water) are typically generated, especially as the well and oilfield
45 matures. The origins of the briny water include flow from above, within and below the targeted
46 hydrocarbon zone as well as flow from injected fluids and additives employed during the
47 production process.[2] OPW is the largest waste stream by volume in the exploration and extraction

48 of oil, and over the lifespan of an oilfield the total volume of produced water can exceed tenfold the
49 volume of hydrocarbons produced.[3] In 2012, for example, California generated an estimated 21
50 billion barrels of OPW, i.e., approximately 15 barrels of OPW for every barrel of oil.[4]

51 In general, OPW is high in salts and dissolved solids and may also contain metals, alkylphenols,
52 trace elements, hydrocarbons, polycyclic aromatic hydrocarbons (PAH), volatile organic compounds
53 (VOC), benzene, toluene, ethylbenzene and xylenes (collectively known as BTEX), naturally
54 radioactive materials (NORM), biocides, and other compounds used during the production
55 process.[3] Concentrations of total dissolved solids (TDS) in OPW range from the low thousand
56 mg/L to several hundred thousand mg/L.[2] In the Western U.S., including several California
57 oilfields, measured concentrations range from 1,000 mg/L to 400,000 mg/L.[5] A case study
58 analyzing the composition of OPW from 630 producing oil and gas wells in California that were
59 stimulated through hydraulic fracturing found that the average concentration of Cr(VI) in the OPW
60 was 8.5 µg/L, while the average concentrations of Sb was 1,100 µg/L, As 1,100 µg/L, Cd 40 µg/L, Pb
61 420 µg/L, Se 1,900 µg/L and Zn 420 µg/L.[4] Many of the elements and compounds used in oil well
62 drilling and production can have varying adverse health effects on the gastrointestinal tract, the
63 nervous system, and the reproductive system when ingested.

64 Although the most common method for managing OPW in California is deep-well injection and
65 disposal in evaporation-percolation pits, accounting for 60% and 18% of total OPW disposal,
66 respectively, a small fraction of California's OPW is applied to land surface.[6] The Clean Water Act
67 stipulates effluent guidelines for on-shore oil and gas extraction facilities that prohibit the discharge
68 of pollutants into surface waters, except for wastewater that is of good enough quality for use in
69 agricultural and wildlife propagation for onshore facilities located in the continental United States
70 and west of the 98th meridian. This creates the possibility for reusing OPW in agricultural crop
71 irrigation, as is the case in California's Central Valley.

72 Although OPW has been reused in surface applications for more than 20 years, it gained wider
73 prominence and interest during the exceptional drought that gripped California between 2011-2016
74 as a means to supplement heavily constrained agricultural freshwater supplies.[7,8] However, the
75 question of whether it can be done safely for consumers, agricultural workers and communities
76 living near OPW application sites, has not been sufficiently answered to date.[9] Among the
77 concerns voiced by public health and environmental groups is whether toxic OPW compounds, such
78 as heavy metals, may become bioavailable and accumulate in parts of the crops intended for human
79 consumption.[9] Concerns about long-term soil health and productivity have also been raised.[9]
80 There is a substantial body of literature, involving field and laboratory studies, regarding the
81 origins, environmental and health effects, and remediation of soil contamination with heavy metals,
82 including Ba, Cd, Pb, and As.[10-16] Studies by Fliessbach et al. (1994) and Khan et al. (2000) show
83 the negative effects on soil productivity resulting from heavy metal contamination.[17-19] While soil
84 accumulation has been widely documented, Toze (2009) found evidence that the seeds and leaves of
85 flax, cotton and rice report store only a small percentage of the heavy metals compared to soils.[20]
86 However, even small concentrations may pose health risks if the crop is consumed frequently, such
87 as rice, or in larger amounts over a short period of time, such as seasonal crops like mandarin
88 oranges. A study in Tongzhou District, Beijing, China in 2007 examined wastewater-irrigated soils
89 used to grow radish, maize, spinach, green cabbage, cauliflower, turnip, and lettuce.[21] It found
90 significant correlations between heavy metal concentration in soils and the edible parts of plants for
91 Cd, Cu, Pb, and Zn. In particular, the concentrations of Cd, Cr, and Ni exceeded regulatory limits set
92 by the State Environmental Protection Administration.[21] Similar issues are known to arise in the
93 proximity of metal ore mines. For example, the soil and corn, jujube, perilla, red pepper, soybean,
94 and spring onions grown near a copper-tungsten mine in South East Korea were found to have
95 elevated concentrations of Cu, Zn, Cd, and Pb.[22] Concentrations varied with crop species and
96 higher concentrations were found in the leafy material of the crops compared with the fruit and
97 seeds. It was also determined that the most dominant factor affecting metal uptake in plants was the
98 metal content in the surface soil, which may be problematic in the long-term application of OPW
99 wastewater.[22]

100 While the existing literature demonstrates the potential for plant uptake of metals through soil
 101 and irrigation water, it is of specific interest to look into the potential soil and crop impacts of OPW
 102 reuse in agriculture in California's Central Valley. As of now, water quality monitoring involving
 103 OPW is limited and not federally regulated.[5] In response to concerns raised by environmental and
 104 public health organizations, the Central Valley Regional Water Board initiated a Food Safety Expert
 105 Panel tasked with examining the available evidence and conducting studies for testing the risk
 106 involved in OPW reuse in the region's agricultural operations, primarily citrus, almond, pistachio,
 107 and table grape farming and to a lesser extent garlic, carrot and tomato plantings.[9] The panel
 108 commissioned sampling campaigns in the spring and summer of 2017 for table grapes, citrus,
 109 almonds, garlic, and pistachios and tested the edible fruit for over 100 organic, inorganic and metal
 110 analytes. A total of 16 compound or elements were detected, among them the metals Ba, Cu, Mo, Ni,
 111 Sr, and Zn, although at mostly low concentrations.[23]

112 The present study complements the work of the Food Expert Panel with an observational,
 113 controlled study of heavy metal accumulation of Ba, Cr, Pb, and Ag in the soil, leaves and fruit of
 114 mandarin orange plants. The goal of the study is to identify if and to what extent the tested heavy
 115 metals accumulate in the soil, leaves, and/or fruit in response to sampling time and the concentration
 116 of heavy metals used in the irrigation water.

117 2. Materials and Methods

118 A total of 23 mandarin orange trees, representing six varieties found in California's Central
 119 Valley and residential gardens, were purchased in 5 gallon pots from the same local nursery in late
 120 spring 2017. The plants were semi-randomly assigned to receiving irrigation water with three
 121 different concentration levels of Ba, Cr, Pb, and Ag (see Table 1) reflecting measurements reported
 122 in the OPW literature.[2,24] Nitrate salts were used due to their high solubilities and the absence of
 123 deleterious effects. The high concentration level was selected to test for substantial plant effects such
 124 as growth decline and mortality. The procedure was semi-random, because fruit-bearing plants
 125 were allocated such that each concentration level had at least two plants with fruits. The tested
 126 metals are found in OPW and are known for their risks to human health and soil productivity.

127 Table 1: Range of concentrations observed in produced water for the tested elements.

Metal salt	low concentration [mg metal/L]	# of plants	medium concentration [mg metal/ L]	# of plants	high concentration [mg metal/L]	# of plants	con c reported in literature [mg metal/L]
Barium nitrate	25.0	* 7	1,500	8	3,000	8	9.65 - 1,740
Chromium(III) nitrate	0.01		0.03		0.10		ND - 0.03
Lead nitrate	5.0		10.0		30.0		<0.2 - 10.2
Silver nitrate	1.0		7.0		15.0		0.047 - 7.0

128 Note: low concentration levels were well above the LOD of the ICP-OES instrument. *A
 129 temporary, region-wide shortage in citrus plants meant that only 23 instead of 24 plants could be
 130 purchased within a reasonable timeframe.
 131

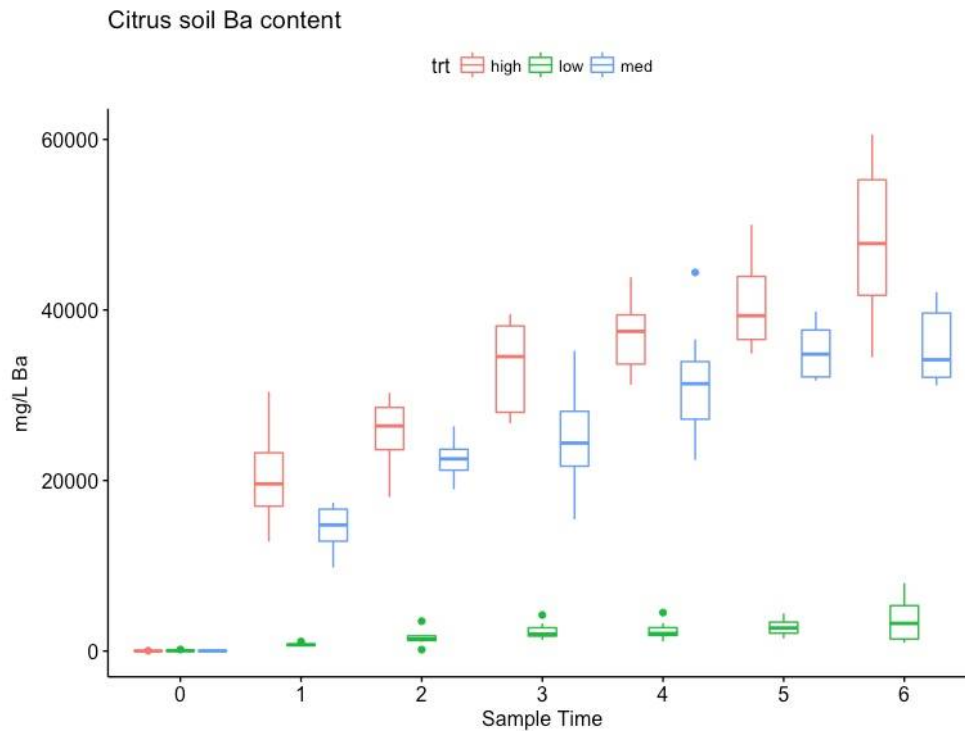
132 Baseline concentrations of the metals in the soil and leaves were determined before watering
133 began. Plants were watered twice (cold season) or three times (hot season) weekly. During August
134 and mid-December through mid-January the plants were watered with freshwater due to breaks in
135 the academic calendar. Samples of soil and leaves were taken every two weeks during the active
136 experimentation period. The fruit was harvested for analysis at the end of the experiment in March,
137 2018. Watering solutions were prepared on the days of irrigation. The soil and leaf samples were
138 then processed by microwave-assisted digestion using EPA Method 3051A. Method blanks were run
139 for each analysis. Specifically, in a typical procedure, 0.300 grams of soil or leaf were dried at 60° C
140 for 3 days, digested with 12 mL of (3:1 HNO₃:HCl, A.C.S. reagent grade, Spectrum Chemical Mfg.
141 Corp.) and filtered through a 2- μ m glass fibre filter (Millipore Sigma AP2501300). The supernatant
142 solution was diluted to 50.0 mL and stored in polypropylene test tubes. The samples were then
143 analyzed using inductively coupled plasma-optical emission spectroscopy (ICP-OES). A
144 PerkinElmer 8300 ICP-OES was used to construct calibration curves over the range of 0 – 400 mg/L
145 Ba; 0 – 5.0 mg/L Cr; 0 – 50 mg/L Pb; and 0 – 0.25 mg/L Ag. The plasma operated at 1500 W RF power
146 with argon gas flows of 10 L/min, 0.4 L/min, and 0.6 L/min in plasma, auxiliary, and nebulizer
147 supplies, respectively. The oranges were analyzed in three parts: the juice, the pith, and the
148 peel. The pith and peels were dried at 60° C for 3 days prior to being processed as described above.

149 3. Results

150 A total of 7 sampling times $t=(0,1,2, \dots,6)$ for soil and leaves were analyzed, with $t=0$
151 representing the baseline sample. While t is a discrete measure of the passage of time, another means
152 to capture the exposure of the plants to the metal-tainted irrigation water is to consider the number
153 of days that plants were watered with the metal solutions between the start of the experiment and
154 the day the respective sample was collected. This watering vector is $w=(0, 20, 26, 31, 35, 39, 43)$, with
155 $w=0$ representing the baseline measurement day. Two types of analysis were conducted to model the
156 change in measured metal concentrations over the course of the experiment. First, analysis of
157 variance (ANOVA) was carried out to determine the effect of *treatment* (low, medium and high metal
158 concentration in irrigation water) and *time* (using the discrete vector t) on the measured
159 concentration of the metal in the soil and leaves. Second, an analysis of covariance (ANCOVA) was
160 run for each metal using the same *treatment* factor and the watering vector, w , as a continuous
161 variable. Since the sampling design was not fully balanced (the low concentration level had only 7
162 replicates), type II and III sums of squares were used.[25] Tukey's honest significant difference test
163 (Tukey HSD) was applied to identify differences in means between factor level combinations. All
164 analyses were carried out in the R statistical language and computing environment, version
165 3.4.3.[26] Results are presented by metal.

166 3.1. Barium

167 Of the tested metals, Ba shows the most pronounced and consistent pattern of increase in
168 accumulation in soil and the leaves for t and w , respectively (see Figure 1). The increase was largest
169 for the high concentration water, followed by the medium and low concentration water,
170 respectively. Concentrations increased from an average baseline level of 17.6 mg/L (sd=11.4 mg/L) to
171 an average of 48,086.5 mg/L (sd=9,270 mg/L) for the high concentration water at time $t=6$. For the low
172 watering concentrations the increase was from 57.6 mg/L (sd=61.8 mg/L) at baseline to 3,674.5 mg/L
173 (sd=2,737.6 mg/L) at $t=6$.



174

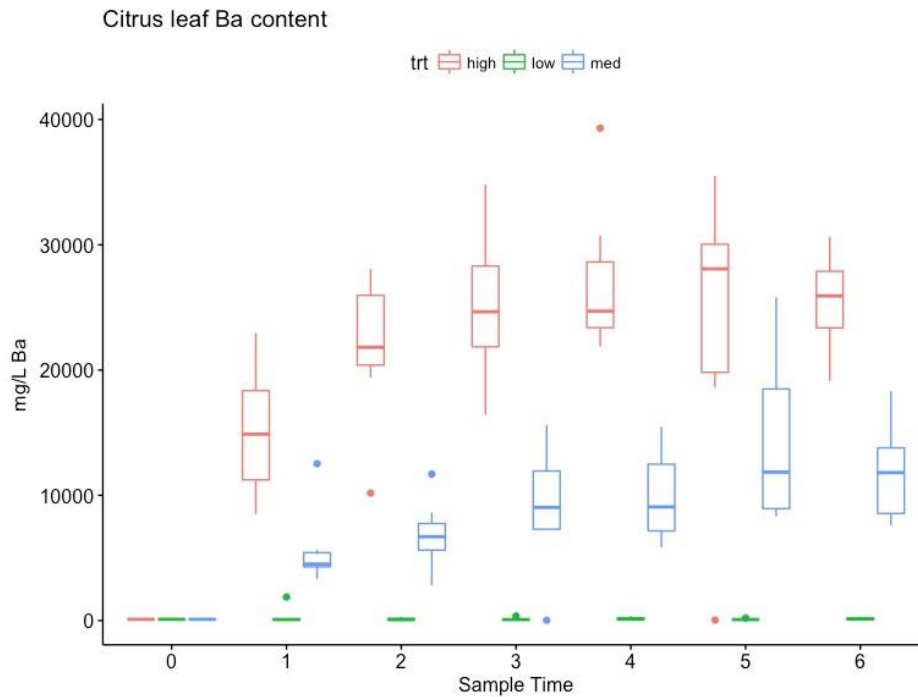
175 Figure 1: Boxplots of measured concentrations of Ba in the soil as a function of Ba concentration in
 176 the irrigation water (low, medium, high) and sampling time, t.

177 The results for the ANOVA are shown in Table 1; both treatment and time factors are statistically
 178 significant as is their interaction, which means that the slopes for the time effects differ across
 179 treatments. Specifically, the higher the Ba concentration in the irrigation water, the faster the soil
 180 accumulates occurs the metal.

181 Table 1: Summary of ANOVA for Ba concentrations in the soil samples.

Factor	F-value	p-value
treatment	444.960	<0.0001
time	135.221	<0.0001
treatment:time	19.669	<0.0001

182 A similar pattern presents for the leaves (see Figure 2 and Table 2), although at a slower rate and
 183 with concentrations leveling off at t=2. No significant accumulation occurs at the low watering
 184 concentration and there is not enough evidence for significant interactions between treatment and
 185 time. The ANOVA results for the leaves are shown in Table 2.



186

187 Figure 2: Boxplots of measured concentrations of Ba in the leaves as a function of Ba concentration in
 188 the irrigation water (low, medium, high) and sampling time, t .

189 Table 2: Summary of ANOVA for Ba concentrations in the leaf samples.

Factor	F-value	p-value
treatment	187.356	<0.0001
time	5.3329	0.0002
treatment:time	1.7141	0.0613

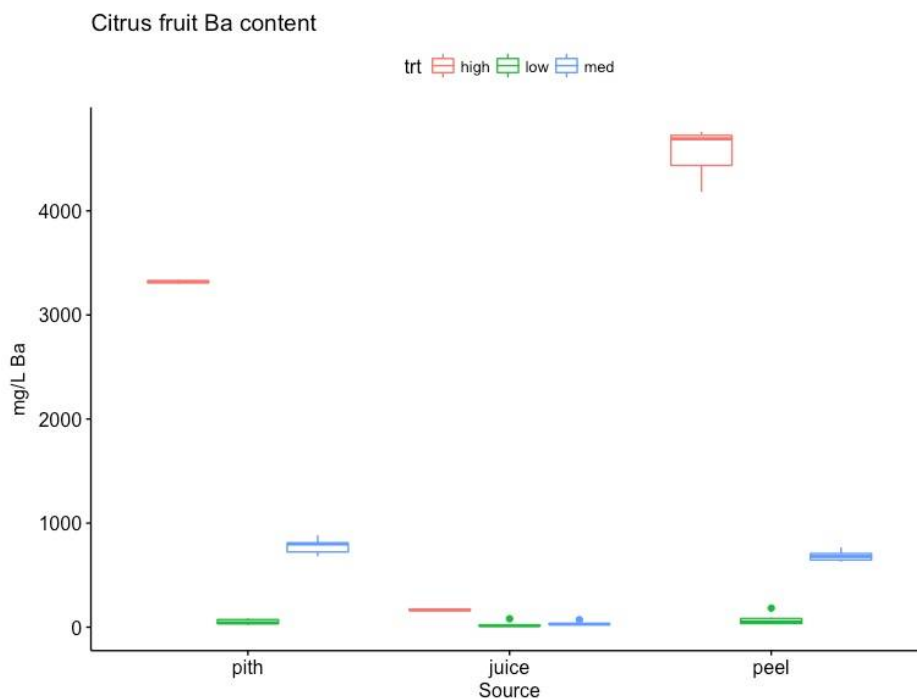
190 When considering watering as a continuous time metric, the results are similar, but the interaction is
 191 now statistically significant and indicates different accumulation rates for the three tested
 192 concentrations (see Table 3).

193 Table 3: Summary of ANCOVA for Ba concentrations in the soil and the leaf samples.

Factor	F-value	p-value
Soil		
treatment	465.40	<0.0001
waterings	845.11	<0.0001
treatment:waterings	105.95	<0.0001
Leaves		
treatment	182.00	<0.0001

waterings	35.02	<0.0001
treatment:waterings	7.72	<0.0001

194 Ba concentrations in the fruit are highest for the peel (mean=1,014.2 mg/L, sd=1,654.1 mg/L) and pith
 195 (mean=834.0 mg/L, sd=1,193.2 mg/L). At the high watering concentration, the peel has a mean
 196 concentration of 4,544.3 mg/L (sd=315.5 mg/L), and the pith a mean of 3,318.4 mg/L (sd=16.8 mg/l).
 197 Concentrations of Ba in the juice remain low regardless of the concentration of Ba in the irrigation
 198 water (mean=52.8 mg/L, sd=56.1 mg/L). The ANOVA of the effect of Ba watering concentrations and
 199 fruit type (peel, pith, and juice) show that both are highly significant and that the effect of treatment
 200 depends on the type of fruit analyzed.



201

202 Figure 3: Boxplots of measured concentrations of Ba in the pith, juice and peel portions of the
 203 harvested mandarin oranges as a function of Ba concentration in the irrigation water (low, medium,
 204 high).

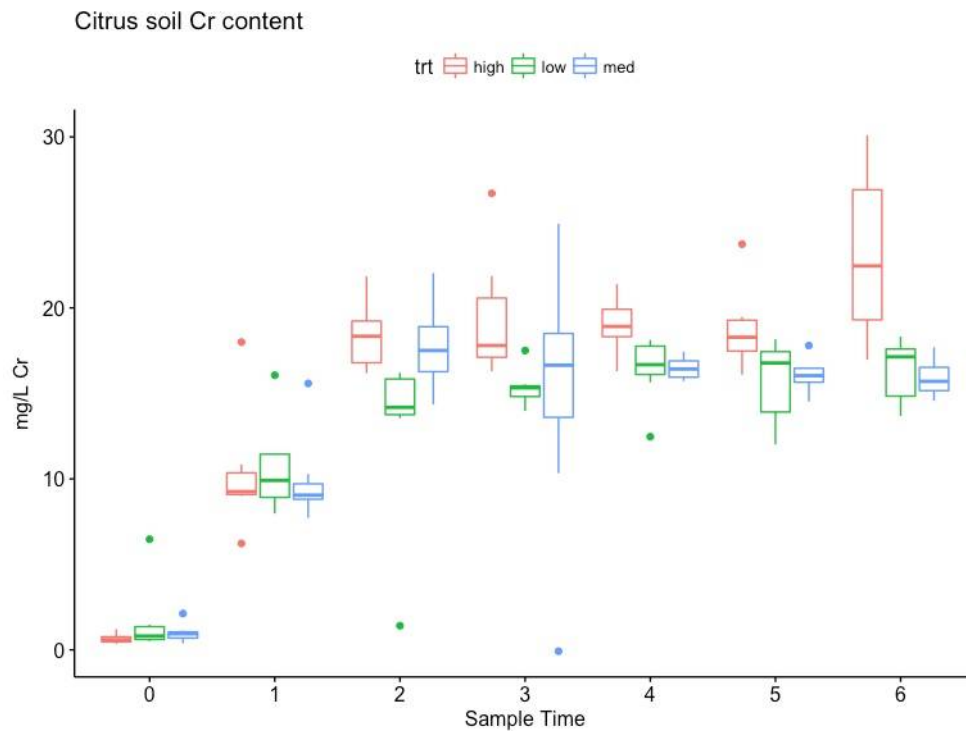
205 Table 4: Summary of ANOVA for the concentrations of Ba in the fruit samples.

Factor	F-value	p-value
treatment	2,469.77	<0.0001
type	735.67	0.0002
treatment:type	603.81	0.0613

206 3.2 Chromium

207 The measurements for Cr in the soil show an increase with time, t , that levels off at $t=2$. As the
 208 boxplots in Figure 4 illustrate, the variation in measured concentrations also increases with time
 209 (sd_{t=0}=1.2 mg/L and sd_{t=6}=5.7 mg/L). There is little to no differentiation in soil accumulation between
 210 the three tested irrigation water concentrations. The results of the ANOVA are shown in Table 5.

211 The results for Cr in the leaves (see Supplementary Material Figure S1) are more volatile with
 212 initial increase up to $t=3$, followed by a sharp drop at $t=4$, substantial rise at $t=5$, and smaller decline
 213 at $t=6$.



214

215 Figure 4: Boxplots of measured concentrations of Cr in the soil as a function of Cr concentration in
 216 the irrigation water (low, medium, high) and sampling time, t .

217 Table 5: Summary of ANOVA for Cr concentrations in the soil samples.

Factor	F-value	p-value
treatment	44.102	<0.0001
time	99.645	<0.0001
treatment:time	3.524	<0.0001

218

219 The results for the ANCOVA with watering days as the continuous variable are shown in Table
 220 6.

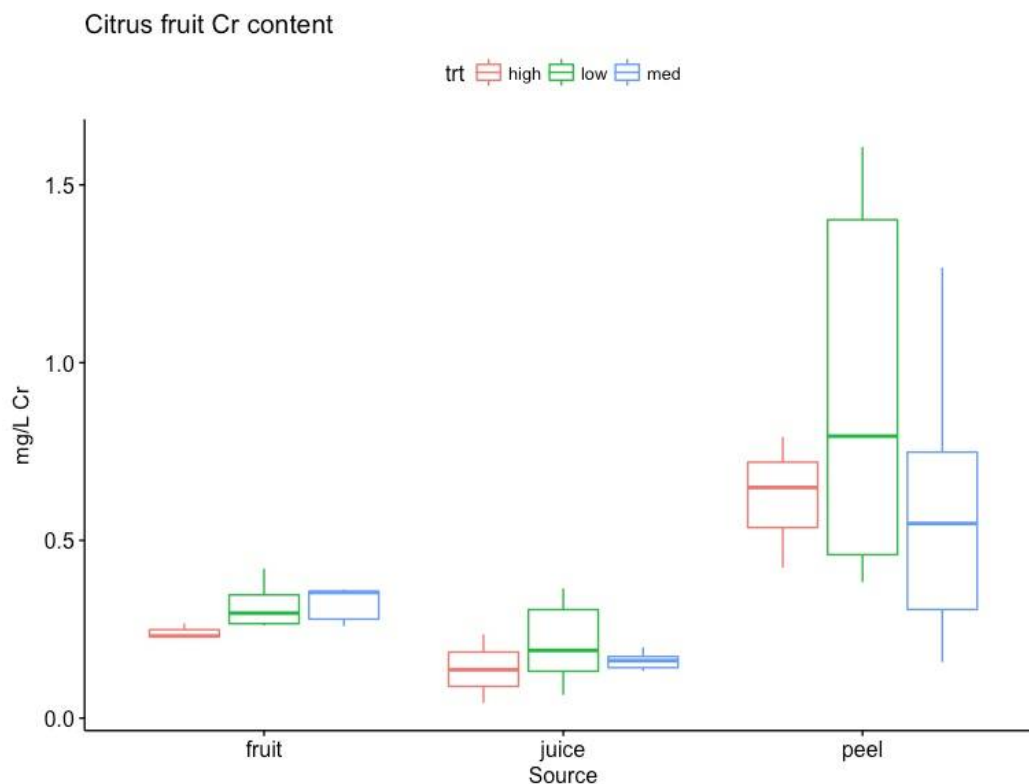
221 Table 6: Summary of ANCOVA for Cr concentrations in the soil and the leaf samples.

Factor	F-value	p-value
<i>Soil</i>		
treatment	28.69	<0.0001
waterings	359.05	<0.0001
treatment:waterings	5.667	<0.001
<i>Leaves</i>		

treatment	3.76	0.0129
waterings	2.88	0.0924
treatment:waterings	0.60	0.6181

222 For the harvested oranges, the highest Cr concentrations are measured in the peel (mean=0.66
 223 mg/L, sd=0.44 mg/L), followed by the pith (mean=0.30 mg/L, sd=0.08 mg/L) and the juice (mean=0.20
 224 mg/L, sd=0.11 mg/L) (see Figure 5 and Table 7). These differences are statistically significant. While
 225 average concentrations increase with the level of watering concentrations for the pith (not for the
 226 peel and juice), F-tests in the ANOVA and Tukey's HSD tests reveal no statistically significant
 227 differences for treatment and treatment:type interaction.

228



229

230 Figure 5: Boxplots of measured concentrations of Cr in the pith, juice and peel portions of the
 231 harvested mandarin oranges as a function of Cr concentration in the irrigation water (low, medium,
 232 high).

233 Table 7: Summary of ANOVA for Cr concentrations in fruit samples.

Factor	F-value	p-value
treatment	1.3142	0.2831
type	15.176	<0.0001
treatment:type	1.4861	0.2077

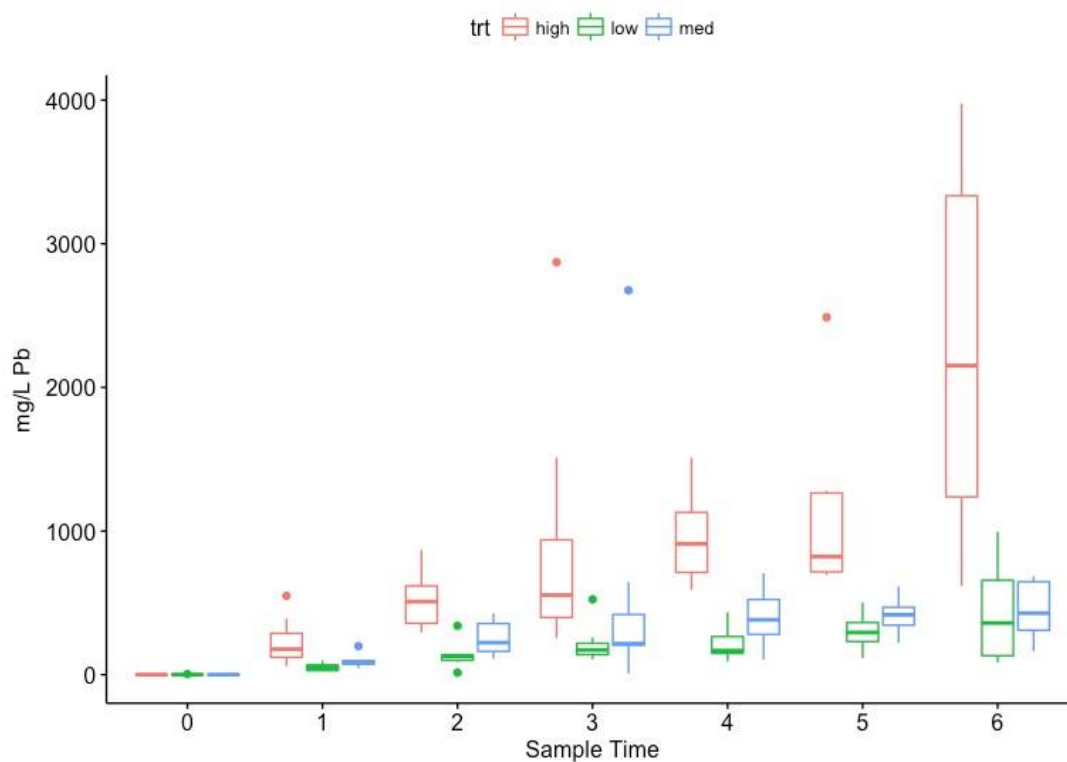
234

235

236 3.3 Lead

237 The results for Pb indicate that accumulation in the soil is associated with time, t , watering days,
 238 w , and the concentration of Pb in the irrigation water (see Figure 6). At low and medium irrigation
 239 water concentrations, the soil concentration of Pb increases slowly and steadily ($\text{mean}_{\text{low}}=194.0$
 240 mg/L , $\text{sd}_{\text{low}}=212.8 \text{ mg/L}$, $\text{mean}_{\text{med}}=309.8 \text{ mg/L}$, $\text{sd}_{\text{med}}=382.9 \text{ mg/L}$), while for the high concentration
 241 water the slope of accumulation is larger, but also subject to greater variation ($\text{mean}=866.8 \text{ mg/L}$,
 242 $\text{sd}=943.5 \text{ mg/L}$). The leaves (see Supplementary Material Figure S2) show little accumulation up to
 243 $t=3$. Beyond this time point, concentrations in the leaves rise but at a markedly increased
 244 between-plant variance. No systematic differences between low, medium and high irrigation water
 245 concentrations are detected using Tukey's HSD test. ANOVA results are shown in Table 8. As with
 246 Ba, the slopes for the soil concentrations for time vary based on treatment.

Citrus soil Pb content



247

248 Figure 6: Boxplots of measured concentrations of Pb in the soil as a function of Pb concentration in
 249 the irrigation water (low, medium, high) and sampling time, t .

250 Table 8: Summary of ANOVA for Pb concentrations in soil samples.

Factor	F-value	p-value
treatment	23.4233	0.2831
type	13.2193	<0.0001
treatment:type	3.3416	0.2077

251 ANCOVA results are shown in Table 9.

252

253 Table 9: Summary of ANCOVA for Pb concentrations in the soil and the leaf samples.

Factor	F-value	p-value
<i>Soil</i>		
treatment	21.208	<0.0001
waterings	57.288	<0.0001
treatment:waterings	9.857	<0.0001
<i>Leaves</i>		
treatment	3.26	0.0238
waterings	11.43	0.001
treatment:waterings	0.87	0.4572

254 Comparatively high levels of variation mark the Pb concentrations in the pith, juice and peel for all
 255 three tested watering concentrations (see Supplementary Material Figure S3). Average
 256 concentrations are highest for the fruit material, but were not statistically significant.

257 3.4 Silver

258 Silver concentrations show little in terms of clear detectable patterns over time or watering
 259 concentrations (see Supplementary Material Figures S4 and S5) for both soil and leaves. While
 260 baseline soil and leaf concentrations are low and subsequent sampling times detect significant
 261 increases, they remain approximately steady over time. Furthermore, the results for the medium and
 262 high watering concentrations are indistinguishable, while low watering concentrations on average
 263 yield lower Ag soil concentrations.

264 For the harvested mandarin oranges, Ag concentrations appear to increase with watering
 265 concentrations in the pith, but not in the juice nor the peel, where they remain well below 1 mg/L
 266 (see Supplementary Material Figure S6).

267 4. Discussion

268 The study presents the result of a controlled experiment of the accumulation of four heavy metals
 269 (Ba, Cd, Pb, Ag) in the soil, leaves and fruit of 23 mandarin orange trees that were irrigated with
 270 water containing three different concentrations of these elements that are representative of OPW
 271 conditions. The study's goals were to detect the soil and plant (leaves, fruit) concentrations of these
 272 elements, which are frequently found in similar concentrations in oilfield produced water (OPW),
 273 and to examine if there are detectable trends over time and as a function of the metals' concentration
 274 in the irrigation water. In the absence of federal (and state) regulations on the permissible
 275 concentrations of these and other compounds found in OPW, it is of interest to learn about their fate
 276 when OPW is reused for agricultural crop irrigation on a frequent basis as is the case in several water
 277 districts in California's agriculturally important Central Valley.

278 The results indicate that soils can accumulate several of the tested elements, in particular Ba, but to a
 279 lesser extent also Pb and Cr, and that generally the accumulated amounts are higher when these
 280 heavy metals are supplied in high concentrations in the irrigation water. The results for Ag were
 281 inconclusive.

282 Compared with the soil samples, the leaf material showed consistently lower concentrations of the
283 tested elements. The most significant uptick was observed for Ba at the high concentration level. This
284 finding is supported by reports in the literature that green, leafy material, with the exception of leafy
285 produce such as spinach and lettuce, tends to store little to no heavy metals.

286 All tested metals have known adverse human health effects that range from mild discomforts to
287 serious organ damage and even death. For example, Ba (in soluble form), Cr and Pb ingested in
288 small to moderate concentrations can impact the digestive and respiratory tracts, the
289 musculoskeletal and neurological systems, and the reproductive organs. The effects depend on the
290 exposure route, frequency and magnitude of exposure, as well as the chemical formulation of the
291 metal (e.g., soluble versus insoluble Ba). Although the OPW in this study is used for crop irrigation
292 and does not directly enter drinking water sources, it is illustrative to compare the measured
293 concentrations with applicable drinking water standards, because of the risk of exposure to
294 consumers of the edible crop. For Ba, the federal Maximum Contaminant Level (MCL) is 2.0 mg/L,
295 for Pb the EPA specified an Action Level of 0.015 mg/L. The MCL for total Cr is 0.1 mg/L, and for Ag
296 the federal Secondary Drinking Water Standard is 2.0 mg/L.[27] The soil, leaf, and fruit
297 concentrations measured for the four metals frequently exceeded these thresholds, noticeably at the
298 medium and high watering concentrations.

299 The exceedance of public health standards is especially concerning for the fruit measurements. The
300 pith appears to be the most likely to accumulate the tested metals at concentrations potentially
301 endangering the health of the consumer with concentrations for Ba ranging from 21.57 mg/L to
302 4,758.25 mg/L, for Cr from 0.13 mg/L to 1.61 mg/L, for Pb from 0.43 mg/L to 8.36 mg/L, and for Ag
303 from 0.10 mg/L to 0.91 mg/L. It is noted, however, that the concentrations measured are for total Ba,
304 Cr, Pb, and Ag and do not distinguish between soluble and insoluble fractions, although the heavy
305 metals were applied in the watering solutions as the soluble species.

306 The study has a few additional limitations. The sample size of 23 citrus plants is relatively small and
307 the experiment was conducted on potted plants and not in an orchard in the Central Valley with
308 actual OPW-supplemented irrigation water. Thus, the typical conditions, including soil cycling, soil
309 types and hydrological movements of irrigation water could not be fully replicated and no soil
310 characterization was done. The study also had a limited duration (June 2017 through March 2018)
311 and was interrupted by two periods of watering with freshwater only. However, this watering, and
312 a small number of precipitation events in the winter months, are assumed to have flushed out some
313 of the metals that had accumulated in the soil as opposed to further enrich it. Finally, due to the time
314 limits on the study, mandarin oranges could only be harvested and tested once.

315 Despite these constraints, the study allows the following conclusions. First, the elevated
316 concentrations of Ba, Cr, and Pb indicate that further research on the potential risks associated with
317 the reuse of OPW in agricultural crop irrigation is needed. The field sampling conducted by the
318 Food Safety Expert Panel for citrus, almonds, pistachios, table grapes and garlic tested for 108
319 elements and chemical compounds and detected 16 in the fruit and seed samples, among them Ba,
320 Cu, Mo, Ni, Sr, and Zn. In several cases, concentrations of these metals in the treated crop were
321 statistically higher than in the controls, but were still judged to be “non-alarming.” The present
322 study indicates that field sampling should be continued at least for the citrus fruit to reaffirm that
323 there is no risk for consumers. The results further signal the need to conduct a systematic soil study
324 to assess the risk of heavy metal accumulation for soil and crop health. The Regional Water Quality
325 Control Board is in the process of planning additional studies for the new growing season to
326 supplement the existing data. These studies aim to include root vegetables such as garlic, potatoes
327 and carrots, which may have a higher propensity to store heavy metals and other contaminants than
328 tree crops. The present study suggests that soil analysis should be included as well.

329 The question of whether OPW use in agriculture poses risk to consumers of the impacted crops also
330 requires testing the long-term effects of OPW application, because crops with very low to low
331 concentrations of compounds with known adverse health effects may pose risks if the crop is
332 consumed in large-enough quantities by consumers (e.g., the increase in consumption of mandarin
333 oranges between October – February).

334 Since OPW chemical composition can change over time, often within short time periods, it remains
335 advisable that OPW water quality testing continues for at least key contaminants that carry
336 substantial health risks at even small concentrations such as heavy metals, BTEX, and radioactive
337 isotopes.

338 Adding more advanced treatment of OPW to regular OPW, crop, and soil quality monitoring can
339 further ensure that harmful and toxic elements and chemicals do not reach agricultural products,
340 especially those aimed for human consumption. With sufficient safeguards in place OPW could be
341 considered a supplemental water resource for California's Central Valley farms and hence make for
342 a mutually beneficial partnership between the oil and agricultural industry in the region.

343 **Supplementary Materials:** The following are available online at www.mdpi.com/link:

344 R Markdown Scripts for Ag, Ba, Cr, and Pb.

345 Data files for Ag, Ba, Cr, and Pb.

346 Figure S1: Boxplots of measured concentrations of Cr in the leaves as a function of Cr concentration in the
347 irrigation water (low, medium, high) and sampling time, t .

348 Figure S2: Boxplots of measured concentrations of Pb in the leaves as a function of Pb concentration in the
349 irrigation water (low, medium, high) and sampling time, t .

350 Figure S3: Boxplots of measured concentrations of Pb in the pith, juice and peel portions of the harvested
351 mandarin oranges as a function of Pb concentration in the irrigation water (low, medium, high).

352 Figure S4: Boxplots of measured concentrations of Ag in the soil as a function of Ag concentration in the
353 irrigation water (low, medium, high) and sampling time, t .

354 Figure S5: Boxplots of measured concentrations of Ag in the leaves as a function of Ag concentration in the
355 irrigation water (low, medium, high) and sampling time, t .

356 Figure S6: Boxplots of measured concentrations of Ag in the pith, juice and peel portions of the harvested
357 mandarin oranges as a function of Ag concentration in the irrigation water (low, medium, high).

358 **Acknowledgments:** The authors wish to thank Penny Manisco and Daniel Guerra of the Harvey Mudd College
359 Chemistry Department, Troy Hansgen and the college grounds crew, Marc los Huertos, Jade Star Lackey, and
360 Kyle McCarty (Pomona College), and Colin Robins and Branwen Williams (Keck Science Department of
361 Claremont McKenna, Pitzer, and Scripps College) for their help in conducting the study. Funding was provided
362 by the Hixon Center for Sustainable Environmental Design and the Office of Community Engagement through
363 the Nathaniel Prize.

364 **Author Contributions:** T.S. and H.v.R. conceived, planned the study, and ran the statistical analysis. T.S.
365 drafted the manuscript and H.v.R. and A.Z. provided edits. A.Z. and H.v.R. processed and analyzed the
366 samples. V.C. and B.P. assisted with solution preparation, watering, and sample preparation.

367 **Conflicts of Interest:** The authors declare that they have no conflicts of interest.

368

369 **References**

- 370 1. Energy Information Administration. Petroleum & Other Liquids, Crude Oil Production.
371 https://www.eia.gov/dnav/pet/pet_crd_crpdn_adc_mbbbl_a.htm (accessed 3 April 2018).
- 372 2. Fakhru'l-Razi, A., Pendashteh, A., Abdullah, L.C., Biak, D.R.A., Madaeni, S.S. and Abidin, Z.Z., 2009.
373 Review of technologies for oil and gas produced water treatment. *Journal of hazardous materials*, 170(2),
374 pp.530-551.
- 375 3. Stephenson, M.T., 1992. A survey of produced water studies. In *Produced Water* (pp. 1-11). Springer US.
- 376 4. Chittick, E.A. and Srebotnjak, T., 2017. An analysis of chemicals and other constituents found in produced
377 water from hydraulically fractured wells in California and the challenges for wastewater management. *Journal*
378 *of Environmental Management*, 204, pp.502-509.
- 379 5. Benko, K.L. and Drewes, J.E., 2008. Produced water in the Western United States: geographical
380 distribution, occurrence, and composition. *Environmental Engineering Science*, 25(2), pp.239-246.
- 381 6. Heberger, M. and Donnelly, K., 2015. Oil, Food, and Water: Challenges and Opportunities for California
382 Agriculture. Pacific Institute, Oakland, CA.
- 383 7. Cawelo Water District, <http://www.cawelowd.org/PrdWater.html> (accessed 3 April 2018).
- 384 8. Gunderson, J., Water Reuse Gains Traction in Conventional Oil and Gas Industry. *Industrial WaterWorld*,
385 vol. 16, issue 5.
386 <http://www.waterworld.com/articles/iww/print/volume-16/issue-5/features/water-reuse-gains-traction-in-conventional-oil-and-gas.html> (accessed 3 April 2018).
- 388 9. California Water Boards, Central Valley Region. Oil Fields – Food Safety,
389 https://www.waterboards.ca.gov/centralvalley/water_issues/oil_fields/food_safety/ (accessed 3 April 2018).
- 390 10. USDA, Natural Resources Conservation Service (2000). Heavy Metal Soil Contamination.
391 https://www.nrcs.usda.gov/Internet/FSE_DOCUMENTS/nrcs142p2_053279.pdf (accessed 3 April 2018).
- 392 11. Oliveira, A. and Pampulha, M.E., 2006. Effects of long-term heavy metal contamination on soil microbial
393 characteristics. *Journal of Bioscience and Bioengineering*, 102(3), pp.157-161.
- 394 12. Obiajunwa, E.I., Pelemo, D.A., Owolabi, S.A., Fasasi, M.K. and Johnson-Fatokun, F.O., 2002.
395 Characterisation of heavy metal pollutants of soils and sediments around a crude-oil production terminal using
396 EDXRF. *Nuclear Instruments and Methods in Physics Research Section B: Beam Interactions with Materials and*
397 *Atoms*, 194(1), pp.61-64.
- 398 13. Mulligan, C.N., Yong, R.N. and Gibbs, B.F., 2001. Remediation technologies for metal-contaminated soils
399 and groundwater: an evaluation. *Engineering Geology*, 60(1), pp.193-207.
- 400 14. Alloway, B.J., 2013. Sources of heavy metals and metalloids in soils. In *Heavy Metals in Soils* (pp. 11-50).
401 Springer Netherlands.
- 402 15. Järup, L., 2003. Hazards of heavy metal contamination. *British Medical Bulletin*, 68(1), pp.167-182.
- 403 16. Sparks, D.L., 2003. *Environmental Soil Chemistry*. Academic Press.
- 404 17. Fliessbach, A., Martens, R. and Reber, H.H., 1994. Soil microbial biomass and microbial activity in soils
405 treated with heavy metal contaminated sewage sludge. *Soil Biology and Biochemistry*, 26(9), pp.1201-1205.

- 406 18. Rattan, R.K., Datta, S.P., Chhonkar, P.K., Suribabu, K. and Singh, A.K., 2005. Long-term impact of
407 irrigation with sewage effluents on heavy metal content in soils, crops and groundwater—a case
408 study. *Agriculture, Ecosystems & Environment*, 109(3), pp.310-322.
- 409 19. Khan, A.G., Kuek, C., Chaudhry, T.M., Khoo, C.S. and Hayes, W.J., 2000. Role of plants, mycorrhizae and
410 phytochelators in heavy metal contaminated land remediation. *Chemosphere*, 41(1), pp.197-207.
- 411 20. Toze, S. Reuse of effluent water—benefits and risks. *Agricultural Water Management*, 2006, 80, 147-159.
- 412 21. Khan, S.; Cao, Q.; Zheng, Y. M.; Huang, Y. Z.; Zhu, Y. G. Health risks of heavy metals in contaminated
413 soils and food crops irrigated with wastewater in Beijing, China. *Environmental Pollution*, 2008, 152, 686-692.
- 414 22. Jung, M. C. Heavy Metal Concentrations in Soils and Factors Affecting Metal Uptake by Plants in the
415 Vicinity of a Korean Cu-W Mine. *Sensors*, 2008, 8, 2413-2423.
- 416 23. Stringfellow, W. *Preliminary Analysis of 2017 Crop Data*, Presentation given at Food Safety Panel, Meeting of
417 January 24, 2018.
418 [https://www.waterboards.ca.gov/rwqcb5/water_issues/oil_fields/food_safety/meetings/2018_0124_crop_sampl](https://www.waterboards.ca.gov/rwqcb5/water_issues/oil_fields/food_safety/meetings/2018_0124_crop_sampling_update.pdf)
419 [ing_update.pdf](https://www.waterboards.ca.gov/rwqcb5/water_issues/oil_fields/food_safety/meetings/2018_0124_crop_sampling_update.pdf) (accessed 3 April 2018).
- 420 24. U.S. EPA, 2016. Hydraulic Fracturing for Oil and Gas: Impacts from the Hydraulic Fracturing Water Cycle
421 on Drinking Water Resources in the United States (Final Report). U.S. Environmental Protection Agency,
422 Washington, DC, EPA/600/R-16/236F, 2016.
- 423 25. Shaw, R. G., and Mitchell-Olds, T., 1993. ANOVA for unbalanced data: an overview. *Ecology*, 74(6),
424 1638-1645.
- 425 26. R Core Team, 2013. R: A language and environment for statistical computing. <http://r-project.org/>
- 426 27. EPA. Drinking Water Contaminants and Standards. <https://www.epa.gov/dwstandardsregulations>
427 (accessed 3 April 2018).