Human, Urban and Environmental-Induced Alterations in Mangroves Pattern along Arabian Gulf Coast, Eastern Province, KSA

Sameh A. Amin; Mai S. Fouad; Wafaa A. Altaisan & Mohammad A. Zyada

Department of Biology, Faculty of Science, University of Dammam, K.S.A.
Department of Botany, Faculty of Science, University of Fayoum, Egypt.
Department of Biology, Faculty of Science, University of Damietta, Egypt.

Abstract: Significant changes have occurred at Arabian Gulf Coast of Saudi Arabia over the last three decades. The area of mangrove was reduced by about 55%. Thus the economic, social, and environmental value of mangroves must be assessed over short-to long-term scales and employ these assessments for the awareness rising at local communities. This study provides a preliminary assessment of the risks on mangrove vegetation; it will provide database to mitigate the tremendous pressures due to coastal development and urban activities. The effects of human development on the mangrove plant cover in Eastern Region of KSA were recognised, during 2013 to 2016. The mean variations of physiochemical characteristics in water and sediment were recorded. With regard to water analyses including; nitrogen, phosphorus, TSS, TDS, BOD, and turbidity were evaluated. On the other hand, for sediment: TDS, nitrogen, phosphorus, sulphate and total organic carbon were assessed. Moreover, the growth parameters: plant height, and size index, of Avicennia marina were recorded and estimated. It is concluded that human impact and urban developments have exerted drastic effects on the coastal ecosystems and its environments.

Key Words: Mangrove; Avicennia marina; Size Index; Sediment; Urban Development; Human Impacts.

Introduction

Mangroves are considered as a crucial component of the world’s coastal ecosystems representing an integrated environmental system to stabilize coastal lands and offer protection against storms and sea-level rise (Mukherjee et al., 2010). They alleviate devastating effects of erosion, storm surges and flooding of coastlines (Zhang et al., 2012). Moreover, The importance of coastal intertidal mangrove habitats comprises their contributions in increasing the sedimentation rate (Thampanya et al., 2006), acting as a physical and biogeochemical barrier for contaminants attacking coastal estuaries and other water bodies (Chowdhury et al., 2016; Qiu et al., 2011) and helping in the accumulation and partitioning of trace elements in the rhizosphere (Zhou et al., 2011). Considered as woody plants mangroves inhabit intertidal zones with high salinity (Shan et al., 2008; Parida and Jha, 2010) and can tolerate a wide range of salinities under natural conditions (Suárez et al., 1998). Also mangroves have wide ecological amplitude ranging from tropical to temperate regions across all continents and dominate large extents of shorelines, estuaries and islands in tropical and subtropical regions worldwide forming biologically important and productive transitional coastal ecosystems (Alongi, 2002; Soares et al., 2012; Wang et al., 2013). As mangroves are considered as assemblages of trees and shrubs, they sustain both ecological and economical services in ecosystems (Alongi, 2009), and playing an important role in both biogeochemical cycles and economic activities (Thua and Populus, 2006). About 80 species of true mangrove trees/shrubs are recognised, from which around 50-60 species make a significant contribution to the structure of mangrove forests. Species diversity is much higher

*e-mail: aminsameh2012@gmail.com
in the Southeast Asian Region, where approximately two-thirds of all species are found, while
approximately 15 species exist in Africa and 10 species in the Americas (FAO, 2007). It is known that
more than 90% of the world’s mangroves are located in developing countries (Duke et al., 2007). The trees of
several genera are economically valuable for timber or fuelwood, especially Rhizophora species. Since,
more than one-third of the global human population lives along coastal areas, their long-term
sustainability depends on the coastal ecosystems (Barbier et al., 2008). However, human activities and
interventions within and near these mangrove areas have led to their degradation and the resources therein
(Alongi, 2002; FAO, 2007). Large areas of mangrove have been converted into fishponds (Ellison, 2008), salt
ponds, agriculture and coastal projects (Ong et al., 1995). Over the globe, due to disturbance of species
distribution, mangroves have been disappearing at an alarming rate worldwide by an annual rate of 1–2%
(FAO, 2007; Lewis, 2009). The loss of mangrove forests has increased from regions of highly
anticipated rise in global temperature (Koch et al., 2015). Thus mangroves may vanish if the
destruction of their ecosystems continues repeatedly (Duke et al., 2007).

Mangrove plant cover are found along some coastlines of the arid Arabian Peninsula
(Balakrishnan, 2012). They are present in the form of fragmented stands in many tidal areas on the
Red Sea and the Arabian Gulf coast, south of latitude 26° north. They consist mainly of Avicennia
marina trees. On the coast of the Red Sea, Avicennia marina is accompanied by a few of Rhizophora
mucronata, however it is very rare in Saudi Arabian Gulf (Thua and Populus, 2006). Mangrove
ecosystems are limited along Arabian Gulf and they are confined to Dammam area (Taraut Bay), with
well-developed communities consisting of Avicennia marina. The inter-tidal mangrove environment
of Eastern Region of Arabian Coast is very important as it supports the local fishing activities, nursery
grounds for many fish, crustaceans and shellfish species, as well as being central for ecotourism
activities. Also due to the consistently diminishing erosion resistance, the extent of mangrove plant
cover along the Saudi Arabian Gulf has been considerably decreased. Moreover, in the Arabian Gulf
area mangrove ecosystems have been principally affected by the large oil spills from the Gulf War
(Saenger, 1993). It is also threatened by the expansion of human settlements, the boom in commercial
aquaculture, the impact of tidal waves and storm surges. Noticeably, in Taraut Bay it has exhibited
drastic reduction due to major stressors including landfilling, dredging, coastal development, solid
and liquid waste disposal (Danish, 2010).

Fortunately, members of the genus Avicennia are dominant within higher latitude forests and
documentary they have expanded their range in recent decades along three continents (Saintilan et
al., 2014). It is not too late to renew the loss in productivity of the mangrove areas, and their plant cover can be
rehabilitated and maintained (Palis, 1998). The present study is conducted to frame the human impact
assessing the physico-chemical parameters on mangrove vegetation in Dammam area (Fig. 1),
especially at Taraut Island. The main objective will be extended to estimate the reduction of
mangrove ratio during the last decades and to throw a highlight spot on mangrove future over the
next few decades.

Materials and Methods

Study Site:

Satellite, Google Earth, pictures, images and geographic information systems provide
collectively useful tools to detect and map the temporal variation in mangroves coverage (Riaza, et
al. 1998; Long and Skewes, 1996; Verstraete et al. 2008). For this purpose the temporal changes in the
geographic distribution of mangroves along the Arabian Gulf Coast at Dammam Region have been
studied by conducting the historical Landsat Multispectral Scanner (MSS) and Landsat Enhanced
Thematic Mapper (ETM) adopted by Khan and Kumar (2009) (Figs 2-4). New changes in these maps
and images were made in the present work through field studies (Fig. 1).
Fig. (1): Map of the sampling sites at the Eastern Region of Arabian Gulf, KSA.

Fig. (2): Satellite image of mangroves area (622 ha) in 1973.
Fig. (3): Satellite image of mangroves area (482 ha) in 1999. After (Khan and Kumar, 2009).

Fig. (4): Satellite image of mangroves area (green color), and red circles are threatened areas endangering their survival during the present decade.

Field Measurements:

a) Vegetation:

Nine locations were selected across mangrove forest. Three transects were outlined and laid out in such a way to represent the variations of mangrove trees at each site. The abundance and growth parameters (the plant height, size index, leaf area, number of main and lateral branches, number and height of aerial roots and number of seedlings/m²) of mangrove were measured at each site to evaluate the growth rate.

B) Water and Sediment Sampling And Analysis:
Across a distance of about 5 to 10 km along the Arabia Gulf coast of Eastern Region, KSA (Fig. 1), monthly visits were arranged to the studied area throughout the period from January 2013 to January 2016 collecting water and sediment samples from the same experimental locations. Surface water and sediment were gathered randomly from a depth of 0–15 cm. Surface water and sediments were analyzed to determine both physical and chemical characteristics. The concentration in water and sediment are expressed in mg/L and in mg/kg dry weight respectively.

The experimental results were extended to measure pH and turbidity of the water in-situ (in duplicate); using the electrometric method for pH and the Nephelometric method for turbidity as per standard methods (APHA, AWWA and WEF 2005). The biological oxygen demand (BOD) was determined following the method described in the APHA, AWWA and WPCF (1985) while TDS is measured according to standard method of APHA (2005).

Consequently, the collected sediment samples were air dried and then crushed for further analysis. A 5 g of field-moist soil was thoroughly mixed with 25 ml distilled water in polyethylene centrifuge tubes and placed on a spinning wheel for 2 h. on termination of the shaking the soil-water slurry was left to settle for 10 min and its pH was measured using an Orion 290A pH meter (Birch et al., 2011). Also, the sediment organic matter (TOC) was determined according to Walkley and Black method (1934).

Phosphorus concentration was estimated using a colorimetric assay as described in Reef & Lovelock (2014). Also the method of Flindt and Lillebo (2005) was used to quantify the Nitrogen content in the soil samples. It has to be noted that this method is a modification of the standard Kjeldhal-N method. Furthermore, soil samples were analyzed for Sulphates estimation according to the method adopted by Jackson (1967) by titration against BaCl₂ in presence of tetrahydroxy-quinone as indicator.

C) Human Impact Assessment:

As Tarout Island is predominantly considered as one of the most ancient sites that were inhabited by humans, moreover, the Island had a significant role in trading purposes in the entire Arabian-Gulf region (https://en.wikipedia.org/wiki/Tarout_ Island). Since the island has a legacy of severe human-induced environmental degradation over its existence, therefore the effect of landfilling, dredging, coastal development and solid waste disposal are presently studied regarding mangrove ecosystem. Therefore, it is mandatory in this present work to quantify anthropogenic disturbance and fouling in terms of degrees (0,-1,-2) on the biodiversity of the sites under investigation.

Statistical Analysis

Analyses of variance (ANOVA) for the water, sediment and mangrove abundance (one-way and two-ways) were carried out. The same analyses was done for mangroves biodiversity. This analysis showed a strong significant difference for both one-way (sites) and two-ways (sites and parameters), P <0.05: 0.006.

Results and Discussion

Water and Sediment Analysis:

The recorded data in Tables (1 and 2), clarifies the values of physico-chemical characteristics of water and soil at the investigated sites of Arabian Gulf coast in Saudi Arabia, respectively. The average measured values of pH are within the alkaline range that enables and supports biological life in agreement with previous studies (GreenTech Consultants 2009; Harris and Vinobaba 2012). Mangroves are well known for their halophytic characteristics (Gong and Ong, 1990), which allow them to survive in high salinity through certain mechanisms of salt tolerance (Wakushima et al., 1994). However, total dissolved solids (TDS) exhibits normal concentrations with relatively steady rates in most samples. Meanwhile, site V showed the highest total suspended solids (TSS) magnitude (5.4 mg/L) followed by site VI. The rest of specimens showed a descending trend. Several studies have reported that Avicennia marina is largely predominant in high saline environments, where salinity is
shown to be ≥ 25 ppt during most of the months of the year (Karunathilake 2003; Jayatissa et al. 2008).

Moreover, Avicennia marina is a facultative halophyte having various adaptations for hypersaline environments (Shete et al. 2007; Jayakody et al. 2008).

Table (1): Some physico-chemical features of water supporting the growth of mangroves at inspected sites of Arabian Gulf Coast, Eastern Region, KSA.

<table>
<thead>
<tr>
<th>Parameter</th>
<th>Site</th>
<th>pH</th>
<th>Turbidity (NTU)</th>
<th>Total Dissolved Solids (TDS) (mg/L)</th>
<th>Total Suspended Solids (TSS) (mg/L)</th>
<th>Total Phosphorus (TP) (mg/L)</th>
<th>Total Nitrogen (TN) (mg/L)</th>
<th>Biochemical Oxygen Demand (BOD) (mg/L)</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>I</td>
<td>8.1±0.5</td>
<td>5.8±0.9</td>
<td>44500±4000</td>
<td>1.8±0.4</td>
<td>2.4±0.4</td>
<td>1.4±0.2</td>
<td>7.9±0.4</td>
</tr>
<tr>
<td></td>
<td>II</td>
<td>8.1±0.4</td>
<td>6.1±0.8</td>
<td>41300±2400</td>
<td>1.5±1.1</td>
<td>2.1±0.2</td>
<td>1.3±0.3</td>
<td>8.8±1.1</td>
</tr>
<tr>
<td></td>
<td>III</td>
<td>8.1±0.4</td>
<td>12.2±1.2</td>
<td>41500±3000</td>
<td>1.8±0.5</td>
<td>2.1±0.3</td>
<td>1.4±0.2</td>
<td>9.8±2.1</td>
</tr>
<tr>
<td></td>
<td>IV</td>
<td>8.0±0.5</td>
<td>8.2±1.8</td>
<td>42000±1700</td>
<td>1.9±1.1</td>
<td>3.1±0.8</td>
<td>1.6±0.2</td>
<td>9.6±1.6</td>
</tr>
<tr>
<td></td>
<td>V</td>
<td>7.9±0.5</td>
<td>8.2±1.1</td>
<td>36500±3100</td>
<td>5.4±2.7</td>
<td>5.1±1.1</td>
<td>6.5±0.9</td>
<td>18.8±3.1</td>
</tr>
<tr>
<td></td>
<td>VI</td>
<td>8.0±0.3</td>
<td>7.2±1.3</td>
<td>37800±2000</td>
<td>3.1±0.9</td>
<td>3.9±0.7</td>
<td>3.1±0.6</td>
<td>10.8±1.3</td>
</tr>
<tr>
<td></td>
<td>VII</td>
<td>8.1±0.4</td>
<td>5.5±0.7</td>
<td>38000±3000</td>
<td>1.6±0.9</td>
<td>3.4±0.9</td>
<td>2.2±0.4</td>
<td>9.2±2.4</td>
</tr>
<tr>
<td></td>
<td>VIII</td>
<td>8.2±0.5</td>
<td>4.5±0.6</td>
<td>40200±2200</td>
<td>2.2±1.1</td>
<td>2.9±0.6</td>
<td>2.1±0.3</td>
<td>8.9±1.1</td>
</tr>
<tr>
<td></td>
<td>IX</td>
<td>8.2±1.1</td>
<td>5.2±1.1</td>
<td>40500±2400</td>
<td>2.1±1.1</td>
<td>2.9±0.7</td>
<td>2.3±0.2</td>
<td>8.8±2.3</td>
</tr>
</tbody>
</table>

(mean values ± SE)

Table (2): Some physico-chemical features of sediment supporting the growth of mangroves at inspected sites of Arabian Gulf Coast, Eastern Region, KSA.

<table>
<thead>
<tr>
<th>Parameter</th>
<th>Sites</th>
<th>pH</th>
<th>Total Dissolved Solids (TDS) (mg/kg)</th>
<th>Total Organic Carbon (TOC) (mg/kg)</th>
<th>Total Phosphorus (TP) (mg/kg)</th>
<th>Total Nitrogen (TN) (mg/kg)</th>
<th>Sulfate (mg/kg)</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>I</td>
<td>8.0±0.3</td>
<td>39500±3000</td>
<td>4.8±0.6</td>
<td>2.2±0.1</td>
<td>1.9±0.3</td>
<td>3500±450</td>
</tr>
<tr>
<td></td>
<td>II</td>
<td>8.0±0.2</td>
<td>38700±2200</td>
<td>4.5±0.7</td>
<td>2.1±0.1</td>
<td>1.8±0.3</td>
<td>3000±400</td>
</tr>
<tr>
<td></td>
<td>III</td>
<td>7.9±0.3</td>
<td>37500±2500</td>
<td>5.8±0.5</td>
<td>2.2±0.2</td>
<td>1.9±0.4</td>
<td>2400±260</td>
</tr>
<tr>
<td></td>
<td>IV</td>
<td>7.9±0.4</td>
<td>37000±2000</td>
<td>7.9±1.1</td>
<td>3.3±0.6</td>
<td>2.3±0.4</td>
<td>2500±200</td>
</tr>
<tr>
<td></td>
<td>V</td>
<td>7.7±0.3</td>
<td>33500±2400</td>
<td>11.4±2.7</td>
<td>5.4±0.9</td>
<td>5.9±0.7</td>
<td>1600±200</td>
</tr>
<tr>
<td></td>
<td>VI</td>
<td>8.0±0.4</td>
<td>34800±2100</td>
<td>7.1±1.9</td>
<td>3.6±0.4</td>
<td>3.2±0.4</td>
<td>1800±200</td>
</tr>
<tr>
<td></td>
<td>VII</td>
<td>8.0±0.3</td>
<td>35000±2300</td>
<td>6.6±0.9</td>
<td>3.1±0.3</td>
<td>2.4±0.3</td>
<td>2100±300</td>
</tr>
<tr>
<td></td>
<td>VIII</td>
<td>8.0±0.4</td>
<td>36200±2000</td>
<td>5.2±1.1</td>
<td>2.8±0.6</td>
<td>2.6±0.4</td>
<td>2900±400</td>
</tr>
<tr>
<td></td>
<td>IX</td>
<td>8.0±0.3</td>
<td>36500±2500</td>
<td>5.1±1.1</td>
<td>2.9±0.4</td>
<td>2.9±0.7</td>
<td>3100±500</td>
</tr>
</tbody>
</table>

(mean values ± SE)

As turbidity measurement could be employed to provide an estimation of total dissolved solids, it has been examined and was found to have the highest values in site III (12.2 NTU) followed by sites IV and V when compared to other locations. Since Phytoplankton, sediments from erosion and re-suspended sediments arising from agitation of the bottom sediments comprising some detritus particles are light they remain suspended in water. Table (3) exhibits a list of anthropogenic actions including waste discharge, urban runoff, and fishing yielding to an aggravated problem. Similar deductions were demonstrated by Guhathakura and Kaviraj (2004).

Regarding the total phosphorus and nitrogen, biochemical oxygen demand BOD and total organic carbon TOC, site V followed by site IV recorded the maximum values when compared to
other sites. Site I showed the lowest concentrations. The largest values of these parameters in sediment could be referred to the increased concentration in water (Zyadah, 2011). Sulfur is an essential chemical component which is used as sulfate ($\text{SO}_4^{2-}$) for the structure of the plants and different biological processes (Hopkins and Hüner, 2004). For all the investigated sites, the amount of sulfates ($\text{SO}_4^{2-}$) exceeds 1600mg/kg, a result that is emphasized by Reddy and D’Angelo (1997) stating that the minimum value of the sulfates in the coastal area must not be less than 2 ppm. On the other hand, the allowable range of nitrogen in the soil is between 10 ppm to 60 ppm (Gong and Ong, 1990). Hence, nitrogen is lower than the normal range over all sites. The distribution of $P$ and $N_2$ are associated with organic matter. Apart from site V, variations recorded for the sites are relatively small, and may be associated with contrasting sedimentation patterns, which is in accordance with Marinho et al. (2012).

Table (3): Anthropogenic impact on the biodiversity at the inspected sites.

<table>
<thead>
<tr>
<th>Site</th>
<th>I</th>
<th>II</th>
<th>III</th>
<th>IV</th>
<th>V</th>
<th>VI</th>
<th>VII</th>
<th>VIII</th>
<th>IX</th>
</tr>
</thead>
<tbody>
<tr>
<td>Diving &amp; Boat anchor</td>
<td>0</td>
<td>-2</td>
<td>0</td>
<td>0</td>
<td>0</td>
<td>-1</td>
<td>0</td>
<td>0</td>
<td>0</td>
</tr>
<tr>
<td>Fishing</td>
<td>-2</td>
<td>-2</td>
<td>-1</td>
<td>0</td>
<td>0</td>
<td>-1</td>
<td>0</td>
<td>0</td>
<td>0</td>
</tr>
<tr>
<td>Sewage drainage water</td>
<td>0</td>
<td>-2</td>
<td>-2</td>
<td>-1</td>
<td>-1</td>
<td>0</td>
<td>-1</td>
<td>-1</td>
<td>-1</td>
</tr>
<tr>
<td>Agricultural waste water</td>
<td>0</td>
<td>0</td>
<td>-1</td>
<td>-2</td>
<td>-1</td>
<td>0</td>
<td>-1</td>
<td>-1</td>
<td>0</td>
</tr>
<tr>
<td>Trampling</td>
<td>-1</td>
<td>-1</td>
<td>-2</td>
<td>-2</td>
<td>-1</td>
<td>-2</td>
<td>-2</td>
<td>-1</td>
<td>0</td>
</tr>
<tr>
<td>Grazing (sheep and camels)</td>
<td>0</td>
<td>-1</td>
<td>-1</td>
<td>-1</td>
<td>0</td>
<td>-1</td>
<td>0</td>
<td>0</td>
<td>0</td>
</tr>
</tbody>
</table>

Impact degree: 0 is neutral, -1 negative impact, -2 much negative impact.

Figure (5) takes into account the human impact illustrating the quantity of waste water drained in the sites under consideration. Thus, the elevated amount of nutrients in water and sediment may be attributed to the increase of sewage and agricultural drainage water at these sites, wastes from ships and boats and in addition to the decomposition of dead plants, particularly at site V which could be referred to the increased levels of organic carbon. TOC levels imply that the sediments inside Mangrove forest could retain organic matter. Additionally, the decayed litter released nutrients into sediments, which would contribute to the increase of organic matter in the sediments. As a consequence, Mangrove plants improve the sediment fertility and promote the physicochemical properties of the intertidal habitat, which would influence the biogeochemical processes of other elements in the sediment (Li et al., 2016). During Mangrove succession, the diversity and biomass of the plant community increase; as a consequence, the matured Mangrove community provides an abundance of sediment biogenic matter, such as SOM, TC, TN, TP and TK (Chen et al., 2016).
Plasticity of mangrove seedlings in environmental conditions is common but has mainly been related to other abiotic factors such as light or water availability (Sultan 2000). Owing to the richness of some investigated locations with nutrients, hence the high abundance and growth “plant height, size, number of main and lateral branches” of *Avicinia marina* at sites V, VII and VIII were noticed as compared to the other sites, while the lowest density was obtained at sites I, IV and IX. Human impact was recorded at all sites, where urban development was clear in sites IV, V, VI and VIII.
Significant changes have occurred in Saudi Arabia Gulf coast over the last three decades. Figs. 2 and 3 showed the Satellite image of mangroves area (622 ha) in 1973 which is reduced to an area of 482 ha in 1999 (Khan and Kumar, 2009). Many aspects of development were recorded in Dammam and Qatif regions that negatively affect the Mangrove plant cover. Fig.(4) shows the last urban disturbance on mangrove which results in plant cover reduction from 482 ha to 390 ha during 1999 to 2009. In a close observation, Fig.(6) shows the landfilling in the studied area during the last four decades. There is a negative impact on the mangrove occurrence as a result of the coastal development that include dredging, filling, and other activities like oil pollution and drainage water (Fig. 5).

In conclusion, the current work is an attempt to demonstrate threats and difficulties that are facing mangroves in their natural habitats. Also, this work outlines the severity of ecological implications that disturb the coastal ecotone where mangrove thrives. In addition, the research emphasizes the ability of *Avicennia marina* to overcome contamination and human disturbances in the ambient environment. The economic, social, and environmental value of Mangrove must be assessed for awareness rising of local communities. However, greater losses in fringe Mangroves and an increased area of basin mangroves can be related to sea-level rise and increased mean air temperatures that in turn related to current climate change (Campbell *et al.*, 2010). Climate change is a futuristic problem creating stress to the entire biosphere. The overexploitations of mangrove forest and oil pollution are considered the main destruction factors in the Arabian Gulf (Zahed *et al.*, 2010). Wide distribution of mangrove is ascribed to its adaptation resilience. A conducted trial In Saudi Arabia, remote sensing data was used to locate suitable sites for Mangrove plantation along the Red Sea Coast (Abd-El Monsef *et al.*, 2013).

Author Contributions: S. A. Ismail, M. Zyadah and M. S. Fouad conceived and designed the experiments. S. A. Ismail and W. Ataisan performed the experiments. S. A. Ismail, analyzed the data. S. A. Ismail, M. Fouad, W. Ataisan and M. Zyadah wrote the manuscript; S. A. Ismail and M. S. Fouad provided editorial advice.
Conflict of Interests: The authors declare that there is no conflict of interests regarding the publication of this paper.

References


Phytoremediation: Management of Environmental contamination, 4, 283-310 Springer.


Verstraete, M. M., Brink, A. B., Scholes, R. J., Beniston, M. and Smith, M. S. (2008). Climate change and desertification: where do we stand, where should we go? Global and Planetary Change, 64, 105-110.


