

1 **Filtration conditions for the removal of organic matter in eutrophic waters**
2 **by freshwater mussels using response surface methodology**

3

4 Hwan-Seok Choi^{1a}, Young-Hyo Kim^{2a}, Hyuk Lee³, David C. Aldridge⁴ and Baik-Ho Kim^{5*}

5

6 ¹Research Institute for Coastal Environment and Fishery-policy, Gwangju 61436, Korea

7 ²Department of Environmental Sciences, Hanyang University, Seoul 04763, South Korea

8 ³National Institute of Environmental Research, Incheon 22689, South Korea

9 ⁴Department of Zoology, University of Cambridge, Downing Street, Cambridge, CB2 3EJ,

10 United Kingdom

11 ⁵Department of Life Science and Research Institute for Natural Sciences, Hanyang University,

12 Seoul 04763, South Korea

13

14 a: co-first authors

15

16 *Correspondence: B.-H. Kim (tigerk@hanyang.ac.kr)

17 Department of Life Science

18 Hanyang University

19 Seoul 04763, South Korea

20 Office: 82-2-2290-0960, C.P: 82-10-7351-2510

21

22

23 **Abstract**

24 In this study, we applied a central composite design to estimate independent variables and
25 establish optimal conditions of filtration rate and feces production that enhance filtration of
26 suspended organic matter by the freshwater mussels *Sinanodonta woodiana*. The results
27 indicated that statistical design methodology offers an efficient and feasible approach for high
28 filtration and low feces production condition optimization. The proposed model equation
29 takes into account the quantitative effect of variables and also the influence of interactions
30 among variables on mussel filtration rate. Under the optimal experimental conditions (mussel
31 size, 13.0 ± 0.2 cm; water current, 17.5 L/h), the experimental filtration rate of 4.47 ± 1.82
32 L/mussel/h showed a degree of correspondence with the predicted value of 8.4 L/mussel/h,
33 which verified the practicability of this optimum strategy.

34

35 **Keywords** : current, filtration rate, freshwater bivalve, *mussel size*, *response surface*
36 *methodology*, *Sinanodonta woodiana*

37

38 1. Introduction

39 Bivalve mollusks often comprise the highest biomass in the benthos of freshwater and marine
40 ecosystems (Newell *et al.* 2004). Their filter-feeding activity removes phytoplankton and
41 other suspended matter from the water column, both through ingestion and sedimentation of
42 particles in feces and pseudofeces (PF). The ecosystem engineering achieved through bivalve
43 filtration can result in improved light penetration within the water, thus, facilitating the
44 growth of bottom-rooting macrophytes, which in turn can provide habitats for other biota
45 (Fanslow *et al.* 1995; Davenport *et al.* 2000). Indeed, studies in freshwater systems have
46 shown that greater biological richness is associated with greater abundances of unionid
47 mussels, both between different river systems (Aldridge *et al.* 2007) and even within the
48 same lake (Chowdhury *et al.* 2016).

49 The potential for harnessing the filtration capacity of freshwater bivalves as biofilters has
50 been recognized with regards to the treatment of drinking water (Lammens *et al.* 2004;
51 McLaughlan & Aldridge 2013). In The Netherlands, introduced zebra mussels (*Dreissena*
52 *polymorpha*) were found to stabilize a phosphorus-enriched lake with clear water (Secchi
53 depth > 1 m) for long periods (Ibelings *et al.* 2007). Zebra mussels have been reported to
54 filter a wide range of plankton, from bacterioplankton to zooplankton, at a rate of
55 approximately 1 L/mussel/day⁻¹ and have also been reported to improve water clarity (Elliot *et*
56 *al.* 2008). The filtration rates of suspension feeders that play an important role in benthic and
57 pelagic coupling by filtering material in the water column vary based on many factors,
58 including species, individual size, water velocity, and water temperatures (Comeau *et al.*
59 2008).

60 Selecting the conditions that can provide the optimal clearance of suspended material by
61 bivalves is an important step in developing effective biofiltration systems. Response Surface

62 Methodology (RSM) is a statistical technique that can be used for designing experiments,
63 building models, evaluating the effects of several factors, and searching for optimum
64 conditions for desirable responses (Jeong *et al.* 2014). Using RSM, the interactions and
65 relative importance of different parameters can be evaluated using a limited number of
66 planned experiments (Wang *et al.* 2007).

67 The main objective of this study was to identify the optimal conditions under which
68 particulate material is removed from the water column by the freshwater mussel *Sinanodonta*
69 *woodiana* Lea. *S. woodiana* is a large mussel species that is native to the Amur and Yangtze
70 river basins. The species is relatively tolerant of poor water quality and has spread throughout
71 much of southeast Asia and South America as a consequence of fish farming (Kim *et al.*
72 2009). In this study, we investigated the relative importance of shell size, rate of water flow,
73 filtration rate (FR), and production of feces/PF using RSM.

74

75 **2. Materials and methods**

76

77 **2.1 Animal collection and experimental design**

78 *S. woodiana* specimens were collected directly from waterways and streams associated with
79 the Geum and Mankyong rivers in Korea and acclimated in laboratory aquaria for at least 3
80 months. The experimental equipment used is described in detail by Lee *et al.* (2009). To study
81 filtration, we used treatment baths of stainless steel (80 × 80 × 145 cm) with a working
82 volume 500 L. Thirty individuals of test mussels were acclimated in holding aquaria for 18
83 days prior to the commencement of experiments. The acclimation and experimental
84 conditions for filtration by *S. woodiana* were a water temperature of $19 \pm 3^\circ\text{C}$, water flow rate

85 of 24 to 48 L/h, and photoperiod of 12 D:12 L.

86

87 **2.2 Measurement of filtration rate and production of bivalve feces and pseudofeces**

88 The ash-free dry mass (AFDM) of each mussel used in this experiment was measured
89 according to the method of Hwang *et al.* (2004). After separating the whole body of the
90 mussel from the shell and weighing, it was transferred to a heat-resistant vessel, desiccated at
91 100°C for 20 min in a drying oven to a constant mass, and then burned in a muffle furnace at
92 500°C for 2 h (APHA 1995). The AFDM of the mussel body was calculated from the
93 difference in dry weight before and after burning. The filtration rates of the mussels (FR:
94 L/mussel/h) in each experiment were determined using the following equation (Coughlan
95 1969):

$$96 \quad FR = V/M \times \ln(T/C)/t,$$

97 where V is the volume of the experimental reactor (L); M is the total AFDM of the mussels; T
98 and C are the concentrations of suspended solids in water passed through the reactor with and
99 without mussels, respectively; and t (hours) is the duration of the experiment.

100 The production of feces and pseudofeces by mussels was measured simultaneously by
101 collecting sediments from mussels at 3-day intervals for 9 days. The sedimented particulate
102 matter was harvested in treated baths and placed in sterilized dishes, and the weight of the
103 pellet after drying at 70°C for 1 h was measured. The pseudofeces production of mussels was
104 calculated by the difference in the dry weights (mg/g AFDM/h) of the sedimented particulate
105 matter in the reactor treatments with and without mussels as follows:

$$106 \quad PFs = V/M \times \ln(T/C)/t,$$

107 where V is the volume of the experimental chamber (L); M is the total AFDM of the mussels;
108 T and C are the total dry weights of the sedimented particulate matter in the reactor with and
109 without mussels, respectively; and t is the duration of the experiment (hours). Water flow in
110 the chamber was adjusted to 24 L/h and 48 L/h using a water pump.

111

112 **2.3 Experimental design and the modeling of filtration by mussels**

113 The experimental design for modeling of mussel filtration condition related to body size
114 aimed to determine the optimal levels of three variables, namely, mussel size (x_1),
115 experimental time (x_2), and water flow (x_3) on filtration rates and production of feces. Each
116 factor in the design was studied at three variable (Table 1). For a 2^3 central composite design
117 (CCD) with three factors, including six center points, a set of 30 experiments was carried out.
118 All the variables were taken at a central coded value considered as zero. The minimum and
119 maximum ranges of variables investigated and the full experimental plan with respect to their
120 values in actual and coded form are listed in Table 1. Upon completion of experiments, the
121 average maximum filtration rate was taken as the dependent variable or response (Y). A
122 second-order polynomial equation was then fitted to the data using the multiple regression
123 procedure. This resulted in an empirical model that related the response measured to the
124 independent variables of the experiment. For a three-factor system, the model equation is as
125 follows:

$$126 \quad Y = \beta_0 + \beta_1 x_1 + \beta_2 x_2 + \beta_3 x_3 + \beta_{11} x_1^2 + \beta_{22} x_2^2 + \beta_{33} x_3^2 + \beta_{12} x_1 x_2 + \beta_{13} x_1 x_3 + \beta_{23} x_2 x_3,$$

127 where Y is the predicted response; β_0 is the intercept; β_1 , β_2 , and β_3 are linear coefficients; β_{11} ,
128 β_{22} , and β_{33} are squared coefficients; and β_{12} , β_{13} , and β_{23} are interaction coefficients. Data
129 were analyzed using the Minitab statistical software package (Minitab Release 14.12.1,
130 Korea).

131

132 3. Results and Discussion

133

134 3.1 Effect of mussel size, water flow, and retention time on *Sinanodonta* filtration rate

135 Among the experimental mussel sizes (mean value \pm SE: 8.5 ± 1.0 to 11.4 ± 1.8 cm), water
136 flow (12 to 48 L/h) and retention time (1.5 to 22.7 h), the filtration rate ranged from $0.87 \pm$
137 0.17 to 4.47 ± 1.82 L/mussel/h (2.67 ± 1.00 L/mussel/h; Table 2). Filtration rate increased
138 with increasing mussel size and with decreasing water flow rate. The larger mussel group
139 (11.4 ± 1.8 cm) had higher filtration rates than the smaller size group (8.5 ± 1.0 cm).
140 Furthermore, filtration rate in the high water current (48 L/h) was reduced relative to that in
141 the lower current (12 L/h).

142

143 3.2 Optimization by response surface methodology

144 The results of CCD experiments for studying the effects of the three independent variables
145 (mussel size, water flow rate, and retention time) on *S. woodiana* filtration rate is shown in
146 Table 3, along with the mean predicted and observed responses. The regression equation
147 obtained after analysis of variance (ANOVA) produced an R^2 value of 0.7625 (a value of $R^2 >$
148 0.75 indicated the adequacy of the model, P value < 0.05), which ensured a satisfactory
149 adjustment of the quadratic model to the experimental data and indicated that 76% of the
150 variability in the response could be explained by the model. The coefficients of the regression
151 equation were calculated using Minitab and the following regression equation was obtained:

$$152 Y = 18.214 - 10.211x_1 + 10.105x_2 + 12.542x_3 - 12.458x_1^2 - 8.243x_2^2 - 9.549x_3^2 + 13.263x_1x_2 + 17.671x_1x_3$$

153 $x_3 - 15.842x_2x_3$

154 Three-dimensional response surface curves were then plotted to determine the interaction of
155 the experimental components and the optimum of each component required for maximum
156 filtration rate. The response surfaces shown in figures 1 and 2 show the relative effect of two
157 variables (mussel size and water flow) with varying retention times. The coordinates of the
158 central point within the highest contour levels in each of these figures corresponds to the
159 optimum filtration rate and feces production of the respective components. Figure 1 shows
160 the response surface for the interactive factors, mussel size (x_1) and water flow (x_2), when the
161 retention time (x_3) ranged from 1.0 to 24.0 h.

162 The maximum filtration rate of mussels under these conditions was predicted to be 8.4
163 L/mussel/h, corresponding to maximum levels (+1) of mussel size (13.0 ± 0.2 cm) and water
164 flow (17.5 L/h; Fig. 1). However, the curve also indicates that the response varies in response
165 to the velocity of water flow. With an increase in water flow (greater than 30 L/h) and a
166 decrease in retention time, the production of feces by mussels further increased to 11.1 g
167 AFDM/ind./h (Fig. 2). However, the response surface curves did not show curvature. Instead
168 they were flattened, with mussel size having relatively little effect, but with greater feces
169 production under conditions of greater flow (Fig. 2). Figure 3 indicates that a greater amount
170 of feces was produced at a higher flow rate, but that mussel size had little effect on feces
171 production. These results are relatively consistent with the estimates of the filtration rate and
172 particle retention efficiency of *Crassostrea virginica* (Brusca 2003), which can process up to
173 37 L/h at 24°C and can capture particles as small as 1 μm in size. In contrast, the filtration
174 rate of marine oysters (up to an approximate valve size of 35 mm) has been reported to be 55
175 L/ind./d (Pietros and Rice 2003). These filtration activities of bivalves demonstrate the
176 difficulty of determining the standard conditions of feeding and excretion. Accordingly,

177 optimization of filtration conditions is the most important factor for the removal of organic
178 matter by freshwater mussels in eutrophic waters. The surface plots are suggestive of a need
179 for a slower water flow and longer retention time to facilitate minimum feces production. On
180 the basis of these results, the model of mussel filtration indicated that the selected water flows
181 and retention times were limiting, and therefore did not result in a significant curved surface
182 in the response surface graph. Thus, a further decrease in water flow, along with an increase
183 in retention time in the system should be implemented for validation. However, because of
184 experimental limitations, simplified water flow, and short retention time interval, the model
185 was validated only with increased mussel body size. In related research (Kim *et al.* 2011),
186 freshwater bivalves with similar body size showed relatively small differences in filtration
187 rate, but in mussels with a limited range of size and density, the filtration rate would further
188 increase with increasing water temperature. We consider that a multifactorial analytical
189 approach, which takes into account the interaction of independent variables (including
190 individual body size, environmental factors, and experimental conditions) provides a basis for
191 models designed to assess the nonlinear nature of the response under limited experimental
192 conditions.

193 With regards to the aforementioned results, Ismail *et al.* (2014) explained that since
194 bivalves are an important food source in the aquatic food web, the kinetic data of filter
195 feeding by bivalves could also be utilized in their research designed to elucidate trophic
196 transfer and the biomagnification of organic matters in an aquatic system. These authors
197 stated that the use of environmentally relevant concentrations and treated wastewater can
198 provide the first indication of the potential efficacy of bivalves in removal of contaminants of
199 emerging concern to improve water quality. We believe that additional studies are needed to
200 determine the concentration dependence of organic matter filtration and correlations with

201 bivalve species, age, and prey competition. Previously reported models relating to the
202 reduction of eutrophication through the use of bivalves in a lake system can provide condition
203 but cannot be applied directly in situ to organic matter removal, since the rates of algae and
204 organic matter removal have not been correlated with mussel filtration rates. Results from
205 algal-based studies have confirmed the high filtration efficiency and use of bivalves for
206 improvement of water quality on a large scale, thereby indicating the possible utility of
207 bivalves for improvement of water quality in engineered systems, or as part of ecological
208 rehabilitation (Gifford *et al.* 2007; McLaughlan and Aldridge 2013). These authors have
209 proposed that the conditions or variables that should be taken into consideration when
210 assessing the application of bivalves for water environment improvement are the selection
211 and maintenance of an appropriate bivalve species and population and the optimization of
212 bivalve filtration rates and feces production.

213

214 **4. Conclusions**

215 (1) The model constructed in the present study indicated that the selected factors of
216 mussel size and water current were limiting, and thus did not result in an adequate
217 surface curvature in the response surface graph. Therefore, a further range of water
218 velocities, along with an increase in retention time, should be assessed for validation
219 purposes. Accordingly, owing to experimental limitations, the model could only be
220 validated with mussel size in the present study.

221 (2) A central composite design was adopted to screen the key factors and identify optimal
222 conditions for filtration rates and feces production that enhance the filtering of
223 suspended organic matter in water by *Sinanodonta woodiana*. The results indicated

224 that statistical design methodology offers an efficient and feasible approach for
225 optimizing the conditions that promote high filtration and low feces production.

226 (3) The proposed model equation illustrated the quantitative effect of variables, and also
227 the interactions among the variables with respect to mussel filtration rate. Under the
228 optimal experimental conditions (mussel size, 13.0 ± 0.2 cm; water currency, 17.5
229 L/h), the experimental filtration rate of 4.47 ± 1.82 L/mussel/h showed a degree of
230 correspondence with the predicted value of 8.4 L/mussel/h, which verified the
231 practicability of this optimization strategy.

232

233 **Acknowledgments:** This study was partly supported by the Hyperspectral Remote Sensing
234 of Algal Distribution of Inherent Optical Properties (NIER–RP2016). The authors thank
235 anonymous referees for their valuable and constructive comments.

236

237 **Author Contributions:** All authors contributed the project. Hwan-Seok Choi wrote the
238 manuscript with collaboration of David C. Aldridge. Hyuk Lee, Baik-Ho Kim and
239 Young-Hyo Kim supported the sample collection and management of mussels and
240 performed the algal and chemical analyses. Baik-Ho Kim and Hwan-Seok Choi
241 conceived and designed the experiments, and co-authors participated in discussions and
242 review of the manuscript.

243

244 **Conflicts of Interest:** The authors declare no conflict of interest

245

246 **References**

- 247 Aldridge, D.C., Fayle, T.M. and Jackson, N. (2007) Freshwater mussel abundance predicts
248 biodiversity in UK lowland rivers. *Aqua. Cons. Mar. Freshw. Ecosyst.*, **17**, 554–564.
- 249 American Public Health Association (APHA) (1995) Standard methods of the examination of
250 water and wastewater (19th ED)., Washington, D.C.
- 251 Chowdhury, G.W., Zieritz A. and Aldridge D.C. (2016) Ecosystem engineering by mussels
252 supports biodiversity and water clarity in a heavily polluted lake in Dhaka,
253 Bangladesh. *Freshw. Sci.*, **38**, 188–199.
- 254 Comeau, L.A., Pernet, F., Tremblay, R., Bates, S.S. and Leblanc, A. (2008) Comparison of
255 eastern oyster (*Crassostrea virginica*) and blue mussel (*Mytilus edulis*) filtration rates
256 at low temperatures. *Can. Tech. Rep. Fish. Aqua. Sci.*, **2810**, 1–17.
- 257 Coughlan, J. (1969) The estimation of filtration rates from the clearance of suspensions. *Mar.*
258 *Biol.*, **29**, 170–180.
- 259 Davenport, J., Smith, R.J.J.W. and Packer, M. (2000) Mussels *Mytilus edulis*: significant
260 consumers and destroyers of mesozooplankton. *Mar. Ecol. Prog. Ser.*, **198**, 131–137.
- 261 Elliot, P., Aldridge, D.C. and Moggridge, G.D. (2008) Zebra mussel filtration and its potential
262 uses in industrial water treatment. *Water Res.*, **42**, 1664–1674.
- 263 Fanslow, D.L., Nalepa, T.F. and Lang, G.A. (1995) Filtration rates of the zebra mussel
264 (*Dreissena polymorpha*) on natural seston from Saginaw Bay, Lake Huron. *J. Great*
265 *Lakes Res.*, **21**, 489–500.
- 266 Gifford, S., Dunstan, R.H., O'Connor, W., Koller, C.E. and MacFarlane, G.R. (2007) Aquatic
267 zooremediation: deploying animals to remediate contaminated aquatic environments.

- 268 *Trends Biotechnol.*, **25**, 60–65.
- 269 Hwang, S.J., Kim, H.S., Shin, J.K., Oh, J.M. and Kong, D.S. (2004) Grazing effects of a
270 freshwater bivalve (*Corbicula leana* PRIME) and large zooplankton on phytoplankton
271 communities in two Korean lakes. *Hydrobiologia*, **515**, 161–179.
- 272 Ibelings, B.W., Portielje, R., Lammens, E.H.R.R., Noordhuis, R., Van den Berg, M.S., Jooisse,
273 W. and Meijer, M.L. (2007) Resilience of alternative stable states during the recovery
274 of shallow lakes from eutrophication: Lake Veluwe as a case study. *Ecosystems*, **10**,
275 4–16.
- 276 Ismail, N.S., Muller, C.E., Morgan, R.R. and Luthy, R.G. (2014) Uptake of contaminants of
277 emerging concern by the bivalves *Anodonta californiensis* and *Corbicula fluminea*.
278 *Env. Sci. Technol.*, **48**, 9211–9219.
- 279 Jeong, Y.S., Kim, J.W., Lee, E.S., Gil, N.Y., Kim, S.S. and Hong, S.T. (2014) Optimization of
280 alkali extraction for preparing oat protein concentrates from oat groats by response
281 surface methodology. *J. Kor. Soc. Food Sci. Nutr.*, **43**, 1462–1466.
- 282 Kim, B.H., Baik, S.K., Hwang, S.O. and Hwang, S.J. (2009) Operation of CROM System
283 and its Effects of on the Removal of Seston in a Eutrophic Reservoir Using a Native
284 Freshwater Bivalve (*Anodonta woodiana*) in Korea. *Kor. J. Limnol.*, **42**, 161–171.
- 285 Kim, B.H., Lee, J.H. and Hwang, S.J. (2011) Inter and intra-specific differences in filtering
286 activities between two unionids, *Anodonta woodiana* and *Unio douglasiae*, in ambient
287 eutrophic lake waters. *Ecol. Eng.*, **37**, 1957–1967.
- 288 Lammens, E.H.R.R., Van Nes, E.H. and Mooij, W.M. (2002) Differences in the exploitation
289 of bream in three shallow lake systems and their relation to water quality. *Freshw.*
290 *Biol.*, **47**, 2435–2442.

- 291 Lauritsen, D.D. (1986) Filter-feeding in *Corbicula uminea* and its effect on seston removal. *J.*
292 *North Am. Benthol. Soc.*, **5**, 165–172.
- 293 Lee, S.H., Baik, S.K., Hwang, S.J. and Kim, B.H. (2009) Comparison of grazing
294 characteristics of a freshwater bivalve *Unio douglasiae* (Unionidae) on the cold and
295 warm phytoplankton communities in eutrophic lake. *Kor. J. Limnol.*, **42**, 115–123
- 296 Lee, J.H., Hwang, S.J., Park, S.G., Hwang, S.O., Yu, C.M. and Kim, B.H. (2009) Continuous
297 removal of organic matters of eutrophic lake using freshwater bivalves: Inter-specific
298 and intra-specific differences. *Kor. J. Limnol.*, **42**, 350–363.
- 299 McLaughlan, C. and Aldridge, D.C. (2013) Cultivation of zebra mussels (*Dreissena*
300 *polymorpha*) within their invaded range to improve water quality in reservoirs. *Water*
301 *Res.* **47**, 4357–4369.
- 302 Newell, R.I.E. and Koch, E.W. (2004) Modeling seagrass density and distribution in response
303 to changes in turbidity stemming from bivalve filtration and seagrass sediment
304 stabilization. *Estuaries*, **27**, 793–806.
- 305 Pietros, J. M. and Rice, M. A. (2003) The impacts of aquacultured oysters, *Crassostrea*
306 *virginica* (Gmelin, 1791) on water column nitrogen and sedimentation: results of a
307 mesocosm study. *Aquaculture*, **220**, 407-422
- 308 Wang, Z.W., Liu, X.L. (2008) Medium optimization for antifungal active substances
309 production from a newly isolated *Paenibacillus* sp. using response surface
310 methodology. *Bioresour. Technol.*, **99**, 8245–8251.

311

312 Table 1. Experimental range and levels of the three independent variables used in response
 313 surface methodology in terms of actual and coded factors

Variables	Range of levels					
	Actual	Coded	Actual	Coded	Actual	Coded
Mussel size (cm)	5.0	-1	8.0	0	12.0	+1
Water flow (L/h)	12.0	-1	24.0	0	48.0	+1
Filtration rate (L/mussel/h)	0.5	-1	1.0	0	2.0	+1

314

315

316 Table 2. Filtration rate of the mussel *Sinanodonta woodiana* according to differences in water
 317 current and retention time

	Shell length (cm)	Flow rate (L/h)	Retention Time (h)	Filtration Rate (L/mussel/h)
<i>Sinanodonta woodiana</i>	8.5 ± 1.0	12	1.5	0.87 ± 0.17
<i>Sinanodonta woodiana</i>	11.4 ± 1.8	24	22.7	4.47 ± 1.82
<i>Sinanodonta woodiana</i>	11.4 ± 1.3	24	3.9	1.23 ± 0.20
<i>Sinanodonta woodiana</i>	10.6 ± 1.9	24	20.8	3.30 ± 1.03
<i>Sinanodonta woodiana</i>	10.8 ± 1.9	48	10.4	2.30 ± 0.97

318

319

320 Table 3. Experimental designs used in response surface methodology using three independent
 321 variables with the center point showing measured and predicted values of *Sinanodonta*
 322 *woodiana* filtration rate

Run order	Shell size (x_1)	Flow rate (x_2)	Production of feces (x_3)	Mean measured response	Predicted response
1	-1	-1	-1	2.36	3.18
2	-1	-1	1	3.40	4.59
3	-1	-1	-1	3.57	4.81
4	-1	-1	1	2.75	3.71
5	-1	1	-1	2.77	3.73
6	-1	1	1	4.06	5.48
7	-1	1	-1	3.21	4.32
8	-1	1	1	1.72	2.33
9	1	-1	-1	2.58	3.47
10	1	-1	1	2.62	3.53
11	1	-1	-1	6.13	8.27
12	1	-1	1	4.78	6.45
13	1	1	-1	2.81	3.79
14	1	1	1	2.47	3.33
15	1	1	-1	13.45	18.14
16	1	1	1	10.71	14.45
17	1	0	0	1.69	2.28
18	1	0	0	6.15	8.29
19	0	-1	0	1.43	1.93
20	0	1	0	4.78	6.45
21	0	0	0	3.97	5.35
22	0	0	0	2.38	3.21
23	0	0	-1	6.53	8.80
24	0	0	1	6.09	8.21
25	0	0	0	6.11	8.24
26	0	0	0	6.45	8.70
27	0	0	0	6.55	8.83
28	0	0	0	6.53	8.80
29	0	0	0	6.28	8.48
30	0	0	0	6.08	8.20

323

324

325

326