

Review

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Review

# Atmospheric Microplastics: Inputs and Outputs

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**Abstract:** The dynamic relationship between microplastics (MPs) in the air and on the Earth's surface involves both natural and anthropogenic forces. MPs are transported from the ocean to the air by bubble scavenging and seaspray formation and released from land sources by wind and human activities. Up to 8.6 megatons of MPs per year have been estimated to be in air above the oceans. They are distributed by wind, water and passive vectors and returned to the Earth's surface via rainfall and passive deposition, but can escape to the stratosphere, where they may exist for months. Anthropogenic sprays, such as paints, agrochemicals, personal care and cosmetic products, and domestic and industrial procedures (e.g., air conditioning, vacuuming and washing, waste disposal, manufacture of plastic-containing objects) add directly to the airborne MP load, which is higher in internal than external air. Atmospheric MPs are less researched than those on land and in water, but, in spite of the major problem of lack of standard methods for determining MP levels, the clothing industry is commonly considered the main contributor to the external air pool, while furnishing fabrics, artificial ventilation devices, and presence and movement of human beings are the main source of indoor MPs. The majority of airborne plastic particles are fibers and fragments; air currents enable them to reach remote environments, potentially traveling thousands of kilometers through the air, before being deposited in the various forms of precipitation (rain, snow, or "dust"). The increasing preoccupation of the populace and greater attention being paid to Industrial Ecology may help to reduce the concentration and spread of MPs and nanoparticles from domestic and industrial activities in the future.

**Keywords:** anthropogenic sprays; sea spray; air conditioning; internal and external air; nanoplastics; textiles

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## 1. Introduction

Studies on microplastics (MPs) in air are recent and few. Possibly the first was published by Dris et al., discussing microplastic contamination in Parisian air; in 2016, they reported that 29% of atmospheric fibers in the air around Paris were petrochemical-based and were more prevalent at urban than non-urban sites. The distribution and movement of these particles through the air is influenced by climatic, meteorological, and human factors. In this review, we are particularly interested in the dynamic relationship that exists between MPs and air and water, initially examining the removal of MPs from the atmosphere by rain events and their addition to the air by aerosols, and then considering artificial, human activities that reduce or increase atmospheric MPs in industrial and domestic situations. Throughout, we emphasize the problems created by the lack of standard methods for MP enumeration and analysis. Sampling, arguably the most important step, is fraught with difficulties, and characterization of MPs in environmental samples is still a challenge. Analytical techniques currently include pyrolysis–gas chromatography–mass spectrometry, matrix-assisted laser desorption/ionisation–mass spectrometry, fluorescence, infrared, Raman, and UV–vis

spectrometries, X-ray photoelectron spectroscopy, energy-dispersive X-ray spectroscopy and atomic force microscopy, scanning electron microscopy, transmission electron microscopy.

For this review, we selected the relevant literature using traditional search engines such as Google Scholar and Scopus, with keywords that included “microplastics” together with one or more of the words “aerial, atmosphere, air, transport, clouds, aerosols, rain, wet deposition, air conditioning, condensates, agricultural, industrial”, as well as serendipitous encounters.

## 2. Atmospheric Circulation of MPs

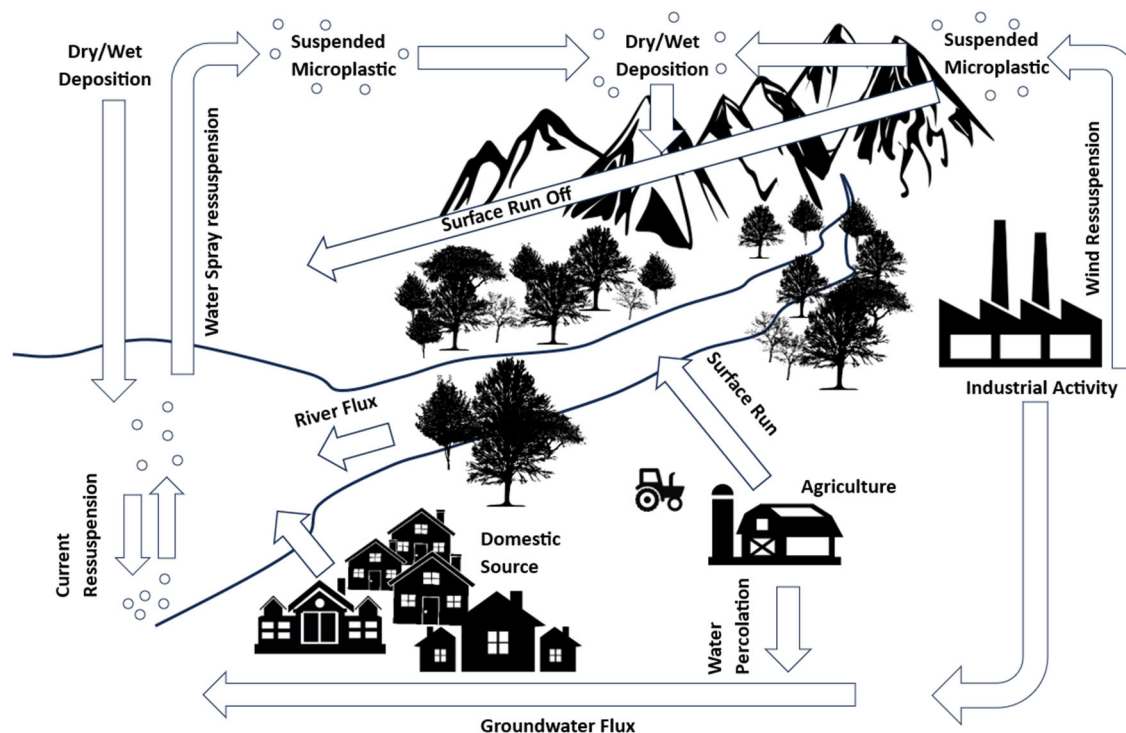
Compared to the availability of MP research in marine and terrestrial environments, studies on atmospheric MPs have only gained focus during the last few years. The air is not a final destination for MPs, but a transport mechanism; nevertheless they may spend a considerable amount of time as aerial particles, especially under appropriate ventilation or wind conditions. It has been calculated that the smallest particles (less than 10  $\mu\text{m}$ ), of moderate hydrophobicity, remain in the still air for an approximate average of  $8.3 \pm 1.0$  days; air movements can prolong this, as well as potentially moving MPs to another location. For instance, a marine wind farm could be responsible for moving MPs not only from the air, but also from seawater, sediment and the ground.

The majority of studies on aerial MPs report results of atmospheric deposition measurements. Quantity, size, morphology, and components have been recorded and evaluated from urbanized, rural and remote locations [1,9–11]. Deposition rates in remote areas, for example, vary from 50 to 700  $\text{MPs m}^{-2}\text{d}^{-1}$  ([2,12–16]).

However, the global migration pathways of MPs from their origins to remote areas are still only superficially understood [12,17–19]. Initial studies mostly considered riverine outflows as the main source of MPs to the open ocean [12,19–23], but latterly studies have reported MPs in remote regions lacking local water discharge sources or riverine influences, including mountainous areas [12,15,25], Arctic and Antarctic snow [26, 13], deep-sea polar environments [27,28] and in the surface air layers over the remote Pacific and Indian oceans [29,30]. These records suggest that MPs may move through the atmosphere over long distances. Figure 1 summarizes the flow of MPs over the land. Measurements of MPs in the Western USA, combined with inverse modelling, have suggested that road MPs are mostly deposited in central Europe, S-E Asia and the East USA coast and that only 1.4% of land MPs are transmitted to the ocean, the latter amount calculated as  $0.418 \pm 0.201 \text{ Tg y}^{-1}$ .

MPs are omnipresent in the atmosphere; their concentrations vary in different areas as a result of altitude, latitude and ecosystem characteristics, fluctuating between seasons. Greater numbers of MPs have been recorded at lower heights than at higher altitudes (a vertical concentration gradient), as a result of atmospheric pressure and gravity. However, the temperature variation in the atmospheric column promotes the ascending circulation of MPs, unless the weather conditions have induced temperature inversion. MPs may also be blocked near the ground, preventing their dispersal.

Chemical transformations affect the dynamic of MPs in the atmosphere. UV rays emitted by solar activity contribute to MP photodegradation, changing their surface properties, generally to become more hydrophilic, and thus altering their diffusion dynamics and behavior towards external agents such as contaminating chemical compounds and potential colonizing microorganisms. The aging process induced by solar rays tends to decrease particle size, making it easier for them to be transported through air fluxes. In parallel, atmospheric  $\text{O}_2$ ,  $\text{SO}_2$  and  $\text{NO}_2$  contribute to the disruption of the polymer chain.



**Figure 1.** Microplastic flux over inland areas, showing terrestrial sources of MPs (domestic, industrial and agricultural), their suspension into the air by wind currents and redeposition onto the ground and into rivers and larger water bodies, whence they can be resuspended in the air. Note that “Industrial activity” includes wastewater treatment plants.

Air flux dynamics around the shore line of water bodies can influence MP distribution; when the air mass comes from onshore (a land breeze), resulting from temperature differences between the continental and water surfaces, there is, apparently, a higher deposition of MPs compared to a predominantly sea breeze. The deposited particles can, of course, be resuspended in the grasshopping process [35,36], a cyclical exchange between air, water and land; in fact, several authors have suggested that the MP atmospheric pool represents the main MP source for the ocean [36–39,4035]. It has been suggested that this atmospheric MP pool comes mainly from textiles [41–45], especially in the case of microfibers, with wearing and washing of clothes the most important source for indoor air [46–48]. Agriculture, road and vehicular emissions, and industrial production of clothing are suggested major sources of MPs for outdoor air [44,48,49]. The release of MPs from industrial processes may become reduced in the future, with the increasing attention to industrial ecology (IE), using sustainable production and distribution, recycling, and end-of-life waste and emission management, particularly if the developing world also accepts the tenets of IE. The situation in the environment is more fragile, with governmental actions required to reduce the use and release of MPs and nanoparticles (NPs).

### 3. Influence of Rain on Atmospheric MPS

Rainfall is an important transport mechanism for the removal of MPs from the air, carrying them to soil or the Earth’s aqueous resources. Hitchcock, for example, found that the levels of atmospheric MPs in Cooks River estuary, Australia, rose from 400 particles/m<sup>3</sup> before a 5-day storm event to up to 17,383 particles/m<sup>3</sup> after the event. It must, of course, be recognized that MPs may also be washed into rivers from the nearby land by heavy rain. The Cooks river is located in Sydney, a large urban conglomerate, where aerial concentrations of MPs would be expected to be high, and release of MPs into local water sources, from surrounding soil, for example, to be under some control. On the other hand, Wei et al., studying atmospheric MPs in the Qing River, Beijing, found that MP concentrations



decreased slightly ( $1164.11/\text{m}^3$  before and  $1037.04/\text{m}^3$  after rainfall), but with rain-induced fragmentation of the larger MPs leading to more small-sized particles.

The importance of rain to the MP air load was demonstrated by Chen et al., who showed that the first rainfall after a long period of dry weather in Beijing contained higher numbers and diversity of morphology and polymer constitution of MPs than the following 7 collections; these MPs were highly fragmented. The wet deposition rate was calculated as  $146\text{--}8629$  items· $\text{m}^{-2}$  per rain event. Rain-driven MP deposition rates have been calculated in various places around the world. MP-laden air liberates much of its particulate burden during the first minutes of precipitation, the so-called “first flush effect” [54,55], and this can lead to mistaken MP deposition rates if rain collections do not include the beginning of the downpour. Bearing in mind this caveat, together with other differences in collection and analysis, as shown in Table 1, we can quote some calculated and published wet deposition rates (Table 1).

**Table 1.** Microplastics (MPs) and nanoplastics (NPs) detected in various locations in wet deposition samples.

Location	MP (NP where stated) concentration (items/ $\text{m}^2$ /day)	Reference	Comments
Urban France (central Paris)	29-280		First atmospheric MP study
Suburban Paris	2-355		One-year uninterrupted collection using stainless steel funnel
Monitoring station in Bernadouze, Central Pyrenees, France	$87 \times 10^3$ ng $\text{m}^{-2}$ $\text{day}^{-1}$ (MPs) $50 \times 10^3$ ng $\text{m}^{-2}$ $\text{day}^{-1}$ (NPs)		Remote mountain location. Samples taken in 5 winter months, collecting for 12-41 days. NPs transported further than MPs.
London, UK	575-1008		Large, polluted city. Rain gauge samples twice/week for 4 weeks; 3-4 days continuous exposure, winter
Coastal rivers catchments of Plymouth and Bristol, S-W England	81.6		Wet deposition levels in Oct-March greater than treated sewage effluent. Lower numbers in rural areas.
Hamburg, Germany	136.5-512		Twice weekly sampling (bulk precipitation samplers) at 6 sites over 12 weeks of winter.
Outskirts of Kassel, Central Germany	$17 \pm 14$		Small city with industry and rural areas. Dry+wet deposition, monthly samples, June-Dec. Custom-made samplers.
River Weser catchment area, NW and Central Germany	$99 \pm 85$		6 varied sites and 2 different collection methods. Higher numbers closer to cities.
Gdynia, Poland	$10 \pm 8$		Small city on sea, Deposition samples taken on 286 days on roof over 2 years.
Spain (and Canary Islands)	5.6 - 78.6		Standard 1 month collections over 10 Spanish towns for 4 consecutive seasons. Higher levels in Barcelona and Madrid and other large cities.

Muskoka-Haliburton, Ontario, Canada	4–9	Data from precipitation monitoring stations in relatively pristine area.
Lanzhou, China	56.97 - 689.05, mean 353.83 (222.25 ± 76.96 during major Covid restrictions)	Sites around the Yellow River during the Covid restrictions. Passive atmospheric deposition sampler used according to standard method for monitoring air and exhaust gas.
Shanghai, China	910 - 3500	Highly polluted city. Collected in 2 stainless steel buckets on a roof on 11 days between Sept. 2019 and June 2020.
Quzhou County (North China Plain)	86–1347 (winter) 892–75,421 (summer)	35 rainfall samples in rural long-term measurement station, August 2020-August 2021. Major fibers Rayon.
Beijing	395.07 ± 41.44 (residential) 180.12 ± 42.22 (agricultural) 133.18 ± 47.44 (forest)	Main sources were textiles. Wind speed was negatively correlated with deposition.
Jakarta, Indonesia	23.422 (rainy season) 5.745 (dry season)	Coastal urban area. Rain gauge used for collections over 12 months.
Parna City, Bihar, E. India	1959.6 ± 205.0 (urban) 1320.4 ± 126.0 (peri-urban).	Wet atmospheric fallout samples in the monsoon period.
Ho Chi Minh City, Vietnam	71–917 (300–5000 μm)	Atmospheric fallouts measured twice per month for a year. Smaller MPs more abundant. No apparent effect of monsoon season.

Li et al., working in the Guangzhou region of China, determined that light rain (less than 2.5mm/h) has a better MP removal effect than heavy (10-50mm/h) rain. In spite of this, the MP atmospheric deposition rate was higher in the wet than the dry season ( $84.00 \pm 6.95$  items/m<sup>2</sup>/d compared to  $47.88 \pm 8.35$  items/m<sup>2</sup>/d), presumably because of greater quantities of rain each day in the wet season. Other workers have also shown higher rain deposition rates in the wet season. For example, in the North China plain 892–75,421 particles/m<sup>2</sup>/day in the wet summer was much higher than the next highest (spring) figure of 735–9428 particles/m<sup>2</sup>/day. It should, perhaps, be borne in mind that rain is often accompanied by wind; the influence of wind velocity on atmospheric MP deposition numbers quoted by various workers [63,69,71–73] could be a function of accompanying rainfall, or vice-versa. On the other hand, in dry conditions the wind may play a more important role. High winds may move surface layers of soil that contain MPs used in fertilizers or as carriers of pesticides into the air, where they will remain until the wind drops, along with any other light MPs. Wind can also erode the surfaces of materials such as discarded plastics in landfills, increasing the aerial load. It is important to note that windy conditions may lead to underestimation of the amount of MPs caught by the sampling apparatus because of decreased collection efficiency.

Österlund et al. suggested that MPs enter remote mountainous lakes mainly through atmospheric deposition (rain/snow). Working in the Tibetan plateau, they found that fibers and films were the major MP morphologies and they could be transported through the atmosphere for up to 800km, the latter figure determined using HYSPLIT simulation. They calculated that, in the monsoon season, the rains could deposit 3.3 tons of MPs. Fibers are the major types of MPs usually reported in the literature on rain washout of atmospheric MPs [69,71,78–80]. This may be because their shape,

with greater surface area, allows more water collisions. Many publications cite clothing textiles as the major source of the atmospheric fibers [12,37,70,80,82,83], though not all of these are petrochemical-based. Some of the textile fibers will have been shed during activities of the local population, although the main outdoor source is likely to be clothing manufacturers; this process has been estimated, by ultrasonic washing of newly produced clothes, to produce 49% of the total fibers released by the procedure, plus those from simulated wet washing and wearing. It has been calculated that about  $0.12 \text{ Tg y}^{-1}$  of synthetic microfibers are released into the environment each year from clothing production (equivalent to one shirt for every 500 manufactured). The recent periods of lockdown imposed on the population during the covid epidemic led to huge reductions in human activity, one of the factors said to influence MP levels in the air. During the Spring 2020 lockdown in France, MP deposition rates were considerably reduced, with median rates of  $5.4 \text{ MP m}^{-2} \cdot \text{d}^{-1}$  in 2020 against  $29.2 \text{ MP m}^{-2} \cdot \text{d}^{-1}$  in 2021. Wet (and dry) deposition rates of MPs are certainly influenced by the human population, not only because of MPs released from their clothes, but also because of their activity-related air movement.

Sun et al. found that MP deposition in Shanghai was correlated particularly with PM<sub>2.5</sub> levels in the air ( $R^2 = 0.76\text{--}0.93$ ), and suggested that these small MPs could be used as a marker for deposition rates. Such small particles, which readily enter the human respiratory apparatus, are normally taken as a measure of indoor air quality, although even in this commonly used determination no standard methods have been published. This is a major problem when comparing MP deposition rates given in the literature (see Table 1). Many factors can influence the results of such measurements, including height of the collector above the ground (potentially collecting splashback) and area of the open mouth of the collector, the material of which the collecting vessel is made (with possible high MP adhesion) and, as previously mentioned, timing of collection within the rain event, considering the “first flush effect”. Collections involving a filtration step will differ from those without and Uddin et al. suggest that filters in sampling devices should be avoided. The problems of sample storage are obvious, and include clumping of particles in static liquid and their adhesion to the internal surface of the storage vessel. Care must be taken in choosing the area of collection, avoiding interfering structures such as high buildings, overhangs or even trees. Huang et al. found that trees with 88% coverage and large three-dimensional spaces formed by leaves could intercept about 16.3% of high-density, small-sized MPs. This could mean that soil beneath forests is protected to some extent from MP pollution, and certainly Zhang et al. found that dust MPs were lower in forest soil in the Ebinur Lake Basin in Northwest China than in construction land or farmland. However the main importance of tree cover is that it should be avoided when taking comparative MP rain deposition rates.

Finally, it must be mentioned that this section concentrates on the removal of atmospheric MPs by rain, but that, in suitable conditions of temperature, hail and snow can also sequester atmospheric particles, with potential future release. This topic is covered in our paper, “Microplastics in the cryosphere - a potential time bomb?”.

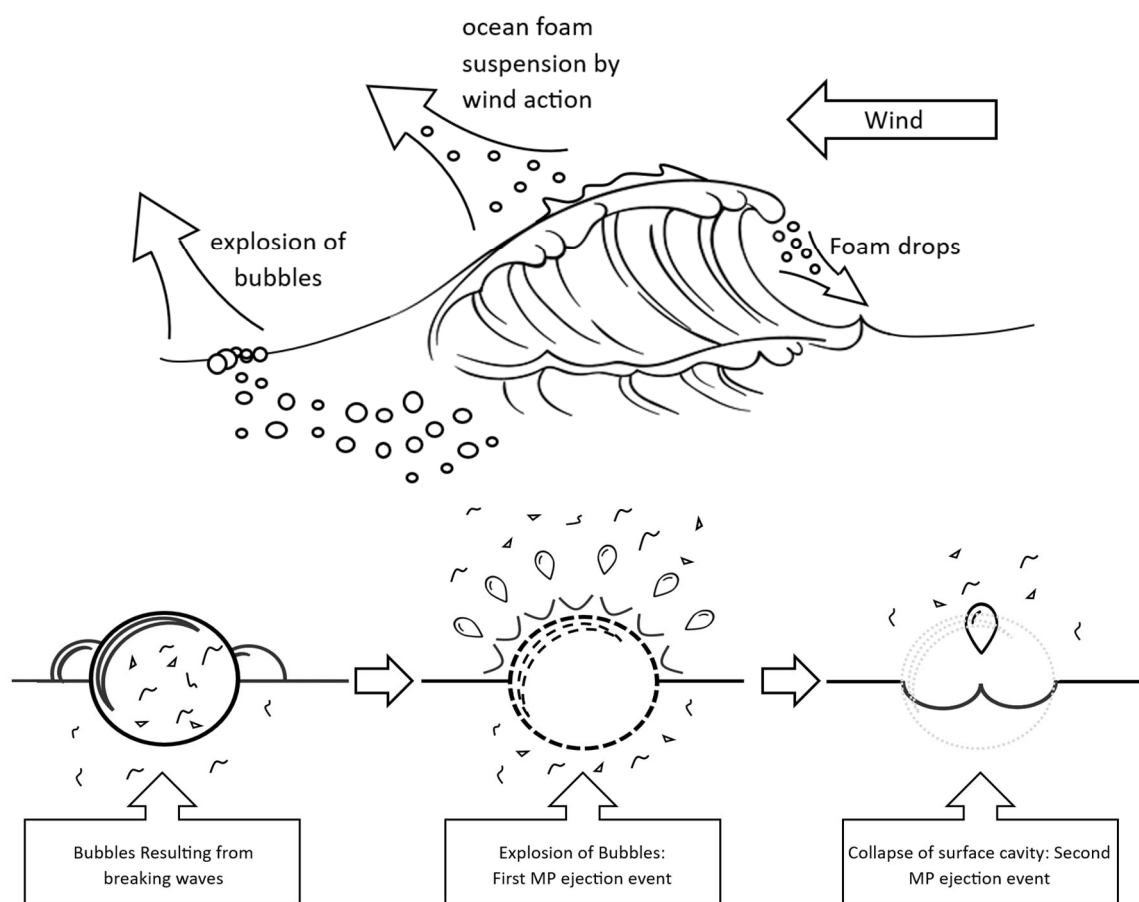
## 4. Aerosols and MPS

It is not within the remit of this article to discuss the inhalation of aerosols and the effects of MPs on human beings. There are many papers and reviews available on these aspects of aerobiology; for example [90–93]. Rather we will discuss the formation of aerosols containing MPs in environmental, domestic and industrial situations and their concentration into water or onto the ground.

### 4.1. Ocean Spray

Breaking ocean waves produce bubbles at the surface; MPs are enriched at the sea surface and even more in these bubbles prior to bursting, due to a process known as bubble scavenging. The bursting of the bubbles produces aerosols that contain inorganic and organic materials. Most particles leaving the sea surface are salt and organic matter, of which  $6700\text{--}7400 \text{ Tg}$  particles (most measuring around  $20 \mu\text{m}$ ) are exported per year by aerosols produced by wave and wind activity through convective updrafts. Micro- and nano-sized salt crystals are expelled from the sea when waves break

and the resulting bubbles burst ; the same scenario can be proposed for MPs (Figure 3), which may have been enriched from the bulk seawater by the bubble scavenging process, especially if they have retained their hydrophobic surfaces. The bursting of the surface bubbles provides nano-sized particles expelled into the adjacent air, able to be carried away by the winds [100,101]. The burst bubble leaves a cavity in the sea surface, which then collapses, generating a second release of particles. This process also exports organic matter that is an acknowledged element for the formation of clouds and rain (particularly in warm air) [101,102] and would lead to the incorporation of MPs into clouds, with subsequent transport through the atmosphere and release in rain (see prior section). The ejection capacity of these processes has been proven by tracking bacteria and virus cells as they travel in the wind and as aerosols across continents and oceans, a process similar to what happens with salt crystals and a blueprint for the release of MPs from marine aerosols. Thus, the much reported ocean “sink” for MPs can become a source of these particles, and evidence for this has been provided by Trainic et al. Noorimotlagh et al., suggest that “bubble bursting” is one of the major sources of MPs to the external air. Sea spray has been stated to be the dominant source of MPs for long-range transport. Harb et al., calculated that aerosolization increases with both ocean concentration and decreasing particle size, while Bucci et al., used a Lagrangian model to determine that a major hotspot for such emissions is the tropics and that MPs released and penetrating into the stratosphere may persist for months. The same authors stated that current observations of MP concentrations above the oceans are underestimated by 1-2 orders of magnitude, though Yang et al., in their review, found the range of published values to vary by 4 orders of magnitude, from  $7.7 \times 10^{-4}$  to 8.6 megatons per year. In spite of variations in degree, the oceans are certainly important sinks of MPs, contributing to their dispersal around the world.



**Figure 3.** Production of sea spray and ejection of MPs by bubble bursting.



#### 4.2. Anthropogenic Aerosols

Can Guven estimates that terrestrial atmospheric MP emissions are much higher than marine concentrations. He calculated a net land-to-sea MP transport of  $25\text{Gg.y}^{-1}$  (compare with Klein and Fisher's determination of  $0.418 \pm 0.201 \text{ Tg.y}^{-1}$ ) and suggested that 99% of MPs are emitted as  $70\mu\text{m}$  particles. There are many sources of terrestrial MPs. Domestic, indoor aerosols containing MPs can result from wearing and laundering of clothes [47,48,112], including tumble dryers, from artificial turf and other flooring materials [114,115], furnishings, unwrapping of plastic packaging, indoor paints and enamels (for example, nail polishing in beauty salons resulted in an air level of  $46 \pm 55 \text{ MPs/m}^3$  compared to  $28 \pm 24 \text{ MPs/m}^3$  in external air). MP counts in indoor air are routinely higher than in outdoor [120,121], and the accident of nature provided by the recent covid epidemic has allowed us to determine that the movement of the populace, both indoors and outdoors, is responsible for much MP prevalence and distribution. In addition, outdoor terrestrial sources include road-related particles, such as tyre-wear MPs and road marking particles [123,124], soil and agriculture related particles (e.g. from plastic mulch, packaging, micro-encapsulated agrochemicals,, landfill sites subject to wind erosion, disposal of wastewater treatment residues/sludge [127,128]), construction materials such as fiber-reinforced concrete and improperly managed public waste. Even the more recently recognized phenomenon of PPE disposal (especially covid-protective masks discarded by the public) is a worrying source of atmospheric MP pollution [131,132,139].

#### 4.3. Air Conditioning

Air conditioners can trap MPs on their filters, initially removing them from the air, but then, when uncleaned for more than 30 days, acting as a source of these particles, as well as of the microorganisms adhering to them. Even clean air conditioners, however, can increase the number of MPs in a room because of an increase in air turbulence, with split air conditioning units increasing MP numbers more than a central AC unit. The latter authors collected and analyzed MP-containing aerosols in domestic and public buildings in Kuwait. In public buildings, the concentrations of MPs were from  $3.2 - 3.7\text{m}^{-3}$  at a time of restricted entry regulations and  $8.7 - 15.3\text{m}^{-3}$  when the restrictions were removed, showing the importance of building inhabitants and their movements on indoor air MP concentrations. Hospital air, in spite of the use of PPE and presence of air conditioning, contained only  $3.9\text{--}4.4 \text{ MPs.m}^{-3}$ , presumably due to increased care with air contamination. A similar low level of atmospheric MPs was found in Al-Hussayni et al.'s study in Mosul, Iraq, in 2023, where the lowest concentration of MPs in 90 buildings under various types of use was found in the doctors' clinics, with the maximum in kindergartens, a clear indication of human activity being a crucial source of MPs in internal air. Torres-Agullo et al. considered that the presence of air conditioning on subway trains in Barcelona was one explanation for lower MP levels here than in buses, suggesting that regular air changes would remove microfibers released by the travelling crowds.

Carpeting, an acknowledged source of microfibers, had no statistically significant effect on MP loading in Kuwait; carpeted offices contained  $8.2\text{--}11.0 \text{ MPs per cubic meter of air}$ . Numbers in residential, carpeted flats were  $10.8\text{--}27.1\text{m}^{-3}$ , while in a mosque, carpeted but also air conditioned, numbers were  $14.3\text{m}^{-3}$ ; these numbers were not significantly different from those determined in uncarpeted rooms. Yasin et al., however, considered that polymer carpets concentrate a significant amount of MPs during their use, with smaller MPs accumulating on vacuuming, the static charge attracting other MPs. Emphasizing the importance of floor coverings, washwaters from carpet-washing shops in Iran were shown to accumulate around  $2000\text{--}3000 \text{ microfibers.m}^{-2} \text{ carpet}$ , with the majority being smaller sizes ( $37\text{--}300 \mu\text{m}$ ).

Washing of synthetic plastics also releases MPs. It has been calculated that washing of clothes is responsible for 28% of the MP microfibers generated by clothing-related processes, 23% produced during their wearing, and the majority (49%) by their manufacture.

#### 4.4. Steam

MPs may be liberated into steam used in plastic repair operations, both in the open air and in industrial situations. For instance, the repair of sewage pipes in the open air can generate MP-containing steam; such repairs often take place in highly populated urban areas, where rapid *in-situ* repairs are necessary. The effect of steam on liberation of MPs and NPs from textiles can be exemplified by the home sterilization of disposable masks, which mainly contain polypropylene (PP). Liang et al used a home cooker to steam for 30 min various types of disposable mask that were used during the covid epidemic, followed by overnight drying at room temperature. The masks were then shaken in deionized water for 24h, after which the water and the mask were examined by microscopy and Raman spectroscopy and for NPs by tracking analysis using ZetaView. Steaming resulted in the release of  $554 \pm 10$  MPs + NPs from each mask, on average, but these particles were detected, not only in the water, but also in the steam, post treatment, at the same level or higher than in the water. Steam treatment resulted in reduction in the size of the released MPs and a concomitant increase in NPs, suggesting that industrial operations employing steam may result in aerial release of more dangerous plastic particles that can penetrate more readily into living tissues.

Domestic procedures also show the effects of steam on plastics. The “sterilization” of rubber teats for baby feeding bottles in boiling water, a common alternative to chemical sterilization, has been shown to release MPs and NPs. Submicrometre-resolved steam etching was shown on the teat surfaces. The authors calculated that MP emission from this source could be as much as  $5.2 \times 10^{13}$  particles per year, globally; a large number, but perhaps less significant than MP release from, for example, road markings and tyres. One of the two common materials used to manufacture teats is silicone, which was not considered an MP-producing material until Hartmann et al.’s 2019 discussion of the classification of MPs and NPs. Indeed, in 2018 Primpke et al. found that silicone interfered with their development of an automated database for FTIR analysis of MPs; removing the silicone cluster from the dendrogram significantly reduced the number of false assignments, confirming that this particular plastic could interfere in MP analysis by FTIR. In 2022, silicone MPs were detected in the dustfall over urban Beijing and in 2023, Fang et al. developed a Raman imaging technique to also reliably detect silicone NPs, with which they successfully demonstrated the potential of silicone sealant in a kitchen to generate both particle sizes. The development of this technique was essential to demonstrate MP production from silicones.

#### 4.5. Paint Sprays

Indoor paints have already been listed as a source of MPs. Spray paints, then, represent a direct source of atmospheric MPs, which will settle not only onto the intended target, but also onto the ground, if not protected; only 50-65% of spray paint finds its intended target, the rest landing on nearby materials. Xu et al., in 2022, found the top layers of soil 2m from a frequently graffitied wall in Berlin contained a maximum of  $2.89 \times 10^7$  MPs.kg<sup>-1</sup> dry soil; this is the highest soil concentration reported up to the time of their publication. Clearly, any soil disturbance, including by wind, will liberate some of these particles once more to the atmosphere.

Automobiles are also generally subject to spray painting. Even polishing the car, an activity indulged in by many on a Sunday afternoon, can remove and disperse MPs. Sobhani et al. estimated that billions or trillions of MPs and NPs (200nm to 7µm) have been produced by the polishing of car hoods over the years. Car washing, either automatically or by hand, will be equally, or more, polluting.

#### 4.6. Body Sprays

Primary MPs are used as microbeads in the production of personal care and cosmetic products (PPCPs), including deodorants, insect repellents, sunscreens, shower gels, shampoos, hair sprays, shaving creams, etc. It has been estimated that, in India alone,  $4.7 \times 10^{10}$  microbeads are released into the environment through untreated sewage, amounting to 3.8 tonnes of microbeads every year, while

the annual per capita emission of MPs from PPCPs in China was estimated to be 2.18 million particles. In a study of body sprays and shower gels, a maximum of 30 different MPs was detected in a single product (a shower gel). Fibrous particles in both types of product were identified mostly as cellulose-derived, but the two body sprays also contained particles that were identified by  $\mu$ -FTIR as PP and PET, and others tentatively assigned to PS, PVC, PTFE, and ethylene vinyl acetate (EVA). Sun et al., in their review, state that approximately 1500 tons/year MPs derived from PPCPs escape from wastewater treatment plants to the surrounding air and water, the global emission per year reaching  $1.2 \times 10^4$  tons. Because of their small size, less than 350  $\mu\text{m}$  in diameter, these MPs can cause considerable damage to marine organisms, as well as being potential carriers of hydrophobic pollutants [157,158] and several countries have banned their use in PPCPs.

Secondary MPs can also be present in domestic spray products. Sprays used to “deodorize” or replace unwanted smells, as well as perfumes for use on the human body or in domestic situations, may contain MPs as materials used to encapsulate, and hence prolong the life of, fragrances. Camerlo et al. used limonene as a test fragrance to evaluate the use of poly(vinyl alcohol) for encapsulation. The PVA fibers held the fragrance for at least 15 days. Many of the MPs are unstable and may be degraded on, or before, release into the environment. The cosmetics industry is attempting to tackle this problem with more “sustainable” cosmetics. More “degradable” plastics are considered options for future products, but truly degradable plastics have yet to be developed. Natural degradable polymers do not have the necessary functionality, mainly strength and toughness [163,164]. The preferred solution is to replace petrochemical-based plastics entirely. Goyal and Jerold consider the options for the production of “green cosmetics”, concluding that advanced technologies and industries are necessary to develop natural, sustainable, and less damaging cosmetics for a more demanding market.

#### 4.7. Agricultural Sprays

Agricultural activities resuspend around  $0.31 \pm 0.13 \text{ Tg y}^{-1}$  of plastics. It used to be considered that the use of waste-treatment effluent (sludge) as a fertilizer was responsible for the presence of most MPs in soil since waste treatment does not remove all MPs [167–169]. MP removal efficiencies in wastewater treatment plants around the world have been calculated at 35–99%. When used to improve agricultural soils, this sludge thus becomes a source of atmospheric MPs [171,172]. However, Radford et al. showed that the difference in MP concentrations between soils could not be explained solely by the use of sludge fertilizer. Other uses of plastics in agricultural practices include plastic mulching, polymer-based fertilizers, and irrigation with untreated wastewater.

Apart from direct spraying for fertilizers and the use of agricultural mulch, a variety of agrochemicals are applied to soil and as crop sprays; pesticides are perhaps the most obvious. Many of these chemicals are plastic-coated, or encapsulated in plastics [176,177] and are thus primary MPs. Coated fertilizers are used on golf courses and in greening projects throughout the world and on rice fields in Japan. Secondary MPs added to agricultural soil, apart from those present in sewage sludge, include mulching and greenhouse films. All these MPs can be removed from the surface layers (upper 30cm) of the soil, where they are at their maximum concentrations, by wind, entering the atmospheric MP dispersion, either with or without attached soil particles. In the former case, they will be rapidly deposited in other areas because of their increased weight, but free MPs may remain as part of the atmospheric MP load for hours, being spread to faraway places, as detailed in the “Atmospheric circulation of MPs” section above.

#### 4.8. Industrial Procedures

Industrial processes used during production of plastic items often involve a cooling phase. Cooling water bleeds and washwaters from various industries (tyre and tube, molded and extruded products, latex industries) can produce MP contamination of surrounding ground and air. Repair procedures used in the open air (cure-in-place technology) can also give rise to aerial MPs, as demonstrated for sewer pipe repairs [182,183]. Machining and processing of plastic-containing

materials in the open air also liberates MPs; examples include carbon fiber-reinforced plastic for automobiles and wind turbine blades.

## 5. Conclusions and Future Perspectives

The threat of MP pollution around the globe is universally accepted; it is strange that the atmosphere should become the last compartment to be studied in this respect, considering its direct impact on human health through inhalation. Perhaps this is due to the circulatory dynamics and low stability of suspended particles in the air. The available research methodologies, developed for the study of MPs in earth and water compartments, are less applicable to the air, often producing unreliable data. Universally accepted standard methods must be developed and employed to allow real MP levels to be determined in the air and in their multiple, often complex, sources. Sampling methods currently used are varied, with low correlation between their numerical results. MP characterization remains a challenge, with limitations, particularly, of throughput. Comparative chemical/instrumental analyses and the use of models must be more developed to further knowledge of the ageing, transport, and distribution in time and space of atmospheric MPs. The atmosphere is the gaseous layer that covers the entire surface of planet Earth, both sea and land, thus constituting a link between all the world's ecosystems. This leads to the conclusion that the omnipresence of MPs on Planet Earth can be attributed principally to atmospheric circulation.

Natural phenomena like ocean sprays and snow/rainfall represent the entry and exit ports for MPs between the several global pools. Their characteristics define the range and distribution of the particles. The majority of atmospheric plastics are fibers and fragments, which are less dense and thus lighter and more able to resist the downward force of gravity, whilst retaining the durability of traditional plastics. Air currents aid them to reach the most remote environments, potentially traveling thousands of miles through the air, before being deposited in liquid (rain or snow) precipitation, or as "dust", under wind-free conditions or because of increased weight by co-adhesion, or adhesion to other aerial particles.

MPs can also be dispersed through the atmosphere by anthropogenic phenomena like wearing and washing of clothing containing artificial fibers, vapors from repairs to plastic bodies, paint sprays used in car or wall painting, agricultural sprays, or steam from plastic softening and sterilization procedures. Primary MPs used in personal care and cosmetic preparations add to the atmospheric load, as can manufacturing procedures, especially those in plastics-production industries that involve cooling steps, and in the clothing industry. Finally, MPs can be released from wastewater treatment plants in aerosols, as well as the better studied liquid effluents and sludge. Improvements in this sector of industry, which implies not only better waste treatment procedures, but also reductions in the amount of processed waste (eg, recycling MP-containing materials that have reached their useful life) are a real possibility for the future health of the Planet.

To reduce the release of MPs into the atmosphere both personal and industrial changes must be made. We already see the increased preoccupation of the populace with plastic-containing products in the search for environmental sustainability. Industry must look to changes in production and maintenance procedures to reduce the dispersion of MPs in the atmosphere. The increasing importance of industrial ecology, seeking to mimic the circular economy of sustainable production and distribution, recycling, and end-of-life waste and emission management, could take control of industry's contribution to release of MPs and NPs into the environment. If this strategy were successful, the job of the world's governments could be considerably simplified; the situation is urgent.

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