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Review

Microplastic Contamination, an Emerging Threat to the Freshwater Environment and Human Health: A Systematic Review

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Abstract: Microplastics have been detected as widespread in an aquatic environment at the microscale, also known as plastic debris. They have continuously increased due to the increase in population, production of synthetic plastics and poor waste management. They are ubiquitous and slowly degrade in soil and water. They are emerging contaminants that have received attention from research communities and public audiences over the last few years. They have high stability and can absorb several other pollutants like heavy metals, pesticides, etc. After entering the environment, they can accumulate and persist for a long time. They can create a serious threat to freshwater ecosystems and human health. These particles can cause physical damage to freshwater organisms. Raman spectroscopy, Fourier-transform infrared spectroscopy, visual identification, density separation, microscopic method, chemical method, thermos-analytical method and hyperspectral imaging method are the commonly known approaches for identification and quantification of microplastics. The noticed concentration of microplastics depends on the analysis method, sampling location and technique. The authors reviewed the sources, health impact, transport and treatment of microplastics in freshwater environments in detail. This study will provide the baseline data for the researchers to do more research on microplastic pollution in the future.

Keywords: plastics; microplastics; freshwater environment; human health; Raman spectroscopy; Fourier-transform infrared spectroscopy

1. Introduction

Plastics are made up of long polymer chains (Cera et al. 2020). Polymers are designed by the polymerization and condensation reaction, such as polyethylene is designed by the reaction of polymerization, while the reaction of condensation designs Nylon. Plastics can be flexible, inexpensive, lightweight, robust and waterproof and act as insulators. They are not biodegradable, but some are biodegradable and can be decomposed by hydrolysis or by the action of microbes or in the occurrence of ultraviolet (UV) light (Bhardwaj and Sharma 2021; Bhardwaj 2022a). They are not usually demolished but are converted from one form to another. They can be classified as mini-microplastics (1×10^{-6} m to $< 1 \times 10^{-3}$ m), microplastics (MPs) (1×10^{-3} m to $< 5 \times 10^{-3}$ m), mesoplastics (5×10^{-3} m to $< 25 \times 10^{-3}$ m), and macroplastics ($\geq 25 \times 10^{-3}$ m) (Figure 1) (Lee et al. 2013).



Figure 1. The Size of the Different Types of Microplastics.

The use of plastic is increasing with the increase in the population. Plastics are regularly used in different types of industries like packaging, electrical, sports, automotive, construction, cosmetics, etc. They are used in film filters of water treatment plants. They are of two types: thermoset plastics and thermoplastics. Thermoset plastics have permanent chemical bonds between the polymers and at the time of heating, they are not softened. It means they cannot be recycled. For example, polyurethane (used in pillows, insulating foams, building insulation, etc.); some polyesters; epoxy resins and some acrylic resins. Thermoplastics get softened at the time of heating while hardening after freezing. It means they can be recycled. For example, polypropylene (used in auto parts, snack wrappers, food packaging, bottle caps, etc.); polyethylene (made up of long ethylene monomer units and is used in several products such as shampoo bottles, toys, plastic bags, pipes, milk bottles, bottle caps, etc.) (Miloloza et al. 2021); polyvinyl chloride (used in frames, pipes, cable insulation, etc.); polyethylene terephthalate (used in water bottles); polystyrene (used in building insulation, eyeglasses, etc.); polycarbonates and polyamides.

Worldwide manufacture of plastics has amplified since 1950 and reached 381×10^9 kg in 2015 (Ritchie and Roser 2019). Plastic manufacture has increased from 322 to 348×10^9 kg from 2015 to 2017 (Europe 2019). The global production of plastics with fibers was estimated to be 381×10^9 kg in 2015 and with additives, it was 407×10^9 kg (Geyer et al. (2017). Between 1950 to 2015, $6,300 \times 10^9$ kg of plastic waste was generated while $4,900 \times 10^9$ kg ended up in landfills and the environment. India is one of the major plastic consumers of the world with the generation of 5.6×10^9 kg of plastic annually (Laskar and Kumar 2019). Ocean and sea serve as the final basins for the world's plastic waste.

Some plastics comprise pro-oxidants that encourage fragmentation and have the potential to form microplastics (Kershaw 2015). Tiny plastic pieces are called microplastics. The term plastic debris was used by Hartmann et al. (2019) for the first time for microplastics. Microplastics are naturally hydrophobic and can go into the freshwater environment via treated and untreated sewage effluent, surface run-off, air deposition, industrial effluent and tainted plastic trash. They are in synthetic clothing, cosmetics and even plastic shopping bags. They've been discovered in places you wouldn't expect, including food, air, beer and tap water. They have been pervasive in the environment for an extensive time and can be swallowed by biota due to their utility, stability, disability and degradation resistance (Peiponen et al. 2019; Bhardwaj 2023).

They can enter tap and bottled water from the water distribution systems. They are present in dust particles and may be a source of air pollution (Bhardwaj and Vikram 2023; Bhardwaj et al. 2023). They differ in size, type, color and density and their physical appearances are strongly related to their fate, toxicity and source (Bhutto and You 2022). The length of these particles is smaller than 5×10^{-3} m (0.0051 m). If the length is less than 1×10^{-6} m then they are termed nanoplastics. They have been categorized into six groups: fragments, pellets, microbeads, fibers, films and foam (Table 1 and Figure 2) (Anderson et al. 2017).

Table 1. Occurrence of Different Types of Microplastics in Freshwater Environment.

S. No.	Sample Types	Locations	Detection Methods	Microplastic's Concentration	Type/Color/Size of Microplastics	References
1	Microbeads from customer products	Great Lakes, USA	EDS and SEM	0.043 particles/m ³	Blue, white and gold; <1×10 ⁻³ m	Eriksen et al. 2013
2	82 % of the fragments and debris, 25 % PE and 19 % PP	Tamar Estuary, Southwest England	Sieving and FTIR spectroscopy	0.028 particles/m ³	Yellow and black; 1–5×10 ⁻³ m PP and < 1 or 1–3×10 ⁻³ m nylon	Sadri and Thompson 2014
3	Fragments and films	Lake Hovsgol, Mongolia	Sieving and light microscopy	0.20 particles/m ³	White and blue	Free et al. 2014

4	Microfibers	Yangtze Estuary System, China	Floatation and stereomicroscope	4137.3 ± 2461.5 and 0.167 ± 0.138 numbers/m ³	Transparent, white and black	Zhao et al. 2014
5	Expanded PS	Pearl River Estuary, Hong Kong	Visual sorting and sieve	Highest 2098 ± 1705, Median 520 ± 688 and lowest 94 ± 44 items/m ²		Fok and Cheung 2015
6	PP and PE	Urban estuaries, China	Micro-Raman spectroscopy, filtration and agitation	Ranged from 10.6 -119.8 %	Transparent, black and white	Zhao et al. 2015
7	PS, PP, polyvinyl chloride and PE	Tibet Plateau Lake, China	SEM and Raman spectroscopy	8 ± 14 to 563 ± 1219 items/m ²	Blue, yellow, white and transparent	Zhang et al. 2016
8	Cellophane, PE, PS and PP	Taihu Lake, China	Micro-FTIR spectroscopy	11.0–234.6 items/kg in	White (29 %) and	Su et al. 2016

			and SEM/EDS	sediment and 3.4–25.8×10 ³ items/m ³ in surface water	transparent (44 %)	
9	Fragments and fibers without plastic pellets	Lagoon- Channel of Bizerte, Northern Tunisia	Stereomicroscop y	3000–18,000 items/kg, dry sediment	Red, white, black, green and blue	Abidli et al. 2017
10	22 % PET, 7 % PP, 22 % fluoro- polymer/Teflo n, microfibers (43 % cotton) and 7 % nitrocellulose	Hudson River, USA	FTIR spectroscopy	0.625 to 2.45×10 ³ fibers/m ³	Blue, black, transparent and red	Miller et al. 2017
11	Secondary microplastics (91 %	River Thames, UK	Sieving, visual inspection and Raman	33.2 ± 16.1×10 ³ particles/100 kg sediment	Yellow and red	Horton et al. 2017a

	fragments)		spectroscopy			
12	PE and PP	Surface water of the urban area, Wuhan, China	Stereoscopic microscopy, SEM and FTIR spectroscopy	1660.0 ± 639.1 to 8925 ± 1591 numbers/m ³	50.4 % to 86.9 % transparent	W. Wang et al. 2017
13	PP and PE	Beijiang River, China	Flotation, SEM and FTIR spectroscopy	178 ± 69 to $544 \pm$ 107 items/kg sediment	Blue and brown	J. Wang et al. 2017
14	Low-density PE	Vembanad Lake, Kerala, India	Raman spectroscopy and wet peroxide oxidation	252.80 ± 25.76 particles/m ²	White and transparent	Sruthy and Ramasamy 2017
15	Polyamides, PVC, acrylics, PS, PET, PP and PE	South Africa, Thailand,	Density separation and FTIR spectroscopy	100 to 1900 pieces/kg dry sediment	Black (14 %), brown (17 %) and white (57 %)	Matsugum a et al. 2017

		Japan and Malaysia				
16	PP (29.4 %), PE (21 %) and PS (38.5 %)	Three Gorges Reservoir, China	Raman spectroscopy and FTIR spectroscopy	25 to 300 numbers/kg wet weight in the sediments and 1597 to 12,611 numbers/m ³ in surface water	Transparent	Di and Wang 2018
17	77.5 % fragments in Winyah Bay and 76.2 % fragments in Charleston Harbor	South Carolina Estuaries, USA	Sieving, H ₂ O ₂ treatment, SEM and FTIR spectroscopy	221.0 ± 25.6 in sediment samples of Winyah Bay and 413.8 ± 76.7 particles/m ² in sediment samples of Charleston Harbor	White, green, grey, blue, black, colorless and red	Gray et al. 2018

18	75.3 % fibers in water and 68.7 % fibers in sediment	Wind Farm, Yellow Sea, China	Sieving, micro- FTIR spectroscopy and density separation	$2.58 \pm 1.14 \times 10^3$ items/kg in the sediment and 0.330 ± 0.278 items/m ³ in the surface water	Black and transparent	Wang et al. 2018
19	PP (15 %), PS (18 %) and PE (45 %)	Italian Subalpine Lakes, Italy	FTIR spectroscopy	4000 to 57,000 particles/km ²		Sighicelli et al. 2018
20	PP	Shanghai, China	Density separation, microscopy and micro-FTIR spectroscopy	$80.2 \pm 59.4 \times 10^3$ items/100 kg dry weight	Red, white, transparent and blue	Peng et al. 2018
21	PA (26.2 %) and cellophane (23.1 %)	Pearl River, China	Micro-Raman spectroscopy	19.86×10^3 microplastic/m ³ for urban and 8.90×10^3 microplastic/m ³ for the estuary	Film, fiber and fragment	Yan et al. 2019

22	PP (37 %) and PE (30 %)	Poyang Lake, China	Micro-Raman spectroscopy	5 to 34×10 ³ microplastic/m ³ for surface water and 54 to 506 microplastic/kg for sediments	Fiber	Yuan et al. 2019
23	PP and PE	Yong River, China	Raman spectroscopy	0.5 to 7.7×10 ³ microplastic/m ³ for surface water and 54 to 506 microplastic/kg for sediments	Fiber	Zhang et al. 2020
24	PP	Danjiangko u Reservoir, China	micro-Raman spectroscopy	0.47 to 15.02×10 ³ microplastic/m ³ in surface water and 15 to 40 microplastic/kg in wastewater	Fiber	Di et al. 2019

25	PS (27.7 %)	Suzhou River and Huangpu River, China	micro-FTIR spectroscopy	0.08 to 7.4×10 ³ microplastic/m ³	Fiber	Luo et al. 2019
26	PP	Yangtze River, China	Raman spectroscopy	4.92 ×10 ⁵ microplastic/k m ²	Fragment	Xiong et al. 2019
27	PP (52.31 %) and PE (27.39 %)	Feilaixia Reservoir, China	micro-FTIR spectroscopy	0.56 microplastic/m ³	Films	Tan et al. 2019

* PE = polyethylene; PP = polypropylene; PET = polyethylene terephthalate; PS = polystyrene; PVC = polyvinylchloride; PA = polyamides; PC = polycarbonates; FTIR = fourier transform infrared; SEM = scanning electron microscope; EDS = energy dispersive x-ray spectroscopy; USA = United States of America; UK = United Kingdom.

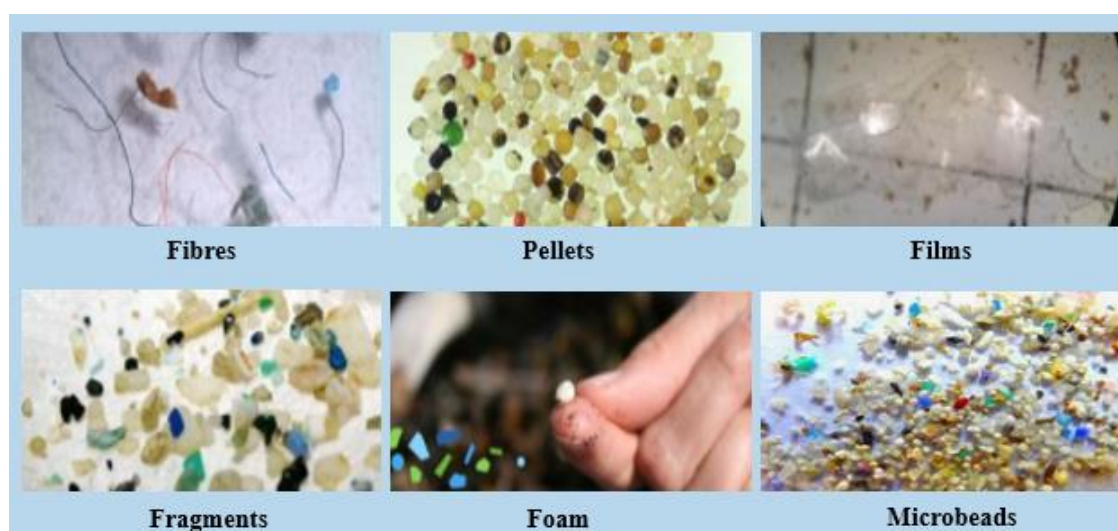


Figure 2. Types of Microplastics.

Microplastics are classified as primary and secondary microplastics based on surface texture and morphology (Bhardwaj 2022b). Primary microplastics are manufactured at the microscale and include plastic fibers, plastic pellets and microbeads. Plastic fibers are used in the textile industry,

plastic pellets are used in the industry and microbeads are used in particular care products. Primary microplastics originate from leakage during the production of plastics and micro-cleansing elements in particular care items (Anderson et al. 2017). Primary microplastics enter the atmosphere through abrasion during washing, unintentional loss from falls during transport or manufacturing and the presence of personal care products in the effluent of households. While secondary microplastics are formed from the bigger plastics products after fragmentation (e.g. bottles, clothes, marine litter, bags, tyres, industrial and agricultural sources, etc.) and this fragmentation occur when larger plastics get exposed to waves, wind and UV radiation (Choudhury et al. 2022). Zhao et al. (2015) described that secondary microplastics are formed by the destruction of bigger plastic particles through photolysis, mechanical forces, thermo-oxidation, thermo-degradation and biodegradation processes.

The existence of microplastics in the freshwater environment is an emerging risk that can affect the capability of persons to preserve biodiversity (Auta et al. (2017a). Several researchers reported the occurrence of different types of microplastics in aquatic environments in different countries and focused on their toxic impact on biota (Triebkorn et al. 2019; Fu and Wang 2019). Koelmans et al. (2019) reported microplastic particles in the range from 0 to 103×10^3 particles/m³ in the freshwater environment. W. Wang et al. (2017) reported the major types of microplastics such as polyethylene terephthalate and polypropylene in the range between 1660.0 ± 639.1 to 8925 ± 1591 numbers/m³ from the inland freshwater of Wuhan, China. Zhang et al. (2021) reported the polypropylene and polyethylene types of microplastics from the Lijiang River, China. Yang et al. (2021) reported the microplastic particles (30.3 ± 15.9 items/kg) from the Ciwalengke River, China. However, the information related to the microplastics in the freshwater environment is still in its initial state as compared to the marine environment. The authors considered approximately 130 to 135 research/review articles for this review and searched these articles from Google Scholar and Research Gate after inputting keywords like plastics, microplastics, freshwater environment, Fourier-transform infrared spectroscopy and Raman spectroscopy. Most of the articles were written from 2000 to 2023. This review article aims to focus on the sources, health impact, transport, and treatment of microplastics in freshwater environments.

2. Health Impacts of Microplastics

The use of plastics is a major threat to the freshwater environment (Sarijan et al. 2021) and the existence of microplastics in drinking water is harmful to the health of humans. This threat exists in three forms: microbiological, chemical and physical. The health impacts, fate and transport of microplastics are not being studied thoroughly yet. However, the presence of microplastics in bottled water, tap water, human tissues, stool and the digestive system of the various invertebrates of freshwater have been reported (Kosuth et al. 2018; Mintenig et al. 2019). Microplastics with low masses ($< 1.0 \times 10^3$ kg/m³) keep floating on the water surface and are consumed by filter-feeding invertebrates (e.g. *Daphnia Magna*) and carnivorous fish (e.g. *Culter Dabryi* and *Culter Alburnus*) (Zhu et al. 2022) while microplastics with high masses ($> 1.0 \times 10^3$ kg/m³) are settled down and consumed by omnivorous fish (e.g. *Sinibrama Wui*) (Zhang et al. 2017).

After exposure to polyvinylchloride or polyethylene, the immune system of fish can be destroyed due to oxidative stress in the leukocytes (Espinosa et al. 2018). Local human activities are the major causes of the accumulation of microplastics in the muscles of fish (Akhbarizadeh et al. 2018). Reduced growth, variation in oxygen consumption, a limited feeding capability, a decreased lifespan and amplified antioxidant-related enzyme action have been reported after the ingestion of microplastics (Windsor et al. 2019). Due to the low feeding capacity of food, less energy is produced to carry out life functions resulting in reproductive and neurological toxicity. Microplastics affect aquatic organisms for several generations due to their slow degradation and stability and also affect the photodegradation of organic mixtures and the poisonousness of metal ions. More than 60 countries have banned single-use plastics and microbeads (UNEP 2018). In 2015, the United States of America (USA) approved an act "Microbead-Free Water Act" which prohibited the manufacture and distribution of cosmetic goods that contain plastic microbeads (Kershaw 2015).

2.1. Physical Hazard

It exists in the form of elements and the harmfulness depends on the shape, size, surface area and surface characteristics. Microplastics greater than 150×10^{-6} m are not absorbed in the body of humans while the absorption of minor elements may be limited. The distribution and absorption of very minor elements of microplastics may be higher. Lu et al. (2016) studied the exposure properties of polystyrene in Zebrafish and confirmed that the poisonousness of microplastics depends on their size. Au et al. (2015) stated that fibers of polypropylene are more poisonous than the spherical elements of polyethylene for the lake's amphipod, *Hyaella Azteca*. Li et al. (2020) stated that fibers are the major type of microplastics in freshwater.

2.2. Chemical Hazard

It exists in the form of polymers such as vinyl chloride, 1,3-butadiene and ethylene oxide. Tetrabromobisphenol A (TBBPA), polybrominated diphenyl ether (PBDEs) and phthalate esters exist in microplastics while they are not bound to the polymer. This chemical threat can simply travel into the atmosphere and the migration rate depends upon the molecular weight of the mixtures. Larger molecular weight particles travel at a slower rate than the smaller particles. Toxic chemicals such as PBDEs, bisphenol A and phthalates get stuck on the microplastics and may encourage their noxious effects after absorption by living organisms (Padervand et al. 2020). These chemicals are endocrinal disruptors and may exhibit their toxic effects on release. After interaction with different types of heavy metals, microplastics can give rise to a serious issue in the freshwater environment (Vedolin et al. 2018). Volatile complexes such as methylene chloride, ethylbenzene, benzene and toluene are released from plastics and can also contribute to long-lasting health effects (Andrady 2017). Microplastics can act as a contaminant transporter for toxic chemicals such as hexachlorobenzene and dichlorodiphenyltrichloroethane (DDT) (Laskar and Kumar 2019; Bhardwaj et al. 2019; Bhardwaj and Jindal 2020; Bhardwaj and Jindal 2022).

2.3. Microbiological Hazard

It's present in the form of biofilm. Biofilm is developed in the water supply when microorganisms start to colonize on the surfaces of microplastics (He et al. 2023). These microbes are harmless and stick to hydrophobic nonpolar surfaces more quickly than they do to hydrophilic ones.

3. Sources and Transport of Microplastics

Polymeric elements from cleaning and cosmetic goods, feedstocks used in the production of plastic goods and plastic powders used for air blasting are the principal sources of microplastics (Jiang 2018). According to Eriksen et al. (2014), microplastics are mostly produced by breaking bigger plastics. Atmospheric deposition, drinking water production, fragmentation, degradation of macroplastics, run-off from land-based sources, wastewater effluent, industrial effluent, combined sewer overflows and human activities such as tourism, sewage treatment plants and distribution are the different major sources of microplastics in freshwater (Cesa et al. 2017).

In addition to this, diverse elements have emerged from diverse foundations like road superficial marking made up of thermoplastic paints, packaging materials, trash of plastic bottles and fibers resulting from textiles (Horton et al. 2017a). The color of microplastics confirms the numerous sources of microplastics and indicates that microplastics originated from synthetic (Yu et al. 2016; Rezanian et al. 2018). Floating macroplastics play the chief source of microplastics in the marine atmosphere.

3.1. Wastewater Effluent

It is a vital collection point of microplastics that are free in daily life and is an extensively recognized cause of microplastics in freshwater (Horton 2017). Microbeads from cosmetic products and synthetic fibers from clothes are the main local inputs into sewage systems. A wastewater treatment plant (WWTP) could act as a pathway for microplastics. There are two paths, direct and

indirect by which the microplastics are out from the WWTPs. By the direct pathway, microplastics are released directly through the effluent of WWTPs and carry high numbers of microplastics. While by the indirect pathway, microplastics are released from the WWTPs into sludge (Gatidou et al. 2019) and this sludge is used as a fertilizer in agricultural lands (Sun et al. 2019). Murphy et al. (2016) reported that 65 million particles of microplastic were out each day from the effluent of WWTPs.

3.2. Run-off from Land-based Sources

Microplastics can create from terrestrial practices, infrastructure, road run-off and tyre debris (Verschoor et al. 2016). City dust is the finest example of a land-based cause of microplastics (Boucher and Friot 2017). Rainstorms and agricultural run-off or farming activities have been recognized as potential causes of microplastics in the freshwater environment (Horton 2017b).

3.3. Combined Sewer Overflows

Horton (2017) described that the barrier of wastewater treatment is temporarily bypassed through heavy rainfall and it is the straight source of microplastics in freshwater.

3.4. Atmospheric Deposition

It has been recognized as an extra possible supplier of microplastics in the freshwater environment through run-off, wet and dry deposition and precipitation (Wright and Kelly 2017). Microplastics that are created from industrial and urban dust can enter in freshwater ecosystems from the atmosphere (Abbasi et al. 2019) and it is an indirect source of microplastics in freshwater. The airborne microplastics originate from waste incineration, buildings, industrial emissions, landfills, fertilizer usage and traffic. These microplastics are a direct health threat to children and building workers through daily ingestion or inhalation (Dehghani et al. 2017). Microplastics in street dust are rich in heavy metals and have a toxic effect on the freshwater environment.

3.5. Industrial Effluent

The involvement of effluents from industries for microplastics in wastewater has yet to be examined (Kooi et al. 2018). However, industrial microplastics have been conveyed in freshwater. Eerkes-Medrano et al. (2015) described the microplastic pollution near the Great Lakes, USA which is situated near the industrial area. Fibers that are out from the textile industries due to tear and washing of clothes can be another important source of microplastics (Henry et al. 2019).

3.6. Drinking-Water Production and Distribution

The treatment of drinking water delivers a wall to microplastics. Some constituents of the treatment plants are fabricated by plastics and their deprivation formed the microplastic particles in drinking water (Mintenig et al. 2019). The bottles and their caps are other sources of microplastic particles in drinking water (Oßmann et al. 2018).

3.7. Fragmentation and Degradation of Macroplastics

After the destruction, macroplastic debris also becomes an important source of microplastics (Morritt et al. 2014) and can enter the freshwater system. Very little research is available on macroplastic disintegration and deprivation in the freshwater ecosystem. Andrady (2007) and Dai et al. (2023) studied the disintegration and deprivation process of macroplastic debris in the aquatic environment. They stated that in the existence of high temperatures and UV light, macroplastics fragmented into microplastics. Zbyszewski and Corcoran (2011) described the degradation of microplastics in freshwater systems by using a microscopic technique. Microplastics can further be split into nanoplastics. The environmental stages of nanoplastics are yet to be measured (Alimi et al. 2018).

The transport pathway of microplastic pollution in the air is not known yet (Horton and Dixon 2018) and the pathway is not separate from terrestrial and aquatic pollution. The route of exposure to microplastics for animals and humans is food (Wright and Kelly 2017). Lau and Wong (2000) reported the presence of polystyrene residual and epoxy resins in food materials having the possibility to enter the human body through food. The diagrammatic representation of the sources, transformation and transport of microplastics is shown in Figure 3.

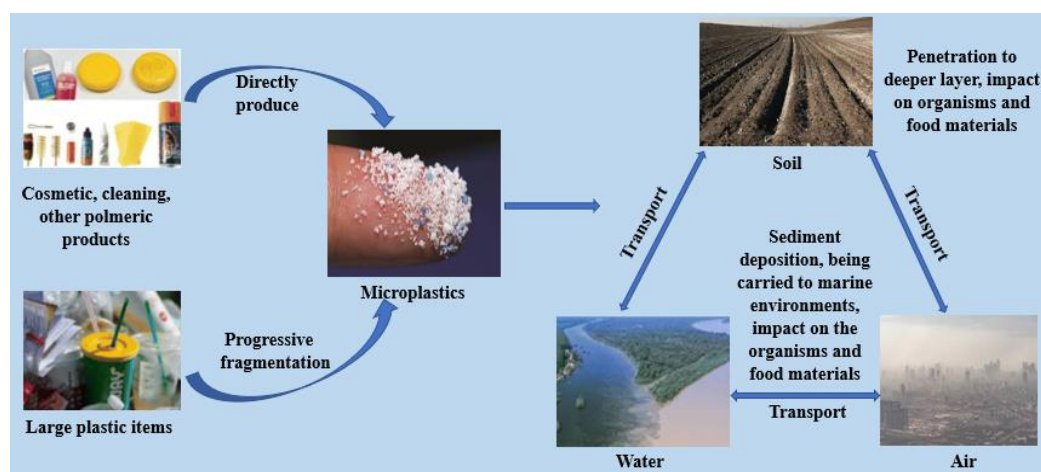


Figure 3. Sources, Transformation and Transport of Microplastics.

4. Analysis of Microplastics

It involves three steps: sampling; extraction and isolation; and identification and characterization (Figure 4).

4.1. Sampling

The samples are collected by using the trawl net (300×10^{-6} m) from the surface of the water.

4.2. Extraction and Isolation

The samples are purified through the filtration technique and then extracted by the density separation method in which the samples are diversified with a liquid of known density. The particles of microplastic float on the surface while other heavy particles sink after that the microplastic particles are collected from the supernatant.

4.3. Identification and Characterization of Microplastics

There are several methods for the identification and characterization of microplastics such as microscopic observation, visual identification, density separation, spectroscopic method, thermo-analytical method, chemical method, hyperspectral imaging (HIS) method and combined method. Microscopic observation is not suitable for the confirmation of plastic particles. Stolte et al. (2015) stated that the identification of secondary microplastics is tough due to the large diversity of pathways and sources.

4.3.1. Visual Identification

This method is obligatory for the parting of microplastics from additional inorganic and organic materials in the residues of samples. The visual estimation can help in the identification of microplastics that originate from laboratory contamination and field samples (Mathalon and Hill 2014). Large microplastics can be detected by this method (Doyle et al. 2011) and particles lesser than 1×10^{-3} m cannot be recognized (Lee et al. 2013). Transparent particles of size 20×10^{-6} m can be identified by this method (Mintenig et al. 2017).

4.3.2. Density Separation

The polymers are split by the differences in their mass. This method is convenient for marine microplastics due to their high density.

4.3.3. Spectroscopic Method

This method is reliable and well-established and is used to recognize the structure of polymers. In this method, the emission or absorption spectra of the particles are matched with reference spectra. If the biofilm is not removed from the surface of the microplastics then it can interfere in the identification and detection of microplastics.

4.3.3.1. Raman Spectroscopy

It is an appropriate method for the identification of microplastics in the aquatic atmosphere (Lenz et al. 2015). Very small plastic particles of size $< 1 \times 10^{-6}$ m can also be measured by this method. By using micro-Raman spectroscopy, small particles of size $< 20 \times 10^{-6}$ m have been detected (Schymanski et al. 2018). The sample (500×10^{-9} m to 800×10^{-9} m) is irradiated with a monochromatic wavelength and the result is compared with polymer spectra libraries to identify the plastic particles (Young and Elliott 2016).

4.3.3.2. Fourier Transform Infrared Spectroscopy (FTIR)

Microplastics of size $10\text{-}20 \times 10^{-6}$ m can be identified using a combination of microscopy and FTIR. It is a dependable and inexpensive method for identifying microplastics (Lusher et al. 2014). It depends on the material, configuration and wavelength. Microplastics can be detected by stimulating molecular vibrations with infrared radiation. Van der Hal et al. (2017) studied the presence of microplastics in aquatic environments and stated that microplastics can be detected at a particular unique infrared spectrum. Löder and Gerdtz (2015) used micro-FTIR spectroscopy for the identification of microplastics of size $< 500 \times 10^{-6}$ m.

4.4. Thermo-analytical Method

In this method, the sample is pyrolyzed under an inert condition and the decomposed product of the individual polymer can be analyzed. This method is used to identify polymer types and requires larger mass particles compared to the spectroscopic method. Hence this method is not suggested for handling huge sample sizes. Gas chromatography-mass spectrometry (GC-MS) is used for the identification of polymers and one particle is analyzed in a single run (Nuelle et al. 2014).

4.5. Chemical Method

It is used for the detection of specific fragments of polymers by inductively coupled plasma mass spectrometry (ICP-MS) (Braun 2018).

4.6. Hyperspectral Imaging (HSI) Method

It is a fast, consistent and non-destructive method for the chemical classification and quantification of microplastics. It was developed by Serranti et al. (2018). There is no need of sample preparation in this method.

4.7. Combined Method

Both spectroscopic and microscopic methods are used to analyze a large quantity of microplastics in water samples. First, spectroscopy is applied for the identification of microplastics and then the stereo microscope is used to count the microplastic particles (Song et al. 2015). A combined method can improve the identification of microplastics in freshwater because the investigation of microplastics is difficult by using a single method (J. Li et al. 2018). Collard et al. (2015) used a new method for the identification of microplastics and it was based on the digestion of

hypochlorite. First, the separation of microplastics from the film was done by sonication after that the analysis was done by Raman spectroscopy.

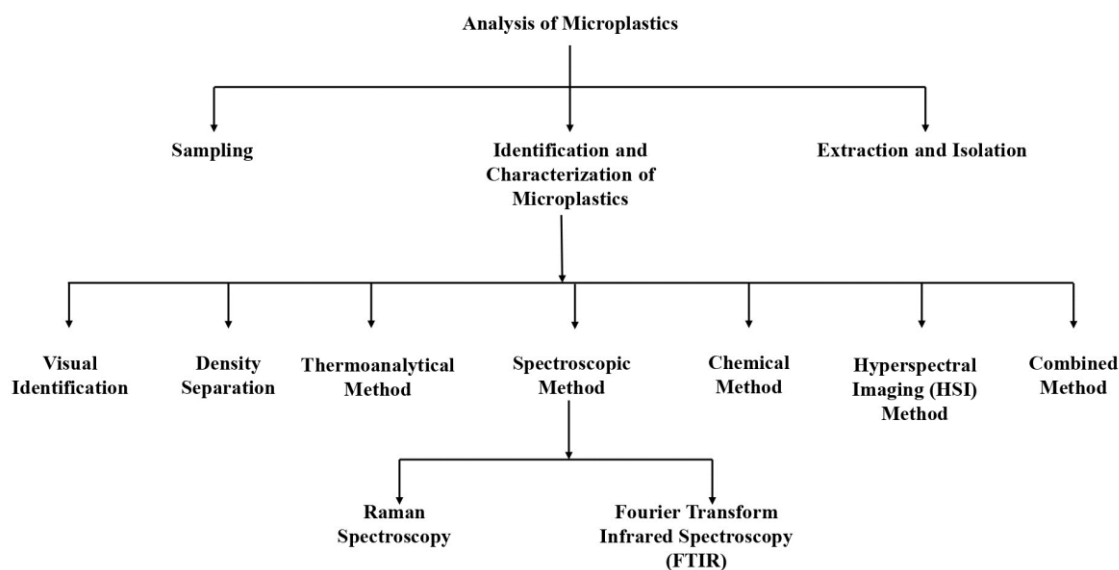


Figure 4. Diagrammatic Representation of the Different Methods for the Analysis of Microplastics.

5. Treatment of Microplastics

There are several methods for the removal/degradation of microplastics such as sorption and filtration methods, chemical methods and biological methods (Figure 5). Electron microscopy (EM), scanning electron microscopy (SEM) and FTIR are used for the study of morphological and structural changes during the degradation of microplastics.

5.1. Sorption and Filtration Methods:

5.1.1. By Adsorption on Green Algae

Microplastic fragments show the devotion behavior on the surface of edible algae and seaweed (Sundbaek et al. 2018). Alginate is a gelatinous substance that is released from the cell walls of the seaweed. Alginate is responsible for the devotion behavior of polystyrene fragments on the surface of seaweed (Martins et al. 2013) and ~ 94.5 % of microplastics are adsorbed by the alginate. The sorption of microplastic particles on the surface of algae differs from the surface charge of particles. Positively charged microplastic particles are more effectively adsorbed on the algae than negative charge microplastic particles (Nolte et al. 2017).

5.1.2. By Using Membrane Technology

Microplastic removal efficiency depends on membrane durability, concentration and the size of the microplastics. L. Li et al. (2018) used membrane technology for the removal of microplastics from polluted water. Horton and Dixon (2018) obtained the filtrate of microplastics within 1200 s by decreasing the turbidity of effluent. Ward (2015) designed a device based on polymer coverings as an extended mesh screen for the removal of microplastics. Membrane bioreactors are suitable for the exclusion of microplastics and these bioreactors can eliminate ~ 99.9 % of microplastic particles per m³ (Lares et al. 2018).

5.2. Chemical Methods

Several researchers used different chemical methods for the elimination of microplastics and stated that the efficiency of the microplastic removal and degradation depends on pH, concentration,

composition of media and type of coagulant. Shirasaki et al. (2016) used coagulation and agglomeration methods for the elimination of microplastics from wastewater. Ariza-Tarazona et al. (2019) used iron and aluminum salt coagulants for the elimination of polyethylene particles and stated that high aluminum doses can increase the elimination efficiency of microplastics. Perren et al. (2018) used the electrocoagulation method for the removal of polyethylene particles from a stirred-tank batch reactor and stated that it is a cost-effective method.

Akbal and Camcl (2011) used metal hydroxide coagulants for the removal of microplastics and reported that these coagulants destroy the colloids and then stabilize the floating microparticles. Polyethylene and polystyrene were degraded by a photocatalyst based on TiO_2 (Wang et al. 2019). The aging of polystyrene and polyethylene was investigated by Liu et al. (2019) while chemical structure degradation of polyethylene and polypropylene was investigated by Brandon et al. (2016). The photocatalytic destruction of low-density polyethylene with the help of ZnO nanoparticles was investigated by Tofa et al. (2019) in aquatic environments.

5.3. Biological Methods

There are several popular biological approaches for the removal and degradation of microplastics from aquatic environments such as by marine organisms, by bacteria and by ingestion. Microorganisms such as fungi, zooplankton and bacteria were found suitable for the removal of microplastics at minimal concentrations. However, the mechanism of the removal and degradation of microplastics through microorganisms is not well understood yet and needs to be explored further.

5.3.1. By Marine Organisms

Due to their small size and lightweight, microplastics are quickly distributed over the ocean surface after traveling a long distance by wind. Ahmed et al. (2018) studied the degradation of natural and artificial microplastics by marine organisms. Dawson et al. (2018) investigated the fragmentation of polyethylene by Antarctic Krill (*Euphausia superba*) in Australia. Cocca et al. (2020) described the harvesting of high-density polyethylene with the help of two marine communities such as Agios and the Souda consortium. Fungi and algae catalyze the reactions of the degradation of microplastics (Urbanek et al. 2018). *Zalerion maritimum* is a naturally occurring fungus that uses microplastics as a nutrient source (Paço et al. 2017).

5.3.2. By Bacteria

Auta et al. (2017b) studied the degradation of polystyrene, polypropylene, polyethylene and polyethylene terephthalate by using bacterial strains such as *Bacillus gottheilii* and *Bacillus cereus*. They reported that *Bacillus gottheilii* is a better microplastic degrader.

5.3.3. By Ingestion

Hall et al. (2015) studied the consumption of microplastic particles in the scleractinian corals and reported polypropylene particles in their gut cavity. They stated that the ingesting rate was 50×10^{-6} m plastic $3600 \times 10^4 \text{ m}^{-2}/\text{s}^{-1}$. Arossa et al. (2019) studied the ingestion of microplastics in the Red Sea giant clam and reported that larger clams ingest higher concentrations of microplastics.

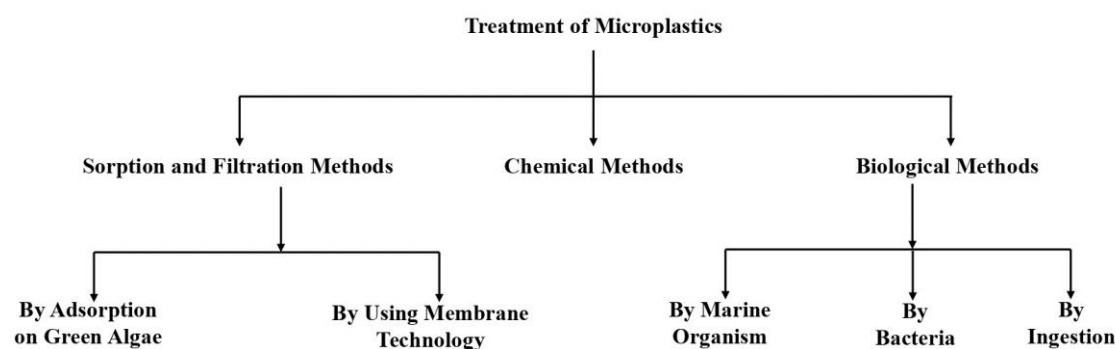


Figure 5. Different Methods for the Treatment of Microplastics.

6. Conclusions and Recommendations

Microplastic pollution is a life-threatening environmental problem. The presence of microplastics threatens the entire freshwater environment. From the previous literature, it can be concluded that different types of microplastics have been detected in freshwater environments by using different methods such as electron microscopy, Raman spectroscopy, Fourier-transform infrared spectroscopy, etc. There is limited study available on the presence of microplastics in aquatic environments. The intensity of microplastic pollution was excessive due to the activity of the inhabitants and industries situated near the freshwater environment. Regarding color transparency, white and blue microplastics were dominant, while polypropylene, polyethylene and fibers were the chief microplastics. Microplastic particles whose size was less than 100×10^{-6} m proved to be most hazardous for the health of humans.

Establishing the standards to determine the ecological risk posed by microplastics is very significant. Researchers believe that the issue of microplastic pollution can be solved through the combined efforts of community enrolment, legislation, and biotechnological and engineering tools. The government and non-government organizations can play an important role in minimizing microplastic pollution by encouraging the recycling of plastics, to use of biodegradable bags and non-plastic resources and to conduct of awareness programs of plastic pollution. In future research, continuous monitoring of the microplastics should be done in regions from where less data has been published like Africa, Asia, and South America. There is a need to develop new cost-effective analytical techniques for the detection and elimination of microplastics from the aquatic environment. New policies should be made worldwide by the authorities for the regular monitoring of plastic pollution in freshwater environment. Industries, non-governmental bodies (NGOs) and government bodies can work together for the reduction/elimination of microplastics from the freshwater environment. People who use items made from plastic waste should be encouraged. Awareness programs like conferences and field activities related to plastic pollution should be conducted by government bodies in a large scale.

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