

River Habitat Survey (RHS) method in the assessment of synecological characteristics of macroinvertebrates: Aquatic Heteroptera of the River Krąpiel valley

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Article

River Habitat Survey (RHS) Method in the Assessment of Synecological Characteristics of Macroinvertebrates: Aquatic Heteroptera of the River Krąpiel Valley

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Simple Summary: There are many methods that can be applied in order to determine the quality of surface water reservoirs, most of which consist assessing hydrochemical properties of the reservoir, as well as studying biological indicators. The River Habitat Survey (RHS) is a method for determining the conditions and the level of anthropogenic alteration in a river based on hydromorphological factors. The aim of this study was to investigate the potential relationship between the composition of aquatic insects fauna and the RHS assessment of the river. We analyzed the distribution and abundance of 14 species of aquatic bugs with different levels of tolerance to environmental gradients in relation to hydromorphological and landscape characteristic of the river Krąpiel. The results showed a link between the species structure of aquatic insect fauna and physical and landscape properties of the reservoir, suggesting that the RHS method can be used to successfully estimate the composition of water bug communities.

Abstract: The River Habitat Survey (RHS) is a method for assessing the conditions of a river channel and valley, taking into account hydromorphological factors, and assessing the degree of naturalness or the degree of anthropogenic alteration within the channel and valley based on physical and landscape data. The hypothesis tested in this study was that the quality of river habitats demonstrated using the RHS should translate into the composition of their fauna. We investigated aquatic bugs (Heteroptera Aquatica), which are a group with considerable dispersal abilities and, at the same time, composed of stenotopic species associated with a particular environment as well as of eurytopic elements inhabiting a wide range of environments (REF). The

research was carried out in a small lowland river, the River Krapiel (north-western Poland) at 6 sites with 43 sub-sites included. Our analyses suggest the RHS method can predict the composition and quality of the aquatic bug fauna inhabiting a given river. Indicators of river habitat quality (HQA and RHQ) will be associated with different habitat types depending on the geographic location and type of river, once indicating habitats associated with the mainstream river and mineral substrate (mountain and small lowland rivers) and at other times habitats covered by vegetation with relatively slow water flow (large and medium lowland rivers). Therefore, RHS may be a good method for estimating the water bug fauna, linking it to the size and character of the river.

Keywords: hydromorphology; hydromorphological indexes; water bugs; landscape scales; physico-chemical parameters

1. Introduction

Determining the quality of surface water reservoirs and their habitats is one of the most important tasks in both basic and applied science; the quality of surface water is one of the main determinants of economic and social development. Therefore, a number of tools have been developed to investigate or assess this factor. Originally, the assessment consisted of the determination of water quality classes and was based on hydrochemical studies (REF). A major shortcoming of this method was that the results described the short-term conditions of the waters (at the time of measurement), giving either inaccurate results or requiring frequent repetition of measurements. Consequently, a number of methods based on the study of biological indicators were introduced: macrophyte method (vascular vegetation), diatom index (diatoms), macroinvertebrates (benthic macroinvertebrates). The River Habitat Survey (RHS) is a method for assessing the conditions of a river channel and valley, taking into account hydromorphological factors, and assessing the degree of naturalness or the degree of anthropogenic alteration within the channel and valley based on physical and landscape data [1–7], thus complementing methods based on biological indicators.

The authors of this study started from the assumption that the quality of river habitats demonstrated using the RHS should translate into the composition of their fauna. Furthermore, the RHS method assesses the parameters of the river channel itself and the landscape parameters of the valley, which are often crucial for the migration of fauna along a river valley or between valleys (REF). To test this hypothesis, we used aquatic bugs (Heteroptera Aquatica), which are a group with considerable dispersal abilities and, at the same time, composed of stenotopic species associated with a particular environment, e.g. lotic waters, as well as of eurytopic elements inhabiting a wide range of environments (REF).

2. Materials and Methods

The small lowland river Krapiel (north-western Poland), which was the areas of a number of ecological studies [1–21,31,45], was sampled. Fieldwork was conducted from May till November, 2011. Samples were taken quarterly from April to October, at 6 sites with 43 sub-sites included (Figure 1). At each site, surveys included a rapids-plo stream pool' system. The "rapids-stream pool" layout included all environments present: rapids - sandy, rocky, with vegetation; stream pool - sandy, silty, rocky, with vegetation. In a given layout, sites from the rapid were always located above sites from the stream pool. In the case of springs, different microhabitats (muddy, sandy, vegetated) and springs located at different distances from the river and in different vegetation communities were included.

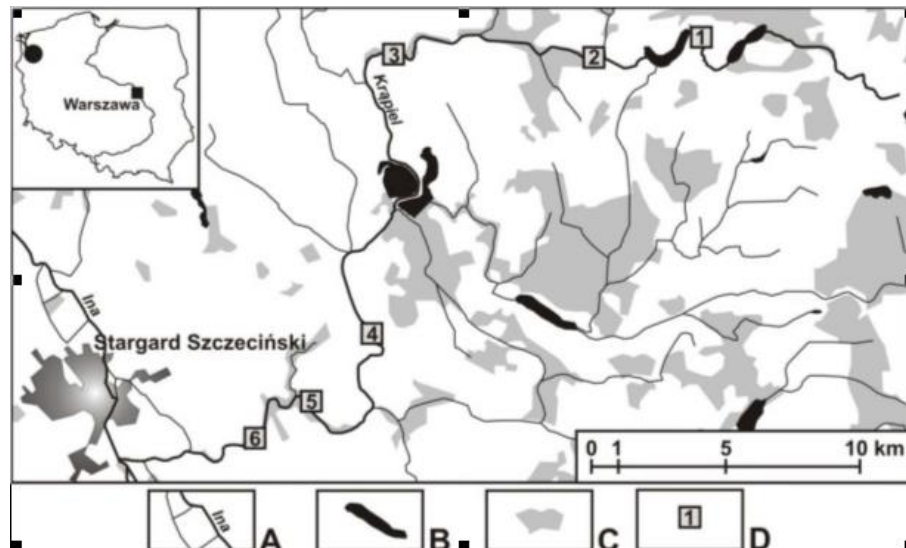


Figure 1. Map of the sites: A - flowing waters, B - lakes, C forests, D - sites.

A hydromorphological assessment of the river was also carried out for each of the study sites. The assessment was based on the River Habitat Survey (RHS) method, which is a standardised method providing results comparable to other surveys [22]. For the purposes of this study, the RHS method was modified and the assessment was based on 100 m river sections instead of 500 m sections. It was considered that 500 m sections would cover too diverse a range of habitats, which would complicate interpretation of the results. On the basis of the field survey, the following indices were calculated [22]:

- Habitat Modification Score (HMS),
- Habitat Quality Assessment (HQA),
- Riverine Habitat Quality Index (RHQ),
- Riverine Habitat Modification Index (RHM).

In addition, the following individual parameters were used: mineralised debris on the channel banks (BN), exposed boulders (OG), channel debris with plants (SU), complex vegetation structure on the top and slope of the bank (Z), fine gravel in the channel (Zr(S)).

The water parameters: temperature (temp.), pH, electrolytic conductivity (cond.) and dissolved oxygen (O₂) content were measured with an Elmetron CX-401 multiparametric sampling probe; water flow (velocity) using a SonTek acoustic FlowTracker flowmeter; BOD₅ by Winkler's method, and the remaining parameters: hardness, NH₄, NO₃, PO₃, Fe, turbidity (turbidite) with the help of Slandi LF205 photometer. Three measurements were performed every time and the median was used for further analyses.

The following parameters were used to describe the habitat: vegetation cover (plants), mineral substrate (mineral), organic substrate (organic), insolation (insolati), mean grain size (M), dispersion of grain parameter values (W).

The landscape analysis was based on the buffer zones and catchment areas defined for each site (macrohabitat) (Figure 2). The buffer zone was defined as a circle around the site with a radius of 500 m. The catchment of a site is defined as the catchment bounded by two consecutive sites (Figure 2).

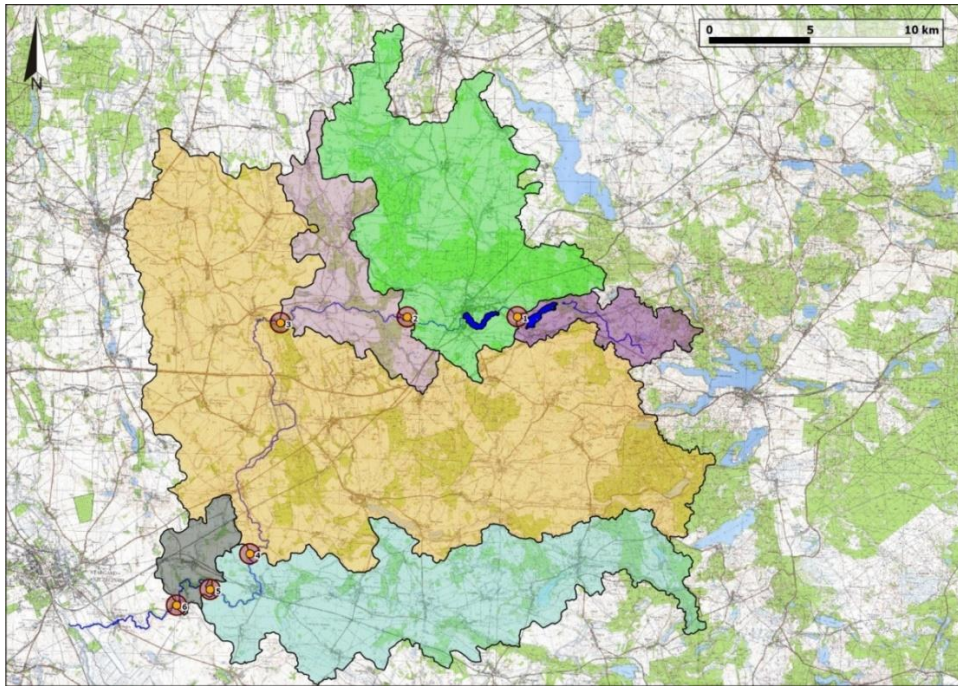


Figure 2. Subcatchments at each locality.

The analysis of the spatial structure of the buffer zones and catchments was based on a set of landscape metrics calculated in TNTmips software (MicroImages™). Classification was based on Landsat TM7 data 28-05-2003, boundaries and linear elements were based on map 1:10 000, classes according to the Corine classification - simplified in the case of meadows (meadows and pastures were combined due to the minimal share of grazing in the surveyed areas), after the initial classification by the isoclass method the extents and correctness of the separation of classes were verified in the field, the isoclass classification was repeated with the combination of classes according to the dendrogram, then the class boundaries were manually drawn in vector form; rivers, roads and drainage ditches appear in the form of polygons, minimum areas of separated patches were equal to ca. 200m. GPS was used to determine the coordinates of the sampling point - the area was delimited with a 500 m radius.

The following measures and indicators were used to analyse landscape structure: 1. measures of patch areas - area (AREA); 2. measures of patch density and size - number of patches (NUP), mean patch size (MPS), median patch size (MEDPS), standard deviation of mean (PSSD), patch density (PD); 3. edge measures - total edge length (TE), edge density (ED), mean edge length (MTE); 4. Shape measures - mean shape index (MSI), mean patch fractal size (MPFD), sum of lobe shape indices (SUM); 5. Diversity and distribution indices - measure of mean distance to nearest neighbour (MNN), measure of boundary diversity (IJI), Shannon's index of patch diversity (SDI), index of spatial distribution and number of patches (SEI), catchment area measured from sources, catchment area, length of catchment boundaries, roughness (Ra), compactness (C), river gradient, distance from sources, distance of individual patches from the river (forests, fields, swamps, buildings, meadows, shrubland, wasteland and water).

After analysing the data, we selected only those landscape parameters, whose effects on faunal distribution were statistically significant.

3. Results

A total of 555 individuals of aquatic bugs belonging to 14 species were caught in the River Krapiel (Table 1).

Table 1. The quantitative composition of aquatic bugs in River Krapiel.

Species	R1 ²	R2	R3	R4	R5	R6	Total	Domination (%)	Frequention (%)
<i>Aphelocheirus aestivalis</i>	f ¹ s		3 1	37 1	77 1	99 153	372	67,0	83,3
<i>Aquarius najas</i>	f s		2	9	34	12	57	10,3	66,7
<i>Callicorixa praeusta</i>	f s		2				2	0,4	16,7
<i>Gerris lacustris</i>	f s			3 1	1 1	10 2	18	3,2	83,3
<i>Hesperocorixa linnaei</i>	f s		1				1	0,2	16,7
<i>Hesperocorixa sahlbergi</i>	f s		3	27	2	7	39	7,0	66,7
<i>Hydrometra stagnorum</i>	f s		2			2	4	0,7	33,3
<i>Nepa cinerea</i>	f s		2	3	8	3 1	17	3,1	83,3
<i>Notonecta glauca</i>	f s		2		1 1	2	6	1,1	66,7
<i>Paracorixa concinna</i>	f s		17		6	2	25	4,5	50,0
<i>Sigara falleni</i>	f s		1		1		2	0,4	33,3
<i>Sigara lateralis</i>	f s					1	1	0,2	16,7
<i>Sigara striata</i>	f s		1 3				4	0,7	16,7
<i>Velia caprai</i>	f s		1 6				7	1,3	16,7

¹ f: flowing zones, s: stagnant zones; ² R1, R2...: sampling sites.

The bugs collected in the Krapiel River were classified into two synecological groups: rheophiles and stagnophiles. The rheophiles were by far the more abundant despite belonging to only three species (*Aphelocheirus aestivalis*, *Aquarius najas*, *Velia caprai*) and constituted 78.6% of the total collected material.

The DCA analysis for the distribution of sites and bug species showed that the length of the gradient represented by the first ordination axis ranks below 3.0, which mandates direct ordination analyses of the RDA type, in order to determine the relationship between species occurrence and the included environmental parameters [23,24].

When analysing the RDA for the dependence of distribution of bugs on the characteristics of the hydromorphological indicators of the riverbed, only one parameter (HQA) was statistically significant (Figure 3). The results of the direct RDA analysis indicate that the variables used in the ordination explain 15.1% of the total variation of species. The habitat quality index was positively correlated with presence of some rheophilic bug species and negatively correlated with presence of eurytopic species.

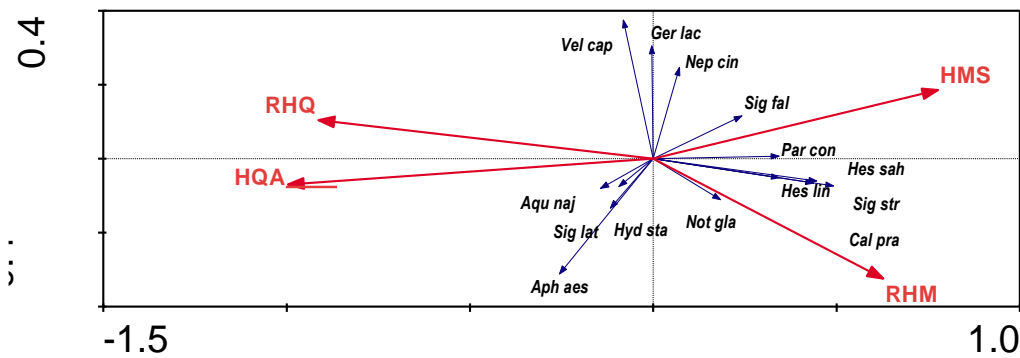


Figure 3. RDA analysis of bug distribution in relation to hydromorphological indicators of the riverbed.

When performing the RDA for the dependence of presence of the aquatic bugs on hydromorphological characteristics of the river channel, collinearity of many variables was detected, so only five that were not colinear were used for further analyses: morphological elements of the channel: exposed boulders (OG), channel debris with plants (SU); bed material: gravel (Zr); morphological elements of the banks: bank debris without plants (BN); vegetation structure: complex (Z). Only one parameter (Zr(S)) was statistically significant (Figure 4). The results of the direct RDA analysis indicate that the variables used in the ordination explain 16.8% of the total variation of species distribution.

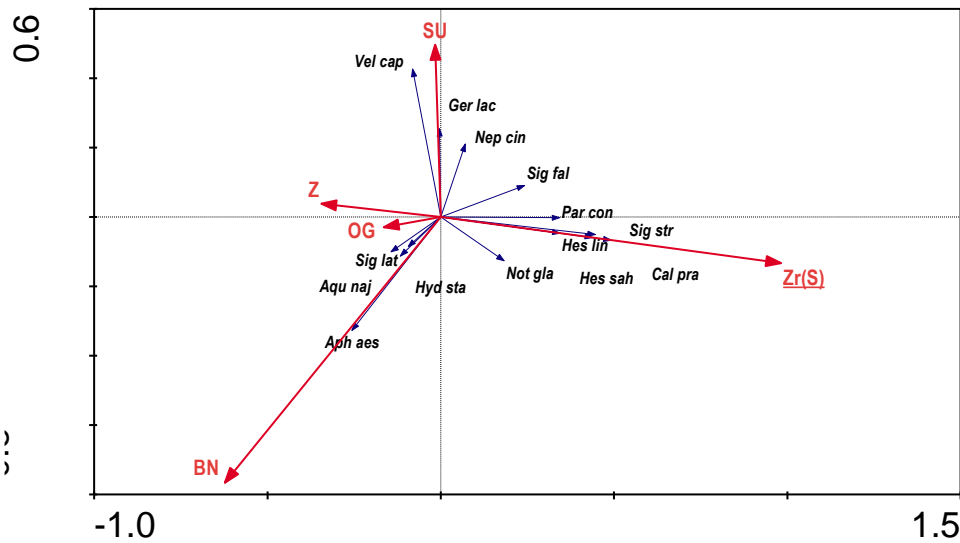


Figure 4. RDA analysis of bug distribution in relation to hydromorphological characteristics of the river channel.

The results of direct RDA analysis for the distribution of bugs as a function of substrate structure, water current velocity, degree of overgrowth by aquatic plants and degree of sunshine indicate that the variables used in the ordination explain 26.2% of the total variation of species distribution. The results of the stepwise selection of environmental variables showed that one variable (organic) is statistically significant (Figure 5) and is connected with Corixidae, which is related to their diet.

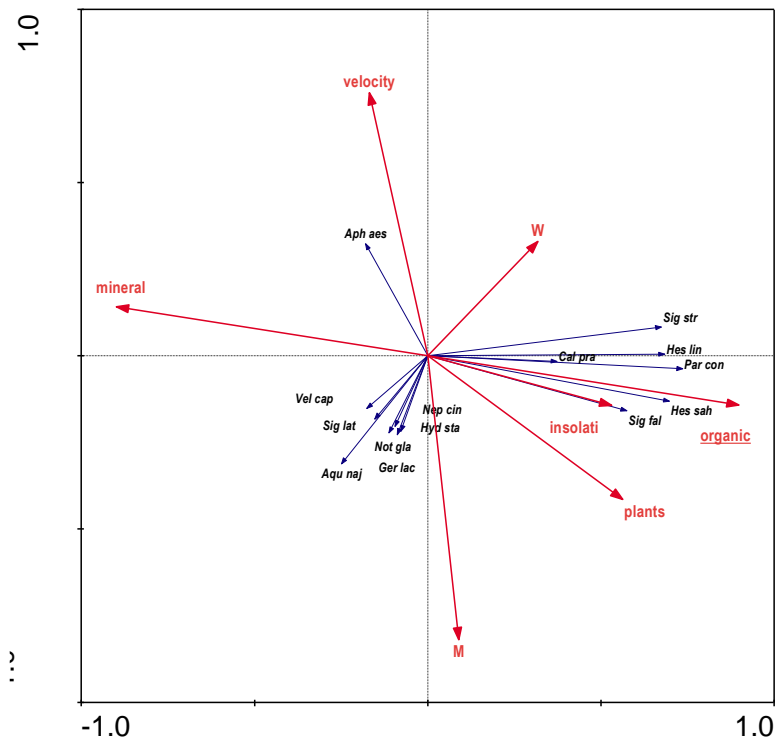


Figure 5. Dependence of bug distribution on river channel structure.

The results of direct RDA analysis for the distribution of bugs in relation to hydrochemical parameters indicate that the variables used in the ordination explained 26.5% of the total species variability. The results of stepwise selection of environmental variables showed that none of the variables explained the range of total species variation in a statistically significant way ($p \leq 0.05$) (Figure 6). In contrast, a correlation can be observed between dissolved oxygen content and *Aphelocheirus aestivalis* abundance.

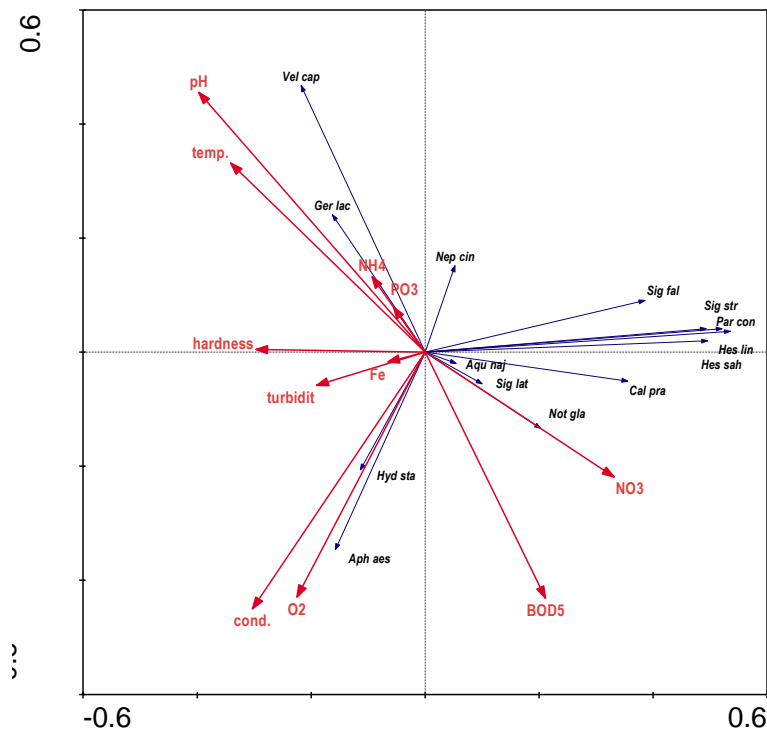


Figure 6. Dependence of bug distribution on hydrochemical parameters.

When analysing the RDA for the dependence of distribution of bugs on buffer zone characteristics, colinearity of many variables was detected, so only five variables that were not colinear were used for further analyses (Figure 7). Only one parameter (MSI) was statistically significant. The variables used in the ordination explain 16.8% of the total variability. Mean shape index was clearly related to the stagnophilous Corixidae and *Notonecta glauca*.

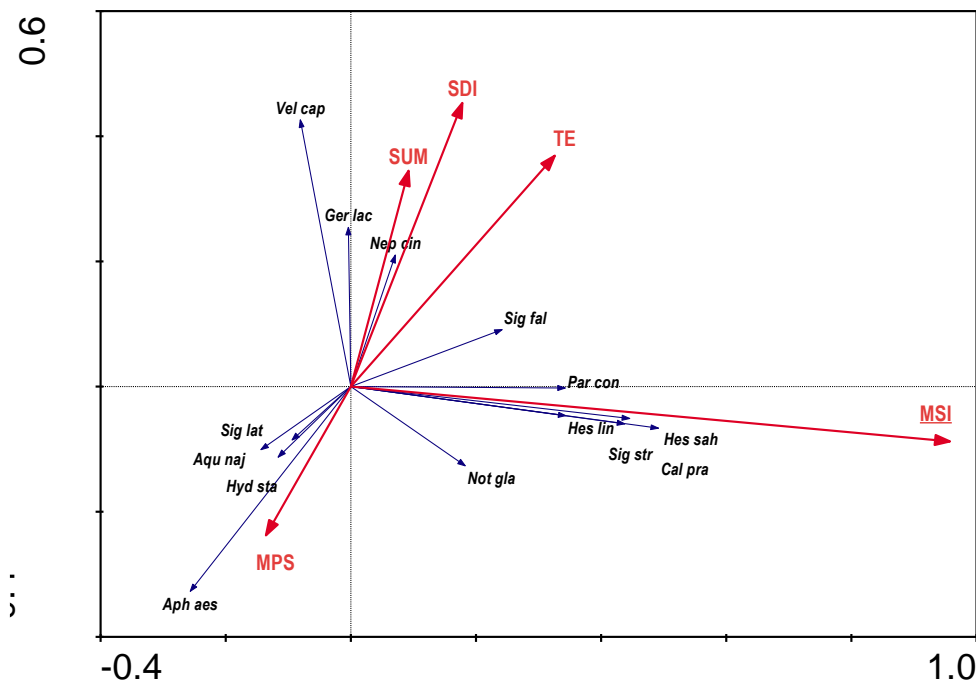


Figure 7. Dependence of bug distribution on landscape characteristics of buffer zones.

When analysing RDA for the dependence of the distribution bugs on the characteristics of individual patches in the buffer zones, colinearity of many variables was also detected, so that only five variables that were not colinear were used for further analyses. These were: total area of patches of a given class: willow thickets - CA (15); mean shape index: arable land - MSI (7); distance from survey point: deciduous forests - L (11); number of patches: compact development - NUNPD (1); mean patch size: deciduous forests - MCA (11) (statistically significant parameters ($p \leq 0.05$) were **bolded** and underlined in the figure) (Figure 8). The variables used in the ordination explain 13.7% of the total variation. The abundance of Corixidae was associated to presence of willow thickets.

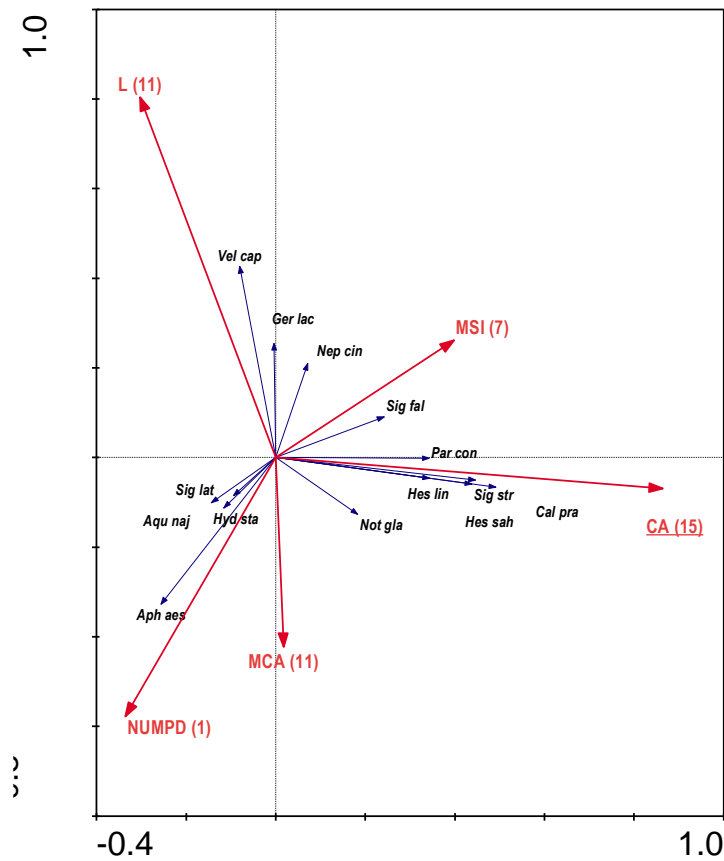


Figure 8. Dependence of bug distribution on patch characteristics in buffer zones.

In the RDA analysis of the influence of catchment elements on the distribution of aquatic bugs, all parameters that did not show colinearity were considered. Parameters that were statistically significant ($p \leq 0.05$) are highlighted (Figure 9). The variables used in the ordination explain 16.8% of the total variation. Presence of the stagnophilous Corixidae and *Notonecta glauca* was associated with area of wastelands.

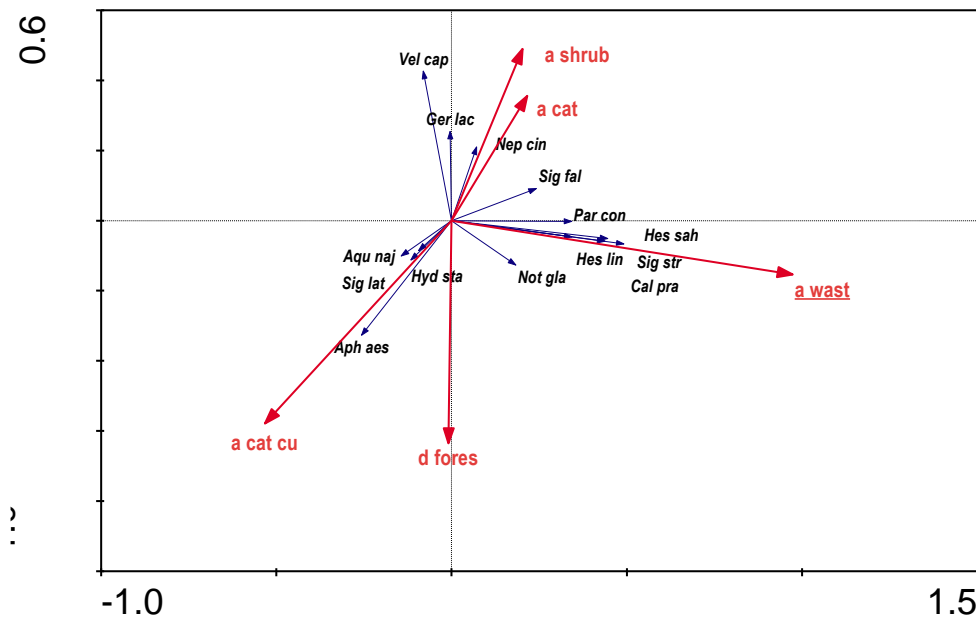


Figure 9. Dependence of bug distribution on catchment characteristics.

4. Discussion

Aquatic bugs are insects commonly found in both flowing and standing waters. Most species are eurytopic, so they inhabit a wide variety of habitats and their possible preferences are difficult to describe. In addition, these species are highly migratory, making them macroinvertebrates that colonise new habitats very quickly. However, they are often associated with a zone along river banks covered by natural vegetation and may therefore be indicative of the naturalness of the river channel, especially in its lower section. Unfavourable habitat conditions including, but not limited to, certain abiotic factors (e.g. high salinity, overfertilisation, low oxygen and others) and biotic factors (e.g. poor development of littoral vegetation, high predation pressure) may significantly reduce both abundance and species richness of aquatic bugs [24–36]. On the other hand, there are also several stenotopic species of aquatic bugs, associated with relatively fast water current and having high oxygen-level demands, making them good indicators of clean running waters [27,38].

Three stenotopic species associated with running water were recorded in our study: *Aphelocheirus aestivalis*, *Aquarius najas* and *Velia caprai*. In the Krapiel River, presence of the first two species, together with *Hydrometra stagnorum* and *Sigara lateralis*, were associated with the middle and lower sections of the river (a cat cu - Figure 8) in open areas (Figure 8) and mineral debris on the banks of the riverbed (BN - Figure 3; mineral - Figure 4) and partly (*Aphelocheirus aestivalis*, *Hydrometra stagnorum*) with highly oxygenated water and increased mineralisation (O₂ and cond.-Figure 5). High values of electrolytic conductivity can significantly limit or even prevent the occurrence of many aquatic bug species, consequently significantly reducing species diversity. Such results were obtained e.g. in the fauna of oligohaline waters discharged from the Bogdanka coal mine [25,26]. Regarding landscape indicators, the species are associated with patches of natural, low vegetation with average patch sizes and undifferentiated boundaries (Figs. 6 and 7). Explaining their association with riverbed habitat and water physicochemical parameters is relatively easy, as the autecological characteristics of these species clearly indicate an association with relatively clean rivers of medium size, not very fast current and coarse mineral substrate [40–43].

The relationship with landscape parameters, on the other hand, is more enigmatic. It seems that in the case of *Aquarius najas* and *Velia caprai*, the open habitats directly next to the river facilitate the dispersal of these species. The presence of these species (with the exception of *Velia caprai*, whose abundance is too low to infer any conclusion) is related to the habitat quality index (HQA). This indicator in the case of the Krapiel River, which is a small lowland river but having a mountainous character over much of its course, is associated with meandering sections, with a relatively fast main river current, numerous channel debris and a gravel bed. In the water current, vegetation is present in insulated places, and the stagnant areas are characterised mostly by mineral (sand and gravel) bottoms, vegetated occasionally.

The second abundant synecological group of aquatic bugs represented in the Krapiel River is the group of eurytopic species (*Callicorixa praeusta*, *Hesperocorixa linnaei*, *H. sahlbergi*, *Paracorixa concinna*, *Sigara falleni*, *S. striata*), which in the study area was associated with organic substrate, covered with vegetation and insulated (Figure 4), water of low turbidity and hardness (Figure 5), and, among landscape parameters, a valley area covered with patches of medium shape variation (Figure 6), with a predominance of willow thickets immediately adjacent to the river (Figure 7) and a predominance of wastelands located in the catchment (Figure 8). Many authors emphasise the positive influence of temperature and insolation on the activity of aquatic bugs, with a particularly strong influence of water thermals, particularly in the shallow littoral, which can locally heat up strongly, creating favourable conditions for the development of thermophilic species [24–27,44]. In the Krapiel River, such habitat and landscape characteristics correspond to transformed sections with a low habitat quality index (HQA) (Figure 2) and with a gravelly bottom outside the main river current (Zr(S) - Figure 3).

Our analyses suggest the River Habitat Survey (RHS) method can predict the composition of the aquatic bug fauna inhabiting a given river. The use of this method for such an estimation can be very convenient as, in addition to the habitat parameters of the river itself, it takes into account the landscape character of its immediate surroundings, which has a major influence on the migration of aquatic bugs along the river valley. However, this estimate is relative to the type and nature of the

river itself. It will depend on the geographical location of the river (the different nature of mountain/submountain rivers and lowland rivers), its size and order. Indicators of river habitat quality (HQA and RHQ) will be associated with different habitat types depending on the geographic location and type of river, once indicating habitats associated with the mainstream river and mineral substrate (mountain and small lowland rivers) and at other times habitats covered by vegetation with relatively slow water flow (large and medium lowland rivers). As such, they will indicate the presence of rheophilic or vegetation-associated fauna. In summary, RHS may be a good method for estimating the water bug fauna, linking it to the size and character of the river. But such estimating is useful of general synecological groups (the number of individuals and species), whereas at the species level, distribution of particular species is better explained by structure of habitats in the river. However, these two components complement one another, and ultimately both of them together explain the nature of the water bugs fauna in the river. Similar relationships were observed for the water mite fauna [45].

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