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Article

Interlaboratory Comparison of ²²⁶Ra and ²²⁸Ra Activity Concentrations in Groundwater and Surface Water

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Abstract: ²²⁶Ra and ²²⁸Ra are typically monitored for groundwater and surface water compliance at legacy U mining and milling sites. Groundwater monitoring results for combined Ra (²²⁶Ra+²²⁸Ra) reported by the existing contract laboratory at a former U mine and mill Site (Wyoming, USA) have been highly variable and with increasing trends at the Site compliance and background wells since the method was changed in 2005. Sample reanalysis has indicated poor reproducibility and significant analytical error in ²²⁸Ra measurements. An interlaboratory comparison was conducted to evaluate the potential causes of the high variability and analytical error. Two different methods were used for ²²⁶Ra (M903.0 and M903.1) and ²²⁸Ra (M904.0 and Ra-05). ²²⁶Ra results were less variable compared to ²²⁸Ra, and ²²⁸Ra data from the existing laboratory were qualified as estimated with high bias due to detection of ²²⁸Ra in the field blank. Compliance with the ²²⁶Ra+²²⁸Ra groundwater standard was either met or not met, depending on which laboratory conducted the analyses. Specific laboratory techniques, rather than the analytical method, are contributing to elevated ²²⁸Ra values being reported. It was recommended that samples whose ²²⁶Ra+²²⁸Ra results exceed the Site standard in the future be reanalyzed by the existing laboratory with a sample split also being sent to an outside laboratory for confirmatory analysis.

Keywords: radium; uranium; groundwater; interlaboratory comparison

1. Introduction

An interlaboratory comparison (ILC) is an exercise carried out by a group of laboratories to compare their performance or to assess a measurement standard, and is typically used to: (1) assess random variation in measurements across a population of laboratories, (2) determine the systematic differences in results among a fixed set of laboratories, or (3) determine the value of a physical property of an artifact or a population of artifacts [1]. An ILC was conducted to evaluate the variability of ²²⁶Ra and ²²⁸Ra activity concentrations (hereafter referred to as concentrations) in groundwater and adjacent surface water impoundment (Impoundment) at a former U mine and mill (Site) located in Wyoming, USA. Assessment of ²²⁶Ra+²²⁸Ra trends at the Site compliance well and selected groundwater monitoring wells has shown that ²²⁶Ra+²²⁸Ra concentrations are highly variable, and there is some uncertainty regarding the quality of isotopic Ra data reported for the Site.

The most common isotopes of Ra are ²²⁶Ra (produced by radioactive decay of ²³⁸U) and ²²⁸Ra (produced during radioactive decay of ²³²Th) [2]. At the Site, ²²⁸Ra is the dominate Ra isotope in groundwater, whereas ²²⁶Ra is the dominant Ra isotope in the adjacent Impoundment. At the Site compliance well (Well 176), the historic concentrations of ²²⁶Ra+²²⁸Ra have periodically exceeded the compliance standard (USEPA Maximum Contaminant Level, MCL) of 5 pCi/L; since approximately 2005, however, the concentrations have shown an increasing trend and with more frequent exceedances of the MCL. In addition to a downgradient well (Well 179), an unexpected increasing trend and high variability in ²²⁶Ra+²²⁸Ra concentrations have also been measured in the unimpacted Site background well (Well 182) since 2005 (Figure 1). The implication of ²²⁶Ra+²²⁸Ra concentrations exceeding the MCL at the compliance location (Well 176) is that the Site boundary may

potentially need to be expanded to demonstrate that downgradient groundwater ²²⁶Ra+²²⁸Ra concentrations will remain protective of human health and the environment.

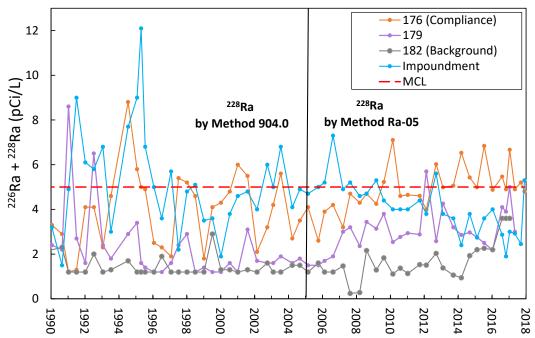


Figure 1. ²²⁶Ra+²²⁸Ra concentrations in Site monitoring wells and the Impoundment.

A preliminary internal evaluation of ²²⁶Ra+²²⁸Ra results for Site groundwater indicated that increasing trends and high variability could be the result of analytical error by the contract laboratory based on several key observations: (1) increases in both ²²⁶Ra+²²⁸Ra and ²²⁸Ra/²²⁶Ra ratios occurred subsequent to a change in the ²²⁸Ra analytical method from Method 904.0 (M904.0) to Method Ra-05 (Ra-05) in 2005 (Figures 1 and 2), (2) ratios of ²²⁸Ra/²²⁶Ra have been increasing in the non-impacted Site background Well 182 concurrently with Well 176 (Figure 2), and (3) re-analysis of samples with unusually elevated ²²⁸Ra has demonstrated at least one occurrence of poor reproducibility and significant analytical error in ²²⁸Ra measurements. For example, the reported ²²⁸Ra value for Site Well 174 (3rd Quarter in 2016) was 8.6 pCi/L, while the field duplicate value was reported as <1.3 pCi/L. Re-analysis results for these two samples were reported as 2.8 pCi/L ²²⁸Ra for the primary sample and 2.7 pCi/L ²²⁸Ra for the duplicate sample. In addition to the poor reproducibility of the field duplicate, additional ²²⁸Ra results appeared to be inaccurate. For instance, ²²⁸Ra for the unimpacted background well (Well 182) was reported as 27 pCi/L which is not consistent with much lower historical results for the background location. Reanalysis of the sample from Well 182 produced a revised result of 2.0 pCi/L. Out of nine total samples which were reanalyzed, the initial results from five samples exceeded the MCL for ²²⁶Ra+²²⁸Ra based on ²²⁸Ra alone, but were reported as below the MCL after reanalysis. The contract laboratory was unable to provide an explanation as to why the reanalysis produced significantly different results from those initially reported.

Therefore, an ILC of ²²⁶Ra and ²²⁸Ra in selected groundwater wells and the adjacent Pit Lake was conducted to evaluate the potential role of random variability and/or systematic error in producing the high variability in ²²⁶Ra+²²⁸Ra concentrations and ²²⁸Ra/²²⁶Ra ratios observed during recent years. Four separate laboratories were evaluated which are designated as Lab A (Site contract laboratory), Lab B, Lab C, and Lab D. The specific objectives were to: (1) evaluate potential differences between ²²⁶Ra results when using a screening method (M903.0) and compare those results to a ²²⁶Ra-specific method (M903.1) and (2) evaluate potential differences between ²²⁸Ra between laboratories using M904.0 and compare those results to ²²⁸Ra when using Method Ra-05. Additional questions include: (1) are there detectable levels of ²³²Th in groundwater which could be migrating and influencing ratios of ²²⁸Ra/²²⁶Ra? (2) are there differences between total (unfiltered)

and dissolved (filtered, $0.45~\mu m$) 226 Ra or 228 Ra, and if so, how do those fractions vary across laboratories?, and (3) does M903.0 for 226 Ra (which measures all alpha-emitting isotopes) overestimate 226 Ra concentrations compared to M903.1 which is specific to 226 Ra? If the presence of additional alpha emitters were to produce higher apparent 226 Ra concentrations, it may be recommended to use M903.1, rather than M903.0, for future compliance monitoring at the Site.

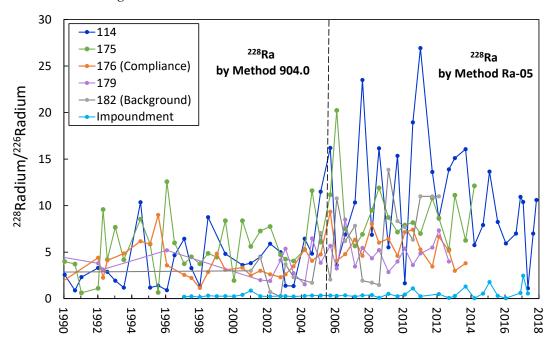


Figure 2. ²²⁸Ra/²²⁶Ra ratios in selected Site monitoring wells and the Impoundment.

2. Materials and Methods

The quality of ²²⁶Ra and ²²⁸Ra data reported by the existing contract laboratory (Lab A) was investigated using an ILC which included analysis of primary samples, a duplicate sample, and field blanks. Groundwater samples were collected from the Site compliance well (Well 176), the Site background well (Well 182), and from an intermediate well located outside of the Site boundary (Well 179). Surface water was collected from the adjacent Impoundment where the U ore was originally mined. Samples were collected for analysis of ²²⁶Ra and ²²⁸Ra, in addition to ²³²Th to evaluate the potential role of ²³²Th as a source of ²²⁸Ra. Water samples were split and processed accordingly to evaluate their dissolved (filtered, 0.45 µm pore-size filter) and total (unfiltered) forms. A duplicate was collected from Well 176, and field blanks (filtered and unfiltered) were prepared using deionized H₂O supplied by each laboratory. All sample collection, preservation, and storage protocols followed standard methods for water sample collection [3]. At each location, the filtered and non-filtered samples were split into four groups and submitted to four laboratories designated as Lab A (existing contract laboratory), Lab B, Lab C, and Lab D. Each laboratory and their corresponding analytical methods are summarized in Table 1.

The methods used for ²³²Th varied among laboratories but are all based on alpha spectroscopy; the methods used for ²²⁶Ra and ²²⁸Ra have been approved by the U.S. Environmental Protection Agency (USEPA) [4] and also varied among the laboratories (Table 2). For ²²⁶Ra, three of the laboratories (Labs A, B, C) used Method 903.0 (M903.0). Although the results are reported as ²²⁶Ra, M903.0 measures the total alpha emitting isotopes of Ra (²²³Ra, ²²⁴Ra, ²²⁶Ra) and is therefore considered a screening method. Method 903.1 (M903.1), which is specific to ²²⁶Ra, was used by Lab D and can be compared to results from M903.0 to determine if other alpha emitting isotopes contribute to the reported ²²⁶Ra concentrations for the Site. For ²²⁸Ra, two of the laboratories used Method 904.0 (M904.0) (Labs B, C), while Lab A and Lab D used Method Ra-05. Each Ra method in Table 2 has been

approved for drinking water compliance analysis by the USEPA and therefore should not produce significantly different results [5].

Table 1. List	of lab	oratories	and metl	hods used	l in	the II	LC.
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Lab	²³² Th	²²⁶ Ra	²²⁸ Ra
A	M908.0	M903.0	Ra-05
В	ESM4506	M903.0	M904.0
С	ASTM D3972	M903.0	M904.0
D	DOE EML HASL-300 Th-01-RC Modified	M903.1	Ra-05

3. Results

The analytical results for total and dissolved 226 Ra and 228 Ra at all locations reported by the four laboratories are provided in Table 3. All dissolved 232 Th results were below the detection limit (0.0131 to 1.11 pCi/L) and only two total 232 Th measurements were detectable at low concentrations (0.0074 and 0.017 pCi/L); therefore, the 232 Th results are not tabulated. This section provides: (1) a summary of the quality assurance/quality control (QA/QC) results (Section 3.1), (2) a general overview of the dissolved and total 226 Ra and 228 Ra results from the various laboratories, (3) a more detailed comparison of dissolved and total 226 Ra and 228 Ra and (4) an evaluation of the combined 226 Ra results between laboratories relative to the MCL.

3.1. Quality Assurance/ Quality Control (QA/QC) Summary

Laboratory data quality was evaluated based on results from method blanks (MB), laboratory control samples (LCS), LCS duplicates (LCSD), matrix spikes (MS), and MS duplicates (MSD) using USEPA guidelines [6]. In the MBs, all ²²⁶Ra and ²²⁸Ra results were below their minimum detectable concentration (MDC). Data accuracy evaluation showed the percent recovery values for the LCS, LCSD, MS, and MSD were within recommended limits. Data precision evaluation showed that the relative percent differences (RPDs) for analytical duplicate samples and field duplicate samples were also within laboratory control limits for all analyses. Total ²²⁸Ra was detected above the MDC in the field blank by Lab A and Lab B, and dissolved ²²⁶Ra was detected above the MDC in the field blank by Lab C. Dissolved ²²⁸Ra was also detected above the MDC in the field blank by Lab D. Samples containing less than 10x the amount found in the field blanks were qualified as J+ (estimated with high bias) (Table 3).

Radiological data such as 226 Ra and 228 Ra are typically reported as the measured value \pm the degree of precision. When a value is below the MDC, the actual value is still reported but the datum is flagged as non-detect and the MDC is provided. The sum of the 226 Ra and 228 Ra is subsequently calculated to determine the combined Ra (226 Ra+ 228 Ra) that is typically used for compliance in the USA. If either of the 226 Ra or 228 Ra values are below detection, the value of the MDC for the non-detect value is summed with the detectable value to calculate the 226 Ra+ 228 Ra concentration. When both 226 Ra and 228 Ra values are below detection, the respective MDC values are summed to obtain the MDC for the combined 226 Ra and 228 Ra, which is then reported as below detection. The sum of individual precision values for 226 Ra and 228 Ra is used as the overall precision for 226 Ra+ 228 Ra.

Table 2. Comparison of USEPA-approved methods for Ra [4].

Analyte	Method /	Method Principle	Method		
Anaryte	\mathbf{MDL}^1	wiethoù i inicipie	Notes		
²²⁶ Ra	903.0 / 0.5	Radiochemical precipitation; alpha counting by scintillation or gas-flow proportional counting.	alpha activity of a drinking water sample is greater than 5 pCi/L, then the ²²⁶ Ra analysis (Method		
²²⁶ Ra	903.1 / 0.5	Radiochemical precipitation; radon emanation; alpha counting by scintillation.	903.1) is required." No radioactive interferences. 226Ra in solution is determined by co-precipitation with BaSO4 and sample analyzed using de-emanation. Measures 228Ra alone or in con-		
²²⁸ Ra	904.0 / 1.0	Radiochemical precipitation; counting by gas-flow propor- tional beta counter	junction with ²²⁶ Ra. Ra in solution from ²²⁸ Ra determination is saved, and the Ra is reprecipitated for ²²⁶ Ra analysis. Ra is pre-		
²²⁸ Ra	Ra-05 \ NR²	O	Ra is precipitated as Ra-BaSO ₄ which is dissolved in a pentasodium diethylenetriamine pentacetate solution. After ingrowth of ²²⁸ Ac, ²²⁸ Ac is extracted with Di 2-ethylhexyl phosphoric acid and back-extracted with HNO ₃ .		

¹MDL = minimum detectable level (pCi/L). ²NR = not reported.

3.2. Summary of ²²⁶Ra Results

The dissolved and total ²²⁶Ra results for all samples are shown on Figure 3. The error bars represent the precision in the analyses. Any reported value which was below the MDC is indicated with the MDC shown in black, and any J+ qualified value (estimated with high bias) is shown in red. None of the dissolved ²²⁶Ra results were significantly different based on their range in precision. The results from Lab C were J+ qualified and were the highest reported at each location, except for the Impoundment. The values reported by Lab D using M903.1, which is specific to ²²⁶Ra, tended to be lower for all samples, with the exception of Well 176 where the Lab A value was slightly lower (Figure 3).

The trend in lower values using M903.1 was not observed for total 226 Ra (Figure 3), except in Well 176 where total 226 Ra was below detection. However, the value of <0.36 pCi/L for total 226 Ra in Well 176 should be equal to or greater than the dissolved value because the sample was not filtered, therefore the value of <0.36 pCi/L is suspect. No 226 Ra was detected in the unfiltered field blanks, and therefore none of total 226 Ra values were qualified as estimated with high bias (Table 3, Figure 3).

Table 3. Comparison of ILC results (pCi/L) for ²²⁶Ra and ²²⁸Ra.¹

Location	Lab	²²⁶ Ra	²²⁶ Ra	²²⁸ Ra	²²⁸ Ra
		(dissolved)	(total)	(dissolved)	(total)
Impoundment	A	1.8	1.5	1.7	3.7 J+
	В	2.2	1.5	< 0.69	1 J+
	C	1.52 J+	1.56	<0.61	< 0.69
	D	1.4	1.66	<0.7	<0.7
	A	0.7	0.8	<1.9	4.1 J+
Well 176	В	0.97	0.74	3.2	3.2 J+
(Compliance)	C	1.37 J+	1.2	4.6	4.8
_	D	0.965	< 0.356	5.2 J+	5.9
	A	0.7	8.0	3	4.1 J+
Well 176	В	1.3	0.83	3.8	4.7 J+
(Duplicate)	C	1.08 J+	0.96	3.9	5.1
- · · · · · · · · · · · · · · · · · · ·	D	0.859	0.777	5.5 J+	7
	A	0.5	0.5	1.9	1.5 J+
147-11 170	В	<0.66	0.58	1.7	4.8 J+
Well 179	C	0.77 J+	0.73	2.07	2.68
	D	0.334	0.639	3.7 J+	5.4
	A	0.3	0.3	2.9	3.3 J+
Well 182	В	<1	< 0.57	<0.75	< 0.55
(Background)	C	0.4 J+	0.32	< 0.62	< 0.64
	D	<0.328	0.487	0.8 J+	1.8
	A	<0.2	<0.2	<1.4	2
Eigld Dlaul.	В	<1.1	<0.55	< 0.63	1.4
Field Blank	C	0.28	< 0.25	<0.51	< 0.63
	D	< 0.249	< 0.234	0.8	<0.7

 $^{^{1}}$ Bold values with a J+ qualifier indicates the value is qualified as an estimate with high bias when detected in the field blank and the sample result is < 10x the amount measured in the field blank.

3.3.Summary of ²²⁸Ra Results

Overall, the ²²⁸Ra results (Figure 4) tended to be more variable across locations and between laboratories compared to ²²⁶Ra (Figure 3). A comparison of these results shows that ²²⁶Ra tends to be higher in the Impoundment, whereas ²²⁸Ra is higher in the groundwater, consistent with historical Site observations. However, no regular trends are apparent based on the method or laboratory used for dissolved ²²⁸Ra. For example, although Lab A and Lab D use Ra-05, the Lab D results were notably higher than Lab A in samples where ²²⁸Ra concentrations were overall higher (≥ 2 pCi/L). The Lab D results were qualified as estimated with high bias due to detection of ²²⁸Ra in the field blank. Although a higher concentration of ²²⁸Ra was reported by Lab A for the field blank, the detection limit (<1.4 pCi/L) was higher compared to other laboratories and methods (Figure 4, Table 2). In samples containing overall lower ²²⁸Ra concentrations (≤ 2 pCi/L), the results from Lab A were notably higher compared to Lab D. The results from Lab B and Lab C using M904.0 tended to be intermediate between those from Lab A and Lab D, where Ra-05 was used (Figure 4).

Similar trends were observed for total ²²⁸Ra compared to dissolved ²²⁸Ra (Figure 4). However, for total ²²⁸Ra, the results from Lab A, rather than Lab D, were qualified as estimated with high bias due to detection of ²²⁸Ra in the field blank (Table 3, Figure 4). Lab A reported 2 pCi/L of ²²⁸Ra in the field blank, and similar to ²²⁶Ra values, results from Lab A were notably higher in samples containing overall low ²²⁸Ra (Impoundment, Well 182).

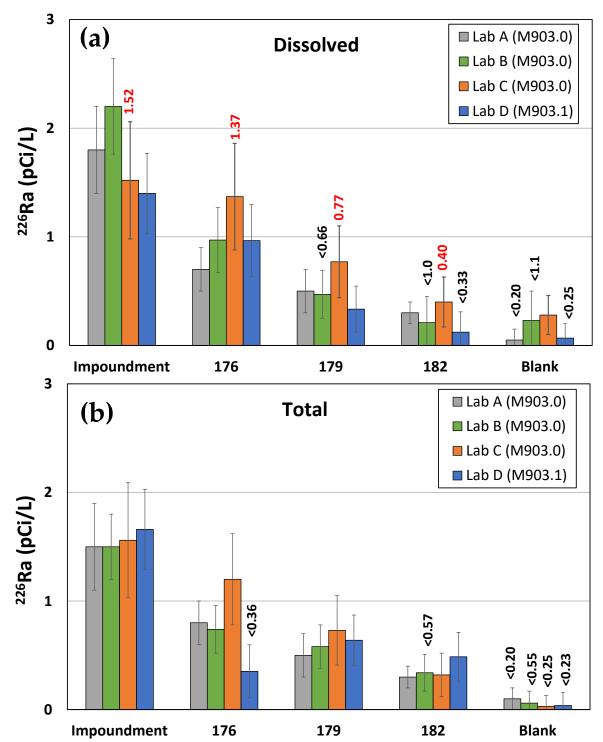


Figure 3. Comparison of: **(a)** Dissolved ²²⁶Ra; **(b)** Total ²²⁶Ra. Values in red are qualified as estimated with high bias (J+) when sample result contained <10x the amount measured in the field blank. Values in black signify the MDC when the reported value is less than the MDC.

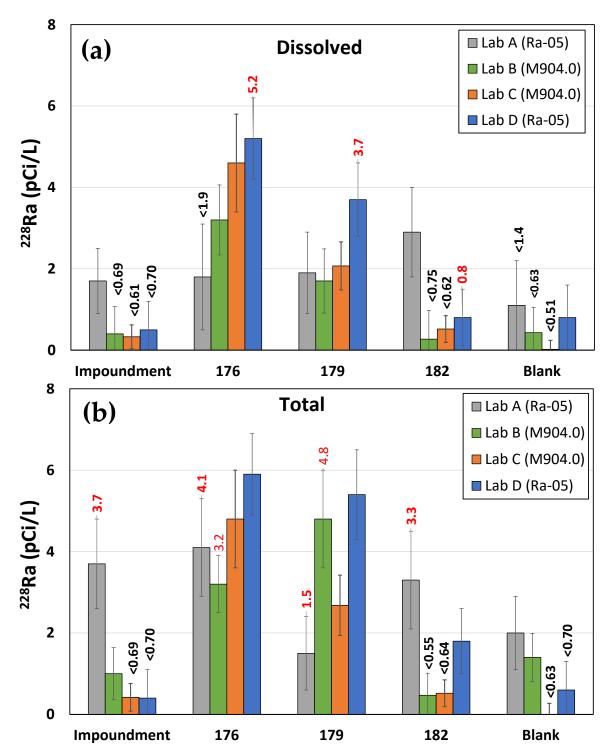


Figure 4. Comparison of: **(a)** Dissolved ²²⁸Ra; **(b)** Total ²²⁸Ra. Values in red are qualified as estimated with high bias (J+) when sample result contained <10x the amount measured in the field blank. Values in black signify the MDC when the reported value is less than the MDC.

3.4. Detailed Comparison of Total vs. Dissolved Ra

A comparison of total and dissolved ²²⁶Ra from each laboratory is shown on Figure 5. The error bars represent the precision in the analyses, with non-detect values shown in black and all J+ qualified values shown in red. Theoretically, total concentrations should be equal to or greater than dissolved concentrations, with the extent of Ra partitioning to suspended particles depending on factors such as total suspended solids concentration and pH [2]. Total and dissolved ²²⁶Ra were generally the same in all samples based on the range in precision of the analyses (Figure 5). However, the difference between the actual reported values in total vs. dissolved at a given location were not consistent between laboratories. For example, the reported total ²²⁶Ra in the Impoundment was higher than dissolved for the Lab D data, lower than dissolved for Lab A and Lab B, and almost equal for Lab C. In the background Well 182, the reported total ²²⁶Ra was greater than dissolved for Lab D and Lab B, less than dissolved for Lab C, and approximately equal for Lab A.

For the ²²⁸Ra data shown on Figure 6, total ²²⁸Ra was more consistently greater than or equal to the dissolved fraction when compared to ²²⁶Ra on Figure 5. It is also apparent that both total and dissolved ²²⁸Ra values from Lab A are elevated in the samples containing low Ra (Impoundment, Well 182, and field blank) compared to the other lab results, including those from Lab D where the same method (Ra-05) was used. In addition, the values reported for total ²²⁸Ra in the field blank were higher compared to non-detectable dissolved ²²⁸Ra from Lab A and Lab B. Differences between total and dissolved ²²⁸Ra in the field blank suggests that processing of the sample for total analysis could have produced false positive results.

3.5. Combined ²²⁶Ra+²²⁸Ra Results

The ²²⁶Ra+²²⁸Ra values for the dissolved and total fractions are shown in comparison to the 5 pCi/L compliance value on Figure 7. The dissolved ²²⁶Ra+²²⁸Ra values reported by Lab C and Lab D exceeded the MCL at Well 176. Total ²²⁶Ra+²²⁸Ra reported by Lab C and Lab D also exceeded the MCL at Well 176, in addition to total ²²⁶Ra+²²⁸Ra reported by Lab A for the Impoundment and total ²²⁶Ra+²²⁸Ra reported by Lab B and Lab D at Well 179. Overall the trends are not entirely consistent between laboratories. For example, there was little difference between dissolved and total ²²⁶Ra+²²⁸Ra for the Impoundment and Well 176 in the data reported by Lab B, Lab C, and Lab D, while the total ²²⁶Ra+²²⁸Ra values reported by Lab A were higher than dissolved for the same two locations. Conversely, values were higher for total ²²⁶Ra+²²⁸Ra compared to dissolved ²²⁶Ra+²²⁸Ra at Well 179 for the data reported by Lab B add Lab D, while total vs. dissolved ²²⁶Ra+²²⁸Ra concentrations reported by Lab A were similar. The dissolved ²²⁶Ra+²²⁸Ra in the field blank was below detection as reported by Lab A and Lab B, while total ²²⁶Ra+²²⁸Ra was detectable; the opposite trend was observed for the data reported by Lab C and Lab D (Figure 7).

4. Summary and Discussion

This ILC of ²²⁶Ra and ²²⁸Ra was conducted to understand the potential reasons for high variability and increasing trends in ²²⁶Ra+²²⁸Ra concentrations that have been observed in Site monitoring wells, including the compliance Well 176 and background Well 182. The results for ²²⁶Ra were fairly consistent between laboratories, and with the exception of dissolved ²²⁶Ra reported by Lab C, all ²²⁶Ra concentrations were below detection in the field blanks. The concentrations of ²²⁸Ra were more variable between laboratories and with a higher frequency of detection in the field blanks.

For both ²²⁶Ra and ²²⁸Ra, there were no consistent trends with respect to the total vs. dissolved concentrations across laboratories, which were essentially equal in most samples when the range in measured precision was considered. The largest differences between reported values for total vs. dissolved Ra were observed in the Lab A data for ²²⁸Ra, where the total ²²⁸Ra data were qualified as estimated with high bias. For both dissolved and total ²²⁶Ra+²²⁸Ra, the results from Lab C and Lab D for the compliance Well 176

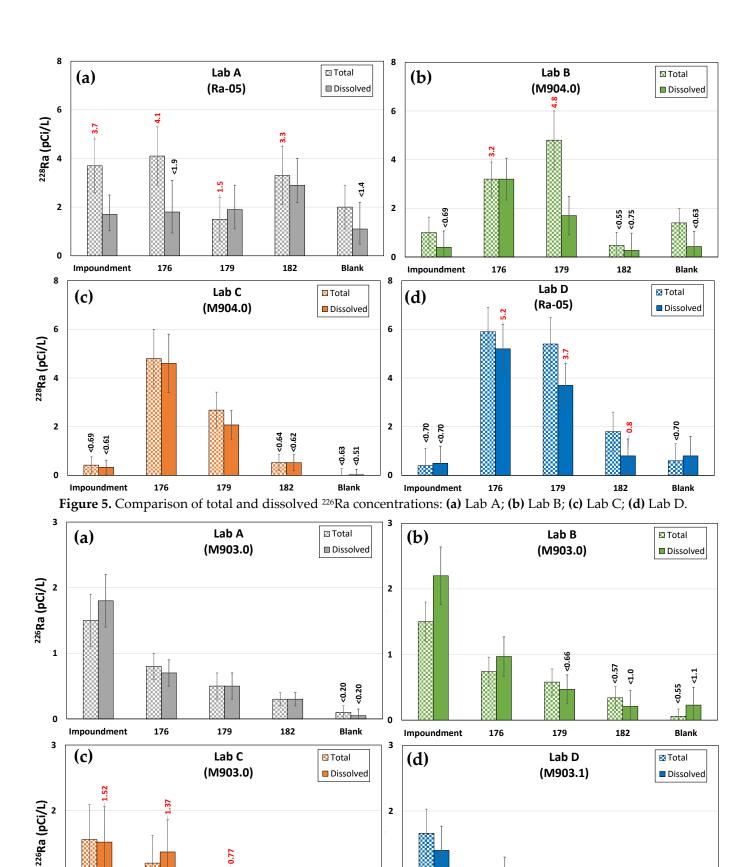


Figure 6. Comparison of total and dissolved ²²⁸Ra concentrations: (a) Lab A; (b) Lab B; (c) Lab C; (d) Lab D.

Impoundment

Blank

182

Impoundment

176

179

<0.36

176

179

182

<0.25

Blank

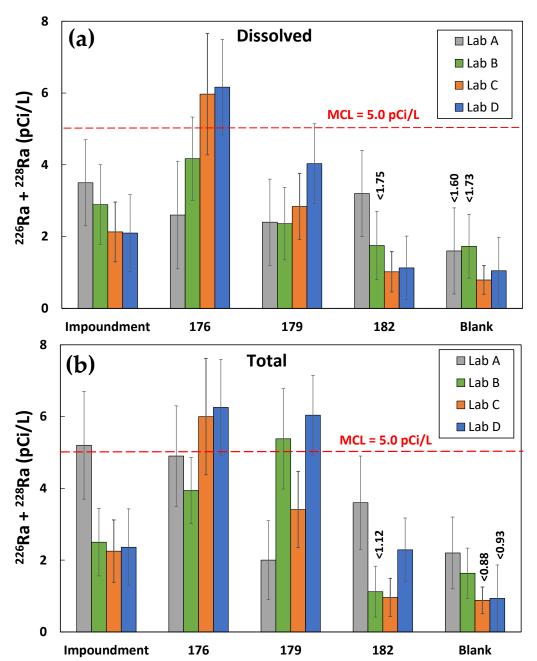


Figure 7. Interlaboratory comparison of: **(a)** Dissolved ²²⁶Ra+²²⁸Ra ; **(b)** Total ²²⁶Ra+²²⁸Ra. Values in black signify the MDC when the reported values were less than the MDC.

exceeded the MCL, whereas the results from Lab A and Lab B were below the MCL. Therefore, compliance with the MCL was either achieved or not achieved, depending on which laboratory conducted the analyses.

Results for ²²⁶Ra obtained using M903.1 (specific to ²²⁶Ra) were usually equal to or slightly lower than those generated using M903.0; however, the differences are insignificant and may be due to analytical variability rather than method specificity. Results for ²²⁸Ra reported by Lab D when using Method Ra-05 were more consistent with those from Lab B and Lab C when using M904.0. The ²²⁸Ra results reported by Lab A when using Method Ra-05 tended to be higher in samples where the remaining laboratories reported low ²²⁸Ra concentrations (< 2 pCi/L), and tended to be lower in samples where the remaining laboratories reported higher ²²⁸Ra concentrations (> 2 pCi/L). Lab A also reported the highest measured ²²⁸Ra concentration of all field blanks (2 pCi/L), and consequently those data for total ²²⁸Ra were qualified as estimated with high bias.

A prior study relating to radionuclide variability in water samples [5] showed that radionuclide activity in groundwater can vary up to four-fold due to seasonal effects, but also noted that the variability could reflect issues related to method and laboratory variability. The greatest source of error for ²²⁶Ra is counting uncertainty, but the overall accuracy and precision for ²²⁶Ra results was acceptable among various laboratories. For ²²⁸Ra, however, the methods are generally not as robust as those used for ²²⁶Ra. No single laboratory was able to provide consistently accurate ²²⁸Ra results, and due to the much greater relative standard deviation (RSD) values, it was concluded that the accuracy of ²²⁸Ra results are dependent on laboratory technique and skill and "may also reflect less rugged analytical methods" [5].

4. Conclusions and Recommendations

The results obtained from this ILC of Site groundwater and surface water samples are generally consistent with the findings from the prior 2011 study [5] with respect to observed variability in ²²⁶Ra compared to ²²⁸Ra. The current ILC indicate that ²²⁸Ra data reported by the existing contract laboratory (Lab A) are biased high relative to those reported by Lab B, Lab C, Lab D, and this may be due to differences in laboratory skill and technique among the various laboratory technicians [5]. However, the resulting upward trends in ²²⁸Ra/²²⁶Ra ratios (observed in both impacted wells and the background location) after the ²²⁸Ra method was changed from 904.0 to Ra-05 are more difficult to explain.

Although the trends were not entirely consistent, the results from this study show a tendency for total Ra to be higher than dissolved Ra, especially for ²²⁸Ra. For samples where the majority of the laboratories reported low ²²⁸Ra concentrations (< 2 pCi/L), it appears that the method and techniques used by Lab A for total ²²⁸Ra is contributing to higher apparent ²²⁸Ra concentrations. Therefore, it was recommended that ²²⁶Ra and ²²⁸Ra analyses (in additional to uranium and other radionuclides) be conducted on a filtered sample for Site compliance evaluations. Analysis for dissolved concentrations would also then be consistent with remaining metals analyses which are also analyzed as dissolved.

For ²²⁶Ra, it was recommended that M903.0 be replaced with M903.1 which is specific to ²²⁶Ra. Method 903.0 includes all alpha emitters and therefore can result in over-reporting of ²²⁶Ra concentrations. The primary additional alpha emitter which could contribute to overestimation of ²²⁶Ra is ²²⁴Ra, which originates from ²³²Th. Although ²³²Th concentrations were very low to non-detect in the Site samples, the presence of ²³²Th in the solid phase of aquifer materials could release ²²⁴Ra into solution. Even though this study did not indicate that M903.0 results in consistent over-reporting of ²²⁶Ra concentrations, use of M903.1 would better ensure that the appropriate analysis for ²²⁶Ra is being conducted. Alternatively, it could be requested that the laboratory employ a 14-day waiting time which will effectively reduce the contribution of ²²⁴Ra to the reported ²²⁶Ra [5].

For ²²⁸Ra, this study indicates that specific laboratory techniques, rather than the method (Ra-05) used by Lab A is potentially contributing to excessively elevated values being reported for ²²⁸Ra at the Site. It should be noted that even though standard methods [4] are being used, many laboratories have developed their own customized Standard Operating Procedures (SOPs) which is often proprietary and difficult to obtain. However, it is recommended that any location whose reported ²²⁶Ra+²²⁸Ra concentrations exceed groundwater or surface water standards in the future be immediately re-sampled and reanalyzed by the existing laboratory, with a sample split also being collected and sent to an outside laboratory for additional confirmatory analysis.

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