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Article

# Suspension Reactor Studies for the Photocatalytic Degradation of Acetaminophen

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## Abstract

Traditional wastewater treatment plants are not generally designed to handle degradation of pharmaceutical-based contaminants, nor do such plants generally incorporate processes that enable complete elimination of pharmaceutical residues. Therefore, pharmaceutical metabolites can be found in groundwater, surface water, and even drinking water in low concentrations. Advanced Oxidation Processes (AOPs) are promising to avoid the contamination of agricultural land and aquifers. Here, we focus specifically on the use of photocatalytic degradation processes via titanium dioxide (TiO<sub>2</sub>) nanoparticles as a photocatalyst, assisted by UV radiation and enhanced by H<sub>2</sub>O<sub>2</sub>. This degradation approach may allow the degradation of pharmaceutical compounds, such as acetaminophen (ACE), in municipal water treatment plants in which a tertiary treatment phase is suitable. We employ a batch, water-cooled jacketed reactor equipped with a UV lamp containing a TiO<sub>2</sub> nanoparticle suspension for the degradation of ACE. We identify favorable conditions for contaminant degradation (~63%) at low concentrations of both ACE and titanium dioxide after 90 minutes of treatment. Several combinations of potential AOPs approaches were compared (e.g., UV-C only, UV-C/H<sub>2</sub>O<sub>2</sub>, UV-C/H<sub>2</sub>O<sub>2</sub>/TiO<sub>2</sub> and H<sub>2</sub>O<sub>2</sub> alone). In general, the combined action of the techniques leads to a better degradation of the contaminant. The results suggest AOPs are promising candidates for removal of pharmaceutical metabolites from wastewater effluents.

**Keywords:** bio-pharmaceutical; advanced oxidation processes; suspension reactors; AOP enhancers

## 1. Introduction

The extensive use of pharmaceuticals for the treatment of both human and animal health conditions has expanded into an emerging environmental problem due to their continuous discharge into water systems and their persistence [1]. These products, sometimes referred to as pharmaceutically active compounds (PhACS), are discharged into wastewater effluents either largely unchanged, i.e. in their original chemical state or in the form of degradation products that are typically not suitable to elimination in traditional water treatment plants [2]. Depending on the nature of the chemical compounds and the treatment used to eliminate them, the PhACS will reach surface and ground water systems and, in the worst cases, will be found in drinking water [3–5].

In general, traditional wastewater treatment plants have not been designed to deal with contaminants such as pharmaceuticals and other related biomedicines [6]. Due to the inability of such wastewater treatment plants to completely eliminate drug residues and their metabolites, the presence of these can be found in groundwater, surface water, and even, in some cases, in drinking water at low concentrations [4,7]. As examples, studies have shown the presence of contaminants such as acetaminophen, ibuprofen, diclofenac, and carbamazepine at levels of 10 µgL<sup>-1</sup>, 112 ngL<sup>-1</sup>, 518 ngL<sup>-1</sup> and 263 ngL<sup>-1</sup>, respectively, in different bodies of water [8]. Even though the concentration of organic compounds is within the currently accepted limits, the tests measuring their concentrations indicate that there is a continuous inflow of these substances into water systems that, eventually, will lead to accumulation [9]. There is, therefore, a need for efficient and environmentally-friendly

technologies to eliminate these pharmaceuticals from water sources for a clean, healthy, and sustainable environment. In summary, evidence indicates an increasingly-alarming picture of bioaccumulation of pharmaceutical and industrial organic matter in water supplies suggesting the potential to have a negative impact on human health and the sustainability of wildlife and entire ecosystems [4,10].

Advanced oxidation processes (AOPs) have the potential to remove pharmaceutical and other organic water contaminants from our water supply [11–13]. AOPs can be used in water purification and treatment, including for the destruction of naturally occurring toxins, contaminants of emerging concern (e.g., estradiol etc.), and pesticides, to name a few [14]. The development of AOPs began in the 1990s and includes a variety of methods to generate hydroxyl radicals and other reactive oxygen species such as superoxide anion radical, hydrogen peroxide and singlet oxygen [15]. Hydroxyl radicals can abstract an electron from electron-rich substrates to form a radical cation which, in an aqueous medium, will hydrolyze and ultimately form an oxidized product. AOPs include but are not limited to UV/O<sub>3</sub>, UV/H<sub>2</sub>O<sub>2</sub>, Fenton, photo-Fenton, non-thermal plasmas, sonolysis, and photocatalysis [16].

This contribution focuses on the study of the degradation of ACE via a selected group of AOPs. In wastewater treatment plants (WWTPs), the tertiary stage of treatment is where such degradation might most likely be achieved. Table 1 briefly describes some of the challenges and potential areas where investigation is needed as related to tertiary treatment processes at municipal WWTPs. As the table shows, different approaches to treat stable contaminants have been taken (such as the use of chlorine), and some of these have unfavorable outcomes. Notably, other techniques that underperform contaminant degradation include the use of UV-based treatment processes, however, these have been utilized without addition of enhancers [14]. The inadequacy of current techniques at removing ACE and similar pharmaceutical contaminants is thus an area of great concern. To this end, we investigate the use of enhancers, including TiO<sub>2</sub> and H<sub>2</sub>O<sub>2</sub>. Our results indicate that enhancers greatly increase the degradation rate of UV-based treatment processes, providing an exciting new approach to enhance the capacity of WWTPs to remove hazardous organic contaminants from our water supply.

**Table 1.** Review Options for Tertiary Treatment Stage.

Tertiary Treatment Process	Comments
<b>Past:</b> Disinfectant - Chlorination	Pharmaceuticals reacting with chlorine could produce products that could pose issues/hazards. It may cause cancer, create unfavorable conditions in the environment when water is discharged.
<b>Current:</b> UV – Treatment	Challenges: Not all compounds can be degraded by this treatment and better conditions and approaches are needed.
<p data-bbox="220 1760 742 2033"><b>Future:</b> UV + TiO<sub>2</sub> /Enhancers</p> <p data-bbox="220 1843 742 2033">This approach is promising based on other efforts reported in the literature. However, further research is needed to identify optimal and other conditions for a particular contaminant.</p>	<ul data-bbox="767 1809 1358 2072" style="list-style-type: none"> <li data-bbox="767 1809 1358 1921">• Promising approach for photocatalysis with potential combination of selected complementary approaches.</li> <li data-bbox="767 1966 1358 2072">• Challenges: Use of suspension reactor for TiO<sub>2</sub> nanoparticles requires a post-filtering stage.</li> </ul>

### 1.1. Model Compound (Acetaminophen)

Acetaminophen (a.k.a., N-acetyl-p-aminophenol, paracetamol) (see Figure 1) is a non-narcotic analgesic drug used for relieving minor aches and pains associated with backache, headache, and arthritis. It is a widely used over-the-counter drug and is a common component of cold and flu medications. The demand for acetaminophen in the United States was estimated to be 30–35 thousand tons per year [17], and dramatic increases in product demand were observed during 2020–2022, associated with SARS-CoV-2 pandemic [14]. Furthermore, demand is expected to continue to increase as the population continues to grow and as availability and access expands to previously underserved global economies [14].

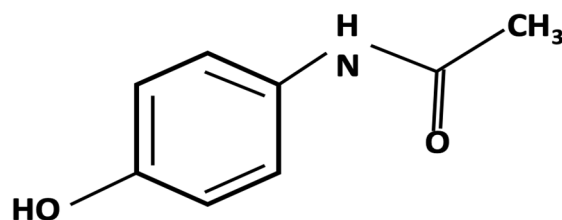


Figure 1. Acetaminophen Molecule.

Acetaminophen is not a naturally occurring compound and was first synthesized by Harmon Northrop Morse in 1877 [18]. Today, it is known, that acetaminophen (or paracetamol, as it is referred to in some parts of the world) is a major product of the metabolism of degradation of phenacetin. The physical and chemical properties of acetaminophen are summarized in Table 2. Due to its widespread use and incomplete degradation, large amounts of acetaminophen have been found in river water, groundwater, sewage, and landfill leachate at levels ranging from a few ng/L to µg/L [19]. Overdose of this compound can lead to the accumulation of toxic metabolites causing fetal hepatotoxicity [20]. In the United States, acetaminophen toxicity is the most common cause of acute hepatic failure, which frequently requires organ transplantation [19].

Table 2. Physical and Chemical Properties of Acetaminophen [14].

<b>Synonyms</b>	Tylenol, Paracetamol, Paracetamolo, 4-Hydroxyaceranilide, Acetaminophen, etc.
<b>Toxicity</b>	Oral rat 2404mg/Kg
<b>Physical State</b>	White Crystalline Power
<b>Melting Point</b>	168°C
<b>Density/Specific Gravity</b>	1.293g/cm <sup>3</sup> at 21°C
<b>Molecular Weight</b>	151.165g/mol
<b>Water Solubility</b>	14000mg/L(25°C)
<b>Half Life</b>	2 to 3 hours
<b>Henry's Law Constant</b>	5.0x10 <sup>-9</sup> Pa m <sup>3</sup> Mol <sup>-1</sup>
<b>pKa Value</b>	9.38
<b>Log K<sub>ow</sub></b>	0.34

## 2. Materials and Methods

### 2.1. Materials

A batch, temperature-controlled Pyrex photochemical glass reactor with a 1000 mL capacity and equipped with a magnetic stirrer, a water circulating jacket, and an opening for aeration was used (see Figure 2). The photoreactor has a T 60/40 center joint and three T 14/20 angle joints. Among the three joints, one is used as a sample port, one is helpful to insert a thermometer to measure the reaction temperature, and the remaining port is maintained airtight by inserting a solid rubber stopper. The reactor vessel holds a Hanovia 450 UV mercury vapor lamp in the reactor central port to radiate UV light to the reaction solution. The lamp was attached to an external power supply (See Figure 2, right side). The list of equipment includes a UV-Visible spectrophotometer, a pH meter, a 28 mm syringe, and Corning 0.45  $\mu\text{m}$  filter paper sheets that were purchased from Fisher Scientific.

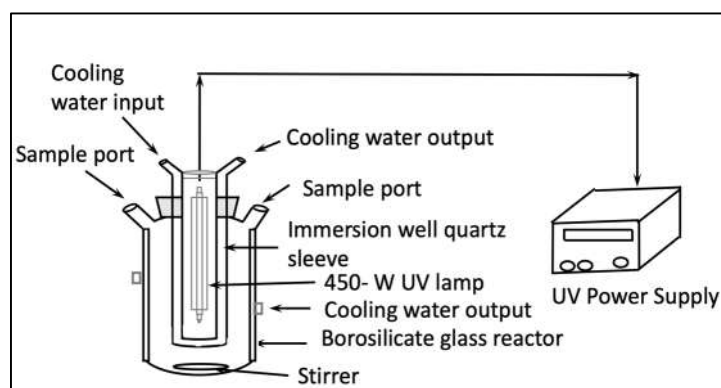


Figure 2. UV-Photocatalytic Set Up for Project [21].

### 2.2. Chemicals

Acetaminophen (N-(4-hydroxyphenyl) acetamide) powder of pure quality (100%) was purchased from Fisher Scientific and used in the reaction. The photocatalyst,  $\text{TiO}_2$  nanoparticles (Titanium (IV) dioxide, 98% anatase powder), was supplied by Sigma-Aldrich and was used to test its effect on the degradation of the pharmaceutical compound at different concentrations. Hydrogen peroxide (30% in water) was purchased from Fisher Scientific. 1M (molar solution) sulfuric acid ( $\text{H}_2\text{SO}_4$ ) and 1M sodium hydroxide (NaOH) were used to change the pH of the solution.

During the suspension study performed in the batch reactor, a number of possibilities (see Table 3) were investigated to check for the most suitable degradation conditions of ACE. The data collected was used for input or guiding the selection of conditions for the proposed technique for the partial or total removal of the contaminant from aqueous solutions.

### 2.3. Photocatalytic Experiments

As indicated in Section 2.1, a cylindrical borosilicate glass photoreactor with a total capacity of one liter was used in the experiments (see Figure 2). During these studies, 0.05 g of acetaminophen was prepared in 500 mL deionized (DI) water and fed into the reactor vessel. The light source used was a medium pressure mercury lamp (Hanovia 450 W UV, Ace Glass, Vineland, NJ, USA) which was placed into the double-wall immersion well containing jacket, with tap water in circulation to cool the reactor (see Figure 2). The light source was connected to a 450 W power supply. Additionally, a magnetic stirrer was utilized to properly mix the suspension.

### 2.4. Factorial Design

In order to determine the optimal number of experiments to be conducted (see Table 3), a  $2^n$  factorial experimental design [14,22] was used. Here,  $n$  corresponds to the primary variables to be

tested in the experiment. The following two variables were used: (i) ACE concentration and (ii) TiO<sub>2</sub> concentration, both with a high (+) level and a low (-) level. As shown in Table 3, four experiments (n=2) were designed for testing.

Table 3: Experimental Design for the ACE Treatment with TiO<sub>2</sub>/UV

Design ACE/TiO <sub>2</sub>	Results/Figures	ACE Concentration	TiO <sub>2</sub> Concentration	UV-Hanovia 450 W-	Media pH	Remarks
High/High	Figure 3	High (+)	High (+)	On	Acidic-4.7-5.5.0	R1
High/Low	Figure 4	High (+)	Low (-)	On	Acidic 4.7-5.0	R2
Low/High	Figure 5	Low (-)	High (+)	On	Acidic 4.7-5.0	R3
Low/Low	Figure 6	Low (-)	Low (-)	On	Acidic 4.8-5.5	R4

A clarification in Table 3: The following values were used for TiO<sub>2</sub>: Low = 0.5 g/ 500 mL; High = 1.00 g/ 500 mL and for ACE: Low = 0.05 g/ 500 mL; High = 0.1 g/ 500 mL

### 2.5. Experimental Procedure

After preparation of the reaction suspension, experiments were performed in the batch reactor. The ultraviolet treatment was conducted for total of 90 minutes with a 5 mL sample taken out from the reactor in ten-minute intervals to check for degradation of ACE. Samples were taken with a 28 mm syringe and then filtered through a 0.45 µm filter in order to remove the solid particles. The pH of the suspension was monitored with a pH meter, and the reaction temperature was maintained at room temperature. For the actual determination of concentration, a 1 mL sample was removed from the 5 mL sample by using a micropipette and diluted to 10 mL in order to determine the particular contaminant concentration by following the procedure outlined in the section below.

### 2.6. Sample Analysis

The spectrophotometric-based process is an excellent analytic technique to quantitatively measure the concentration of different organic compounds present in aqueous media. The photocatalytic degradation of acetaminophen was analyzed by using a Perkin Elmer Lambda 2 UV spectrophotometer at a wavelength of 243 nm, that corresponds to the maximum absorbance of acetaminophen. A 1 mL sample was removed from the samples taken from the batch reactor at a given time, by using a micropipette and diluted to 10 mL with deionized water for accurate detection to help dissolve other compounds (such as intermediaries that do not completely degrade) that may be present in that suspension samples.

Absorbance was converted to concentration by using the Beer- Lambert Law (see Appendix 1) yielding the following relationship:

$$C_1 = C_0 \frac{A_1}{A_0}$$

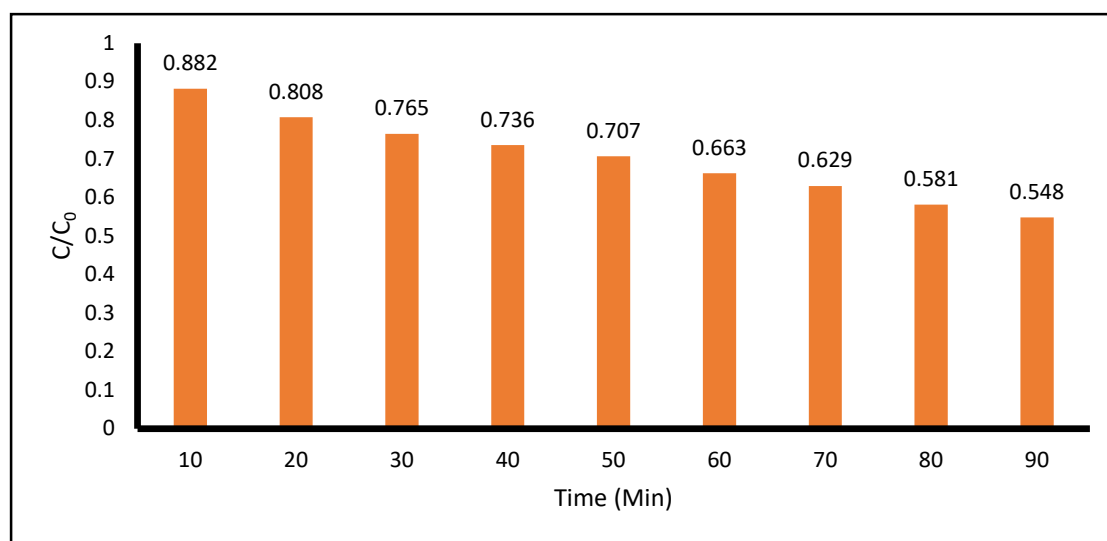
## 3. Data Collected and Brief Descriptions and Observations of Results

Based on the implementation of the experimental strategy described in Section 2.4 (See Table 3), a series of experiments were conducted, and degradation results were successfully collected for description and analysis (see Section 3.2). A description of the different experiments is offered in Section 3.1, and a discussion of the key trends found in the results collected is included in Section 3.2.

### 3.1. Description of Experiments and Results Collected

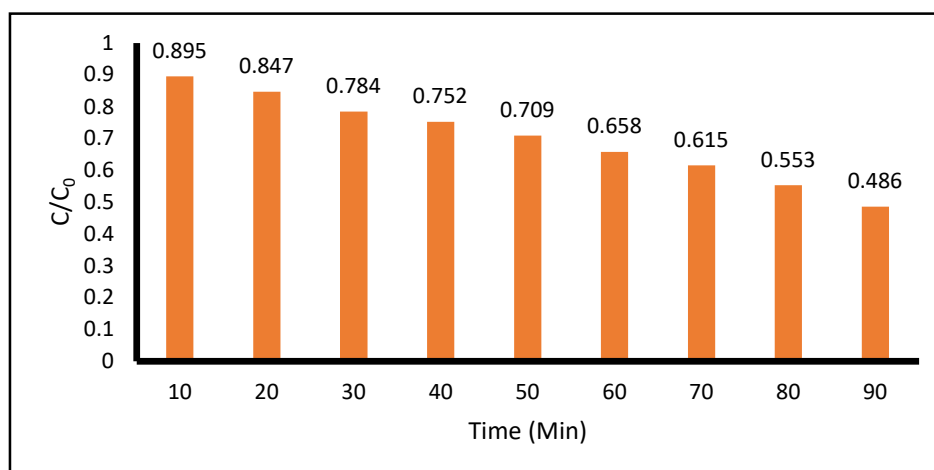
Degradation of ACE as a model pharmaceutical contaminant in an aqueous solution with the aid of TiO<sub>2</sub> as a photocatalyst used as a suspension of nanoparticles in the batch reactor was monitored over the period of 90 minutes. The experimental conditions followed a 2<sup>n</sup> factorial design described in the Section 2.4, (See Table 3). The absorbance of all the samples was measured under the fixed wavelength of 243 nm from the spectrophotometer used to determine the concentration of ACE as a function of time. The total treatment time was 90 minutes for the various cases studied.

Figure 3 shows the photocatalytic degradation of ACE for the case with the highest concentrations of both the model contaminant and TiO<sub>2</sub>. The results show a degradation of 11.8% and 45.2% degradation at 10 and 90 minutes, respectively. During the experiment, the pH level in the reaction vessel was also monitored indicating that the pH remains relatively constant in the acidic domain between 4.7 and 5.5.



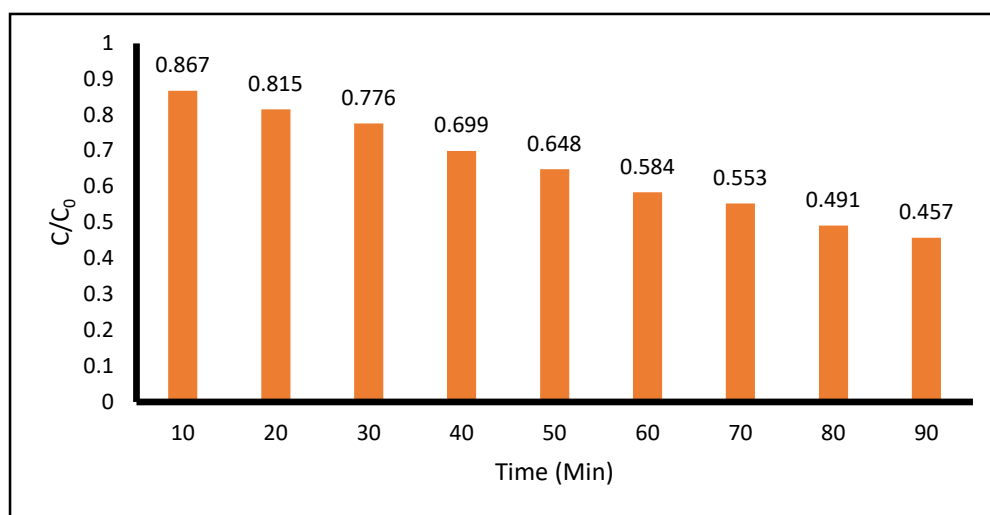
**Figure 3.** Photocatalytic Degradation of High Concentration of Acetaminophen VS High Concentration of TiO<sub>2</sub> measured at 243 nm by the spectrophotometric technique. (Acetaminophen = 0.1 g, TiO<sub>2</sub>=1.07 g, Total Volume = 500 mL).

Figure 4 shows the degradation of the model contaminant for the case of high concentration of ACE, and low concentration of TiO<sub>2</sub> during the total 90 minutes of treatment time. The levels of degradation, at 10 and 90 minutes of treatment time, were 10.5% and 51.4%, respectively, as measured by the spectrophotometric technique indicated above. With the conditions of the experiment, the pH remains within a relatively small acidic range with values between 4.7 and 5.0.



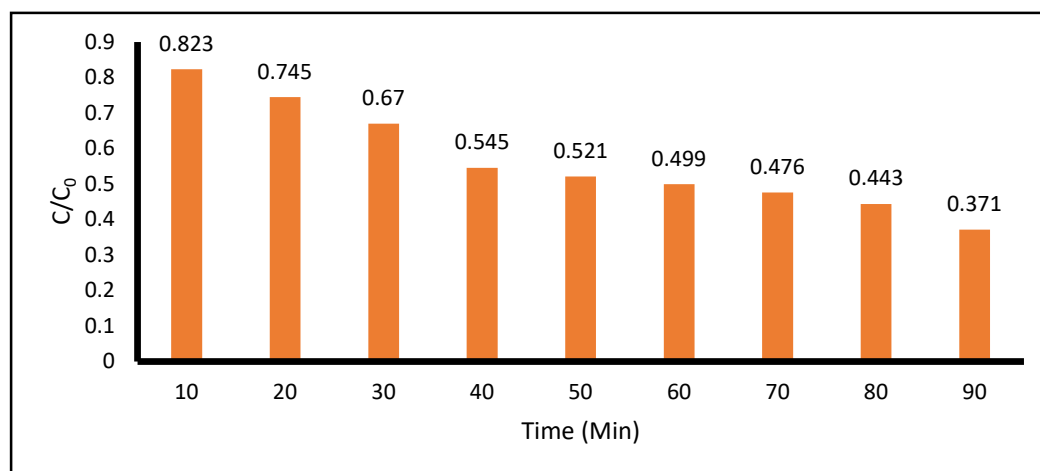
**Figure 4.** Photocatalytic Degradation of High Concentration of Acetaminophen VS Low Concentration of TiO<sub>2</sub> measured at 243 nm by spectrophotometric technique. (Acetaminophen = 0.1 g, TiO<sub>2</sub> = 0.5 g, pH= 5.0, Total Volume = 500 mL).

Figure 5 shows the degradation of the model contaminant for the case of low concentration of ACE and high concentration of TiO<sub>2</sub> during 90 minutes of total treatment time. The levels of degradation at 10 and 90 minutes of treatment time, were 13.3% and 54.3%, respectively, as measured at 243 nm by the spectrophotometric technique. In this condition, the pH remained acidic with the values between 4.5 and 5.5.



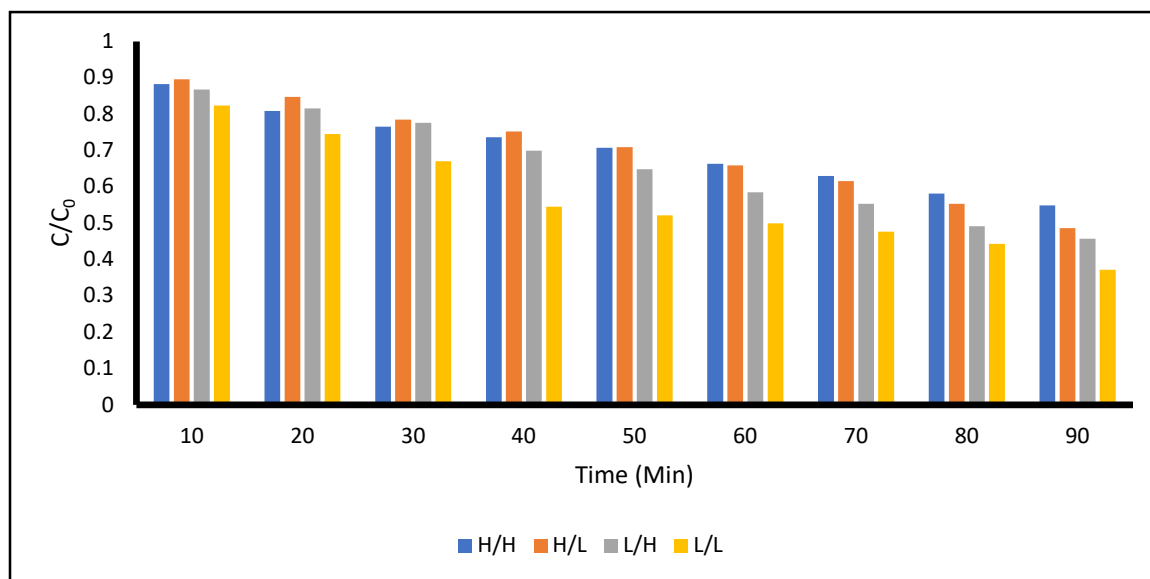
**Figure 5.** Photocatalytic Degradation of Low Concentration of Acetaminophen VS High Concentration of TiO<sub>2</sub> measured at 243 nm by spectrophotometric technique. (Acetaminophen =0.05 g, TiO<sub>2</sub> =1.05 g, pH= 5.5, Total Volume = 500 mL).

Figure 6 shows the degradation of the model contaminant for the case of low concentration of ACE and low concentration of TiO<sub>2</sub> during 90 minutes of total treatment time. The degradation at 10 and 90 minutes of treatment time was 17.7% and 62.9%, respectively, as measured at 243 nm by the spectrophotometric technique. As before, in this condition, the pH remained acidic, with values between 4.8 and 5.5.



**Figure 6.** Photocatalytic Degradation of Low Concentration of Acetaminophen vs Low Concentration of TiO<sub>2</sub> measured at 243 nm by spectrophotometric technique. (Acetaminophen=0.05 g, TiO<sub>2</sub> =0.5 g, pH=5.5, Total Volume = 500 mL).

Finally, Figure 7 provides a comparison of the degradation of ACE at different times of reactor exposure within the total time of 90 minutes frame for different ratios of concentration of ACE and concentration of the photocatalyst. A quick observation is that the case of low concentration of ACE treated with the low concentration of photocatalyst seems to be the most effective for all the times that measurements were taken. On the contrary, the case of highest concentration of ACE treated with the highest concentration of the photocatalyst shows that the degradation is less effective. Data from samples collected during the low ACE concentration and high photocatalyst concentration studies as well as the high ACE concentration and low photocatalyst concentrations are intermediate to the low-low and high-high design experiments. Additional comments will be offered in the section below.



**Figure 7.** Comparison of the Degradation of Acetaminophen on different Time Intervals with different Concentrations of Photocatalyst ( $\text{TiO}_2$ ) and Model Compound ACE.

### 3.2. Discussion of Degradation Results

From Figure 7, as was indicated in Section 3.1, it can be observed that favorable conditions for the ACE degradation (~63%) were identified at low concentrations of the contaminant treated with a low concentration of titanium dioxide at 90 minutes of treatment. Although more studies are needed, a preliminary explanation of the results reported can be suggested. The photocatalytic efficiency depends on the interactions of radiation with the compound on the photocatalyst surface. It seems that incident light with this wavelength having low concentrations of the ACE and photocatalyst has a better light-surface contact and less light dispersion. Therefore, more energy may be used for a greater photocatalytic activity producing more electron ( $e^-$ ) ( $h^+$ ) pairs on the  $\text{TiO}_2$  surface, which is helpful for the degradation of the ACE compound [23].

In the case of high substrate (ACE) concentration, photonic efficiency diminishes because more organic substances are absorbed on the  $\text{TiO}_2$  surface. However, the intensity of radiation and irradiation time are constant, which possibly limits the formation of  $\text{O}_2^-$  and  $\text{OH}^-$  on the  $\text{TiO}_2$  surface, which ultimately decreases the photo-degradation [24]. In the suspension medium, the catalyst particles are freely dispersed in fluid phase, therefore, the photocatalyst is fully integrated into the liquid mobile phase. One potential aspect limiting efficiency that has been indicated in the literature is the light scattering effect [14]. Therefore, only the proper amount of photocatalyst as it is needed should be added to the solution due to the increase in light scattering and adsorption by particles in the suspended medium. This aspect was not studied in the project, and it will require further work to completely determine the optimal level of nanoparticles needed in the solution.

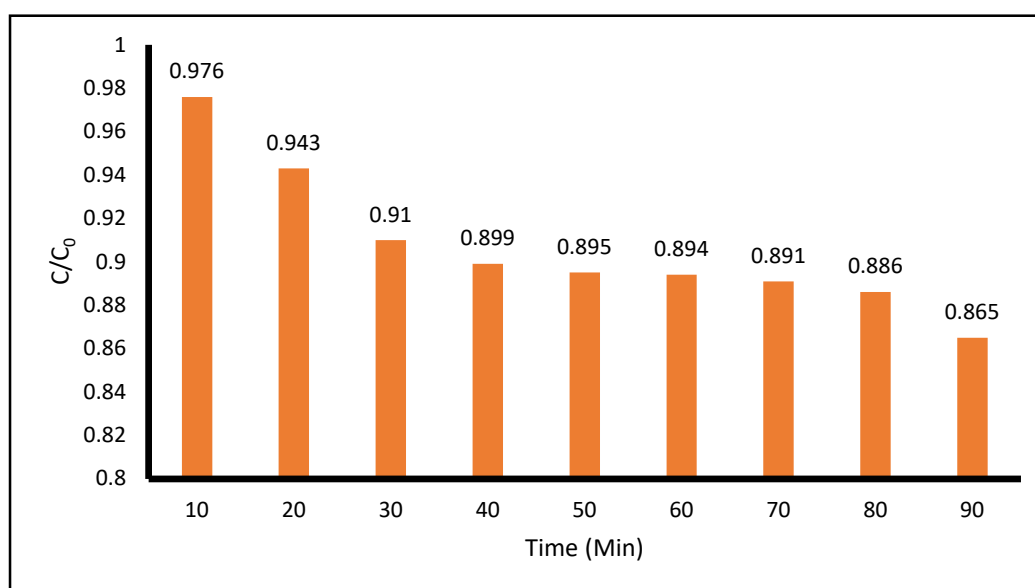
#### 4. Different Degradation Approaches

In order to gain a better understanding of the different degradation techniques associated with AOPs and their impact on the extent of conversion of ACE, additional studies were performed. The aim of these experiments was to compare the degradation levels achieved with different AOPs including the use of UV (alone), H<sub>2</sub>O<sub>2</sub> treatment, dual UV/H<sub>2</sub>O<sub>2</sub>, and UV/H<sub>2</sub>O<sub>2</sub>/TiO<sub>2</sub> combined. The effect of H<sub>2</sub>O<sub>2</sub> volume (i.e., 1 mL and 5 mL) was also used to investigate degradation of ACE. Furthermore, data collected was used for comparisons with the experiments already conducted for the case of UV/TiO<sub>2</sub> (See Section 3.1). Table 4 gives the experimental design matrix of the different AOPs for different experiments performed and selected remarks are also included.

Table 4 -Experimental Design for ACE Treatment with Enhancers

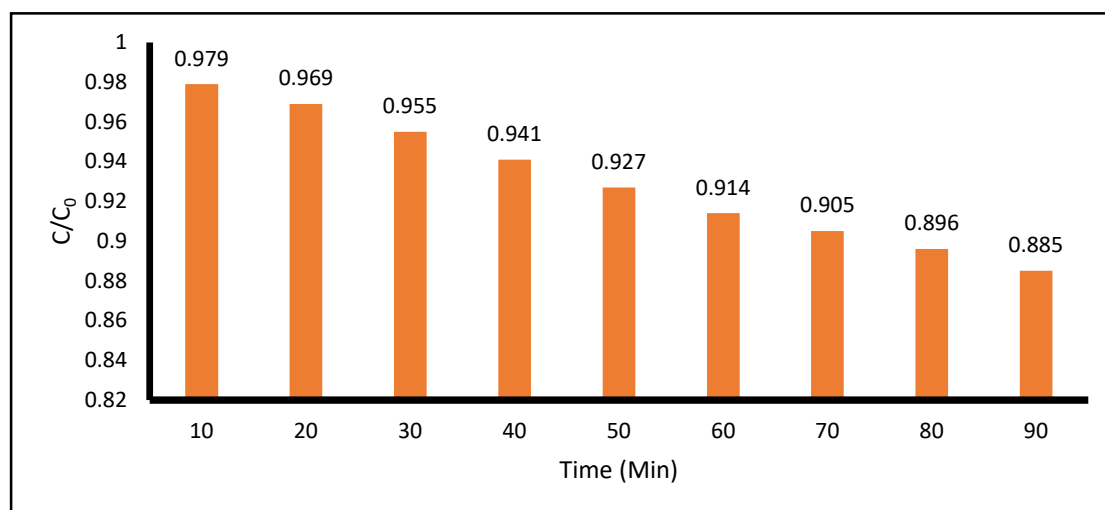
Design	Results	ACE Concentration	TiO <sub>2</sub> Concentration	UV-Hanovia 450 W-	H <sub>2</sub> O <sub>2</sub>	Media pH	Remark
UV-Only	Figure 8	High (+)	None	On	None	Acidic-	R1
H <sub>2</sub> O <sub>2</sub> -Only	Figure 9	High (+)	None	Off	5mL	Acidic	R2
UV+ H <sub>2</sub> O <sub>2</sub>	Figure 10	High(+)	None	On	1mL 5mL	Acidic	R3/R4
UV+ H <sub>2</sub> O <sub>2</sub>	Figure 11	High(+)	None	On	1mL 5mL	Basic	R5
UV+ H <sub>2</sub> O <sub>2</sub> +TiO <sub>2</sub>	Table 6 Case F	High(+)	Low (-) 0.5 g	On	1mL	Acidic 4.8-5.5	R6

R1: Figure 8 shows the degradation of ACE with the application of *only* UV-C radiation during 90 minutes of treatment. The degradation of ACE at the 10- and 90-minute points of the photocatalytic process was 2.4% and 13.5%, respectively. The pH measured during the entire process remained between 4.7 and 5.



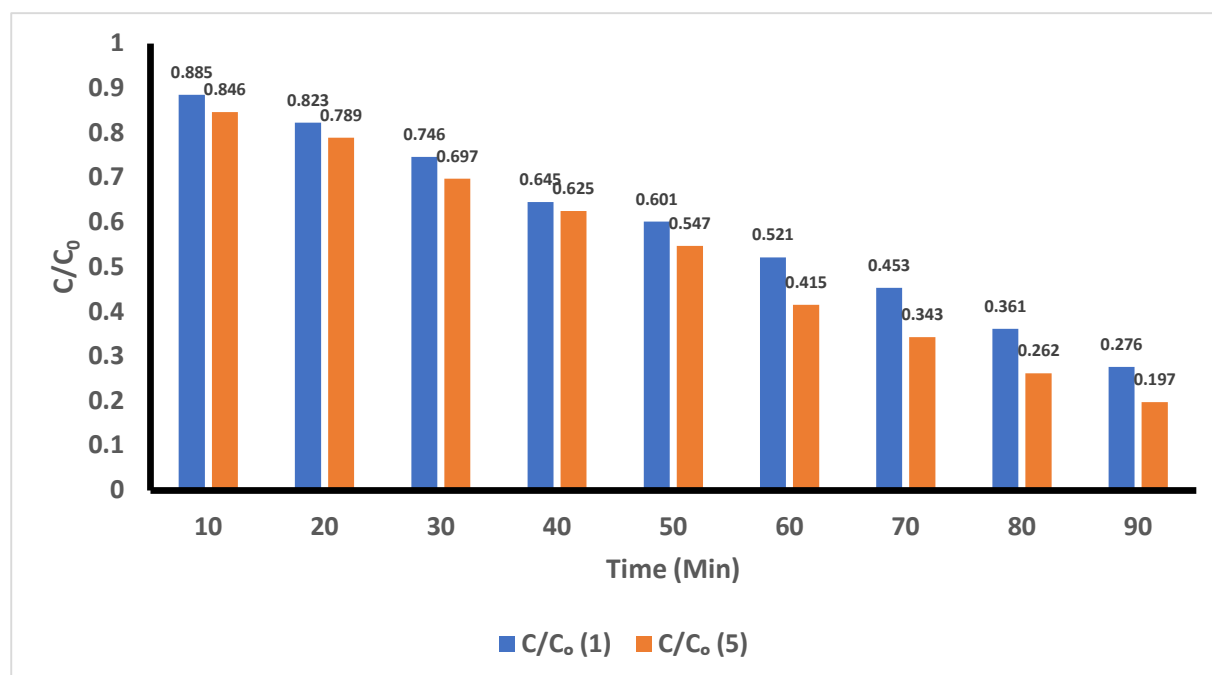
**Figure 8.** Photocatalytic Degradation of Acetaminophen (ACE) with the Application of UV light measured at 243 nm by spectrophotometric techniques. (Acetaminophen = 0.05 g, pH = 4.8, Total volume = 500 mL).

R2: Figure 9 summarizes the degradation of ACE achieved by the treatment with the application of *only* H<sub>2</sub>O<sub>2</sub>. Spectroscopic analysis was performed on the samples of ACE treated solution as a function of time. During the 90 minutes of reaction time, only 11.5% of ACE degradation was achieved in the acidic medium (pH = 4.7 to 5.1).



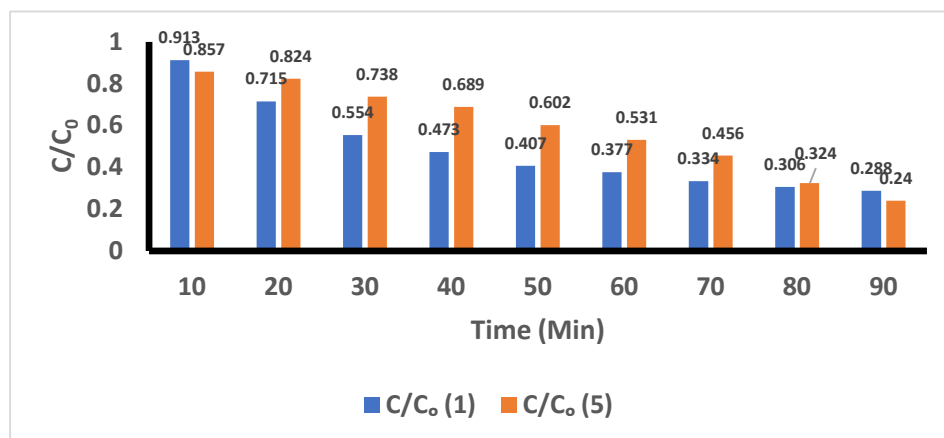
**Figure 9.** Photocatalytic Degradation of ACE with the Application of H<sub>2</sub>O<sub>2</sub> (only) measured at 243 nm by spectrophotometric techniques. (Acetaminophen=0.05 g, pH=4.9, volume of H<sub>2</sub>O<sub>2</sub> = 5 mL, Total Volume = 500 mL).

R3/R4: In Figure 10, a comparison of the degradation of ACE in acidic medium when using different volumes of H<sub>2</sub>O<sub>2</sub> (1 mL and 5 mL) and while applying UV-C radiation is shown. The results reveal degradation values of 11.5% and 72.4% were achieved by applying 1 mL of H<sub>2</sub>O<sub>2</sub> after 10 minutes and 90 minutes of treatment, respectively. Similarly, we found that by applying 5 mL of H<sub>2</sub>O<sub>2</sub> during the testing, degradation levels of ACE were 15.4% and 80.3% after 10 minutes and 90 minutes of treatment, respectively.



**Figure 10.** Comparison of the Degradation of Acetaminophen when Using Different Volumes of H<sub>2</sub>O<sub>2</sub> (1 mL and 5 mL) and with Application of UV Light and an Acidic Medium (pH= 4.8) (Acetaminophen = 0.05 g, Total Volume = 500 mL).

R5: In Figure 11, a comparison of the degradation of ACE in basic medium when using different volumes of H<sub>2</sub>O<sub>2</sub> (1 mL and 5 mL) and while applying UV-C radiation is shown. The results show that after addition of 1 mL H<sub>2</sub>O<sub>2</sub> that 8.7% and 71.2% of ACE was degraded after 10 minutes and 90 minutes of treatment, respectively. Similarly, by applying 5 mL of H<sub>2</sub>O<sub>2</sub> to the solutions, the results indicate that the ACE degradations achieved were 14.3% and 76.0% at 10 and 90 minutes, respectively.



**Figure 11.** Comparison of the Degradation of Acetaminophen when Using Different Volumes of H<sub>2</sub>O<sub>2</sub> (1 mL and 5 mL) and with Application of UV Light and Basic Media (pH= 9.0 and 11.3) (Acetaminophen = 0.05 g, Total Volume = 500 mL).

The effect of pH on ACE degradation is summarized in Table 5. As expected [4], the largest conversion of ACE is found for the neutral medium (pH=7.3). For example, at 10 minutes of treatment, the conversion was 15.6%, compared to 11.5% in the acidic medium (pH=5) and 8.7% in the case for basic medium (pH=9). When the treatment time increases to 90 minutes, the values change to 75.7% (neutral), 72.4% (acidic) and 71.2% (basic). The volume of H<sub>2</sub>O<sub>2</sub> that was used in all three cases was 1 mL. The effect of increasing the volume of H<sub>2</sub>O<sub>2</sub> was also tested by using 5 mL. The results of the degradation of ACE for the case of 10 minutes of treatment show that the degradation reaches 19.7% (neutral), 15.4% (acidic) and 14.3% (basic) while for 90 minutes of treatment the degradation increased to 81.1% (neutral), 80.3% (acidic) and 76% (basic).

**Table 5.** Effect of pH on the Degradation of ACE under UV Radiation.

Volume of H <sub>2</sub> O <sub>2</sub> (mL)	Medium (Acidic) pH=5		Medium (Basic) pH=9		Medium (Neutral) pH=7.3	
	Time (Min)					
	10 min	90 min	10 min	90 min	10 min	90 min
1mL	11.5%	72.4%	8.7%	71.2%	15.6%	75.7%
5mL	15.4%	80.3%	14.3%	76.0%	19.7%	81.1%

R6: 6 The effect of adding TiO<sub>2</sub> along with the UV+H<sub>2</sub>O<sub>2</sub> treatment was also tested, please see case F in Table 6. Also, in Table 6, a comparison of the degradation of ACE with the application H<sub>2</sub>O<sub>2</sub> (1 mL and 5 mL), TiO<sub>2</sub> (0.5g and 1.0g), and UV light are shown. The results indicate that by applying 1.0 mL H<sub>2</sub>O<sub>2</sub> and 0.5g TiO<sub>2</sub>, degradation rates were 23.7% and 82.8% during 10 and 90 minutes of treatment, respectively. Contaminant degradation after application of 5 mL H<sub>2</sub>O<sub>2</sub> and 1g TiO<sub>2</sub> were 15.2% and 67.1% degradation during 10 and 90 minutes of treatment, respectively. .

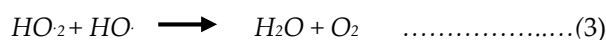
A comparison of all results is provided in Table 6. The results show the most favorable conditions for ACE degradation were achieved with the application of UV/H<sub>2</sub>O<sub>2</sub>/TiO<sub>2</sub> after 90 minutes of treatment, yielding a degradation of 82.8%. In general, one can observe the beneficial action of the mutual work of two or more of the degradation techniques working together. For example, the use of UV-only and H<sub>2</sub>O<sub>2</sub>-only produces similar degradation and relatively low values (i.e. at 90 minutes of treatment, 13.50% vs 11.50%) while when the actions of UV-TiO<sub>2</sub> are together the degradation reaches 62.90% for the same treatment time; this increases to 82.80% with the addition of 1 mL of H<sub>2</sub>O<sub>2</sub>. Additional comments and a statistically based analysis is provided in Section 5, below.

**Table 6.** Comparison of the Degradation of Acetaminophen at different time intervals with the application of UV, H<sub>2</sub>O<sub>2</sub>, UV/H<sub>2</sub>O<sub>2</sub>, UV/TiO<sub>2</sub> and UV/H<sub>2</sub>O<sub>2</sub>/TiO<sub>2</sub>

Reduction of ACE	UV Only	H <sub>2</sub> O <sub>2</sub> Only	UV/H <sub>2</sub> O <sub>2</sub> (1mL)	UV/H <sub>2</sub> O <sub>2</sub> (5mL)	UV/TiO <sub>2</sub>	UV/H <sub>2</sub> O <sub>2</sub> /TiO <sub>2</sub>
Time VS. Group	A	B	C	D	E	F
10 min	2.40%	2.10%	15.60%	19.70%	17.70%	23.70%
20 min	5.70%	3.10%	19.60%	30.20%	25.50%	34.30%
30 min	9.00%	4.50%	26.50%	39.70%	33.00%	39.10%
40 min	10.10%	5.90%	33.80%	47.30%	45.50%	45.30%
50 min	10.50%	7.30%	43.10%	55.90%	47.90%	53.70%
60 min	10.60%	8.60%	55.60%	63.40%	50.10%	58.20%
70 min	10.90%	9.50%	60.00%	72.50%	52.40%	63.90%
80 min	11.40%	10.40%	67.70%	76.80%	55.70%	76.30%
90 min	13.50%	11.50%	75.70%	81.10%	62.90%	82.80%

## 5. Comparative Results: Discussion on Different AOPs

The research reported in this contribution is aimed to increase understanding of the role of the different AOPs in degrading ACE. For example, with the application of only UV-C radiation, ACE was not degraded efficiently, consistent with results reported elsewhere. It has been postulated that the presence of -NH<sub>2</sub> groups stabilizes the compound under these conditions [25]. However, the degradation of ACE significantly increases with UV-C and H<sub>2</sub>O<sub>2</sub> or TiO<sub>2</sub>, or H<sub>2</sub>O<sub>2</sub> and TiO<sub>2</sub> combined (see Table 6). Addition of these enhancers most likely helps to generate hydroxyl radicals (HO·), which are highly reactive, have a short lifespan, and are capable of oxidizing different organic compounds [25]. The formation of radicals takes place when H<sub>2</sub>O<sub>2</sub> is combined with UV-C radiation according to the following possible reaction mechanism [26].



### 5.1. Effect of H<sub>2</sub>O<sub>2</sub> Volume

When small amounts of H<sub>2</sub>O<sub>2</sub> are introduced into the UV-C/TiO<sub>2</sub> system, the rate of degradation is increased due to the considerable generation of hydroxyl radicals by UV-C/H<sub>2</sub>O<sub>2</sub>. (See Table 6, Case F; and Figure 10). Since H<sub>2</sub>O<sub>2</sub> is a strong electron acceptor, oxygen reacts with electrons which are coming from the valence band of the photocatalyst to generate HO· and OH·. If the reactive species, such as oxygen, and H<sub>2</sub>O<sub>2</sub> are not present near the surface of the TiO<sub>2</sub>, the electron-hole pairs will recombine and the energy absorbed will degenerate as heat [27]. As a result, if H<sub>2</sub>O<sub>2</sub> is low, the photocatalyst is capable of suppressing radical formation, leading to reduced contaminant degradation. Conversely, excess H<sub>2</sub>O<sub>2</sub> can act like an electron hole, or it may react with H<sub>2</sub>O<sub>2</sub> which ultimately forms peroxide compounds and hampers the degradation process [28]. Therefore, optimal concentrations of H<sub>2</sub>O<sub>2</sub> are needed to attain maximum degradation results.

### 5.2. Effect of pH on Acetaminophen Photodegradation

Results with the effect of solution pH are summarized in Table 5. The pH of the solution determines the surface charge properties of TiO<sub>2</sub>, the size of aggregates formed, the charge of organic molecules, adsorption of molecules on the catalyst, and the concentration of the hydroxyl radicals [29]. In an acidic or neutral medium, ACE is in its non-ionic form, where water solubility is the highest and the adsorption on the photocatalyst is increased. At a nearly neutral pH, the degradation reaches the highest efficiency. This is most likely due to the fact that a larger number of OH· radicals are generated since, as pH increases, more hydroxide ions on the active sites of TiO<sub>2</sub> can be oxidized to form more OH·. However, as the pH increases, electrostatic repulsion occurs between the photocatalyst surface and ACE (pK<sub>a</sub> = 9.5). Faster destruction of oxidizable compounds present in alkali medium could compete for the hydroxyl radical. Therefore, it leads to lower ACE degradation and minor mineralization.

At a high pH level (pH > 9), in the UV-C/H<sub>2</sub>O<sub>2</sub> system, the degradation of ACE is slow due to the self-decay of H<sub>2</sub>O<sub>2</sub>. This would foster the waste of H<sub>2</sub>O<sub>2</sub>. The molar extinction coefficient of HO<sub>2</sub>· on UV-C absorption will be higher with H<sub>2</sub>O<sub>2</sub> in the system, so less HO· will be generated. With the increase of the pH level, a fraction of HO<sub>2</sub>· will also increase and possibly scavenge more H<sub>2</sub>O<sub>2</sub> and HO· in the system. Therefore, the overall degradation rate would decrease [27].

### 5.3. Effect of Photocatalyst Dosage

Results with TiO<sub>2</sub> addition are included in Section 3.2. For these studies, photocatalytic degradation increases with the increase of TiO<sub>2</sub> dosages. However, further increase of TiO<sub>2</sub> may lead to a detrimental effect on degradation [9]. The increase in the degradation rates is due to the increase in total surface area, i.e. number of active sites, available for the photocatalytic reaction. When the dosage of TiO<sub>2</sub> is in excess, the intensity of incident light will be reduced because of the decrease in light penetration and the rise of light scattering. This will lessen the overall photo degradation process [30].

### 5.4. Effect of Acetaminophen Concentration

The photocatalytic degradation depends on the initial concentration of ACE. The degradation rate will increase with the increasing concentration of ACE, however, further increase in the concentration of reactants will show a negative trend on degradation [31]. This most likely occurs due to the number of photocatalyst active sites already occupied by the increased amount of ACE present in the solution, thereby blocking the formation of oxidants [32]. Additionally, higher contaminant concentrations of ACE can compete with the absorbance of photons to activate and the photocatalyst and produce electron-hole pairs, that would hinder the activation of photocatalyst causing the lower degradation of ACE [33].

## 6. Statistical Analysis of the Degradation Results

An ANOVA (no replication) method has been applied to the data collected for the degradation of ACE at different time intervals with the application of UV, H<sub>2</sub>O<sub>2</sub>, UV/H<sub>2</sub>O<sub>2</sub>, UV/TiO<sub>2</sub> and UV/H<sub>2</sub>O<sub>2</sub>/TiO<sub>2</sub> as it is shown in Table 6. Each degradation method has been considered a different group (A-F). Furthermore, two hypotheses for the ANOVA can be stated as: 1- H<sub>0</sub> (Null Hypothesis): The null hypothesis posits that there is no significant variance between the methods utilized for the degradation of ACE. It specifically assumes equal means for the response across all groups. 2. H<sub>1</sub> (Alternative Hypothesis): The alternative hypothesis proposes that there exists a significant difference among the methods applied for the degradation of ACE. It suggests that at least one group's mean differs from the others.

To study the behavior of the case of UV only vs UV/TiO<sub>2</sub>, (columns A-F in Table 6), we summarize the results of the ANOVA test in Table 7. For this case the F-statistics are high while the p-value is less than 0.05. Therefore, it can be stated that a statistically significant difference exists between the groups; the asterisk (\*) count represents the level of intensity; model is successful. The rejection of the null hypothesis (H<sub>0</sub>) suggests a notable dissimilarity among the various groups.

Table 7. Results of the ANOVA Test for A-F.

Table Analyzed	Col: One-way ANOVA
Data sets analyzed	A-F
<b>ANOVA summary</b>	
F	15.88
P value	<0.0001
P value summary	****
Significant diff. among means (P < 0.05)?	Yes
R squared	0.6233

Therefore, it can be inferred that the different applied methods significantly impact the degradation of ACE, thereby supporting the alternative hypothesis (H<sub>1</sub>). Also, for multiple comparison and to further comprehend the relationships between groups, post-hoc tests are necessary. In this study, Tukey's test has been applied to the data of Table 6 for this purpose, and the results are included in Table 8. Some useful observations can be made from this table. For example, when comparing all pairs of groups, if the obtained p-value is below 0.05, it indicates a statistically significant difference between the two distinct degradation methods. In Table 8, The notation "ns", signifies that there is no statistically significant difference between the two groups being compared. Furthermore, from the values shown in Table 8, there are significant differences between following groups:

- UV Only vs. UV/H<sub>2</sub>O<sub>2</sub>(5 mL)
- UV Only vs. UV/H<sub>2</sub>O<sub>2</sub>/TiO<sub>2</sub>
- H<sub>2</sub>O<sub>2</sub> Only vs. UV/H<sub>2</sub>O<sub>2</sub>(5 mL)
- H<sub>2</sub>O<sub>2</sub> Only vs. UV/H<sub>2</sub>O<sub>2</sub>/TiO<sub>2</sub>

In addition, the percentage reduction of ACE for these top two groups is as below, based on Table 6 are:

- UV/H<sub>2</sub>O<sub>2</sub>(5 mL) : 81.10%
- UV/H<sub>2</sub>O<sub>2</sub>/TiO<sub>2</sub> : 82.80%

Conversely, by comparing the p-values of UV/H<sub>2</sub>O<sub>2</sub>(5 mL) with UV/H<sub>2</sub>O<sub>2</sub>/TiO<sub>2</sub>, no statistically significant differences were observed in the outcome.

- UV/ H<sub>2</sub>O<sub>2</sub>(5 mL) vs. UV/ H<sub>2</sub>O<sub>2</sub>/ TiO<sub>2</sub> (ns)

This result indicates that by adding TiO<sub>2</sub> to the treatment method no significant change has been achieved with respect to the UV/H<sub>2</sub>O<sub>2</sub>(5 mL) method. There is no added value. Analysis similar to this can help in identifying the most promising combinations of AOPs in order to optimize resources and investment when trying to design a new AOP implementation for the degradation of ACE.

**Table 8.** Results of the Tukey's Multiple Comparison Test for the Data Collected.

Tukey's multiple comparisons test	Mean Diff.	95.00% CI of diff.	Below threshold?	Summary	Adjusted P Value	
UV Only vs. H <sub>2</sub> O <sub>2</sub> Only	2.356	-20.16 to 24.88	No	ns	0.9996	A-B
UV Only vs. UV/H <sub>2</sub> O <sub>2</sub> (1mL)	-34.83	-57.35 to -12.31	Yes	***	0.0004	A-C
UV Only vs. UV/H <sub>2</sub> O <sub>2</sub> (5mL)	-44.72	-67.24 to -22.20	Yes	****	<0.0001	A-D
UV Only vs. UV/TiO <sub>2</sub>	-34.07	-56.59 to -11.55	Yes	***	0.0006	A-E
UV Only vs. UV/H <sub>2</sub> O <sub>2</sub> /TiO <sub>2</sub>	-43.69	-66.21 to -21.17	Yes	****	<0.0001	A-F
H <sub>2</sub> O <sub>2</sub> Only vs. UV/H <sub>2</sub> O <sub>2</sub> (1mL)	-37.19	-59.71 to -14.67	Yes	***	0.0002	B-C
H <sub>2</sub> O <sub>2</sub> Only vs. UV/H <sub>2</sub> O <sub>2</sub> (5mL)	-47.08	-69.60 to -24.56	Yes	****	<0.0001	B-D
H <sub>2</sub> O <sub>2</sub> Only vs. UV/TiO <sub>2</sub>	-36.42	-58.94 to -13.90	Yes	***	0.0002	B-E
H <sub>2</sub> O <sub>2</sub> Only vs. UV/H <sub>2</sub> O <sub>2</sub> /TiO <sub>2</sub>	-46.04	-68.56 to -23.52	Yes	****	<0.0001	B-F
UV/H <sub>2</sub> O <sub>2</sub> (1mL) vs. UV/H <sub>2</sub> O <sub>2</sub> (5mL)	-9.889	-32.41 to 12.63	No	ns	0.7818	C-D
UV/H <sub>2</sub> O <sub>2</sub> (1mL) vs. UV/TiO <sub>2</sub>	0.7667	-21.75 to 23.29	No	ns	>0.9999	C-E
UV/H <sub>2</sub> O <sub>2</sub> (1mL) vs. UV/H <sub>2</sub> O <sub>2</sub> /TiO <sub>2</sub>	-8.856	-31.38 to 13.66	No	ns	0.8501	C-F
UV/H <sub>2</sub> O <sub>2</sub> (5mL) vs. UV/TiO <sub>2</sub>	10.66	-11.86 to 33.18	No	ns	0.7242	D-E
UV/H <sub>2</sub> O <sub>2</sub> (5mL) vs. UV/H <sub>2</sub> O <sub>2</sub> /TiO <sub>2</sub>	1.033	-21.49 to 23.55	No	ns	>0.9999	D-F
UV/TiO <sub>2</sub> vs. UV/H <sub>2</sub> O <sub>2</sub> /TiO <sub>2</sub>	-9.622	-32.14 to 12.90	No	ns	0.8005	E-F

## 7. Summary, Concluding Remarks and Future Work

The efforts of this research project focused on two key aims: The first one was the testing of the degradation of acetaminophen with the use of TiO<sub>2</sub> photocatalyst in the form of a nanoparticle suspension added to a batch photoreactor. The second aim was focused on testing a few selected options of Advanced Oxidation Processes and collecting data for comparison purposes. These included UV-C only, UV-C/H<sub>2</sub>O<sub>2</sub>, UV-C/H<sub>2</sub>O<sub>2</sub>/TiO<sub>2</sub> and H<sub>2</sub>O<sub>2</sub> alone. For the planning of both sets of experiments, a 2<sup>n</sup> factorial design matrix was utilized and implemented in order to guide the collection of data. The total treatment time was set at 90 minutes, and ten-minute intervals were used to collect samples of treated solution for analysis. All the concentration analyses were measured under the fixed wavelength of 243 nm from the spectrophotometer used to identify the concentration of ACE as a function of time.

For the first aim, a successful degradation of ACE in aqueous solution in the batch photoreactor was achieved for two different concentrations of the TiO<sub>2</sub> photocatalyst. The encouraging circumstances for a contaminant degradation of approximately 63% were identified at low concentrations of ACE versus low concentration of titanium dioxide at 90 minutes of treatment at pH 5.5. When either higher concentration of photocatalyst or higher concentration of ACE was used, the degradation efficiency was tampered either by potential dispersion of light due to the increase in nanoparticles or competing adsorption of the contaminant sites with the photocatalyst sites which affected the utilization of the UV-light. These types of results are useful outcomes for the potential use of thin-film photocatalyst approaches that will avoid the filtration stage which is usually costly in the implementation of suspension reactors [4].

For the second aim of comparing selected AOPs, an effective comparison was performed to see the degradation of ACE with different AOPs (e.g., UV-C only, UV-C/H<sub>2</sub>O<sub>2</sub>, UV-C/H<sub>2</sub>O<sub>2</sub>/TiO<sub>2</sub> and H<sub>2</sub>O<sub>2</sub> alone). The most favorable condition for the acetaminophen degradation of 81.5% was found to be the application of UV/H<sub>2</sub>O<sub>2</sub>/TiO<sub>2</sub> at 90 minutes of treatment. The research also indicated that the treatment using UV/H<sub>2</sub>O<sub>2</sub>(5 mL) produced an 81.10% degradation of ACE compared to the treatment of UV/H<sub>2</sub>O<sub>2</sub>/TiO<sub>2</sub> that led to an 82.80% degradation of ACE, with no statistically significant differences observed. This type of result implies that more work needs to be conducted in order to advance the technology for the degradation of ACE. Among potential efforts, the thin-film approaches with the addition of H<sub>2</sub>O<sub>2</sub> and/or other potential enhancers should be an excellent target of future efforts [4,14,34,35].

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## Appendix A

=According to the Beer- Lambert Law, the absorbance and the concentration are related as:

$$A = \epsilon b C \dots\dots\dots A1$$

where A = absorbance, b = path length, C = concentration and,  $\epsilon$  = molar absorption coefficient.

Therefore, with a fixed path length and constant molar absorption coefficient, it can be inferred that the absorbance is directly proportional to the concentration (C).

$$A \propto C \dots\dots\dots A2$$

Considering two different measurements, it can be shown that

$$\frac{A_0}{A_1} = \frac{\epsilon b C_0}{\epsilon b C_1} \dots\dots\dots A3$$

where  $A_0$  and  $C_0$  are absorbance and concentration, respectively, at time zero, while  $A_1$  and  $C_1$  are the absorbance and concentration, respectively, at time  $t_1$ . Since  $\epsilon$  b are constant values independent of the concentration, the following relationship is achieved:

$$C_1 = C_0 \frac{A_1}{A_0} \dots\dots\dots A4$$

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