

1 Article

2 Delineation of Cocoa Agroforests Using Multi- 3 Season Sentinel-1 SAR Images: Low Grey Level 4 Range Reduces Uncertainties in GLCM Texture- 5 Based Mapping

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13 **Abstract:** Delineating the cropping area of cocoa agroforests is a major challenge for quantifying the
14 contribution of the land use expansion to tropical deforestation. Discriminating cocoa agroforests
15 from tropical transition forests using multi-spectral optical images is difficult due to a similarity in
16 the spectral characteristics of their canopy; moreover, optical sensors are largely impeded by the
17 frequent cloud cover in the tropics. This study explores multi-season Sentinel-1 C-band SAR image
18 to discriminate cocoa agroforests from transition forests for a heterogeneous landscape in central
19 Cameroon. We use an ensemble classifier, random forest, to average SAR image texture features of
20 GLCM (Grey Level Co-occurrence Matrix) across seasons; next, we compare classification
21 performance with results from RapidEye optical data. Moreover, we assess the performance of
22 GLCM texture feature extraction at four different grey level quantization: 32bits, 8bits, 6bits, and
23 4bits. The classification overall accuracy (OA) of texture-based maps outperformed that from an
24 optical image; the highest OA of 88.8% was recorded at 6bits grey level. This quantization level, in
25 comparison to the initial 32bits in SAR images, reduced the class prediction error by 2.9%. Although
26 this prediction gain may be large for the landscape area, the resultant thematic map reveals the
27 decrease and fragmentation of forest cover by cocoa agroforests. According to our classification
28 validation, the Shannon entropy (H) or uncertainty provides a reliable validation for class
29 predictions and reveals detail inference for discriminating inherently heterogeneous vegetation
30 categories. The texture-based classification achieved a reliable accuracy considering the
31 heterogeneity of the landscape and vegetation classes.

32 **Keywords:** mapping cocoa agroforests; Congo Basin rainforest; sentinel-1; SAR; GLCM textures;
33 grey level quantization; random forest algorithm; machine learning; classification uncertainty

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36 **Conference paper:**

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40 2018.

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43 1. Introduction

44 The mapping of cocoa commodity cropland is essential to quantify its ecosystem services, and
45 as well the disservices related tropical forest cover loss. Agricultural land expansions, predominantly
46 for oil palm, rubber, and cocoa plantations, contribute significantly to tropical deforestation [1–3].
47 Moreover, these commodity cropping lands, amongst others, provide different ecological services in
48 terms of carbon sequestration, habitat provision, conservation of biodiversity [4,5]. Thus, a reliable
49 and recurrent mapping of such cropping area is crucial for customizing forest landscape management
50 to the respective land use expansion.

51 Agroforestry has been suggested as an agricultural option for sustainable cocoa production;
52 Cocoa Agroforestry refers to the system of growing cocoa tree crop in the understorey of multi-strata
53 canopy trees [6], which comprise a diversity of timber, fruits, and NTFP (Non-Timber Forest Product)
54 producing tree species [5,7,8]. Cocoa is a tree crop of high economic importance, and particularly in
55 tropical sub-Saharan Africa [2,9] that contributes about 70% of global cocoa dry beans export [10].
56 Regrettably, the expansion of cocoa production lands contributes significantly to the loss of forest
57 cover [11,12]. Such expansions are somewhat specific to countries and production landscapes [13–
58 16]; therefore, some are more destructive to forests than others. On a global scale, cocoa production
59 was responsible for 57% of the global agricultural land expansion rate of 132,000ha y⁻¹ in the period
60 2000–2013 [17]. However, such figures need validation, at national levels, through mapping of actual
61 cropping lands.

62 From an ecological standpoint, compared to the intensive mono-stratum cocoa plantations and
63 other high canopy commodity crops such as oil palm, rubber, etc., cocoa agroforests sustain
64 ecosystem services at a scale that is considered second to transition forests [18–20]. Regarding
65 management in most cocoa-producing nations, available literature does rarely address spatial
66 mapping of cocoa production area. Management projections on production area are based on the
67 FAO's (Food and Agricultural Organization) database on crops, FAOSTAT. This database depends
68 on sporadic Country reports of annual harvest area, and projection from these reports, which may
69 not represent the actual ground reality [21] In the case of Cameroon, cocoa is predominantly grown
70 in small-scale agroforests of 1-3ha [5,7,8]. Thus, the cocoa area of 123,120 ha in Cameroon's Centre
71 Region production hotspot for example [22], is in effect the harvested area, which is based on seasonal
72 records by local farmers' and Cocoa cooperatives. Thence, the national statistics is monitored and
73 published by the National Cocoa and Coffee Board (NCCB). Depending on the variety and
74 propagation technique, established cocoa farms require about 3 or more years of crop tending, before
75 the harvest of first produce [23]. Consequently, the FAO records of the harvested area may be, at the
76 minimum, 3 years short of possible expansions in cocoa farms. On the assumption of continuous
77 expansion of cocoa production land, therefore, the associated impact on forest cover is far greater
78 than management extrapolations made solely from published data on harvested areas. For a
79 sustainable management of cocoa production landscapes, national government programs that
80 stimulate export of dry cocoa beans [2,9,22] need support for reliable and updated estimations of both
81 harvest and expansion areas for cocoa agroforests.

82 The application of earth observation data provides large-scale mapping of commodity cropping
83 area. Unfortunately, the discrimination of cocoa agroforest areas with multi-strata canopy, using
84 optical reflectance and vegetation indices is yet not successful [17,24]. Cocoa agroforests have similar
85 canopy structure as transition forests [16]. In moist tropical zones, a high frequency of clouds and
86 atmospheric aerosols hampers application of optical satellite data. SAR images on the contrary,
87 provide cloud and season independent information about land surface features. Based on texture
88 information extraction, the analysis of SAR images has been used for discrimination of cropland
89 [25,26] and forest biomass estimation [27]. Unlike optical imagery that captures reflectance of trees
90 and forest canopies, SAR data capture the water content (a dielectric property) and structure
91 (geometric property) of target features. The later information is only provided if the targets' size is
92 lower or close to the wavelength of the SAR sensor. Thus, use of SAR imagery, e.g. for vegetation
93 mapping, is predetermined by sensor wavelength, and necessitates image processing procedures that
94 vary with vegetation type and scale of assessment.

95 Long wavelength SAR such as L-band ($\lambda \approx 25$ cm) provide details on volume scattering from
96 branches and stems, which are essential for aboveground biomass estimation [28]. A SAR-based
97 index, Radar Vegetation Index (RVI), was developed for biomass monitoring using L-band data.
98 However, the application of such an index requires removal of the contribution from soil surface
99 backscatter [29]. Although such bands may improve discrimination of vegetation with high tree
100 canopy, they are less reliable for mapping features of low or no vegetation. In the latter scenario,
101 therefore, other SAR wavelengths may be equally reliable.

102 The C-band sensor is the wavelength ($\lambda \approx 5$ cm) with the largest temporal series in SAR remote
103 sensing. Its utility has been less explored for mapping tropical land cover, and especially so in
104 commodity croplands in the predominantly heterogeneous farming conditions in sub-Saharan
105 Africa. The C-Band SAR penetrates the vegetation canopy only to a limited extent. However, as in
106 the case of settlement and grassland land cover classes, forests have a high temporal stability of SAR
107 backscatter signals. To Thiel et al. [30], the contrast between these land cover classes and agricultural
108 land is high in the cross-polarized (HV or VH) SAR image bands. Besides, Stimulus et al. [31] reported
109 that texture measures are needed to discriminate settlement areas from forests. Thus, considering the
110 seasonal changes in structure and water content of vegetation canopy elements, a temporal metric of
111 texture from C-band SAR images may be able to discriminate perennial agroforestry land cover.

112 Texture measures from Grey Level Co-occurrence Matrix (GLCM) provide reliable information
113 on the spatial relationship of images pixels [32]. The GLCM provides a joint probability distribution
114 or co-occurrence frequency of grey levels (or intensity tones) in an image based on three parameters:
115 the pixel(s) distance, angular displacement and image sub-region - analysis window size. Several
116 second order ,between two pixels, statistics from the GLCM were proposed [33] to describe the
117 texture in an image. The use of GLCM texture measures depends on the geometry of target features
118 and their characteristic spatial structure in the landscape[32] . For land cover classification in a
119 heterogeneous landscape, Mishra et al. [34] observed that texture information was more valuable to
120 improve classification accuracy in SAR image than for optical image. Yet, the authors [34] suggested
121 that an optimum combination of texture features is needed for the specific type of landscape
122 heterogeneity. Land cover classification using GLCM texture extraction have focused on scale or
123 window size [32,34–37]. However, the importance of grey level quantization in GLCM texture
124 analysis has been emphasized [38–40]. Moreover, for texture feature extraction, use of image grey
125 levels beyond the depth of pixels (range of values) may increase uncertainty in results [41]. Thus,
126 application of GLCM matrix for land use and land cover (LULC) classification [30,32,37] do not take
127 grey level quantization in account, which may be vital in mapping heterogeneous agricultural
128 landscapes - in particular, the inherently heterogeneous LULC categories.

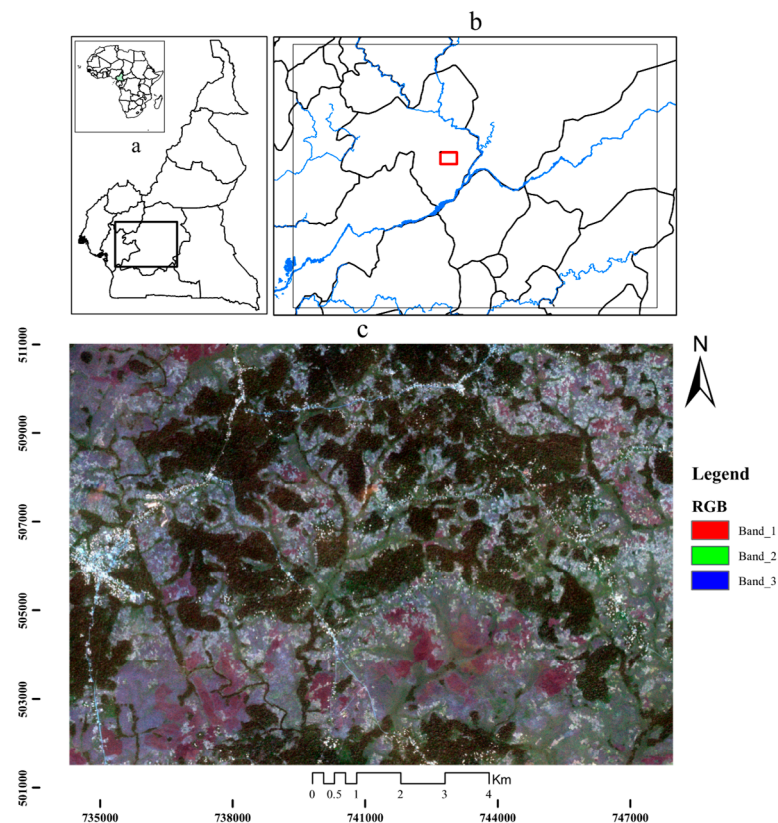
129 The purpose of this study was to assess the temporal contribution of SAR volume scattering,
130 essentially by vegetation canopy, in discriminating perennial cocoa agroforests land use. We explored
131 multi-seasonal multi-polarization Sentinel-1 C-band SAR images for the following objectives: (1)
132 Evaluate the performance of GLCM texture-based discrimination of cocoa agroforest land use from
133 transition forests cover in comparison to typical classification from multi-spectral optical image –
134 using a RapidEye image; (2) Assess the contribution of grey level quantization in improving texture-
135 based classification performance. We compared four different grey level quantization or dynamic
136 pixel range: 32bits, 8bits, 6bits and 4bits; (3) Assess the information gain from Shannon Entropy (H)
137 or uncertainty as a classifier performance estimator.

138 We applied the GLCM to estimate four selected texture statistics based on [32]: contrast, entropy,
139 variance, and correlation, which provided texture information on structure and water content of
140 vegetation – volume scattering. Then, conducted an averaging of SAR volume scattering across
141 seasons by using a machine (random forest) learning classification algorithm; we include other land
142 cover classes in classification analysis to derive a thematic land cover map of the heterogeneous
143 landscape

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147 2. Study site

148 This study was conducted in the landscape of Bakoa (32N 734280m E 510975m N and 747435m
 149 N 501480m E, 123.28 km²), which is located in the Bokito District of the Mbam and Inoubou
 150 Department, in the Centre Region of Cameroon (Figure 1). This area is classified as a savannah-forest
 151 transition zone. The topography features a rolling terrain and the altitude ranges between 500 – 900m
 152 a.s.l. The vegetation is a mosaic of bush-savannah, subsistence farming, and perennial cocoa
 153 agroforests. These perennial agroforests are established mainly within or along patches of transition
 154 and gallery forests. The study area is situated in the humid forest bimodal agro-ecological zone,
 155 which is characterized by two dry and wet seasons. The total annual rainfall ranges between 1300 –
 156 1500mm with a long rainy season from August to November. The main dry season lasts about 5
 157 months from November – April. The mean annual temperature is 25°C.
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161 **Figure 1.** (a) Study area located in the centre region of Cameroon; (b) Study landscape, in red
 162 footprint, within the Bam and Inoubou administrative department; (c) RapidEye false colour image
 163 (RGB: Blue, Green, and Red spectral bands) revealing a mosaic of forest and savannah vegetation in
 164 the landscape.

165 3. Materials and methods

166 3.1 Satellite data: optical and radar imagery

167 We acquired a multispectral optical image of 5m spatial resolution from RapidEye, which was
 168 recorded in the dry season of 2015. The image comprised five spectral bands in the Blue (400-510 nm),
 169 Green (520-590 nm), Red (630-685 nm), RedEdge (690-730 nm), and Near Infrared (760-850 nm) range
 170 of the electromagnetic spectrum. Four different image tiles, acquired on the same date, were needed
 171 to cover the study landscape.

172 We accessed Sentinel-1A C-band ($\lambda=5.5$ cm) SAR images for the study area from the Sentinel
 173 Scientific Data Hub of the European Space Agency (ESA). The SAR data were acquired in dual (VV
 174 and VH) polarization, the Interferometric Wide swath (IW) observation mode, and pre-processed to

175 Level-1 or ground range detection level (GRDH) of 10m spatial resolution. We selected a temporal
 176 series of 50 images acquired between March 2015 and April 2017 that covered both the dry and wet
 177 seasons. Using the image processing tools of Sentinel Application Platform (SNAP) version 5.0, we
 178 prepared image subsets and pre-processed them sequentially from radar backscatter intensity values
 179 to sigma naught (σ_0) backscatter coefficients: thermal noise removal, filtering with orbit file,
 180 radiometric calibration, geometric rectification, and terrain correction. The digital elevation model
 181 (DEM) of the Shuttle radar topographic mission (SRTM) was applied, in SNAP with Sentinel-1
 182 toolbox (S-1TBX), for terrain correction and geometric rectification of SAR images. We used both the
 183 co- (VV) and cross-polarised (VH) bands of all images. We then projected the pre-processed 10m
 184 resolution images in WGS 1984 UTM Zone 32 N. A list of the remote sensing data is presented in
 185 Table 1.

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Table 1. Used remote sensing data: mono-date RapidEye (5m) and multi-date Sentinel-1 SAR (10m) data.

Satellite mission	Scene ID(s)	Acquisition date (DD/MM/YYYY)	Sensing stop time (HH:MM:SS UTC)	Acquisition mode (Polarization)	Data level
RapidEye: RE-3	3241224_				
	3241225_	09-JAN-2015	10:35:41.00	MSI , Optical	L3A
	3241124_				
	3241124_				
Sentinel-1A	_006256_008304_78DE	06-JUN-2015	17:28:11.147769		
	_007306_00A05D_2111	17-AUG-2015	17:28:14.323283		
	_007831_00AE86_4926	22-SEP-2015	17:28:15.577539		
	_008706_00C641_B612	21-NOV-2015	17:28:15.454239		
	_010456_00F838_64CF	20-MAR-2016	17:28:13.302867	IW Ascending	Level1
	_011156_010D64_7E35	07-MAY-2016	17:28:15.219784	(Dual: VV,VH)	GRD
	_012031_012962_8F08	06-JUL-2016	17:28:18.678294		
	_012906_01465C_878C	04-SEP-2016	17:28:21.557129		
_014831_0182BC_16C4	14-JAN-2017	17:28:19.166952			
_015706_019D94_BC50	15-MAR-2017	17:28:18.681981			

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190 3.2 Field campaigns

191 During the field campaigns conducted in 2015, 2016 and 2017, we collected ground information
 192 on land use and cover. The field data comprised ground GPS information and inventory of
 193 representative areas that characterize the different land cover and uses in the landscape (see Figure 2
 194 and Table 2).

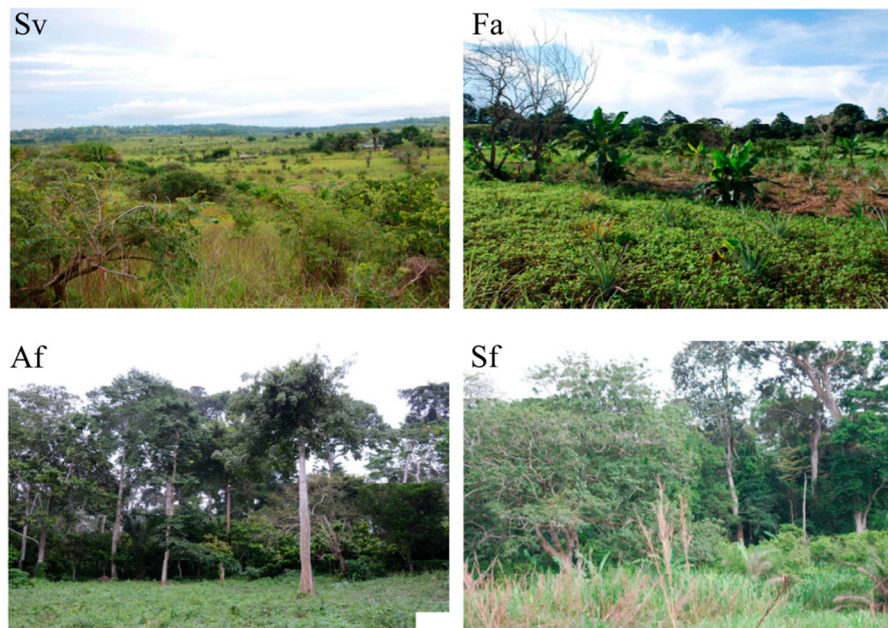
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Table 2. Description of the thematic land cover types used for classification of land cover (Figure 2).

Class acronym	Class Name	Description
Bu	Built up	Residents, commercial/markets, industrial, administrative settings
Es	Earth road/bare soil	Land areas of exposed soil, bare rocks
Sv	Shrub/grassland Savannah	<i>Imperata sp.</i> savannah land: Shrubby and grassland areas which have not been converted to farm
W	Water	River, ponds, seasonal and permanent swamps

Af	Perennial cocoa agroforests	Land areas used for cocoa production with various degrees of canopy stratification: canopy/shade trees are mainly deciduous
Fa	Subsistence farming	Savannah and forest land areas that have been converted essentially for permanent or seasonal subsistence crop production; including farm fallows
Sf	Transition/Secondary forest	Disturbed and gallery forest patches, secret/cultural forest, hunting forest: have a more permanent and less stratified canopy structure

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Figure 2. The range of vegetation land cover differ mainly in the density of woody biomass, which changes with season or phenological period. Class acronyms are described in Table 2.

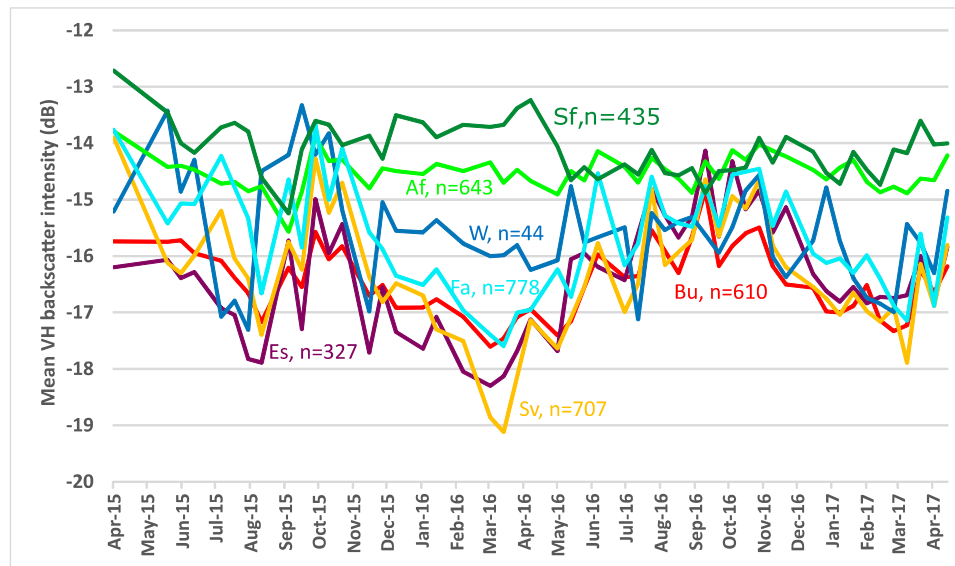
206 3.3 Image processing workflow

207 For RapidEye images, the following pre-processing protocol was conducted: atmospheric
208 rectification by dark object subtraction (DOS), radiometric calibration to reflectance values, geometric
209 correction, and finally computing different vegetation indices from a mosaicked image. After
210 subsetting the SAR images, we used the batch processing mode of SNAP for the following pre-
211 processing steps: radiometric calibration to Sigma0 (decibels), and geocoding with SRTM 3sec DEM
212 using RangeDoppler Terrain Correction. We used intensity backscatter profile (Figure 3) and the
213 Random Forest (RF) important variables criterion to remove noisy SAR images, and selected a
214 subsample of 10 (of the 50) important images that represent six wet and four dry seasons between
215 2015 and 2017 (see summary description in Table 1 and Figure 4).

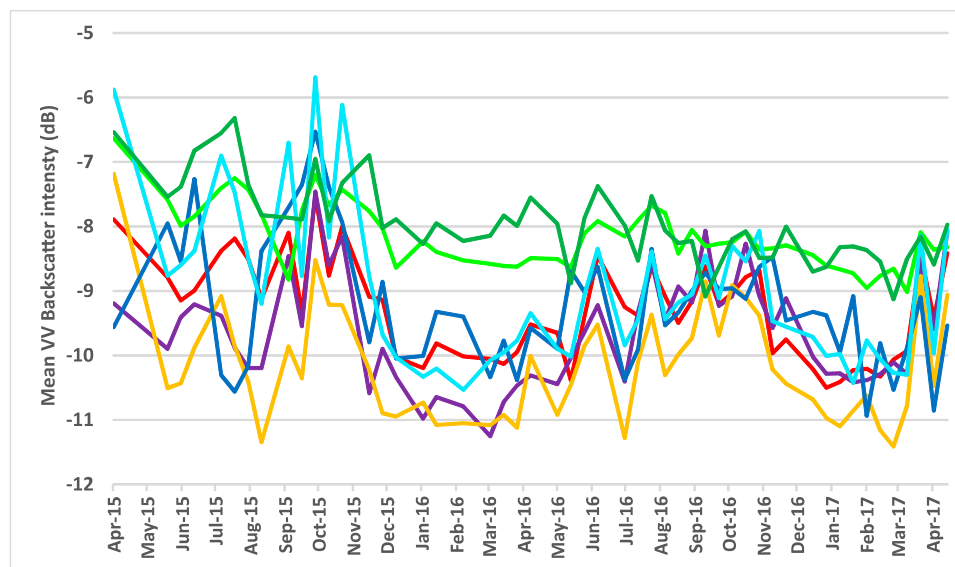
216 The image processing steps, detailed in following subsections, comprised of three major
217 categories: (a) Feature extraction: this consisted of computing images of vegetation indices and
218 GLCM texture images; (b) Image classification: we co-registered the vegetation index and texture
219 images into the separate stacks or models described in Table 3. Then we ran eight RF ensemble
220 (machine learning) classification algorithms, using the image stacks as input (Table 3), and (c) Post-
221 processing: estimation of uncertainties in classified maps, in addition to accuracy metrics, as the basis
222 for validating the classifier models. Finally, for the texture-based model with the highest overall

223 accuracy, we evaluated GLCM texture images at four different grey level quantization, to improve
 224 classification uncertainties.

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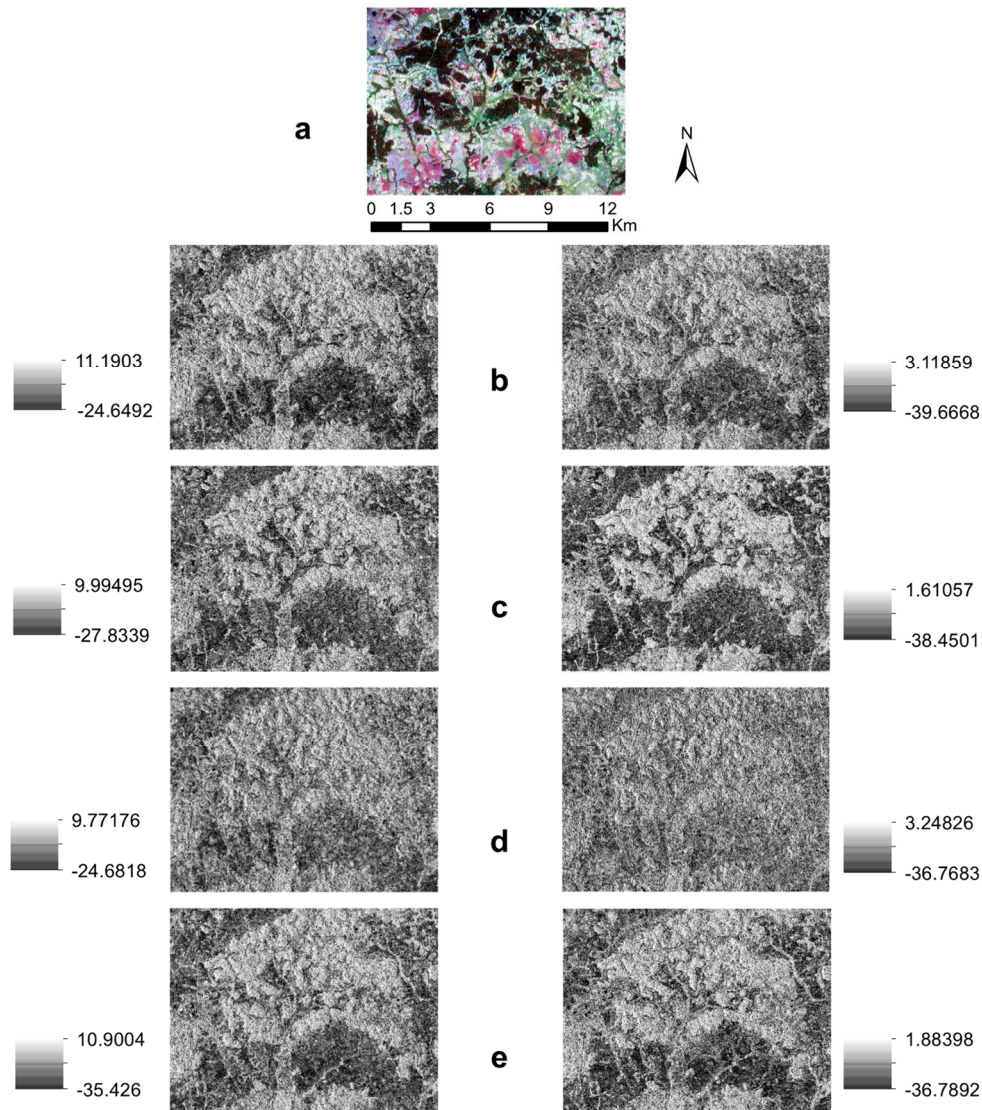
(a)



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Figure 3. Radar backscatter intensity temporal profiles for the different land use/cover types using the SAR images. (a) Vertical transmitted, horizontal receive backscatter; (b) Vertical transmitted, vertical received backscatter. The number, n , for each label pertains to the amount of the sample pixels ($10 \times 10 \text{m}^2$)



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Figure 4. Radar backscatter intensity (dB) images of study landscape for key seasons from 2015 to 2017. (a) RapidEye false colour (RGB: bands 1, 2, and 3) composite image; (b) start of wet season: 17 August, 2015; (c) peak dry season: 20 March, 2016; (d) peak wet season: 4 September, 2016; (e) mid of dry season: 14 January, 2017. The left and right columns are, respectively, the VV and VH backscatter for each image. The North and scale bar is applicable to all images.

243 3.3.1. Feature extraction: Vegetation Indices (VIs) and GLCM texture features

244 The monitoring of vegetation status and extent is often based on normalisation ratios of spectral
245 bands, in the Visible and Near Infrared (NIR) spectrum [42] in spaceborne imagery. These ratios are
246 based on the contrasting spectral response of vegetation to the Red and NIR wavelengths.

247 The application of indices such as NDVI (Normalised Difference Vegetation Index) for
248 vegetation monitoring have faced several challenges [43], notably issue of saturation for biomass
249 above certain thresholds, and which is common in moist tropical vegetation. And, although
250 saturation may not be an issue over agricultural landscapes, reflectance from soil background often
251 perturb discrimination of sparse vegetation or cropland from bare soil [44]. In this study, we used
252 VIs whose values indicate the status and abundance of vegetation and biomass, and that minimise
253 the effect of soil background on vegetation reflectance values [45]: NDVI, gNDVI (green NDVI), EVI2
254 (Enhanced Vegetation Index), SAVI (Soil Adjusted Vegetation Index), and MSAVI (Modified SAVI)
255 [46–50]. However, to provide additional information on vegetation characteristics and vitality, recent

256 optical sensors include an additional spectral band - the red edge band featured in RapidEye and
 257 Sentinel2 [51]. This band is located between red absorption (by chlorophylls) zone, and the NIR
 258 waveband. Since radar backscatter signals from a ground resolution cell are pseudorandom, the
 259 interaction of microwaves with terrain objects may be difficult to predict. Moreover, SAR images
 260 have speckle effect because the response signal of a resolution cell is a coherent interference from
 261 multiple scattering elements within the cell. Based on texture information extraction, the analysis of
 262 SAR images has been used for discrimination of cropland [26] and forest biomass estimation [27].
 263 Often, the GLCM statistical approach is used in analysing SAR textures. The GLCM is a sparse matrix
 264 that stores co-occurrence probabilities of inter-pixel grey levels in an image [33]. These probabilities
 265 provide a second-order measure for texture features in an image: they represent conditional joint
 266 probabilities of all pairwise combination of grey levels (G) in the spatial window of analysis, and
 267 depend on both the spatial orientation (θ) and displacement distance (δ). Computation of GLCM is
 268 faster for images with fewer grey levels, because the matrix is dimensioned to G. The conditional
 269 probabilities are estimated as follows:

$$270 \Pr(x) = \{C_{ij} | (\theta, \delta)\} \quad (1)$$

271 Where, C_{ij} = co-occurrence probability between grey level i and j ; and is defined by

$$272 \quad \quad \quad C_{ij} = P_{ij} / \sum_{i,j=1}^G P_{ij} \quad (2)$$

273 Where, P_{ij} = number of occurrence of grey levels i and j within the given window, for a certain
 274 (θ, δ) pair; G = the quantized number of grey levels. The denominator sums up to the total number
 275 of grey level pairs (i, j) within the analysis window.

276 Although different second-order statistics are commonly used to classify single images [52],
 277 some GLCM texture measures are auto-correlated [33]: a selection of a few texture measures may be
 278 reliable in achieving specific image analysis objective(s) [32]. We assess the accuracy of SAR images,
 279 covering several seasons, in discriminating perennial agroforestry land cover, using four less
 280 correlated GLCM texture measures: *Contrast*, *Entropy*, *Correlation*, and *Variance*. We estimated the
 281 textures measures from GLCM using a 5×5 moving window, an aggregate orientation of four
 282 directions ($0^\circ, 45^\circ, 90^\circ$, and 135°), and one-pixel displacement (inter-pixel distance).

$$283 \quad \quad \quad \text{Contrast} = \sum_{i=0}^{G-1} \sum_{j=0}^{G-1} P_{i,j} (i - j)^2 \quad (4)$$

$$284 \quad \quad \quad \text{Entropy} = \sum_{i=0}^{G-1} \sum_{j=0}^{G-1} P_{i,j} \log_2 P_{i,j} \quad (5)$$

$$285 \quad \quad \quad \text{Variance} = \sigma_i^2 = \sum_{i,j=0}^{G-1} P_{i,j} (i - \mu_i)^2 \quad (6)$$

$$286 \quad \quad \quad \text{Correlation} = \sum_{i,j=0}^{G-1} P_{i,j} \left[\frac{(i - \mu_i)(j - \mu_j)}{\sigma_i \sigma_j} \right] \quad (7)$$

287 where $P_{i,j}$ is the joint probability distribution of the grey levels i and j at two ends of a displacement
 288 vector in the assessment window, and G is the number of rows or columns. Since we considered a
 289 symmetrical GLCM, $\mu_i \equiv \mu_j$ and $\sigma_i^2 \equiv \sigma_j^2$. And for Entropy, $0 \times \ln(0) = 0$, since $\ln(0)$ is undefined.

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291 3.3.2. Classification: Random Forest ensemble algorithm

292 In this study, we used a non-parametric machine learning algorithm, the Random Forest (RF)
 293 ensemble as an image classifier. This algorithm builds multiple decision trees for the same dataset
 294 based on random bootstrapping of sample training data [53]. The random forest classifier is less
 295 influenced by the common issue of over-fitting and is able to handle a large number of variables [54].
 296 Firstly, each tree is built from a random subset (n) of two thirds of the original samples (N) – the ‘in-
 297 bag’ data; and secondly, from a subset (m) randomly selected from the total (M) variables in the
 298

306 dataset – *mtry*, in each decision tree nodes are split using a best split variable – the one that yields the
 307 highest decrease in impurity [54]. The algorithm is a soft classifier on the basis of the probability
 308 voting of pixels belonging into the respective classes considered (Table 2). Compared to other non-
 309 parametric classification algorithms, it is less constrained by the need of extensive training and test
 310 data samples; this is due to an integrated out-of-bag (OOB) error estimation and accuracy test
 311 following a bootstrap sub-sampling on input data. Several sources provide additional details on the
 312 random forest algorithm [54,55].

313 We ran eight RF models for the different images stacks as classifier input (Table 3). For each
 314 model, we evaluated the OOB error curve and *mtry* to prune decision trees to an optimal number.
 315 For a spatially explicit and unbiased representation of each land cover class in the RF models, we
 316 divided the extracted pixel information for each class into a stratified random sample of 70% and 30%
 317 pixels respectively for training and testing the models. Image classification was conducted using the
 318 random forest package [56] of R programming software 3.4.3.

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 320 **Table 3.** The respective image stacks, used to compare the Random Forest (RF) classification
 321 accuracy.

Data Categories	Model	Image stack
RE: Dry season Multi-spectral RapidEye Image	RE1	TOA Reflectance of B, G, R, Red Edge, and NIR: 5 Bands
	RE2	TOA Reflectance and Vegetation Indices (VIs): 10 Bands
GL: Multi-date and season SAR GLCM Textures	GL1	Multi-date VV GLCM Textures: 40 Bands
	GL2	Multi-date VH GLCM Textures: 40 Bands
	GL3	Multi-date VV and VH GLCM Textures: 80 Bands
GLI: Multi-date and season SAR intensity and GLCM Textures	GLI1	Multi-date SAR VV Sigma0 intensity and VV GLCM Textures: 50 bands
	GLI2	Multi-date SAR VH Sigma0 intensity and VH GLCM Textures: 50 bands
	GLI3	Multi-date SAR VV plus VH Sigma0 intensity and, VV plus VH GLCM Textures: 100 bands

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 323 Several studies have used RF for classification of forest cover [57] and cropland [26]. Although
 324 data over-fitting and poor prediction with RF approach has been reported [58] contrarily, it has been
 325 shown to provide relatively better classification of croplands [26] and mangrove vegetation [52]. Thus
 326 the performance of the RF ensemble classification algorithm may vary for different landscapes and
 327 cropping systems. For example, Loosvelt et al. [59], observed a high classification uncertainties for
 328 mixed pixels, at the heterogeneous boundaries of internally homogeneous cropping fields. Likewise,
 329 Van Tricht et al. [60], reported low classification confidence at such crop boundaries. Mixed
 330 cropping systems are common and the norm in moist tropical landscapes. However, reported
 331 research on the processing and use of SAR images for mapping of tropical heterogeneous cropping
 332 land, such as perennial agroforestry, is scarce.
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334 3.3.3. Post-processing and classification uncertainty assessment

335 In remote sensing mapping, the validity and reliability of classified maps are often decided on
 336 basis of estimated overall accuracy and kappa coefficient [26]. Values such as user's and producer's
 337 accuracy are prone to errors and uncertainties [61]. As a soft classifier, however, the RF algorithm
 338 provides the possibility for assessing data- and computation-related uncertainties [59]. In our
 339 analysis, we used user's accuracy (omission error), producer's accuracy (commission error), overall
 340 accuracy and Kappa statistics – which compares results of a chance classification versus our RF model

341 accuracy. However, the pixel-based classification methods are prone to uncertainties coming from
 342 either the use of unreliable data [61]. Thus, RF algorithm, as a soft classifier, provides a vector (P_u) of
 343 classification probability or votes for each image pixel - $P_u = P_1, P_2, P_3, \dots, P_n$ for a classification with
 344 n categories, and P_i denotes the probability of belonging to class i (Table 2).

345 In this study, in addition to model OOB error estimation, we evaluated classification
 346 uncertainties of RF models using the maximum classifier probability (U), and a weighted uncertainty
 347 measure entropy: the Shannon entropy (H) (Shannon, 1948; Vajapeyam, 2014). These uncertainties
 348 were calculated as:

$$349 \quad U = 1 - P_{max} \quad (8)$$

$$350 \quad H = - \sum_{i=1}^N P_i \cdot \log P_i \quad (9)$$

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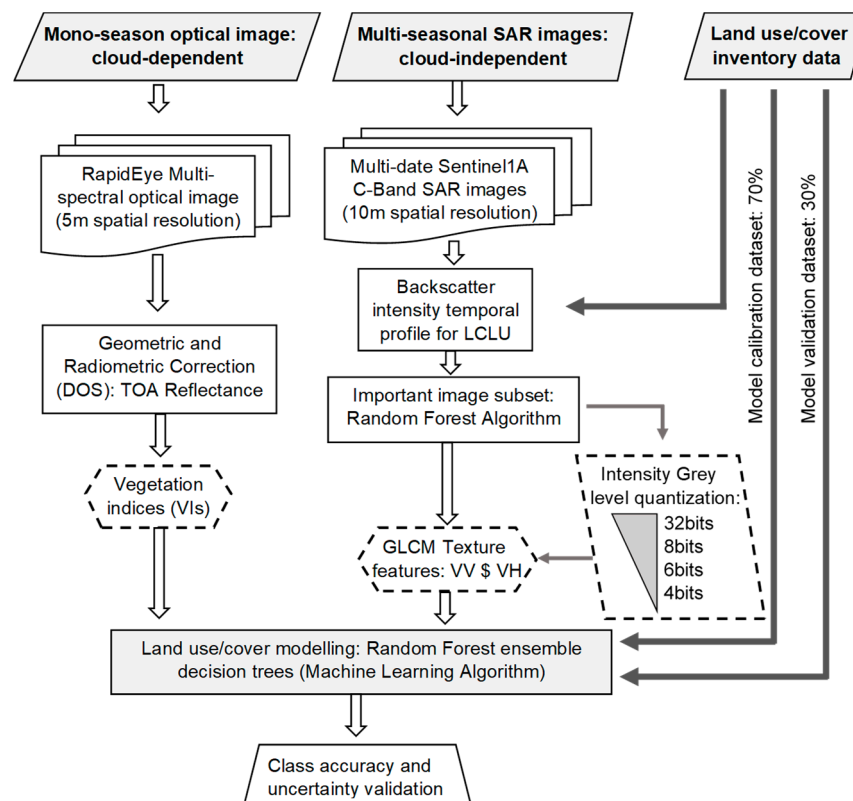
352 Where, P_i = probability of belonging to class i and P_{max} = maximum probability vote for a pixel's
 353 class. N = the total number of classes considered for analysis.

354 The maximum probability class assignment, by the soft classifier, for a pixel does not always
 355 result in assigning the true class label to the concerned pixel. Thus, by considering the entire range of
 356 values in a pixel's probability vector, H , compared to U that only makes use of P_{max} , provides a more
 357 robust measure of uncertainty; it has a maximum value at highest entropy – equal probability votes
 358 for all classes considered.

359 Loosvelt et al. [59], showed that H is reliable for evaluating uncertainties in mapping cropland
 360 from SAR images. However, our study area is characterised by heterogeneous cropping systems and
 361 is located in a tropical landscape (Figure 2). For the best-performing RF models, based on kappa
 362 accuracy, we computed and analysed U and H uncertainties for the classified maps, and the land
 363 cover classes considered in the study area (Table 2). The uncertainty estimations and analysis were
 364 conducted in Spyder IDE (Integrated Development Environment) of Anaconda distribution for
 365 Python programming software version 3.0 (Anaconda 3)

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370 **Figure 5.** Schematic outline of the image processing and texture feature extraction, and land
 371 use/cover (LCLU) delineation by a machine learning algorithm. Input and outputs are shaded, and
 372 feature extraction in broken border lines. DOS: Dark object subtraction, TOA: Top of Atmosphere
 373

374 4. Results

375 All the RF models had classification accuracies above 70%. Classification error and the
 376 sensitivity in discriminating land cover classes was different for each model. The models with the
 377 highest classification reliability, in increasing order of importance, were RE1, GLI3, and GL3.

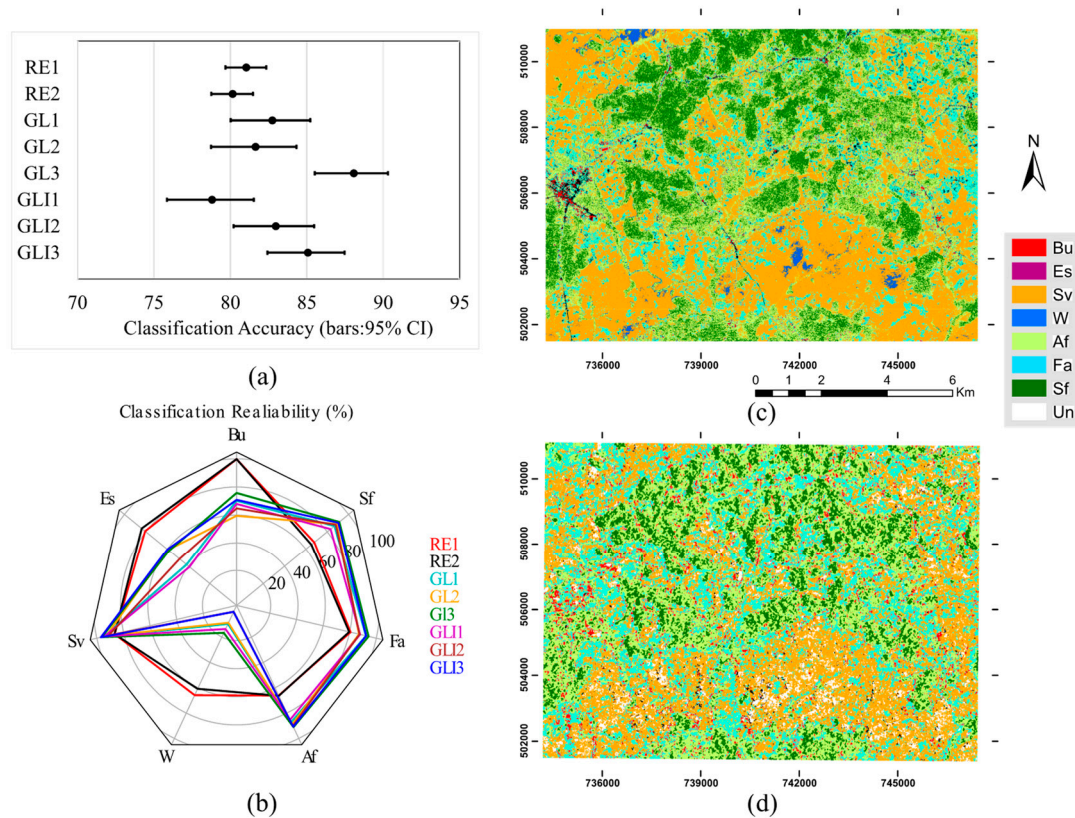
378 4.1 Classification accuracy

379 Table 4 summarizes the classification results for all eight RF models. All models had a reliable
 380 overall accuracy (OA) above 70%. However, compared to using VV or VH bands separately, the use
 381 of both co- and cross-polarization bands (GL3) resulted in the highest classification accuracy. The
 382 GL3 model had the highest overall accuracy of 88.1% and kappa of 0.85; and, compared to other
 383 models, the OOB error estimate was the least with 12.8%. Also, classification from the multi-spectral
 384 optical image (RE1 model) had a reliable overall accuracy of 81.1%, but it had a lower kappa 76.9%.
 385 Compared to the GL3 model, the OOB error difference of +7% was observed for the RE1. The GLCM
 386 textures are reliable for discriminating the land cover/uses. And considering the heterogeneous and
 387 dynamic vegetation in the landscape, an improvement feature selection using the GLCM approach is
 388 necessary to reduce class uncertainties.
 389

390 **Table 4.** Classification accuracies of different feature models based on the Random Forest (RF)
 391 classifier algorithm.
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Model	Overall Accuracy-OA % (95% CI)	Kappa	OOB Error %
RE1	81.04 (79.68, 82.35)	0.769	19.18
RE2	80.15 (78.76, 81.48)	0.757	19.46
GL1	82.74 (80.02, 85.23)	0.787	17.12
GL2	81.65 (78.74, 84.32)	0.773	18.47
GL3	88.07 (85.52, 90.31)	0.853	12.85
GLI1	78.80 (75.85, 81.53)	0.738	19.66
GLI2	82.97 (80.21, 85.48)	0.789	18.71
GLI3	85.07 (82.42, 87.47)	0.817	13.69

393
 394 The thematic land cover map from RE1 and GL3 models are shown in Figure 6. Separately, both
 395 VV and VH GLCM derived texture measures were poor in the prediction of non-vegetated land
 396 covers, and more so when both bands were included in the same model (Figure 6b). When included
 397 as input layers, the SAR backscatter intensity did not improve classification accuracy. Likewise, the
 398 inclusion of vegetation indices from the multispectral optical image, taken during a dry season, did
 399 not improve classification accuracy (Figure 6a). The texture measures from both VH and VH
 400 backscatter provide comparable, and may be complementary, LULC mapping accuracy to the
 401 commonly used vegetation indices of optical image.
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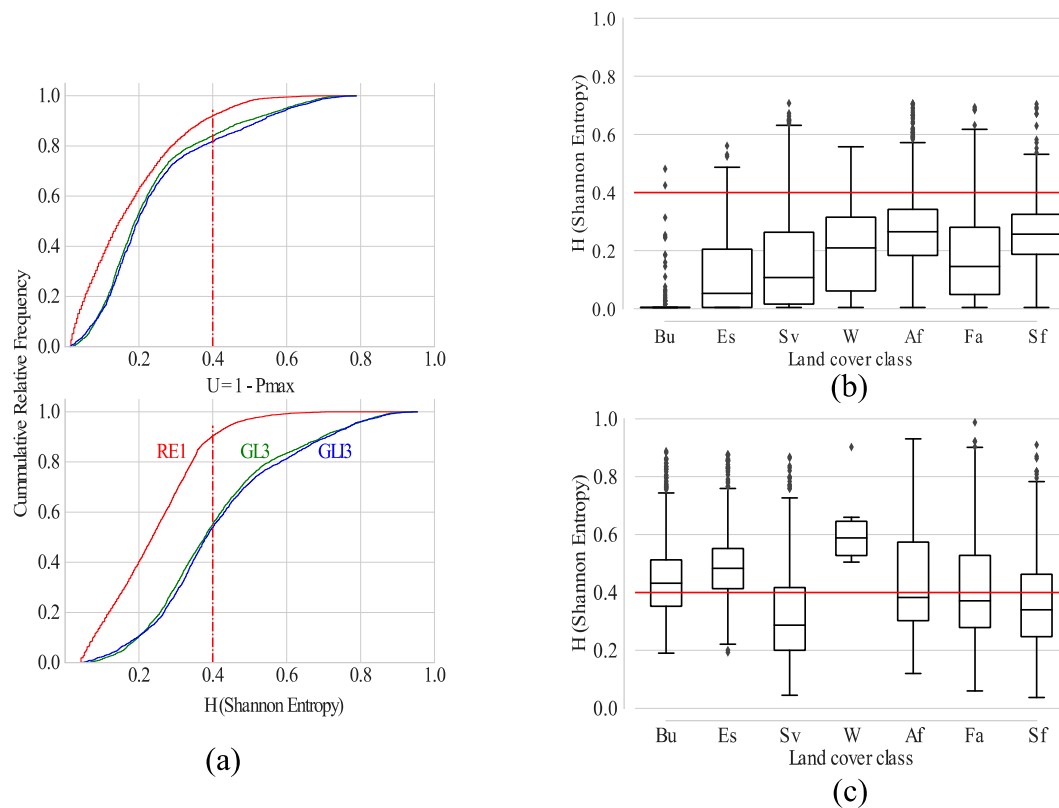
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Figure 6. Pixel-based classification result for the eight models evaluated by random forest ensemble algorithm. (a) Classification accuracy; (b) Class reliability estimates; (c) Thematic land cover/use map from model RE1; (d) Thematic map model GL3. The scale, legend and north arrow apply to both c and d.

410 Visually, RE1 map shows a relatively intact and continuous expanse of transition forest patches
411 (Figure 6c). Contrarily, the classified map from GL3 revealed that transition forest cover is highly
412 fragmented by cocoa agroforests into smaller patches (Figure 6d). Also, from classification
413 reliability estimates (100 - commission error) in Figure 6b, the RE1 model was more reliable in
414 delineating non-vegetation land features. SAR-based texture images had a high reliability in
415 delineating vegetation landscape features (Sv, Af, Fa, and Sf). Thus, although multi-spectral optical
416 image had a better classification prediction of land cover classes in general, it was less reliable in
417 discriminating perennial agroforests from transition forest land cover.

418 4.2 Uncertainty in discriminating vegetation land cover

419 The classification results from dry season RapidEye multi-spectral optical image (RE1) had a low
420 overall and class uncertainties. From the cumulative estimates of class probabilities in Figure 7,
421 classification uncertainty from the RE1 converges at a probability of around 0.6 for both U and H,
422 whereas the uncertainty from GL3 map converges at higher probabilities – 0.7 and 0.9 respectively
423 for U and H (Figure 7a). About 90% of pixels classified by RE1 had H uncertainties below 0.4,
424 compared to about 50% of pixels for GL3. This difference is less obvious in the cumulative plot of U.
425 Thus, uncertainty difference between RE1 and GL3 was better revealed by Shannon entropy or H
426 uncertainty.
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430 **Figure 7.** Classification uncertainties as validation of models with highest accuracy. (a) The
431 Shannon entropy (H) clearly reveals uncertainty in classification accuracy validation. For the thematic
432 maps from RE1 and GL3 R models: As example, the proportion of pixels with uncertainty below 0.4;
433 (b) Individual class uncertainty, (H), for model RE1; (c) Class uncertainty, (H), for model GL3.

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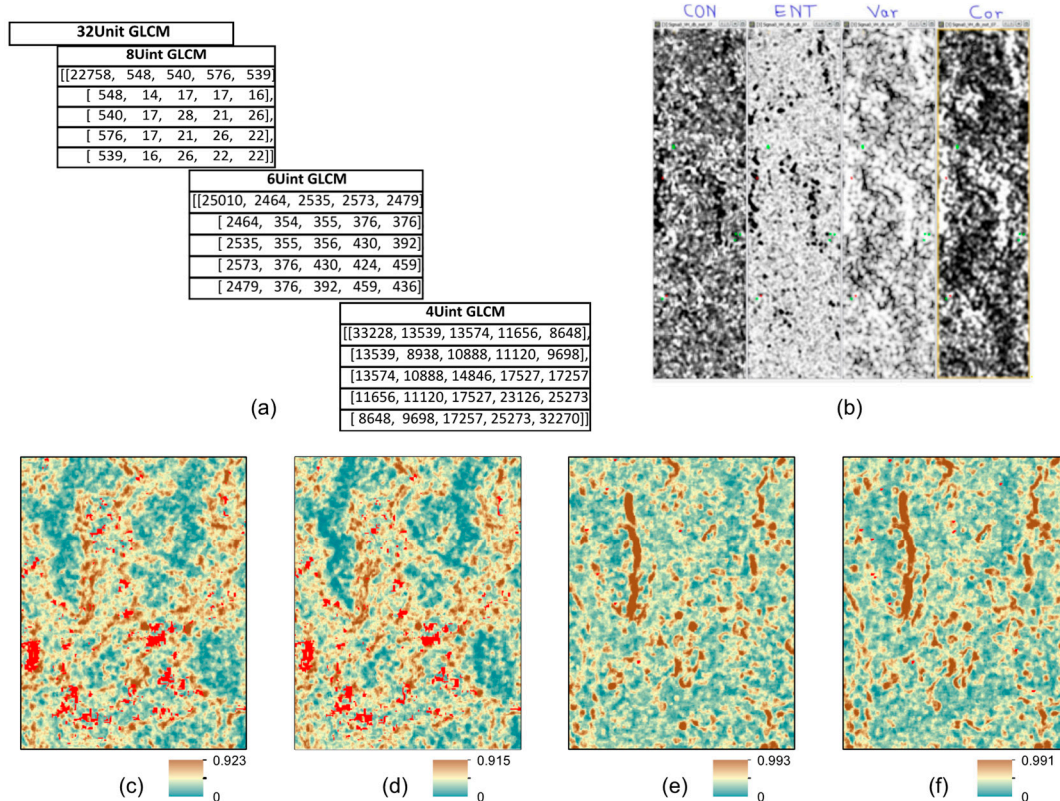
435 The individual class uncertainties are compared in Figure 7. Although the classified map from
436 the multi-spectral image (RE1 model) had a lower accuracy, the class uncertainty was, compared to
437 other land cover types, high for perennial cocoa agroforests and transition forest cover (Figure 7b).
438 In comparison to RE1, the multi-seasonal SAR image textures, from the GL3 model, had a high overall
439 uncertainty of a pixel's class prediction. However, perennial agroforests and transition forests were
440 discriminated with relatively lower individual class uncertainty (Figure 7c): The median of class
441 uncertainties were in a range between 0.2 and 0.4, which is comparable to those obtained from the
442 single date multi-spectral image (RE1). The uncertainties in land cover/use discrimination by RE1
443 may reflect vegetation status/phenology - canopy greenness; while, in the GL3 model, volume
444 scattering of radar signal reflects changes in water content and structure in different vegetation
445 canopy.

446 4.3 Contribution of pixel depth to texture feature extraction

447 The SAR backscatter intensity images have a pixel quantization of 32 bits. The likelihood of grey
448 level co-occurrences was lower at such high dynamic pixel range, as observable from unclassified
449 pixels, *Un class*, in Figure 6c. In order to improve classification uncertainty from SAR image textures,
450 we computed and compared three (3) different image pixel quantization or grey levels; 32bits – GL3
451 (as in original SAR intensity image), 8bits – GL3_B8, 6bits – GL3_B6, and 4bits – GL3_B4.

452 The contribution of grey levels on GLCM feature co-occurrence and classification probability is
453 shown in Figure 8 and Table 5. Pixel co-occurrence was low at high pixel dynamic ranges (Figure
454 8a), which resulted in a large clustering of features with low and no predictions in Figures 8a and d.
455 The dynamic pixel range of 64 grey level significantly reduced the prediction error (Figure 8e). The
456 class prediction probability is improved to a maximum of 0.99 in GL3_B6, with a difference of 7%
457 from 0.92 recorded in GL3. The prediction did not improve with further reduction of grey levels to 4

458 bits (Figure 8d). Optimizing the image pixel depth resulted in a marginal improvement of
 459 classification accuracy to a kappa value of 86.22 (Table 5). Though, the OOB error of prediction is
 460 reduced remarkably from the initial 12.8% (GL3) to 9.9% in GL3_B6. These results show that the
 461 dynamic pixel range was vital in feature selection (co-occurrence between pixels), and texture feature
 462 extraction by the ensemble algorithm.
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Figure 8. Classification probability maps reveal improved pixel classification with at lower grey level quantization. (a) Co-occurrence values in the first 5×5 cells of GLCM; (b) Sample details of the four GLCM texture measures; (c) to (f) are snips showing the respective details of the classification probability map for GL3, GL3_B8, GL3_B6, and GL3_B4 models. Unclassified areas are shown as red pixels in (c) and (d)

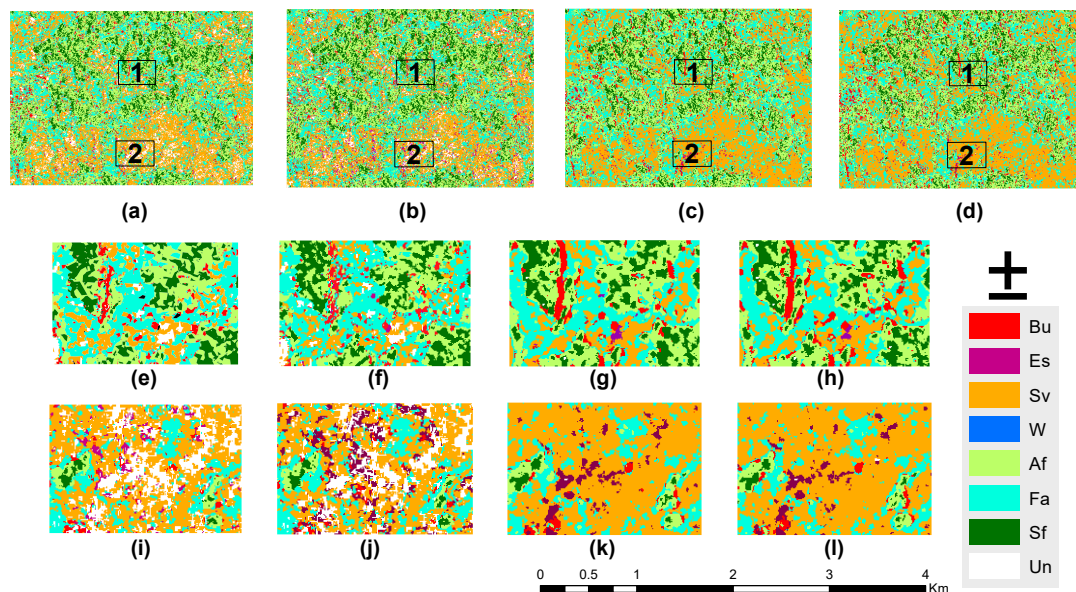
Table 5. Classification results of different grey level GLCM models and land cover/use surface area estimates. See Appendix A for details of class errors.

Model	Overall Accuracy-OA % (95% CI)	Kappa	OOB Error %
RE1	81.04 (79.68, 82.35)	0.769	19.18
GL3	88.07 (85.52, 90.31)	0.854	12.85
GL3_B8	88.23 (85.74, 90.41)	0.854	11.84
GL3_B6	88.83 (86.48, 90.90)	0.862	9.92
GL3_B4	88.86 (86.50, 90.94)	0.862	10.38

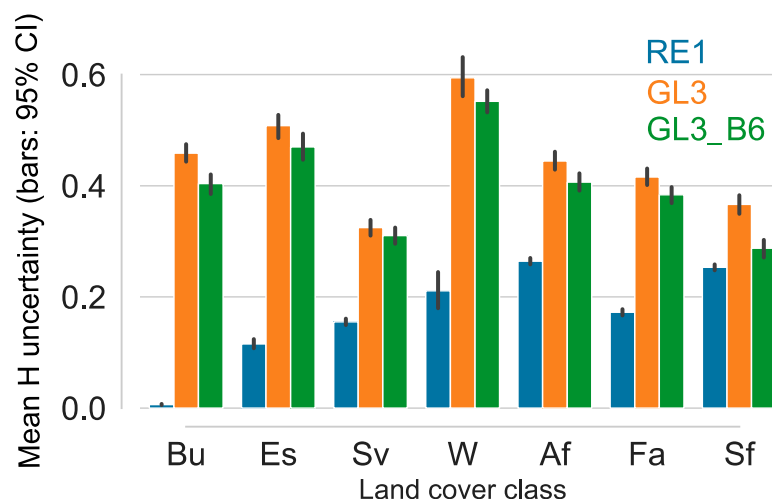
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The improvements in the predicted land cover/use maps, after optimizing the dynamic pixel ranges, are illustrated in Figures 9 and 10. Figure 9 shows the predicted maps for the different texture-based models. Compared to the classification from GL3, at 32 bits pixel depth, with an estimated 647.3 ha as unclassified area (Figure 9a), the analysis at 6bits dynamic pixel range show less mosaic in discriminating class areas and reduced the unclassified land area to about 7.4 ha (Figure 9c). See Table A3 in Appendix.

481 The model validation results, by Shannon Entropy estimates, are shown in Figure 10. The difference
 482 in class uncertainty between GL3 and GL3_B6 is most evident for the following classes; Built up (Bu),
 483 Cocoa agroforests (Af), subsistence farms (Fa), and transition forests (Sf). The GL3_B6 had a
 484 comparable lower class error for these classes: respectively 16.4%, 4.4%, 5.3%, and 6.1% (see Figure
 485 A1 in Appendix). Remarkably, the class uncertainty estimates for Sf is low and comparable for
 486 GL3_B6 and RE1. Unlike the RE1 model, with no difference in class uncertainty between Af and Sf,
 487 the significant distinction between their class uncertainties in the GL3_B6 confirms the reliability of
 488 the model to discriminate these two vegetation categories.
 489

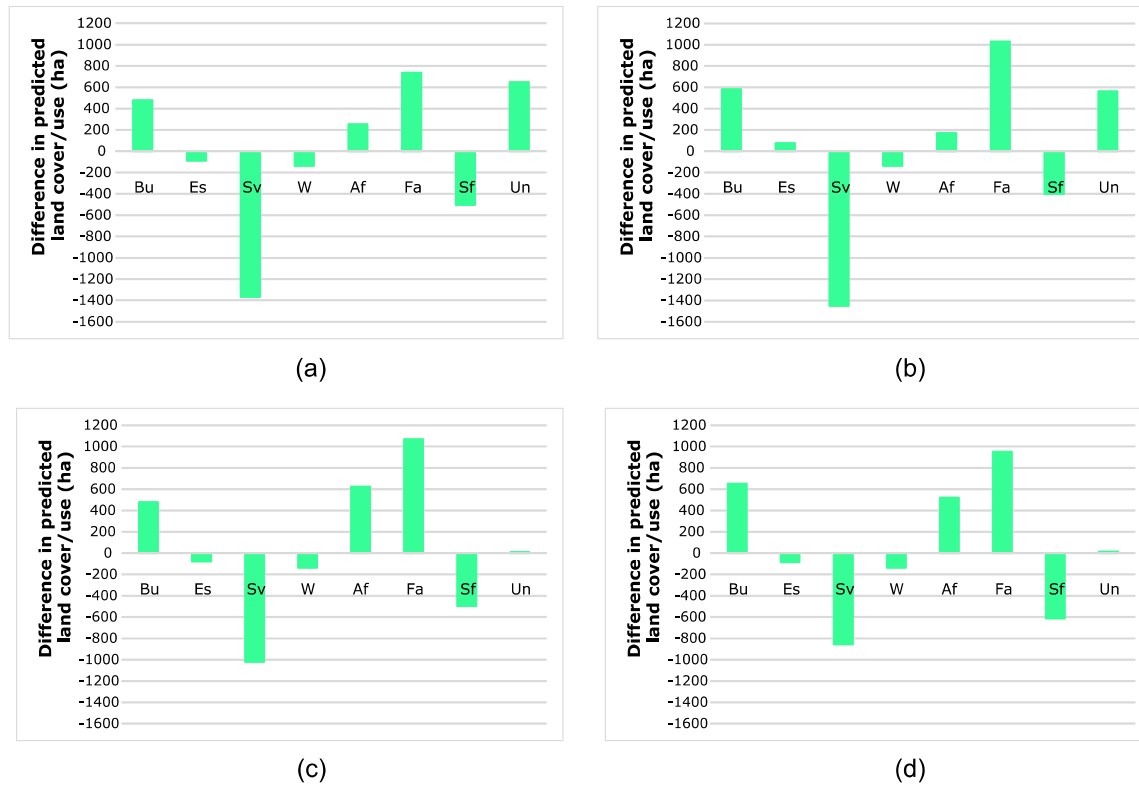


490
 491 **Figure 9.** Classification prediction, based on Random forest algorithm, for the GLCM texture-
 492 based models. (a) GL3; (b) GL3_B8; (c) GL3_B6; (d) GL3_B4; (e) to (h) are corresponding details of the
 493 prediction maps in area 1 ; (i) to (l) are corresponding details in area 2. The legend and north arrow
 494 apply to all images and the scale bar to images (e) to (l).
 495



496
 497 **Figure 10.** Comparison of validation by Shannon Entropy (H) estimation. Class uncertainty estimate
 498 is significantly improve by the GL3_B6, as observed for Built-up areas (Bu), Cocoa agroforests (Af),
 499 subsistence farmlands (Fa) and transition forests (Fs).
 500
 501

502 The estimated land cover/use against the reference model (RE1) is shown in Figure 11. The total
 503 area for each class is summarized in Table A3 (see Appendix). A significant reduction in
 504 unclassified area is shown in 11c and d. The area of transition forest is lower for all models than
 505 estimated by RE1. Though the class error and uncertainty for GL3_B6 is comparable to that of RE1
 506 (Figure 10). The land area for subsistence farming, cocoa agroforests and built up is remarkably larger
 507 in the texture-based predicted maps.
 508



509 Figure 11. Comparison of the predicted land cover/use from RE1 model (optical image), in the
 510 landscape surface area of 11344.61ha, with: (a) GL3; (b) GL3_B8; (c) GLC_B6; and (d) GL3_B4. See
 511 Table A2.
 512

513 5. Discussion

514 Land use and land cover (LULC) classification of satellite image is commonly applied to mono-
 515 date images; we use a rapidEye image that was acquired in the dry season – during minimal or no
 516 hindrance from cloud cover. In LULC change detection, a temporal series of images is often used,
 517 which require a reliable consistency in the characteristics of the processed images. For SAR image
 518 processing, therefore, image filtering or multi-looking is often applied to reduce speckle noise in SAR
 519 images [60]. However, such pre-processing also reduces the resolution of images. Thus, considering
 520 the landscape structure and the inherently heterogeneous vegetation categories in this study, we did
 521 not consider image speckle as noise. The selected SAR images were recorded during the ascending
 522 satellite overpass. We conducted a temporal average of measured textures across seasons, which,
 523 reduced any potential noise from individual image pixels; meanwhile, the seasonal differences in
 524 volume scattering over vegetation cover provided the texture details for discriminating the
 525 vegetation types. The temporal SAR data and the texture measures were, however, less sensitive to
 526 mapping non-vegetation land cover - notably the water cover. The low classification sensitivity for
 527 water areas can be explained by the following: variability in water cover as a result of seasonal
 528 swampy areas, the seasonal conversion of some swampy areas into subsistence farms of adapted
 529 crops, and the low backscatter intensity over water, which culminates in low or no co-occurring pixels

530 in the GLCM. Consequently, the GLCM had a low likelihood to extract texture for the water class.
531 Nonetheless, this land cover was not of main interest in this study.

532 Compared to a “business as usual” classification using a mono-season multi-spectral optical
533 image (RE1), a combination of texture from both VV and VH bands and the 6bits grey level
534 quantization prior to GLCM texture classification had the highest OA of 88.1% and kappa of 0.85.
535 Moreover, this accuracy resulted in, respectively, 3% and 9.3% reduction in prediction error over the
536 GLCM texture at the default 32bits grey levels and the optical-image. Albeit the relatively low high
537 prediction error of the RE1 model, the class prediction error was low for land cover with no- and low-
538 vegetation: built up, bare soil, savannah and subsistence farmlands; these have rather distinctive
539 optical spectral signatures. For the vegetation cover with high canopy, Table A1 shows the high
540 confusion between cocoa agroforests and transition forests, both of which have high canopy. Since
541 their canopy structure is similar [16], they may not be reliably discriminated using their spectral
542 signatures [17], and less so during the dry season - of low leaf proliferation. Therefore, the
543 classification performance from optical data is logical, and for the vegetation classes, reflects the
544 vegetation status per season and phenological cycle; consequently, the optical reflectance in the dry
545 season was less distinctive for cocoa agroforests versus secondary forests.

546 In the texture-based classification, we averaged volume scattering across seasons, which
547 captured differences the vegetation types. As seen in Table A2, the confusion between cocoa
548 agroforests and transition forest is low, compared to other classes. This indicates that optimizing the
549 grey level improved the classification and helped to distinguish especially the vegetation classes with
550 high heterogeneous canopy. For the study landscape the range of backscatter intensity, for VH and
551 VV bands, is on average above 34 dB. The 6bits grey level quantization was consistent with the range
552 of backscatter intensity in both VV and VH bands, and our results on grey level quantization confirms
553 other reports [39–41]. Therefore, our land cover/use classification performance at grey level
554 quantization of 6bits or 64 levels (GL3_B6) were optimal for discriminating different vegetation,
555 particularly those featuring a high canopy. Other studies on heterogeneous cropland mapping
556 recorded an accuracy of 71% using C-band SAR intensity images [26]. However, in terms of OA, our
557 result is in line with the accuracy observed in different heterogeneous cropping landscapes using a
558 combination of C-band SAR and optical data though [26,34]. The authors [26,34], mapped cropping
559 lands with inherent homogeneous canopy. The landscape in the study is, however, tropical and
560 dominated by vegetation and cropping fields with internally heterogeneous canopy.

561 Our texture-based classification result shows spatial fragmentation of forest cover by cocoa
562 agroforests land use. These transition forest patches are consistent with field observations; the
563 transition forest patches are mostly owned by families and community groups for hunting,
564 performing traditional rituals, and serve as potential cocoa agroforests parcels. The similarity in
565 canopy structure of cocoa agroforests and transition forests is explained by their matching class
566 uncertainty estimates from the optical data i.e. RE1 model. Moreover, the confusion matrix in Table
567 A1 reveals a high commission error between the two classes. Thus, by averaging the seasonal radar
568 volume scattering, the GL3_B6 model discriminated the two vegetation cover with significantly
569 different class uncertainties: the mean class uncertainty for transition forests was 0.28 in GL3_B6
570 based on GL3_B6 model, and unlike in RE1 model, this was significantly different from the 0.4 class
571 average for cocoa agroforests. Following the high overall accuracy and a corresponding low
572 individual class uncertainty, the multi-date texture information from SAR images provided a reliable
573 classifier input for discriminating of perennial agroforestry land cover from transition forest.

574 Classification validation based on accuracy metrics as overall, user, and producer accuracies are
575 influenced by sample class distribution in training data [64]. The differences in pixel resolution
576 between images types (5 m for RapidEye and 10 m for SAR C-Band) resulted in a different number
577 of reference pixels; of which lower numbers result in low estimates of misclassification probability
578 (commission or omission errors). This explains the high overall accuracy for SAR image. The sample
579 size is not as influential in classification validation by estimates of entropy or uncertainty. Entropy is
580 a classic metric for biodiversity in ecology; thus, for this study, it describes the likelihood of a pixel
581 belonging to either one of the classes considered [59]. Considering the heterogeneous landscapes,

582 however, our results contradict other reported uncertainty in croplands mapping [59]; the trend of U
583 and H uncertainties were different for our study landscape. At the uncertainty of 0.4 for example, the
584 entropy or H, compared to U, revealed better the difference in the cumulative proportion of pixels
585 between RE1 and GLE models. The slope of cumulative Shannon entropy (H) was linear for RE1 and
586 sigmoidal for GL3. The former reflects a high likelihood of a single land cover class for each pixel in
587 the multi-spectral image. The sigmoid curve represents a complete range of probabilities for each
588 image pixel- a soft boundary between land cover classes. Compared to perennial agroforests, the low
589 individual class uncertainty for transition forest is logical and consistent, as there is a high temporal
590 stability of radar volume scattering over forest area [31].

591 The classification uncertainty, measured by Shannon entropy (H), for transition forests was
592 comparable for both RE1 and GL3_B6: the likelihood of a pixel being classified only as transition
593 forest was similar for both RE1 and GL3_B6 models. Noteworthy is the quantization level used for
594 GLCM computation in the later, which it reduces unclassified areas to only 7.4 ha. Consequently, the
595 forest area, estimated to cover 1706.9 ha (mean H = 0.25, class error = 0.32) using the RE1, was 500 ha
596 (mean H = 0.28, class error = 0.06) less following the estimates from GL3_B6. Although the uncertainty
597 for cocoa agroforests was significantly reduced by the GL3_B6, their relatively high entropy (mean
598 H = 0.40, class error) indicate a high likelihood of classifying cocoa agroforests as one of the other
599 classes. This distinction between the average entropy for cocoa agroforest (Af) and transition forest
600 (Sf) for GL3_B6 model, confirms a reliable discrimination of the two vegetation types. The high
601 variability in tree density and structure between cocoa agroforests may result in high intra-class
602 variations in SAR backscatter intensity and confusion with other vegetation types, which might
603 explain their high H uncertainty. Likewise, the diversity in farm subsistence crop types and stages
604 may explain the high H for the class. As well for savannah lands, there is a frequent change of canopy
605 volume and structure along with seasonal phenological cycles, except for areas converted to
606 subsistence farming or previous farms that have been left fallow. Such physical changes results to
607 unavoidable classification confusions [60]. The high class uncertainties validate the variability within
608 land cover types and land use dynamics in the landscape; therefore, the uncertainties may not be
609 easily reduced in a texture-based classification of SAR data.

610 The results of this study provide a new insight into the application of GLCM for SAR image
611 processing for mapping tropical croplands of heterogeneous and multi-strata canopy. We highlight
612 the important consideration of grey level quantization in heterogeneous land cover/use classification.
613 Depending on the landscape structure and vegetation types, though, application of SAR imagery for
614 cropland mapping may warrant different analysis procedure [26,34,60]. Thus there may be a need to
615 customised analysis parameter depending on objective and landscape characteristics. The texture
616 measures were pre-selected following the recommendation in reference [32], which is based on
617 inference from optical image classification results, are tailored for vegetation mapping. This pre-
618 selection of texture feature introduced bias in discriminating areas without vegetation. The GLCM
619 texture measures were tailored to discriminate vegetation types, and less sensitive to non-vegetation
620 cover. Texture feature extraction procedure are well established in the literature, with varying
621 opinions. We highlight here, some issues which limit this study. We estimated the GLCM textures
622 using a 5 × 5 window scale, considering the 10m pixel resolution for sentinel1 SAR. However,
623 different window sizes may influence both texture values and classification accuracy. The analysis of
624 different window sizes was not the aim of this study and may be a subject for further investigated
625 for this and other landscapes. We perceive that a fusion of texture information with optical images
626 [26,34], at a temporal scale may improve the uncertainty in discriminating vegetation classes. Also,
627 considering the diversity of vegetation across cocoa production landscapes, the complex structure
628 and canopy of cocoa agroforests, the continuous archive and improvement in SAR image resolution
629 can train deep learning procedures for land cover/use classification at the scale of several landscapes.

630 6. Conclusions

631 This study, to our knowledge, is the first to explore multi-polarization and multi-temporal

632 C-band SAR for discriminating cocoa agroforest cropping land in a heterogeneous landscape. We use
 633 seasonal differences in volume scattering from the dielectric (water content) status and structure of
 634 vegetation canopy as a metric to discriminate vegetation types. We make the following conclusions:
 635 1. For the same window size and an invariant direction, reducing the grey level quantization
 636 improved classification accuracy marginally, but significantly reduced the uncertainty in
 637 discriminating cocoa agroforests from other vegetation covers.
 638 2. Classification validation by estimates of Shannon entropy (H) reveal subtle differences in
 639 individual class prediction and provide reliable information for making inferences in
 640 heterogeneous landscape mapping.
 641 3. The magnitude of forest fragmentation by cocoa agroforest, which is concealed by vegetation
 642 indices from spectral reflectance, is mapped using GLCM texture measures from C-band SAR
 643 images.

644 We suggest an approach for mapping cocoa agroforests in tropical heterogeneous cropping
 645 landscapes using C-band SAR imagery. The approach shows the reliability of Sentinel1 SAR image
 646 archive in landscape monitoring; especially for mapping cocoa agroforests expansion, and their
 647 contribution to the loss of transition and primary forest cover. Therefore, it has potential application
 648 in, for example, estimating the contribution of agroforestry to national and regional REDD+
 649 (Reducing Emission from Deforestation and Forest Degradation and the role of conservation,
 650 sustainable management of forests and enhancement of carbon stocks in developing countries)
 651 strategies. However, there is a need to assess classification uncertainties in different agroforestry
 652 dominant landscapes – for an operational regional mapping in the Congo Basin sub-region.
 653

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 655 Frederick N. Numbisi and Frieke M.B. Van Coillie; field investigation, Frederick N. Numbisi; data curation,
 656 Frederick N. Numbisi; data analysis, Frederick N. Numbisi; original draft preparation, Frederick N. Numbisi.;
 657 review and editing, Frederick N. Numbisi, Frieke M.B. Van Coillie and Robert De Wulf; supervision, Frieke M.B.
 658 Van Coillie and Robert De Wulf.

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668 **Conflicts of Interest:** The authors declare no conflict of interest.

669

670 Appendix A

671 **Table A1.** Classification confusion matrix of RE1 Model. OA = Overall Accuracy, UA = User

672

		Accuracy, PA = Producer Accuracy.								
		nTrees = 250, mTry = 2, OOB error = 19.2%, OA = 81.0%								
		Reference							Class Error	PA
		Bu	Es	Sv	W	Af	Fa	Sf		
Predicted	Bu	1238	0	0	0	0	1	0	0.001	99.9
	Es	2	529	8	0	0	117	0	0.194	80.6
	Sv	0	0	1469	1	52	172	2	0.134	86.6
	W	0	0	9	52	16	0	2	0.342	65.8

Af	0	0	64	8	855	31	335	0.339	66.1
Fa	0	57	233	0	55	1447	3	0.194	80.6
Sf	0	0	3	0	309	4	665	0.322	67.7
UA	99.8	90.3	82.3	85	66.4	81.7	66.0		

673

674

Table A2. Classification confusion matrix of GL3_B6 Model.

nTrees = 550, mTry = 8, OOB error = 9.9%, OA = 88.8%										
	Reference							Class Error	PA	
	Bu	Es	Sv	W	Af	Fa	Sf			
Predicted	Bu	244	2	5	0	23	17	1	0.164	83.6
	Es	14	89	19	0	6	21	0	0.403	59.7
	Sv	0	0	424	0	0	12	0	0.028	97.2
	W	2	0	3	0	6	7	0	1.000	0
	Af	3	0	0	0	325	8	4	0.044	95.6
	Fa	1	0	18	0	3	415	1	0.053	94.8
	Sf	0	0	0	0	16	0	246	0.061	93.9
UA	92.4	97.8	90.4	0	85.8	86.5	97.6			

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Table A3. Predicted surface area (ha) of land cover/use for optical multi-spectral image and GLCM textures at four different grey level quantization

LULC Class	Model				
	RE1	GI3	GI3_B8	GI3_B6	GI3_B4
Bu	68.89	547.38	650.35	545.89	719.07
Es	160.73	73.41	235.14	83.14	76.15
Sv	4485.2	3118.63	3035.37	3465.95	3629.72
W	137.08	0.27274	0.15079	0.16927	0.16446
Af	2733.63	2986.84	2904.18	3355.87	3254.77
Fa	2052.11	2787.95	3081.47	3120.90	3002.42
Sf	1706.94	1202.86	1306.70	1210.16	1094.61
Un	0	647.32	562.62	7.42	12.60

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