

1 Article

2 Modelling wetland growing season rainfall 3 interception losses based on maximum canopy 4 storage measurements

5
6 Wojciech Ciężkowski ^{1,*}, Tomasz Berezowski ², Małgorzata Kleniewska ¹, Sylwia Szporak-
7 Wasilewska ³ and Jarosław Chormański ¹

8 ¹ Department of Hydraulic Engineering, Faculty of Civil and Environmental Engineering, Warsaw
9 University of Life Sciences, Nowoursynowska 166, 02-787 Warsaw, Poland

10 ² Gdansk University of Technology, Faculty of Electronics, Telecommunication and Informatics, Department
11 of Geoinformatics, Gabriela Narutowicza 11/12, 80-233 Gdansk, Poland

12 ³ Water Center Laboratory, Faculty of Civil and Environmental Engineering, Warsaw University of Life
13 Sciences, Nowoursynowska 166, 02-787 Warsaw, Poland

14 * Correspondence: w.ciezkowski@levis.sggw.pl

15

16 **Abstract:** This study estimates rainfall interception losses from natural wetland ecosystems based
17 on maximum canopy storage measurements. Rainfall interception losses play an important role in
18 water balance, which is crucial in wetlands, and has not yet been thoroughly studied in relation to
19 this type of ecosystem. Maximum canopy storage was measured using the weight method. Based
20 on these measurements, daily values of interception losses were estimated and then used to
21 calculate long term interception losses based on precipitation and potential evapotranspiration data
22 for the 1971–2015 period. Depending mainly on the number of days with precipitation, the results
23 show that total interception losses for the growing season as well as monthly interception losses are
24 around 13% of gross rainfall. This value is similar to the values observed for some forests. Hence,
25 interception losses should not be disregarded in hydrologic models of wetlands, especially because
26 data trends in meteorological conditions (mainly number of days with precipitation) show that
27 interception losses will increase in the future if those trends stay the same.

28 **Keywords:** interception losses; water balance; water storage capacity; wetland; sedges; Biebrza river

29

30 1. Introduction

31 Rainfall interception is the process where ground surfaces, whether vegetated or sealed because of
32 human construction, catching and retaining water during precipitation [1]. It is of considerable
33 importance in water resource management [2], and in the context of climate change [3]. Phytoelement
34 wetness caused by rainfall interception also plays an important role in many areas such as the control
35 of plant disease [4,5], plant susceptibility to dry and wet acid deposition [6,7], in plant photosynthesis
36 and in plant yield [8,9]. However, the lack of knowledge about rainfall interception means that this
37 process has been disregarded or underestimated in rainfall-runoff analysis [10]. Some models
38 disregard it completely, e.g. the THALES model [11], the DBSIM model [12], the REW approach [13],
39 and the Topkapi model [14]. Other models simplify interception, and lump its value with
40 transpiration: Topmodel [15], SHE [16], Wetspa [17].

41 The rain that falls and hits plant surfaces is temporarily retained on the plant surface and,
42 ultimately, either evaporates or makes its way to the ground by throughfall or by flowing down
43 branches and stems [3]. Storage water can be divided into transistor storage (water that drips down
44 immediately), conditional storage (drip out by leaf movements caused, e.g., by wind) and residual
45 storage (water which can be removed only by evaporation) [18]. Residual storage is the most widely

46 examined and it can be represented as: canopy storage [18], water storage capacity [19], canopy
47 rainfall capacity [20] and rainfall interception [1,21].

48 Most research on rainfall interception by plant canopies has focused on different kind of trees in
49 different climatic zones, for instances tropical rain forests in Guyana [22], Brazil [23] and eastern
50 Amazonia [24]; tropical montane cloud forest of Guatemala [25]; laurel forest in Canary Island [26];
51 lowland tropical rainforest in Brunei [27]; boreal forest in Sweden [28]; arid zone pine forest [29]; and
52 deciduous Mediterranean forest in Slovenia [30]. The number of studies on rainfall interception by
53 lower vegetation are inadequate with focusses mainly on: grasslands [19,21,31], drylands [1,20]
54 shrublands [31,32], or croplands [33]. There are two reasons for this. The first is that interception
55 losses from forest are generally greater in absolute volume than those from lower vegetation.
56 However some studies [31] showed that interception of grasslands can reach approximately 30% of
57 annual rainfall, which is similar to many forests [34]. The second reason is the difficulties in
58 measuring rainfall interception in lower vegetation, especially in natural heterogeneous ecosystems.

59 Lately, a few studies on rainfall interception have been made in a natural wetland ecosystem in
60 Biebrza Basin (NE Poland). Those studies focused on determining the spatial distribution of
61 interception capacity based on ground truth measurements of maximum canopy storage and ALS
62 (airborne laser scanning) [35], also ground truth data was used with Landsat 7 ETM+ image [36].
63 There have also been numerous studies conducted on estimation of interception using remote sensing
64 techniques that have focused on forest [37,38], croplands [39] or grasslands [38]. Remote sensing
65 techniques show great potential in estimating the spatial distribution of interception, but ground
66 measurements are mandatory for studies on less well-recognised ecosystems such as wetlands. This
67 study follows increasing interest in interception on wetlands ecosystems and on growing season
68 interception in this area.

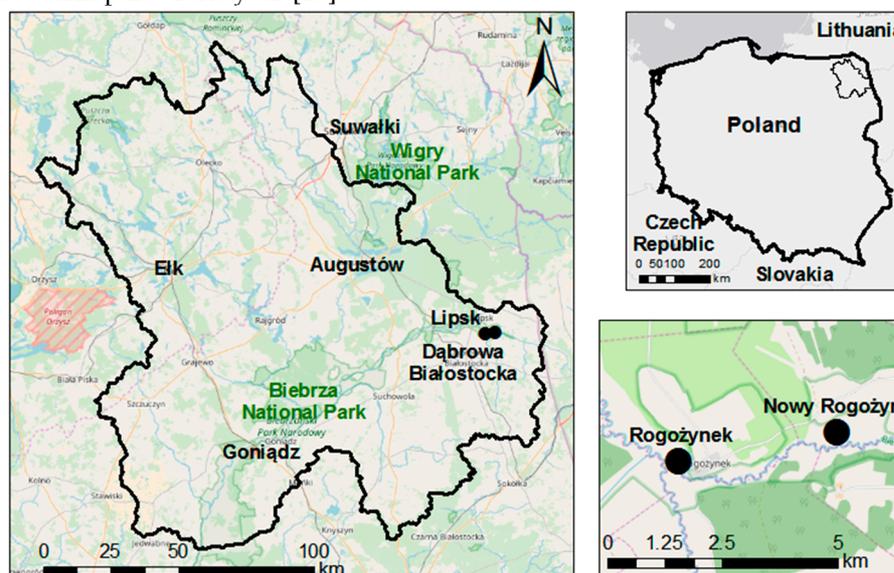
69 Studies on the hydrologic cycle has shown a rising interest in rainfall interception phenomena.
70 Even in urban areas where rainfall interception losses seem to be insignificant has begun to be
71 investigated for examination of its influence on water balance [40,41]. The main aim of this study is
72 to investigate the influence of hydrological processes and meteorological conditions (precipitation,
73 evapotranspiration, length of growing season, number of days with precipitation) on the amount of
74 intercepted water for the typical land covers in the Biebrza Valley. The second aim is to estimate
75 interception losses for long time series based on ground measurements of maximum canopy storage.
76 The research focuses only on the growing season due to its length (perennial average 200 days) and
77 the fact that the height and size of analysed plants outside the growing season is negligible. Relation
78 with meteorological conditions together with investigations of the perennial trends of meteorological
79 conditions in year 1971-2015 can show potential influence of climate change on rainfall interception
80 in future. These results might improve hydrological modelling on wetlands and the understanding
81 of wetland ecohydrological processes.

82 2. Materials and Methods

83 2.1. Study site

84 The maximum canopy storage measurements were taken in the Upper Biebrza Basin (NE
85 Poland) (Figure 1). The study area is part of a river valley and is dominated by peat soils (mainly peat
86 moss with the exception of areas near the river). The area is inundated due to irregular floods. The
87 dominated plant community is *Carici canescentis-Agrostietum caninae*, creating a belt of up to 0.6 km
88 wide on both banks along the river. The area is also covered by *Caricion lasiocarpae*, *Molinio-*
89 *Arrhenatheretea*, as well as *Phragmitetum* communities which grows along the Biebrza river. In sections
90 of the valley abandoned by agriculture, as it is typical of natural river valleys with wetland
91 vegetation, sedges will form in hummock patterns. Research has been conducted on areas covered
92 with phytocenoses of semi-natural plant communities created by a single, extensive, mechanical
93 mowing followed by biomass removal, typically carried out after July 1st, depending on whether soil
94 moisture conditions in the valley is allowing for the entry of field equipment or not. This type of
95 usage is dominant in the study area of the Biebrza valley (Figure 2).

96 The climate of the Biebrza River Valley is temperate transitional, with evident continental
 97 influences. The mean annual air temperature in the basin was 6.6 °C in the years 1961–2000. However,
 98 in individual years the mean values can vary between 4.8 °C and 8.3 °C. The mean annual
 99 precipitation ranged from 536 mm to 610 mm, and the mean annual wind speed was 3.8 m·s⁻¹ [42].
 100 The diverse topography and morphology lead to variations in local and macro-scale climate
 101 conditions and high biodiversity, which are the main cause of differences in temperature and
 102 humidity in the warm part of the year [43].



103
104 **Figure 1.** Study site area (location in the Biebrza River Basin – left).



105
106 **Figure 2.** Meadows under extensive agricultural use (in Summer – left; in Autumn –right).

107 2.2. Maximum canopy storage field measurements

108 Field measurements focused on determining maximum canopy storage (I_{max} [mm]) were conducted
 109 on two meadows (Figure 1). The two selected meadows (in Nowy Rogożyn and Rogożynek) are
 110 mowed one time per year and are representative for the whole Upper Biebrza Basin wetland
 111 meadows. Acquired phytosociological relevés from the selected meadows show high variability of
 112 vegetation with up to 60 different plant species per meadow.

113 Measurements are made during three field campaigns in 2014: 26–29 June (1–2 months into the
 114 growing season), 28–30 July (peak of growing season) and 4–5 September (plant weeding, 1–2 weeks
 115 before the end of the growing season). The dates of field campaigns were chosen to represent changes
 116 in plant phenology from the beginning of growth through to the state of full development and plant
 117 weeding till the death of the plants. Field campaigns were organized on sunny days in hours when
 dew was not present. First, in order to characterise plant communities presented in the research area,

118 phytosociological data were collected in relevés. Then samples were collected by cutting all the
119 above-ground vegetations within a 0.25m by 0.25m square. In total, 68 samples were collected, with
120 each sample consisting of 90–250 plants.

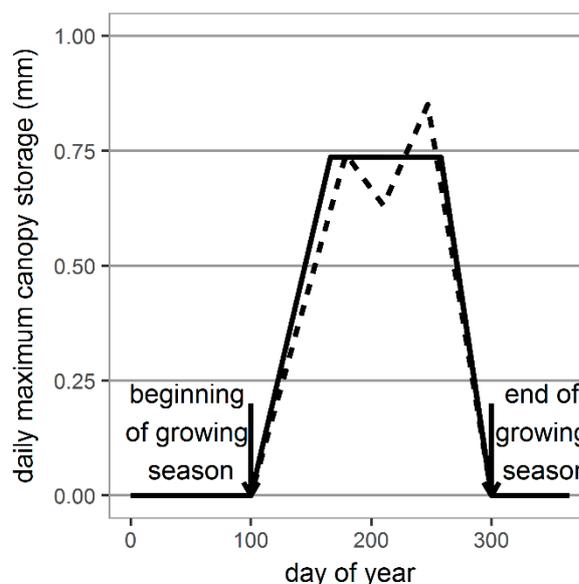
121 All collected plants were immediately transported to a laboratory, where further measurements
122 were conducted in a natural state. I_{\max} was determined by weighting. I_{\max} is calculated as the
123 difference between the freshly collected and the weight of a soaked plant. There are two methods of
124 soaking plants with water: spraying [21,31,32,44] and submerging [1,19,45]. In this study plants were
125 submerged because it is less ideal with spraying where the droplet size will cause variations [46].
126 Results obtained by submersion are less dependent on human factors and are more reproducible.

127 All plants from each square were weighed one by one to an accuracy of 0.01 g using a WPS2100/C
128 scale (RADWAG). Then, to avoid interactions between plants, the plants were submerged in a tank
129 of water one by one. After a few seconds after submersion (to allow the excess water to drip from the
130 leaves) each plant was weighed. The difference between the weight of plants before and after
131 submergence is a factor of I_{\max} in grams per sampling area (0.0625 m²). This value was recalculated
132 into mm of H₂O per unit area by dividing the volume of intercepted water (calculated assuming water
133 density equal to 1 g/cm³) by the sampling area.

134 2.3 *Generating daily values of maximum canopy storage*

135 All measurements of maximum canopy storage were taken in three campaigns during a single
136 growing season, and the means of I_{\max} for the three campaigns were tested for statistical significance
137 in their differences using the ANOVA test. We find the difference not statistically significant with
138 $p=0.141$ (section 3.1). Hence, to obtain perennial daily values of I_{\max} the Monte Carlo method was
139 used. Based on the 2014 measurements ($n=68$), 100,000 repetitions of daily unbiased maximum
140 canopy storage samples were generated for the 1971–2015 period. Our measurements distribution
141 was fitted with gamma distribution using the maximum-likelihood estimation method. Gamma
142 distribution was chosen because it is often used to model maximum discharge or precipitation.
143 Maximum canopy storage values can be modelled using gamma distribution. Our tests with other
144 distributions, such as X^2 , lognormal and normal, showed that gamma distribution fits better to the
145 data. The fitted gamma distribution parameters were: shape=4.704 and rate=6.386. The fitted
146 distribution is good according to the Kolmogorov–Smirnov test ($D=0.0617$, $p=0.95$). The generated
147 values were assigned to the mid-day of the 15th days of June, July, August and September. Other
148 values were linearly interpolated assuming that I_{\max} is 0 at the beginning and end of the growing
149 season. Previous research on wetland meadow canopy storage in this area [35,36,45,47] also focused
150 only on the growing season. This is due to the negligible height of sedge communities before and
151 after the growing season in this area. Hence, this study focused on rainfall interception losses in the
152 growing season.

153 The results of measurements and the example of generated daily maximum canopy storage
154 values are show in Figure 3. For diurnal calculation, daily values of maximum canopy storage depend
155 on the beginning and end of the growing season, which was calculated based on air temperature
156 (section 3.2).



157

158

Figure 3. Measured (dashed line) and mean of generated (solid line) daily maximum canopy storage.

159

2.4 Interception losses model

160

Daily rainfall interception losses (I_{day}), meaning water that was retained by plants during a given period, were calculated with equation (1) based on daily precipitation (P), potential evapotranspiration (ET_0) (calculated according to [48]) and maximum canopy storage (I_{max}):

163

$$I_{day} = \begin{cases} I_{max} & \text{if } P - ET_0 > I_{max} \\ P - ET_0 & \text{if } P - ET_0 \leq I_{max} \end{cases} \quad (1)$$

164

Calculations of I_{day} were performed for data from the nearest non-urban meteorological station located in the Biebrza Valley (in Pieńczykówek) for the years 1971 to 2015. Then, it is repeated for all 100,000 simulated values of maximum canopy storage. The meteorological data for Pieńczykówek station were made available by the Institute of Meteorology and Water Management: National Research Institute (IMGW-PIB) in Poland.

169

This model of rainfall interception is a simple one as there are no well-defined physically-based interception model of wetland vegetation in the study area. The most common physically-based models are Gash [49] and Rutter [34], which were developed for forest ecosystems, and applying them to the ecosystems of our research area was difficult due to the issue of distinguishing certain model parameters (such as trunk storage capacity or free throughfall coefficient, which are technically difficult to define for low vegetation). For this reason, rainfall interception models for grasslands or wetland meadows have not been properly configured. And of the nearly 70 applications we found in the literatures, only 4 were agricultural [3], and these models are difficult to implement due to issues with parametrisation and measurements of rainfall interception. Because of similar reasons, and adding the diversity in natural heterogeneous wetland ecosystems, it is even more difficult to model their rainfall interception.

180

3. Results

181

3.1. Field measurements results

182

The results of the three field campaigns are compared in Table 1 and Figure 4. The numbers of samples per campaign varies from 19 in July to 29 in June, but the ANOVA test show no statistically significant relation between the means of maximum canopy storage for these campaigns ($p=0.141$). In all months the standard deviation was about 50% of the mean value, which is caused by high biodiversity (phytosociological relevés were recorded from 24 to 39 different species; in the same sampling areas the number of species was constant during the growing season). The highest (1.61

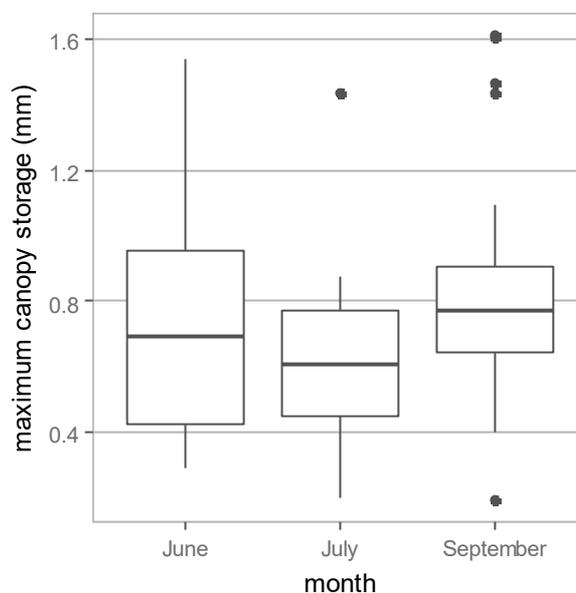
187

188 mm) and the lowest (0.19 mm) values of I_{\max} were measured in September. Other than high
 189 biodiversity which is causing the large variation in values of I_{\max} , plants in September are also going
 190 through different stages of their life cycle (some were still growing, while others had already dried
 191 out) which is another reason for the big range of observed values in this month. In June and July all
 192 plants are in the same phenological phase. From all campaigns, the mean values of I_{\max} are the same
 193 at 0.75 mm.

194 **Table 1.** Maximum canopy storage in mm H₂O per unit area and in g H₂O per sampling area
 195 measured in 3 field campaigns

date	number of samples	maximum canopy storage					
		mm H ₂ O per unit area			g H ₂ O per sampling area		
		mean ± SD	min	max	mean ± SD	min	max
26-29.06	28	0.74 ± 0.35	0.29	1.54	46.4 ± 21.9	17.9	96.2
28-30.07	19	0.63 ± 0.27	0.20	1.44	39.5 ± 16.9	12.2	89.7
3-5.09	21	0.85 ± 0.39	0.19	1.61	53.2 ± 24.4	12.1	100.6

196



197

198

Figure 4. Measured values of maximum canopy storage for the 3 field campaigns

199

3.2 Meteorological element trends, 1971–2015

200

201

202

203

204

205

206

207

208

209

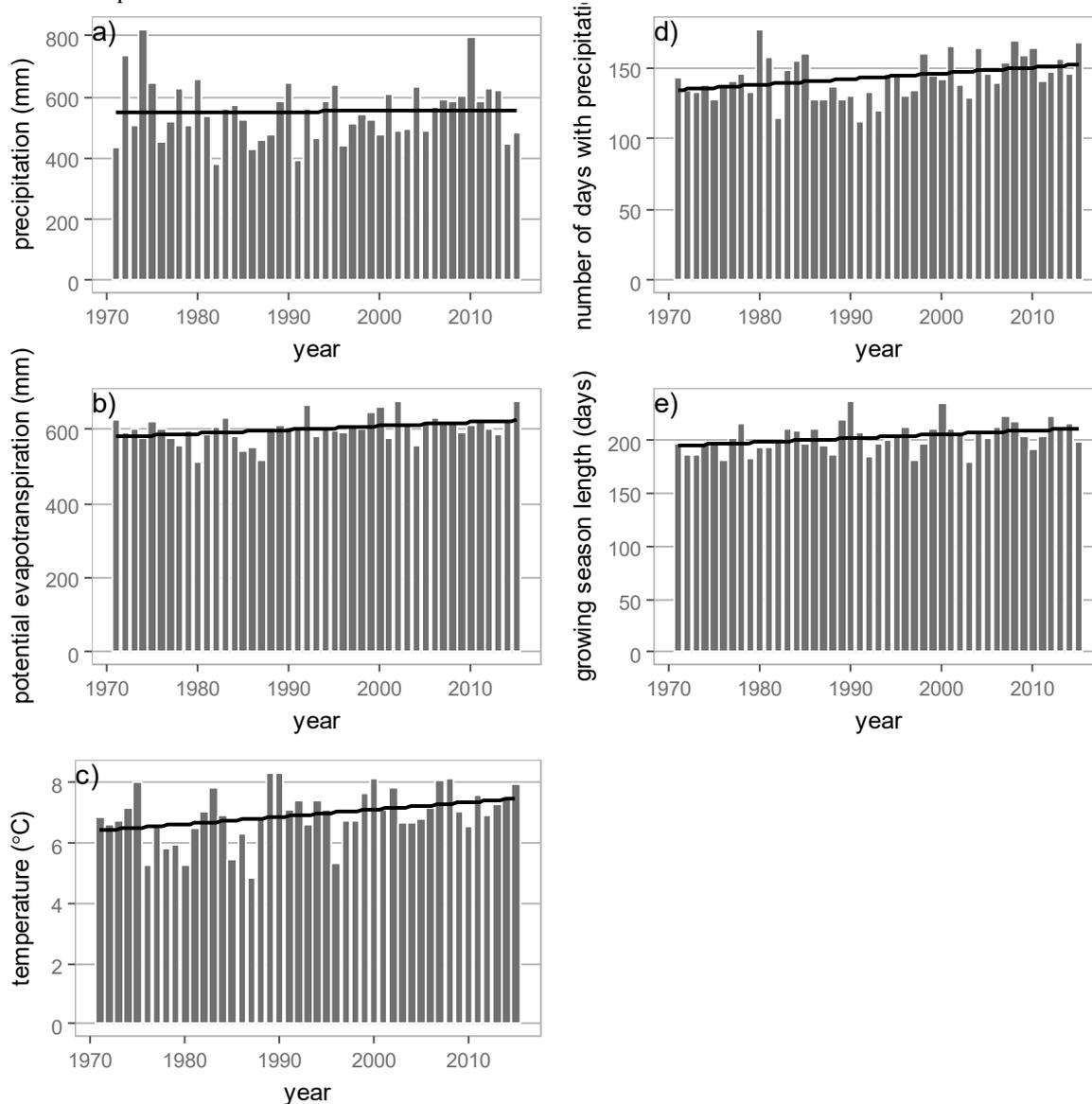
210

211

212

Based on daily values of air temperature (mean, minimum and maximum), relative humidity and wind speed, potential evapotranspiration was calculated using the Penman–Monteith equation with the FAO modification [47]. Potential evapotranspiration and daily precipitation are used to investigate perennial trends of precipitation sum, potential evapotranspiration sum, mean temperature and days with precipitation in the whole year and during the growing season. The length of the growing season was also investigated. The growing period in Poland is closely related to the dates on which the +5°C daily mean air temperature threshold value is permanently crossed. Because the ecological factor which affects the rate of plant life processes is thermal conditions, the duration of the growing period is approximately equal to the thermal growing season [50]. Results for perennial trends are showed in Figure 5. Mean perennial values of meteorological elements and slope of the trend (for whole year and for only growing season) were calculated, and the results are summarized in Table 2. Apart from precipitation and number of days with precipitation in the growing season, all elements show slightly growing trends which are significant (p -value ≤ 0.05). The

213 non-parametric Mann–Kendall trend test was also performed, using the Kendall R package [51].
 214 Results are compared in Table 2 and it shows the same trends as the linear trends.



215
 216 **Figure 5.** Perennial trends of a) precipitation, b) potential evapotranspiration, c) temperature, d)
 217 number of days with precipitation, e) length of growing season

218 **Table 2.** Annual and growing season means of measured meteorological elements, and slope of their
 219 trends

meteorological parameter	annual mean	trend slope	growing season mean	trend slope	annual M-K tau
precipitation (mm)	552**	+0.15 mm/year	383**	-0.02 mm/year	0.062**
potential evapotranspiration (mm)	600*	+1 mm/year	530*	+1.76 mm/year	0.259*
temperature (°C)	6.9*	+0.02 °C/year	13.1*	+0.02 °C/year	0.259*
number of days with precipitation (-)	143*	+0.4 day/year	78**	+0.18 day/year	0.274*
growing season length (-)	204*	+0.37 day/year			0.272*

220 * p-value ≤ 0.05

221 **p-value > 0.05

222 3.3. Relations between interception losses and meteorological elements

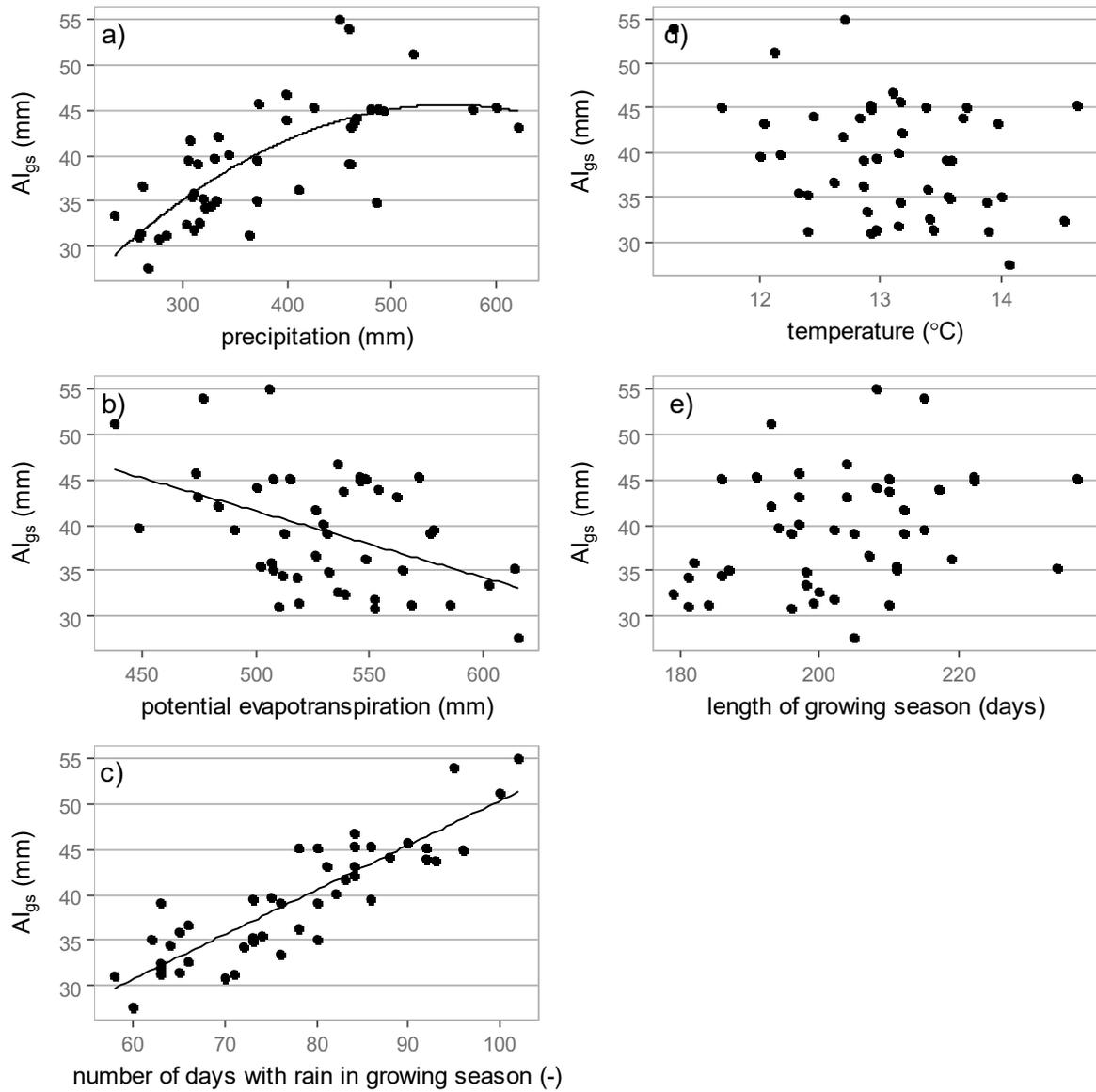
223 The daily interception losses calculated in Section 2.4 was used to compute the growing season and
224 monthly sums of interceptions, hereinafter referred to as 'absolute growing season' (AI_{gs}) and
225 'monthly' interception (AI_{month}) expressed in mm. Absolute interception was also divided by growing
226 season or monthly sum of precipitation, which hereinafter are referred to as 'relative growing season
227 interception' (RI_{gs}) and 'relative monthly interception' losses (RI_{month}), and are both expressed as
228 percentage of precipitation. Additionally, for both RI_{gs} and RI_{month} the relationship with
229 meteorological elements was investigated in the monthly and growing season timescales.

230 3.3.1. Growing season interception

231 Calculated absolute interception in the growing season ranged from 27.6 mm to 55.0 mm, with a
232 mean of 39.3 mm. The part of the rain that was intercepted (relative interception) ranged from 9.0%
233 to 16.3%, with a mean of 12.6%. Extreme values were observed in different months for both type of
234 interception.

235 The relation between absolute interception in the growing season and hydrological and climatic
236 variables is presented in Figure 6. The relation between the number of days with rain in the growing
237 season and the length of the growing season is positive, while, the relation between potential
238 evapotranspiration and temperature is conversely negative. This relation is observed because in
239 years with high potential evapotranspiration, the sum of precipitation is low, and the total amount
240 of water which can be intercepted by plants understandably drops. Correlation coefficients (R) are
241 compared in Table 3. The highest correlation was obtained for number of days with rain in the
242 growing season ($R=0.88$), while a weak but significant correlation was also obtained for precipitation
243 ($R=0.72$). The quadratic relation with precipitation shows that in wet years (precipitation above 550
244 mm), intercepted water cannot evaporate between rain episodes (because of high humidity which
245 leads to slow evaporation) and interception rates start to drop after the threshold precipitation value
246 is exceeded. Other meteorological variables (potential evapotranspiration, temperature and the
247 length of the growing season) were not significantly correlated with the absolute rainfall interception
248 in the growing season.

249 The relationship of relative interception in the growing season to hydrological and climatic
250 variables is showed in Figure 7. All relations are negative. Coefficients of correlation (R) are compared
251 in Table 3. The highest correlation was obtained for precipitation ($R=0.75$). Other meteorological
252 variables (potential evapotranspiration, temperature, number of days with precipitation in the
253 growing season, and length of growing season) were not significantly correlated with the relative
254 interception in the growing season.



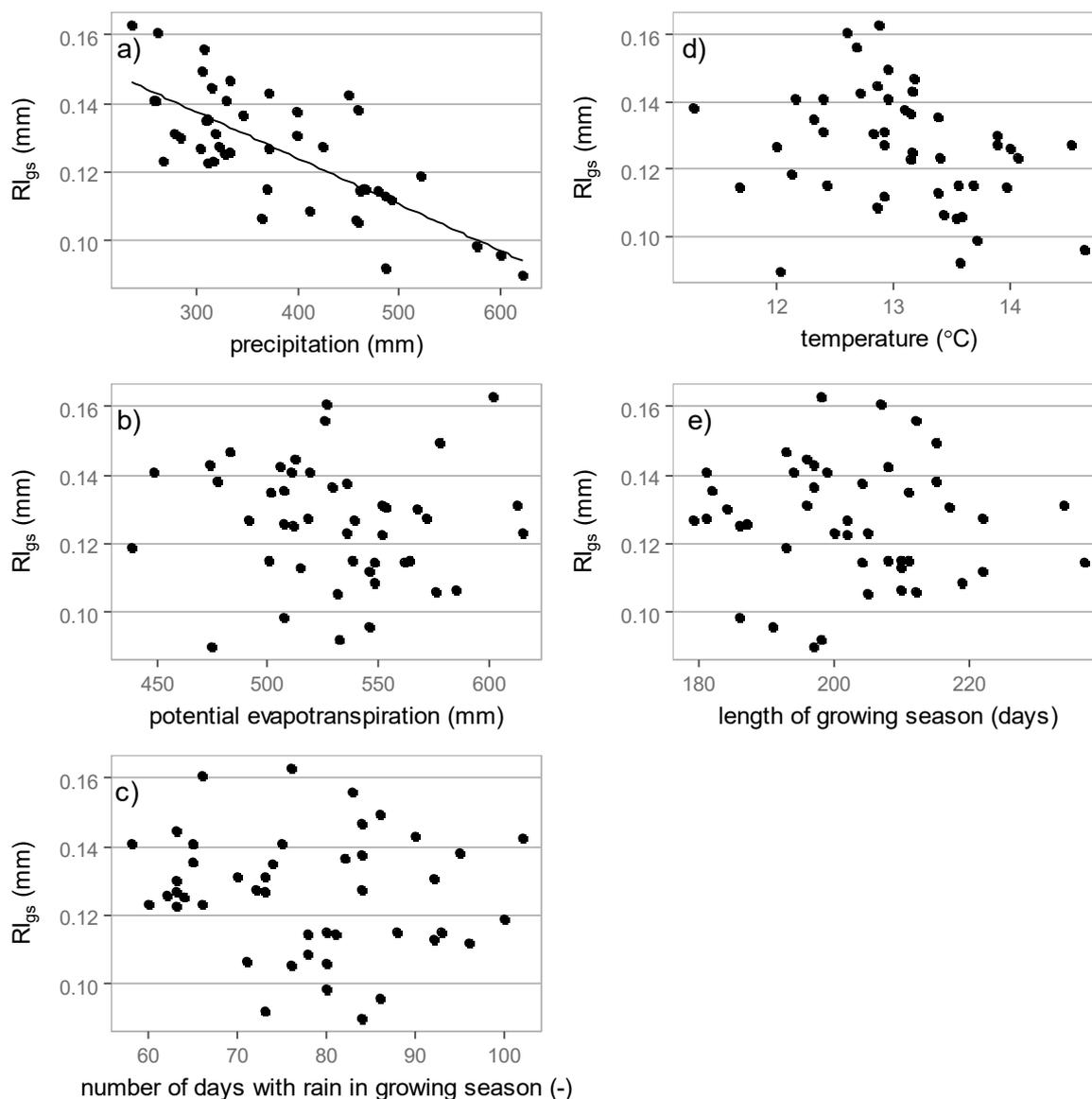
255

256

257

258

Figure 6. Relationship between absolute growing season interception (A_{lgs}) and: a) sum of precipitation, b) sum of potential evapotranspiration, c) number of days with rain, d) mean temperature, e) length of the growing season



259

260

261

262

Figure 7. Relationship between relative growing season interception and: a) sum of precipitation, b) sum of potential evapotranspiration, c) number of days with rain, d) mean temperature, e) length of the growing season

263

264

Table 3. Correlation coefficients for relation between growing season interception (absolute and relative) and meteorological elements

meteorological element	absolute interception (mm)	relative interception (-)
precipitation	0.72*	0.75*
potential evapotranspiration	0.45*	0.05**
numbers of days with rain	0.88*	0.13**
temperature	0.36*	0.28*
growing season length	0.30*	0.08**

265

* p-value ≤ 0.05

266

**p-value > 0.05

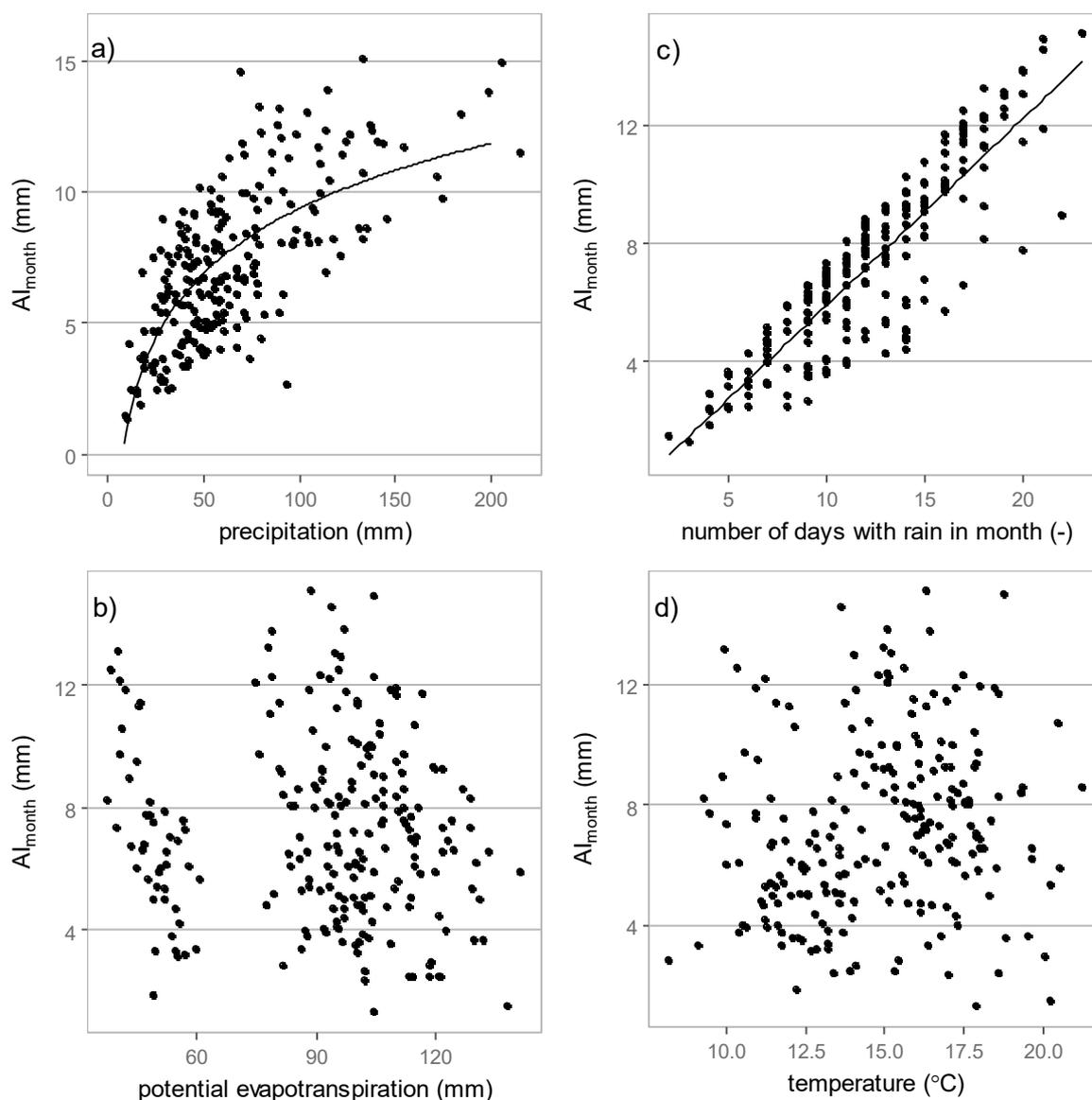
267

3.3.2. Monthly interception

268 Calculated monthly absolute interception in the growing season ranged from 1.32 mm to 15.07 mm
 269 with a mean of 7.25 mm. The part of the rain that was intercepted (relative interception) ranged from
 270 3.8 % to 38.2 % with a mean of 13.0 %.

271 The relation between monthly absolute interception and observed meteorological elements is
 272 shown in Figure 8. Correlation coefficients (R) are compared in Table 4, separately for single months
 273 and overall for all months together. A high correlation was observed for the number of days with
 274 precipitation per month. Overall R was 0.88, but for single months, R is higher with figures ranging
 275 from 0.92 in May to 0.98 in August. Also significant, but much weaker, was the correlation of monthly
 276 AI_{month} with monthly sum of precipitation ($R=0.69$). In September and July, R is 0.71 and 0.74,
 277 respectively, while for other months the R values are slightly lower. The overall correlation with
 278 potential evapotranspiration is irrelevant, but for single months the trend is significant and is up to
 279 0.73 in July.

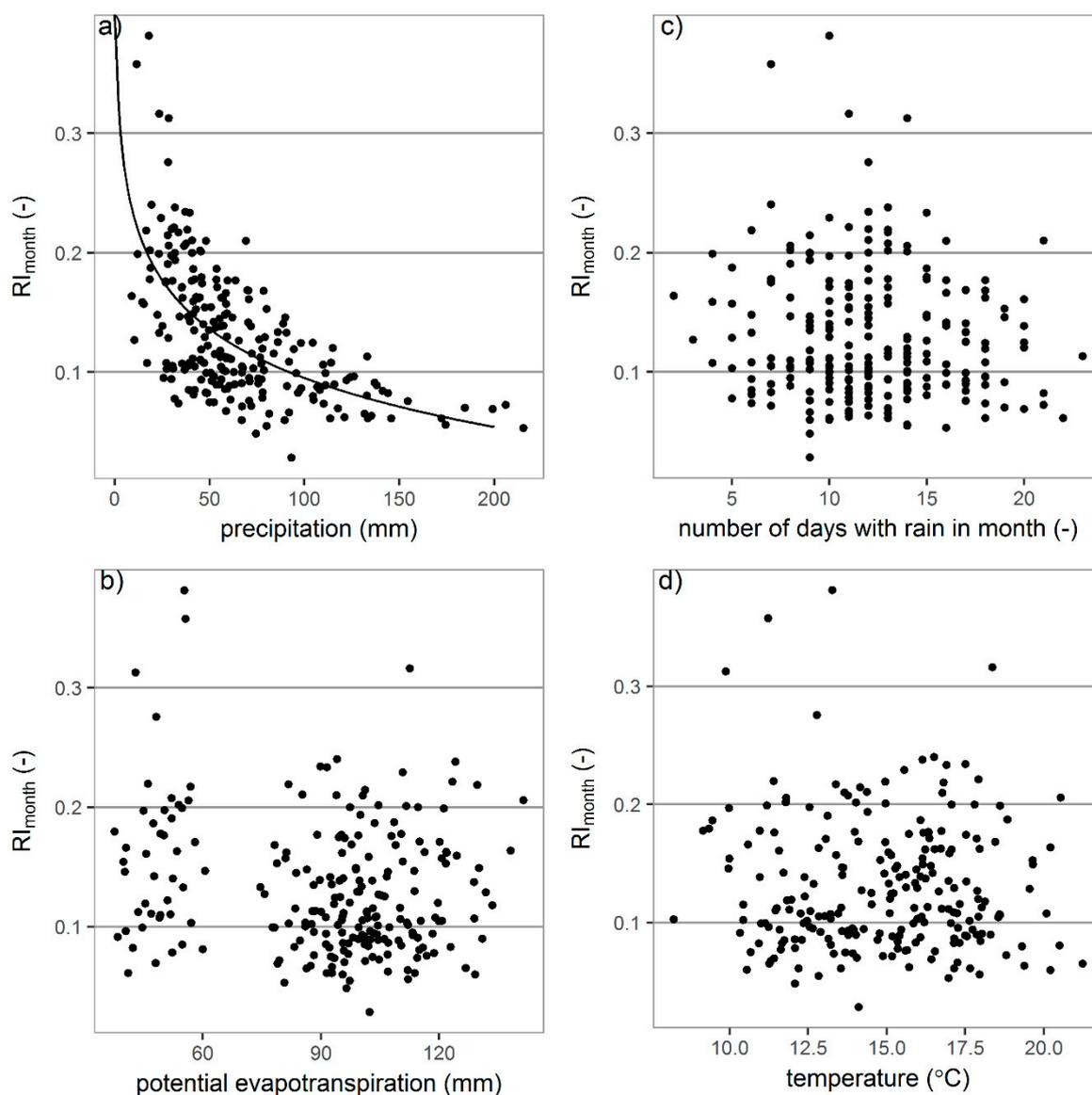
280 The relation between relative interception and observed meteorological elements is shown in
 281 Figure 9. Correlation coefficients (R) are compared in Table 4, separately for single months and
 282 overall for all available data. A significant correlation of relative interception was observed only for
 283 precipitation, but overall R is only 0.62, with a maximum of 0.77 in June.



284

285 **Figure 8.** Relationship between absolute month interception (AI_{month}) and: a) sum of precipitation, b)

286 sum of potential evapotranspiration, c) number of days with rain, d) mean temperature



287

288
289

Figure 9. Relationship between relative month interception (RI_{month}) and: a) sum of precipitation, b) sum of potential evapotranspiration, c) number of days with rain, d) mean temperature

290
291

Table 4. Statistically significant ($p\text{-value} < 0.05$) correlation coefficients for relation between growing season absolute interception (AI), relative interception (RI) and meteorological elements

292

		May	June	July	August	September	All data
precipitation	AI	0.68	0.62	0.74	0.67	0.71	0.69
	RI	0.60	0.77	0.70	0.65	0.59	0.62
potential evapotranspiration	AI	0.40	0.59	0.73	0.64	0.69	*
	RI	*	*	*	0.37	*	*
numbers of days with rain	AI	0.92	0.98	0.97	0.98	0.97	0.88
	RI	*	*	*	*	*	*
temperature	AI	*	0.35	0.50	0.38	0.31	*
	RI	*	*	*	*	*	*

293

*statistically insignificant relation ($p\text{-value} > 0.5$)

294

295 **4. Discussion**296 *4.1 Maximum canopy storage*

297 Investigation of maximum canopy storage on wetlands is not common. Based on what was observed
298 in the literature, values between 0.2 mm and 1 mm are used, for example in the WetSpa model [17].
299 In this research the smallest value of maximum canopy storage is 0.19 mm, but the upper limit of the
300 result is higher than 1 mm, reaching 1.61 mm (Table 1). Another research conducted in the Lower
301 Biebrza Valley (to the south of the study site) shows that for selected species maximum canopy
302 storage can range from 0.87 mm to 7.37 mm [45]. When compared with our work, results presented
303 in [42] seems to be overestimated. This large variability is caused by the high biodiversity in the
304 natural wetland ecosystem. We also would like to note that the sampling sizes and submerging
305 techniques for the other study are different. In this study, each plant sample of 25 cm ×25 cm grid has
306 been split for individuals plants and measurement taken one by one, but in [45] the whole plant
307 sample in the 10 cm ×10 cm square has been submerged, which leads to a better generalisation of
308 results in this study.

309 Most of the related works have been focused on grasslands and meadows, where plants are
310 similar in morphology. For 27 alpine plant species found in the Glacier Lakes in Wyoming the
311 maximum canopy storage was maximally 0.05 mm [52]. On the other hand, for 9 herbaceous species
312 in the Alps the observed values ranged from 0.13 mm to 3.14 [19]. In the Bodega Marine Reserve
313 (California) for 18 clover species the maximum canopy storage ranged from 0.13 mm to 0.36 mm [5].
314 In comparison, the results for wetland vegetation in this study fall in the middle.

315 Maximum canopy storage is in general lower for wetland plants than for forest. For example in
316 [53], the maximum canopy storage for a Douglas-fir forest (>450 years old) located within the Gifford
317 Pinchot National Forest (southern Washington) equals 3.32 mm, but in the same study for a young
318 Douglas-fir forest (20–25 years old) the value was much smaller at 1.4 mm, which is comparable to
319 the extreme values shown in this study. In [54] a Douglas-fir forest in Speulderbos in the centre of the
320 Netherlands has obtained higher values, such as 2.4 mm. However, some researchers have reported
321 similar results of wetland vegetation for *P. sitchensis* in Great Britain (0.75 mm [55]) or even smaller
322 for *Pinus sp.* in the Bray forest located 20 km southwest of Bordeaux, with a range of 0.50 to 0.55 mm
323 [56]. Hence, this study shows that maximum canopy storage for wetland plants communities is
324 smaller than for forests, but for some specific wetland sites it intercepts more water than forest.

325 *4.2 Interception losses*

326 Maximum canopy storage provides information on the maximum amount of water that can be held
327 by plants during a rainfall event. To investigate the influence of interception on water balance we
328 need the meteorological information about precipitation and evapotranspiration. With these values
329 it is possible to estimate interception losses in a given period of time.

330 Most research on interception losses is focused on forests [57]. Generally, interception losses in
331 tree canopies represent 9% to 36% of the gross rainfall [58,59]. It has been estimated the rainfall
332 interception losses is up to 48% of gross rainfall for the Corsican pine stands [34]. For upland forests
333 in England the interception losses amount to 35% in areas with an annual rainfall of more than 1000
334 mm, but it is reported that the loss is higher in areas with lower rainfall with 500–600 mm, amounting
335 to about 40–50% [60]. For the mountainous areas in Poland, interception losses amount to 30% for
336 both coniferous and deciduous forests where the annual rainfall ranged from 560 mm to 740 mm [61].
337 For wetland plants investigated in this study the amount of interception losses is lower than values
338 for forests (mean equal to 12.6% of rainfall) but as with trees, higher gross rainfall was associated
339 with lower interception losses. In the wettest growing season (sum of precipitation 621 mm)
340 interception losses amount to only 9% of gross rainfall, while in the driest growing season (sum of
341 precipitation 235 mm) interception losses are almost twice as high and amount to 16.3% of the gross
342 rainfall.

343 There are fewer studies conducted on interception losses for grass or other herbaceous plants,
344 and with those few studies, their results show very high variability. In one study conducted in Texas,
345 the interception losses for *Curlymesquite* grass were estimated as 10.8% of annual rainfall, while
346 interception losses for *Sideoats grama* grass were estimated at 18.1% of annual rainfall [21]. In another
347 study conducted in Australia, higher values were observed for *Curly Mitchell* grass at 32% of annual
348 rainfall [31]. The study area in [21] has a mean annual precipitation of 609 mm, while in [31], it is 255
349 mm. In our study for interception losses values of wetland plants, that area has a mean annual
350 precipitation of 552 mm, lies between values of the two previously mentioned studies. These suggest
351 that annual rainfall is an important factor and can lead to big variations in estimation of interception
352 losses.

353 Our study confirmed that there are different driving factors of maximum canopy storage and
354 interception losses. Maximum canopy storage depends on plant morphology and is highly varied
355 across the same plant types, which can be seen by high standard deviation of field measurements and
356 it is confirmed by different studies carried out in Biebrza Valley [36,45]. In addition, when maximum
357 canopy storage was calculated for the entire plant communities (in the measuring squares) for the
358 selected times of growing season, high variability is still observed where the mean values do not
359 differ significantly. Meanwhile, interception losses are driven by hydrological and climatic variables,
360 not just plant morphology and phenology (expressed by maximum canopy storage).

361 Interception losses was calculated with daily time step, based on precipitation, potential
362 evapotranspiration and maximum canopy storage. So the obtained values of interception losses must
363 be affected by those processes. However, daily values were then used to calculate relationship of
364 interception losses in growing season and month time scales. For these two time scales, more factors
365 were taken into account and this factor: maximum storage capacity and precipitation distribution
366 over time play an important role in values of monthly or growing season interception losses. Our
367 results show there is a high correlation between growing season and monthly interception losses and
368 the number of days with precipitation in a given period, also precipitation is highly correlated with
369 interception losses.

370 Based on the data from 1971-2015, we observe a trend of increasing number of days with
371 precipitation, at a rate of 0.4 days/year. The R of days of raining and absolute interception losses is
372 very high at 0.88. Therefore, it is reasonable to expect the interception losses will be higher in the
373 future. Our projection is it will be increased by 7.2 mm till 2050 and by 17.4 mm till 2100.

374 If the trends in our collected meteorological data remain the same in the future, a great impact
375 on interception is anticipated. Potential huge increase in interception losses will in turn affects water
376 balance. In our study area, there is no significant increase in precipitation over the years. However,
377 for study areas where change in precipitation is significant, then this change needs to be considered
378 in the projection/estimation of interception.

379 5. Conclusions

380 Based on one year's measurements of maximum canopy storage for wetland plant communities,
381 interception losses for the 1971–2015 period were estimated using meteorological elements. For both
382 months and growing seasons, the mean value was around 13% gross precipitation, which shows that
383 interception cannot be neglected in such areas and should be taken into account with more care in
384 hydrological modelling.

385 One essential process in interception losses estimation is precipitation. The total sum of
386 precipitation in the investigated period highly affected the estimation of interception. However, more
387 important is the time distribution of precipitation, and the results show that the number of days with
388 precipitation in the growing seasons is most strongly correlated with annual absolute rainfall
389 interception. Even higher correlations are observed for monthly absolute interception. The fact that
390 precipitation characteristics are driven by climate change makes rainfall interception an important
391 parameter to consider in water-balance studies.

392 Because interception losses are calculated based on daily values of climate and hydrological
393 variables, its dependence on the latter is natural. However, relation between interception losses with

394 parameters is made in different time scales (in months and growing season) and we should look at
395 all other processes such as distribution of rainfall over time and changes in maximum storage
396 capacity. These observations enable us to project future trend of rainfall interception.

397 For annual and monthly relative interception, no strong correlations with meteorological
398 elements were observed. Nonetheless, interception is a complex phenomenon which depends not
399 only on meteorological conditions, but also on plant morphology. Hence, still more comprehensive
400 research should be performed to fully understand the interception process itself and its influence on
401 the whole ecosystem.

402 **Acknowledgments:** The research was conducted within the scope of research grant NCN2012/05/B/NZ9/03467
403 "INterception-TRanspiration-EVaporation; interdependencies of hydrological processes in WETland
404 ECOSystems (INTREV-WetEco)". The authors would like to thank Jonathan C-W Chan from Vrije Universiteit
405 Brussel for his help in preparing the final version of this manuscript.

406 **Author Contributions:** Wojciech Ciężkowski conducted the field measurements, prepared the manuscript and
407 analyzed the data; Tomasz Berezowski conducted the field measurement, supervised the first author, suggested
408 the methodology, cooperated in data analysis and reviewed the manuscript; Małgorzata Kleniewska conducted
409 the field measurements, processed the data, discussed methodology, and reviewed the manuscript; Sylwia
410 Szporak-Wasilewska and Jarosław Chormański conducted the field measurements, discussed methodology and
411 reviewed the manuscript.

412 **Conflicts of Interest:** "The authors declare no conflict of interest."

413 References

- 414 1. Wood, M.K.; Jones, T.L.; Vera-Cruz, M.T. Rainfall interception by selected plants in the chihuahuan
415 desert. *Journal of Range Management* **1998**, *51*, 91-96.
- 416 2. van Dijk, A.; Bruijnzeel, L.A. Modelling rainfall interception by vegetation of variable density using an
417 adapted analytical model. Part 1. Model description. *Journal of Hydrology* **2001**, *247*, 230-238.
- 418 3. Muzylo, A.; Llorens, P.; Valente, F.; Keizer, J.J.; Domingo, F.; Gash, J.H.C. A review of rainfall
419 interception modelling. *Journal of Hydrology* **2009**, *370*, 191-206.
- 420 4. Campbell, C.L.; Madden, L.V. *Introduction to plant disease epidemiology*. John Wiley & Sons.: 1990.
- 421 5. Bradley, D.J.; Gilbert, G.S.; Parker, I.M. Susceptibility of clover species to fungal infection: The
422 interaction of leaf surface traits and environment. *American Journal of Botany* **2003**, *90*, 857-864.
- 423 6. Brueggemann, E.; Spindler, G. Wet and dry deposition of sulphur at the site melpitz in east germany -
424 in memorium dedicated to wolfgang rolle. *Water Air and Soil Pollution* **1999**, *109*, 81-99.
- 425 7. Wesely, M.L.; Sisterson, D.L.; Jastrow, J.D. Observations of the chemical-properties of dew on
426 vegetation that affect the dry deposition of so₂. *Journal of Geophysical Research-Atmospheres* **1990**, *95*,
427 7501-7514.
- 428 8. Brewer, C.A.; Smith, W.K. Patterns of leaf surface wetness for montane and subalpine plants. *Plant Cell*
429 *and Environment* **1997**, *20*, 1-11.
- 430 9. Hanba, Y.T.; Moriya, A.; Kimura, K. Effect of leaf surface wetness and wettability on photosynthesis in
431 bean and pea. *Plant Cell and Environment* **2004**, *27*, 413-421.
- 432 10. Savenije, H.H.G. The importance of interception and why we should delete the term evapotranspiration
433 from our vocabulary. *Hydrological Processes* **2004**, *18*, 1507-1511.
- 434 11. Grayson, R.B.; Moore, I.D.; McMahon, T.A. Physically based hydrologic modeling .1. A terrain-based
435 model for investigative purposes. *Water Resources Research* **1992**, *28*, 2639-2658.
- 436 12. Garrote, L.; Bras, R.L. A distributed model for real-time flood forecasting using digital elevation
437 models. *Journal of Hydrology* **1995**, *167*, 279-306.
- 438 13. Reggiani, P.; Rientjes, T.H.M. Flux parameterization in the representative elementary watershed
439 approach: Application to a natural basin. *Water Resources Research* **2005**, *41*, 18.

- 440 14. Liu, Z.Y.; Todini, E. Towards a comprehensive physically-based rainfall-runoff model. *Hydrology and*
441 *Earth System Sciences* **2002**, *6*, 859-881.
- 442 15. Beven, K.; Kirkby, M.J. A physically based, variable contributing area model of basin hydrology/un
443 modèle à base physique de zone d'appel variable de l'hydrologie du bassin versant. *Hydrological Sciences*
444 *Journal* **1979**, *24*, 43-69.
- 445 16. Abbot, M.; Bathurst, J.; Cunge, J.; O'Connell, P.; Rasmussen, J. An introduction to the european
446 hydrologic system-système hydologique europeen." She". 1: History and philosophy of a physically
447 based, distributed modelling system. *Journal of Hydrology* **1990**, *87*, 45-59.
- 448 17. Liu, Y.; De Smedt, F. Wetspa extension, a gis-based hydrologic model for flood prediction and
449 watershed management. *Vrije Universiteit Brussel, Belgium* **2004**, *1*, e108.
- 450 18. Grah, R.F.; Wilson, C.C. Some components of rainfall interception. *Journal of Forestry* **1944**, *42*, 890-898.
- 451 19. Wohlfahrt, G.; Bianchi, K.; Cernusca, A. Leaf and stem maximum water storage capacity of herbaceous
452 plants in a mountain meadow. *Journal of Hydrology* **2006**, *319*, 383-390.
- 453 20. Yu, K.L.; Pypker, T.G.; Keim, R.F.; Chen, N.; Yang, Y.B.; Guo, S.Q.; Li, W.J.; Wang, G. Canopy rainfall
454 storage capacity as affected by sub-alpine grassland degradation in the qinghai-tibetan plateau, china.
455 *Hydrological Processes* **2012**, *26*, 3114-3123.
- 456 21. Thurow, T.L.; Blackburn, W.H.; Warren, S.D.; Taylor, C.A. Rainfall interception by midgrass,
457 shortgrass, and live oak mottes. *Journal of Range Management* **1987**, *40*, 455-460.
- 458 22. Jetten, V.G. Interception of tropical rain forest: Performance of a canopy water balance model.
459 *Hydrological Processes* **1996**, *10*, 671-685.
- 460 23. Germer, S.; Elsenbeer, H.; Moraes, J.M. Throughfall and temporal trends of rainfall redistribution in an
461 open tropical rainforest, south-western amazonia (rondonia, brazil). *Hydrology and Earth System Sciences*
462 **2006**, *10*, 383-393.
- 463 24. Czikowsky, M.J.; Fitzjarrald, D.R. Detecting rainfall interception in an amazonian rain forest with eddy
464 flux measurements. *Journal of Hydrology* **2009**, *377*, 92-105.
- 465 25. Holder, C.D. Rainfall interception and fog precipitation in a tropical montane cloud forest of guatemala.
466 *Forest Ecology and Management* **2004**, *190*, 373-384.
- 467 26. Aboal, J.R.; Jimenez, M.S.; Morales, D.; Hernandez, J.M. Rainfall interception in laurel forest in the
468 canary islands. *Agricultural and Forest Meteorology* **1999**, *97*, 73-86.
- 469 27. Dykes, A.P. Rainfall interception from a lowland tropical rainforest in brunei. *Journal of Hydrology* **1997**,
470 *200*, 260-279.
- 471 28. Grelle, A.; Lundberg, A.; Lindroth, A.; Moren, A.S.; Cienciala, E. Evaporation components of a boreal
472 forest: Variations during the growing season. *Journal of Hydrology* **1997**, *197*, 70-87.
- 473 29. Shachnovich, Y.; Berliner, P.R.; Bar, P. Rainfall interception and spatial distribution of throughfall in a
474 pine forest planted in an arid zone. *Journal of Hydrology* **2008**, *349*, 168-177.
- 475 30. Sraj, M.; Brilly, M.; Mikos, M. Rainfall interception by two deciduous mediterranean forests of
476 contrasting stature in slovenia. *Agricultural and Forest Meteorology* **2008**, *148*, 121-134.
- 477 31. Dunkerley, D.L.; Booth, T.L. Plant canopy interception of rainfall and its significance in a banded
478 landscape, arid western new south wales, australia. *Water Resources Research* **1999**, *35*, 1581-1586.
- 479 32. Tromble, J.M. Interception of rainfall by tarbush. *Journal of Range Management* **1983**, *36*, 525-526.
- 480 33. Kołodziej, J.; Liniewicz, K.; Bednarek, H. „intercepcja opadów atmosferycznych w łąkach zbóż.”. *Acta*
481 *Agrophysica* **2005**, *6*, 381-391.

- 482 34. Rutter, A.J.; Robins, P.C.; Morton, A.J.; Kershaw, K.A. Predictive model of rainfall interception in
483 forests, .1. Derivation of model from observations in a plantation of corsican pine. *Agricultural*
484 *Meteorology* **1972**, *9*, 367-&.
- 485 35. Berezowski, T.; Chormanski, J.; Kleniewska, M.; Szporak-Wasilewska, S.; Ieee. Towards rainfall
486 interception capacity estimation using als lidar data. *2015 Ieee International Geoscience and Remote Sensing*
487 *Symposium (Igarss)* **2015**, 735-738.
- 488 36. Suliga, J.; Chormanski, J.; Szporak-Wasilewska, S.; Kleniewska, M.; Berezowski, T.; van Griensven, A.;
489 Verbeiren, B. In *Derivation from the landsat 7 ndvi and ground truth validation of lai and interception storage*
490 *capacity for wetland ecosystems in biebrza valley, poland*, Conference on Remote Sensing for Agriculture,
491 Ecosystems, and Hydrology XVII part of the International Symposium on Remote Sensing, Toulouse,
492 FRANCE, Sep 22-24, 2015; Spie-Int Soc Optical Engineering: Toulouse, FRANCE, 2015.
- 493 37. de Jong, S.M.; Jetten, V.G. Estimating spatial patterns of rainfall interception from remotely sensed
494 vegetation indices and spectral mixture analysis. *International Journal of Geographical Information Science*
495 **2007**, *21*, 529-545.
- 496 38. Gomez, J.A.; Giraldez, J.V.; Fereres, E. Rainfall interception by olive trees in relation to leaf area.
497 *Agricultural Water Management* **2001**, *49*, 65-76.
- 498 39. Hoyningen-Huene, J.v. Die interzeption des niederschlages in landwirtschaftlichen pflanzenbeständen.
499 Arbeitsbericht deutscher verband für wasserwirtschaft und kulturbau, dvwk, braunschweig, germany.
500 DVWK Braunschweig: 1981.
- 501 40. Verbeiren, B.; Khanh Nguyen, H.; Wirion, C.; Batelaan, O. An earth observation based method to assess
502 the influence of seasonal dynamics of canopy interception storage on the urban water balance. *Belgeo.*
503 *Revue belge de géographie* **2016**.
- 504 41. Wirion, C.; Ho, K.N.; Bauwens, W.; Verbeiren, B. Using remote sensing to describe urban surface
505 properties for improved hydrological modelling.
- 506 42. Górnaiak, A. Klimat i termika wód powierzchniowych kotliny biebrzańskiej. W: Kotlina biebrzańska i
507 biebrzański park narodowy. Aktualny stan, zagrożenia i potrzeby czynnej ochrony środowiska. Red.
508 H. Banaszuk. Wydaw. *Ekonomia i Środowisko*, Białystok **2004**.
- 509 43. Kossowska-Cezak, U.; Olszewski, K.; Przybylska, G. Climate of the biebrza valley. *Zeszyty Problemowe*
510 *Postepow Nauk Rolniczych* **1991**, *372*, 119-158.
- 511 44. West, N.E.; Gifford, G.F. Rainfall interception by cool-desert shrubs. *Journal of Range Management* **1976**,
512 *29*, 171-172.
- 513 45. Ignar, S.; Węglewska, A.; Szporak-Wasilewska, S.; Chormański, J. Spatial and temporal variability of
514 the interception in the natural wetland valley, the lower biebrza basin case study. *Annals of Warsaw*
515 *University of Life Sciences-SGGW. Land Reclamation* **2013**, *45*, 111-119.
- 516 46. Calder, I.R.; Hall, R.L.; Rosier, P.T.W.; Bastable, H.G.; Prasanna, K.T. Dependence of rainfall
517 interception on drop size .2. Experimental determination of the wetting functions and two-layer
518 stochastic model parameters for five tropical tree species. *Journal of Hydrology* **1996**, *185*, 379-388.
- 519 47. Szporak-Wasilewska, S.; Szatyłowicz, J.; Okruszko, T.; Ignar, S. Application of the surface energy
520 balance system model (sebs) for mapping evapotranspiration of extensively used river valley with
521 wetland vegetation. *Towards Horizon* **2013**, *2020*, 929-942.
- 522 48. Allen, R.G.; Pereira, L.S.; Raes, D.; Smith, M. Crop evapotranspiration-guidelines for computing crop
523 water requirements-fao irrigation and drainage paper 56. *FAO, Rome* **1998**, *300*, D05109.

- 524 49. Gash, J.H.C. Analytical model of rainfall interception by forests. *Quarterly Journal of the Royal*
525 *Meteorological Society* **1979**, *105*, 43-55.
- 526 50. Kępińska-Kasprzak, M.; Mager, P. Thermal growing season in poland calculated by two different
527 methods. *Annals of Warsaw University of Life Sciences, Land Reclamation* **2015**, *47*, 261-273.
- 528 51. McLeod, A. Package 'kendall.'kendall rank correlation and mann-kendall trend test, cran,
529 classification/msc 62m10, 91b84. 2011.
- 530 52. Monson, R.K.; Grant, M.C.; Jaeger, C.H.; Schoettle, A.W. Morphological causes for the retention of
531 precipitation in the crowns of alpine plants. *Environmental and Experimental Botany* **1992**, *32*, 319-327.
- 532 53. Pypker, T.G.; Bond, B.J.; Link, T.E.; Marks, D.; Unsworth, M.H. The importance of canopy structure in
533 controlling the interception loss of rainfall: Examples from a young and an old-growth douglas-fir
534 forest. *Agricultural and Forest Meteorology* **2005**, *130*, 113-129.
- 535 54. Klaassen, W.; Bosveld, F.; de Water, E. Water storage and evaporation as constituents of rainfall
536 interception. *Journal of Hydrology* **1998**, *212*, 36-50.
- 537 55. Gash, J.; Wright, I.; Lloyd, C.R. Comparative estimates of interception loss from three coniferous forests
538 in great britain. *Journal of Hydrology* **1980**, *48*, 89-105.
- 539 56. Loustau, D.; Berbigier, P.; Granier, A. Interception loss, throughfall and stemflow in a maritime pine
540 stand. Ii. An application of gash's analytical model of interception. *Journal of Hydrology* **1992**, *138*, 469-
541 485.
- 542 57. Dunkerley, D. Measuring interception loss and canopy storage in dryland vegetation: A brief review
543 and evaluation of available research strategies. *Hydrological Processes* **2000**, *14*, 669-678.
- 544 58. Hormann, G.; Branding, A.; Clemen, T.; Herbst, M.; Hinrichs, A.; Thamm, F. Calculation and simulation
545 of wind controlled canopy interception of a beech forest in northern germany. *Agricultural and Forest*
546 *Meteorology* **1996**, *79*, 131-148.
- 547 59. Levia, D.F.; Keim, R.F.; Carlyle-Moses, D.E.; Frost, E.E. Throughfall and stemflow in wooded
548 ecosystems. In *Forest hydrology and biogeochemistry: Synthesis of past research and future directions*, Levia,
549 D.F.; CarlyleMoses, D.; Tanaka, T., Eds. Springer: Dordrecht, 2011; Vol. 216, pp 425-443.
- 550 60. Calder, I.R. *Evaporation in the uplands*. Wiley: 1990.
- 551 61. Kozłowski, R.; Józwiak, M. Transformacja opadów atmosferycznych w strefie drzew wybranych
552 ekosystemów leśnych w górach świętokrzyskich= the transformation of precipitation in the tree canopy
553 in selected forest ecosystems of poland's świętokrzyskie mountains. *Przegląd Geograficzny* **2017**, *89*, 133-
554 153.